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Dependence of urban heat islands on anthropogenic heat emissions: A synthesis from seven typical cities in China

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13

14 Abstract

15 Anthropogenic heat (AH), an important urban heat source, is often overlooked or simplified in research on the driving mechanisms of the urban heat island effect (UHI), and case studies 16 investigating the impacts of different AH estimation methods are scarce. This study estimated 17 the AH in seven typical Chinese cities based on a remote sensing surface energy balance model 18 (AH_{seb}) and an energy consumption inventory-machine learning model (AH_{inv}). The intensity 19 20 of the surface UHI was extracted using land surface temperatures, and then the linear mixed-21 effects model and geographic detectors were used to analyze the driving effect of AH on the 22 UHI. Despite the similar shapes of the spatial profile curves, the AH derived from the two models differed in both temporal and spatial characteristics, which was more typical in winter 23 and in urban centers, and AH_{inv} had a more notable central spread feature than AH_{seb}. The 24 25 AH driving effects on UHI were notably influenced by spatial and temporal heterogeneity, particularly in regions with distinct background climates. However, after controlling for the random effects of the background climate, AH still exhibited a considerable enhancing effect on the UHI. AH_{seb} outperformed AH_{inv} in terms of linear positive correlation and interpretation rate for UHI. Meanwhile, interactions with other potential factors enhanced AH driving effects. This study offers guidance for simulating and analyzing urban climates to propose practical and effective measures for optimizing urban thermal environments, with a focus on AH or and energy consumption control.

33 Key words: Anthropogenic heat, Urban heat islands, Driving analysis, Spatiotemporal
34 heterogeneity

35

36 1. Introduction

37 Urbanization has had the greatest impact on the transformation of the Earth's environment, 38 as evidenced by the growth in urban area and population in recent decades (Chapman et al., 39 2017; Zhou et al., 2018). During the rapid process of urbanization, natural landscapes are replaced with impervious surfaces, changing the properties and geometries of the ground 40 41 surface and influencing energy absorption, storage, and emission, as well as releasing additional anthropogenic heat (Firozjaei et al., 2020; Meng et al., 2018; Mirzaei and Haghighat, 2010; 42 43 Voogt and Oke, 2003). Urban heat islands (UHIs) are a major environmental issue caused by urbanization, with substantially impacts on urban climate, ecology, and human health 44 (Mohajerani et al., 2017; Ulpiani, 2021; Yang et al., 2022b). Understanding the underlying 45 46 mechanisms of UHI is crucial for developing effective measures to optimize the quality of 47 human living space (Meng et al., 2022).

Land surface temperature (LST) obtained through thermal infrared remote sensing has 48 broader applicability compared to site-based atmospheric temperature measurements, which 49 50 have spatial limitations (Bahi et al., 2019; Dewan et al., 2021). LST can determine the spatial 51 and temporal characteristics of UHI and enhance our understanding of urban climate dynamics 52 (Hu et al., 2020; Ward et al., 2016; Wu et al., 2022). Therefore, surface UHI based on LST has become a crucial aspect in urban thermal environment studies (Fu et al., 2022; Peng et al., 2012). 53 54 The formation of UHIs is a result of a complex interaction between multiple factors affecting surface energy balance (Rizwan et al., 2008; Zhao et al., 2014); in recent years, there has been 55 extensive exploration of the patterns driving UHIs in multiple regions and time periods. To 56 57 identify feasible measures for thermal environment mitigation under specific climatic and geographic conditions, a wide range of potential UHI drivers related to urbanization are 58 considered using statistical or machine learning methods. These drivers include factors such as 59 vegetation and water distribution, the characteristics of artificial surfaces, landscape patterns, 60 three-dimensional building structures, and human activity (Dewan et al., 2021; Fu et al., 2022; 61 62 Hu et al., 2020; Hu et al., 2022; Liang et al., 2020; Meng et al., 2022; Mohajerani et al., 2017; 63 Ramirez-Aguilar and Lucas Souza, 2019; Wang et al., 2021b). However, most current studies 64 focus on the difference in energy balance between urban and rural areas due to changes in the physical properties of the land surface resulting from urbanization, often overlooking or 65 simplifying anthropogenic heat (AH), which is a key energy source and driver of the urban 66 67 thermal environment (Chapman et al., 2017; Peng et al., 2012; Yang et al., 2022a). Instead,

socioeconomic variables such as nighttime lighting, population density, gross domestic product,
and road networks have been used as proxies for AH in UHI studies (Liang et al., 2020; Raj et
al., 2020; Wang et al., 2021b; Wang et al., 2021c; Zhou et al., 2014), although these factors are
not directly correlated with LST. The difficulties in obtaining AH data with spatial and temporal
heterogeneity may be an important impediment to relevant studies (Dong et al., 2017; Sun et
al., 2018).

To support the exploration of the impact of AH on the urban thermal environment, 74 75 numerous implementable AH estimation methods, models, and datasets have been proposed. These can be divided into three categories: energy consumption inventory methods, surface 76 energy balance methods, and building energy simulation methods (Sailor, 2011). Building 77 78 energy simulation is considered the most accurate method for determining building heat 79 emissions, but it has a high demand for data and cannot be applied in large-scale studies (Alhazmi et al., 2022; Nie et al., 2014; Vahmani et al., 2022). The energy consumption 80 81 inventory method is the most common method for estimating AH and is highly applicable and 82 scalable in various studies (Quah and Roth, 2012; Smith et al., 2009; Varquez et al., 2021). With 83 advancements in communication and network technologies, big data with location-based 84 semantic information, spatial interaction information, and real-time dynamic information have 85 great potential in representing and explaining the spatiotemporal characteristics of human 86 activities (Chen et al., 2020; Gao et al., 2017); combining these data with inventory methods 87 can effectively improve the resolution of AH estimation (Liu et al., 2021; Ming et al., 2022; Xu 88 et al., 2021). Furthermore, the combination of machine-learning algorithms allows for more

accurate and applicable AH models (Chen et al., 2020; Qian et al., 2022; Wang et al., 2022b). 89 90 The surface energy balance (SEB) method uses micro-meteorological observation techniques 91 such as the eddy covariance system to determine the terms in the SEB equation, with the residual term considered as the AH (Pigeon et al., 2007). The combination of SEB with 92 93 quantitative remote sensing techniques (RS-SEB) allows for application on larger spatial scales 94 (Kato and Yamaguchi, 2005), making it one of the most effective methods for obtaining high-95 resolution AH and a common means of validating the results of AH estimation (Chow et al., 96 2014; Peng et al., 2021; Zhou et al., 2012). Due to the strengths and weaknesses of various methods, multi-method mixed modeling and cross-validation have become an important 97 98 direction for future AH studies (Wang et al., 2022a; Zheng and Weng, 2018). However, further 99 investigation is necessary to examine the differences in the spatial and temporal distribution of 100 AH estimated by different methods and its impact on the urban thermal environment.

101 The study of the impact of AH on the urban thermal environment is still in the early stages 102 of numerical simulation experiments (Molnar et al., 2020; Singh et al., 2022; Tao et al., 2022; 103 Zhan and Xie, 2022) and lacks systematic analysis in real cases. Based on refined AH data, the 104 role of AH in the urban thermal environment can be reflected more intuitively and convincingly 105 through a mature statistical-based analysis from UHI driving studies (Hu et al., 2020), which 106 also provide example arguments for numerical simulation studies. In this study, AH was modeled in seven cities in different geographic regions of China using two methods. The 107 objectives of this study were to 1) investigate the differences in the spatial and temporal 108 109 characteristics of different AH results, 2) explore the important relationships between AH and 110 UHI, and 3) examine the impacts of AH on UHI and its interactions with other potential factors.111

112 **2. Study area and dataset**

113 **2.1. Study area**

114 Since the late 1970s, China has undergone unprecedented economic development and urbanization (Cao et al., 2016; Schneider and Mertes, 2014; Yang et al., 2019), resulting in a 115 116 dramatic increase in energy consumption and associated AH emissions. This has altered the 117 energy flow in urban ecosystems and affected urban ecological processes such as urban regional 118 climate and atmospheric environment, causing frequent extreme heat events, decreased air 119 quality, and serious impacts on the health of residents in Chinese cities (Cong et al., 2022; Gu 120 et al., 2016). One representative central city from each of the seven natural geographic regions 121 of China (Zhao et al., 2015) was selected as the target area (Fig. 1). These cities serve as the 122 economic, cultural, and transportation hubs of their respective regions and as well as the entire 123 country, and they exhibit considerable variations in climate, topography, and other natural 124 conditions. Hence, studying the driving mechanisms of AH and UHI in these cities will provide 125 valuable information for urban thermal environment management. Due to their close integration, 126 Foshan and Guangzhou are treated as one region in this study and are referred to as Guangzhou 127 in the following text.



Fig. 1. Geographical location, terrain, and administrative extents of the study areas.



128 129

131 **2.2. Dataset**

132 To calculate UHI intensities and obtain the surface parameters required for the RS-SEB 133 model, the following data were used: MOD11A1 1000m resolution daily surface temperature and emissivity product, MOD09GA 500m resolution daily surface reflectance product from 134 NASA (https://ladsweb.nascom.nasa.gov), ALOS 30m resolution global digital surface model 135 (DSM) dataset (Tadono et al., 2016), 2017 10m resolution global land cover data, and 30m 136 resolution global artificial impervious surface data (Gong et al., 2020; Gong et al., 2019). The 137 meteorological parameters required for the RS-SEB model were obtained from the National 138 139 Centers for Environmental Information site data (NCEI GIS Team, 2021) and ERA5 140 atmospheric reanalysis data (Muñoz Sabater, 2019) corresponding to the Terra satellite crossing
141 moment.

142 The energy consumption inventory-machine learning model used in this study is based on 143 a previous study (Qian et al., 2022) and requires a large and diverse dataset including energy 144 consumption, socioeconomic, point-of-interest, road/railway, nighttime illumination, surface 145 temperature, and meteorological, and topographical data. The data sources and preprocessing 146 are described in the original study (Qian et al., 2022). Additionally, population heat data 147 obtained from cell phone user location information from the Baidu Huiyan big data platform 148 (https://huiyan.baidu.com) were also included in this study to describe the dynamic population 149 aggregation within the city to estimate the hourly gridded AH. The monthly average data for 150 July and December 2017 (2016 for Guangzhou) in the study area were obtained based on the 151 above data sources.

152

153 **3. Method**

This study was divided into two major parts (**Fig. 2a**): AH estimation and UHI driving analysis of the mean satellite transit moment in summer (July) and winter (December). First, the AH was estimated using a machine learning model that combined the energy inventory method and a remote sensing surface energy balance model. After the UHI intensity was calculated, a linear mixed-effects model and geographic detectors were used to examine the important relationships between AH and UHI.



Fig. 2. Workflow for (a) the whole study and (b) the energy inventory-machine learning model.
 AH: anthropogenic heat, SEB: surface energy balance, UHI: urban heat island.

163 **3.1 AH estimation**

164 3.1.1. Inventory based-machine learning method

165 The most common AH estimation method is the top-down energy consumption inventory method, which is based on large-scale energy consumption data and is downscaled step-by-step 166 167 using specified spatial-temporal rules. In our previous study (Qian et al., 2022), we constructed 168 an efficient AH estimation model combined with top-down inventory and machine learning 169 (ML) algorithms. Further details on the data preprocessing, AH sample estimation, and model 170 training evaluation can be found in Qian et al. (2022). Therefore, this study improved the 171 training process of ML models using the stacking integration framework (Wolpert, 1992) to 172 fully exploit the advantages of various ML algorithms. Furthermore, the monthly AH from the 173 ML model is the average of the whole-day AH for that month, which may be further refined in 174 time scales to obtain the AH at the Terra satellite transit moment for that month (Fig. 2b). The 175 intraday variations in multi-source AH were estimated using population heat (PH) data within 176 each grid and the intraday variation curve of industrial heat proposed in previous studies (Liu et al., 2021; Zheng and Weng, 2018). The AH obtained using this method is denoted as AH_{inv}
and is calculated as follows:

179
$$AH_{inv}^{h} = f_{h1} \cdot AH_{1}^{m} + f_{h2} \cdot AH_{2}^{m}$$
(1.)

180
$$f_{h2} = \frac{d_1 \times ph_{h1} + d_2 \times ph_{h2}}{d_1 \times ph_{d1} + d_2 \times ph_{d2}}$$
(2.)

where AH_{inv}^{h} denotes the mean AH ($W \cdot m^{2}$) at time h of satellite transit for the corresponding 181 month, AH_1^m is the monthly mean industrial heat from the ML model, f_{h1} is the intraday 182 variation coefficient of industrial heat (%), AH_2^m is the monthly mean transportation heat, 183 184 building heat, and metabolic heat from the ML model, and f_{h2} is the intraday variation 185 coefficient of population (%). ph_{h1} and ph_{h2} are the PH values at time h in that month for weekdays and weekends, d_1 and d_2 are the number of weekdays and weekends in that month, 186 187 and ph_{d1} and ph_{d2} are the average PH values for the entire day on weekdays and weekends. 188 Detailed information about the stacking framework is provided in the Appendix. For the sake 189 of brevity, AH derived using this method is hereafter referred to as AH_{inv}. Both the calculation 190 process and results have a resolution of 500 m.

191 3.1.2. Remote sensing-surface energy balance model

AH has been recognized as an important source term in urban multiscale energy systems (Chapman et al., 2017; Pigeon et al., 2007), whereas the sum of the energy received and released from the surface system remains constant during transformation and transfer, and the surface energy balance method considers the residual term of the energy budget as the AH-induced disturbance. Kato and Yamaguchi (2005) first proposed the remote sensing surface energy balance (RS-SEB) model for AH estimation. This model, which applies quantitative remote sensing techniques to estimate surface energy, has been widely adopted in urban-scale studies
(Firozjaei et al., 2020; Wong et al., 2015; Yu et al., 2021). In this study, a similar RS-SEB model
was constructed to estimate the daytime AH following the previous modeling process:

201
$$AH = R_n - (H + LE + G)$$
 (3.)

where R_n denotes the net radiation, H represents the sensible heat flux, LE is the latent heat flux, and G indicates ground or conductive heat flux at the material surface. However, due to the uncertainty in the calculation of parameters at night, G and AH are combined as the principal surface heat sources during nighttime to represent the release of heat storage at night (Kato and Yamaguchi, 2007):

207

$$\Delta S = AH - G = H + LE - R_n \tag{4.}$$

208 where ΔS represents the intensity of heat storage release at night and its magnitude reflects the 209 intensity of AH at the city scale. Considering the coarse spatial resolution of thermal infrared 210 remote sensing, the influence of mixed pixels on the calculation of surface parameters required 211 for RS-SEB cannot be overlooked (Liu et al., 2020). Therefore, in this study, fixed surface 212 parameters such as aerodynamic resistance were computed using high-resolution DSM and land 213 cover data, and their pixel values at coarse resolution (500 m) were calculated using sub-pixel 214 aggregation. For specific information on RS-SEB parameter calculation and model construction, 215 please refer to Kato and Yamaguchi (2005) and Zhou et al. (2012). For brevity, the AH with a 216 spatial resolution of 500 m based on this method is referred to as AH_{seb}.

3.2. UHI intensity

218

In this study, urban heat island intensity (UHII) was defined as the difference between the

urban LST and the mean LST of the surrounding rural areas (Clinton and Gong, 2013; Zhou et
al., 2014), Therefore, the boundaries of the urban built-up areas and rural areas of the city must
be determined. The impervious surface distribution density (ISDD) was calculated to describe
the extent and density of impervious surface in a specified radius around a pixel (Firozjaei et
al., 2020; Meng et al., 2018):

224
$$ISDD_{s}(r) = \frac{\sum_{i=1}^{n} B_{si} \cdot \left(1 - \frac{D_{i}}{2r}\right)}{\sum_{i=1}^{n} \left(1 - \frac{D_{i}}{2r}\right)}$$
(5.)

where *s* represents the center pixel, B_{si} is the value of the *i*th pixel within radius *r* (impervious pixel = 1; permeable pixel = 0), D_i is the distance between pixel *i* and the center pixel, and *n* is the total number of pixels within a circle of radius *r* (1,000 m). Urban built-up area boundaries were extracted using the city clustering algorithm (CCA) based on the obtained ISDD (Rozenfeld et al., 2008). The rural area in this study is defined as a buffer zone around an urban area that is the same size as the urban area (Dewan et al., 2021; Zhou et al., 2014); the UHII can be calculated using the following equation:

232 $UHII = LST_u - \overline{LST_r}$ (6.)

where LST_u represents the surface temperature within the urban area of the city and $\overline{LST_r}$ is the average surface temperature in rural areas.

235 **3.3. Relationship between AH and UHI**

Linear mixed-effects models and geographic detectors were used to elucidate the relationship between AH and UHI. AH estimation models are the empirical and physical expressions of this intrinsic relationship; however mathematical and statistical approaches can provide a more intuitive and interpretable result.

240 3.3.1. Mixed effects model

This study aimed to determine the practical contribution of AH to UHII by using a linear 241 242 mixed-effects model. In multi-regional and multi-temporal studies, UHII is largely determined 243 by climatic context, including geographical area, season, and diurnal factors. Thus, the role of 244 AH may not be significant. Mixed-effects models (Bollen and Brand, 2010; Sheiner and Grasela, 245 1991) could be used to address this issue. We treated the variation in UHII with AH as a fixed effect, while regional, seasonal, and diurnal factors were considered random effects and were 246 247 included in the analysis as categorical variables (Table A1). By excluding the random effects, we could determine the actual impact of AH in a multi-spatial and temporal environment, and 248 249 also reflect the important role of climate context. In this study, nested effects between different 250 random effects were considered and implemented in R, and the optimal mixed-effects model 251 was selected based on the Akaike information criterion.

252 **3.3.2.** Geographic detectors

253 Geographic detectors are a set of statistical methods for detecting spatial differentiation and revealing the driving forces behind it. They are divided into four modules: factor detection, 254 255 interaction detection, risk detection, and ecological detection (Wang and Hu, 2012; Wang et al., 256 2010). The basic assumption is that if an independent variable has a substantial effect on a 257 dependent variable, the spatial distributions of the independent and dependent variables should 258 be similar. Geographic detectors can identify complex linear and nonlinear relationships between independent and dependent variables, and the analysis results are not affected by 259 260 multivariate covariance (Wang et al., 2021a). This study employed the factor detection module

to measure the UHI spatial heterogeneity induced by AH, which is quantitatively described bythe q-statistic:

263
$$q = 1 - \frac{\sum_{h=1}^{L} N_h \sigma_h^2}{N \sigma^2}$$
(7.)

264 where h is the layering of the dependent variable Y or factor X, namely the classification or zoning, σ_h^2 and σ^2 are the variances of the Y values for layer h and the entire region, 265 266 respectively. The value range of q is [0,1], and a larger value indicates a more notable spatial 267 heterogeneity of Y or a stronger interpretation of Y by X. Interaction detection assesses the 268 interaction of the two factors on the target variable, represented by the stacked q-value. In 269 addition to AH, this study calculated UHI drivers that were commonly used in previous studies (Peng et al., 2012; Wu et al., 2022; Yang et al., 2017) to determine the interaction between AH 270 271 and other drivers (Table A2), including landscape pattern index, class level index of impervious surface and vegetation, enhanced vegetation index (EVI) and surface albedo. 272

273

4. Results

275 4.1. Characteristics of AH distribution

The AH and nighttime ΔS values were obtained for the seven cities during spring and summer using both models (**Fig. 3**). The AH_{inv} was notably stronger during the day than at night in the same season, and it was stronger in summer than in winter in the southern regions (Shanghai, Wuhan, Chengdu, and Guangzhou), where summer temperatures were high and winter temperatures were relatively mild, but in the northern regions (Beijing, Shenyang, and Lanzhou), where winters are extremely cold, the winter AH_{inv} was stronger. For daytime

 AH_{seb} , the AH values in the urban built-up area were much larger in summer than in winter, 282 283 while the spatial distribution of ΔS at night in winter was more consistent with the urban 284 boundaries. Despite the similar distribution ranges, the AH obtained by the two models differed 285 in both temporal and spatial details. The extent and intensity of AH_{seb} were stronger than 286 AH_{inv} in the summer in some cities, whereas AH_{seb} was generally weaker in the winter. The high values of AH_{inv} were remarkably clustered in the urban center and radiated outward, 287 whereas the high values of AH_{seb} were mainly distributed in the low-rise areas and industrial 288 289 areas around the urban areas, and were weak in the urban center with dense high-rise buildings.



290

291 Fig. 3. Spatial and temporal distribution of anthropogenic heat and nighttime heat storage: (a)-

292 (d) AH_{inv} in summer daytime, winter daytime, summer nighttime and winter nighttime, 293 respectively; (e)–(f) AH_{seb} in summer and winter daytime, respectively; (g)–(h) heat storage 294 (ΔS) in summer and winter nighttime based on RS-SEB, respectively.

To better reflect the spatial and temporal differences in the results of the different models, 295 296 the AH in Beijing and Shanghai were selected for the spatial profile visualization of urban areas 297 (Fig. 4). AH exhibited a roughly peaked curve, with high values in the central region where the 298 impervious aggregation density was the strongest. The summer daytime curves of AH for the 299 two models were similar, particularly in regions dominated by built-up areas, whereas there 300 were large differences in peripheral regions dominated by rural areas, primarily in the 301 anomalously low and high values of AH_{seb} in the periphery regions. During the winter, AH_{seb} 302 was lower than AH_{inv} during the daytime, especially in metropolitan areas. The AH_{inv} in Beijing was slightly lower at night than during the day, whereas the AH_{inv} in Shanghai 303 304 decreased more notably at night. In contrast, the nighttime ΔS in Beijing was very strong, 305 whereas the nighttime ΔS in Shanghai was closer to the daytime AH values. The spatial 306 distribution of ΔS was smoother than that of the AH profiles.



307

308 **Fig. 4.** Spatial profiles of AH and nighttime heat storage (Δ S) based on different methods in (a) 309 Beijing and (b) Shanghai. Each unit of "Row" and "Col" refers to one grid (500 m) of rows and

309 Beijing and (b) Shangh310 columns, respectively.

311 **4.2. Relationship between AH and UHI**

312 The spatial and temporal characteristics of UHI were determined using the LST map (**Fig.**

5 and **Table A3**). UHI was strongest during summer daytime in most regions, with the exception of Lanzhou, where an urban cold island was observed during the daytime in both summer and winter due to its arid and sparsely vegetated conditions. UHI was weaker during winter daytime compared to summer, but in Guangzhou, which is located in the tropics, the UHI was much stronger during winter daytime than in other regions, although the nighttime UHI was not pronounced. However, in most regions, a strong UHI was observed during both summer and winter nighttime, with a distinct central radiative distribution.



Fig. 5. Surface temperatures in major built-up and rural areas in each city at different timeperiods.

320

In general, there was a clear positive relationship between the AH and UHII (**Fig. 6**). The positive relationship between AH_{inv} and UHII was strong during summer daytime in most cities, but weak during the winter daytime and relatively substantial at night. However, the relationship between AH_{inv} and UHII in Lanzhou and Guangzhou was different: notable in 327 Lanzhou only at night and in Guangzhou only during the daytime. Compared with AHinv, 328 AH_{seb} and nighttime ΔS had stronger linear correlations with UHII, and this relationship was 329 highly significant (p<0.001) in most cases. The results of the mixed-effects model (Table 1) 330 indicated that random effects, such as regional, seasonal, and diurnal effects, had high variance, 331 suggesting a substantial impact on the UHI. After excluding random effects, the positive linear 332 contribution of AH to UHI remained significant (p<0.001). This shows the fixed warming effect 333 of AH in the urban thermal environment, and the slope is a measure of the impact of AH on 334 UHII. The greater slope of AH_{seb} indicated a stronger influence than AH_{inv}, but the larger 335 random effects also implied that AH_{seb} was more susceptible to interference from the urban 336 background climate.







341

	AH type	Random effect		Fixed effect	
		Random term	Variance explanation	Intercept	Slope
	AH _{inv}	AH Region/Season/Time	39 %	-0.02	0.49***
	AH _{seb}	AH Region/Season	74 %	-0.04	0.66***

342 **Table 1.** Results of the mixed-effects model based on AH and UHII (Z-Score normalization).

343

344 **4.3.** The driving effects of AH on UHI

345 The q-value of the geographic detectors captured the extent to which the independent 346 variables accounted for the target variables and explored the role of AH in driving surface UHI 347 (Fig. 7). The interpretation rate for AH_{inv} was higher during summer and winter nights. In all 348 cities except Guangzhou and Lanzhou, the interpretation rate of AH_{inv} was not greatly 349 different between summer daytime and nighttime. The interpretation rate of AH_{inv} was much 350 stronger during winter nighttime than daytime in most cities, but the opposite was true for 351 Guangzhou. The explanatory rate of daytime AH_{seb} was generally higher than that of AH_{inv} 352 and was stronger in summer than in winter. The explanatory rate of nighttime ΔS was the 353 strongest for UHI, particularly during winter nighttime. However, the interpretation rate of 354 nighttime ΔS in Guangzhou was lower than that of AH_{seb} in daytime for both winter and 355 summer. Overall, the results from the different AH models showed notable differences in terms 356 of the driving effects on UHI in multiple spatial and temporal scenarios.



358 Fig. 7. Interpretation rate (q-value) of AH and ΔS on UHII calculated using factor detection.

357

The interaction between AH and all the factors improved the interpretation rate for UHI (Fig. 8). Generally, the interactions were similar to the results of the separate effects of AH, with the interaction of AH having the lowest interpretation rate during winter daytime and a relatively high rate during summer daytime, except for Guangzhou and Lanzhou. The effects of these factors did not differ greatly, but were influenced by spatial and temporal heterogeneity. For the interaction of AH_{inv}, the impervious surface and vegetation factors differed 365 considerably, particularly at night with clear dividing lines in the interpretation rate plot, 366 whereas albedo was a more prominent factor, often exhibiting large differences from the other 367 factors and was the most dominant enhancer during winter daytime when the overall interpretation was low. The interaction characteristics of the other factors with $\,AH_{seb}\,$ and $\,\Delta S$ 368 369 were similar to AH_{inv}, but the interaction was stronger due to the higher explanatory rate of AH_{seb} and ΔS , which could even reach 100% in Beijing and Lanzhou during winter nights. 370 The role of albedo, however, was not as prominent in AH_{seb} and ΔS and the interaction 371 372 interpretation rate of the landscape-level factors was weak.



Fig. 8. Interaction interpretation rate for UHII (q-value) between other driving factors and (a) AH_{inv} , (b) AH_{seb} and nighttime ΔS.

376

373

377 **5. Discussion**

378 **5.1. Implications of the AH from different methods**

379 The distinction between the inventory-based method and the energy balance method was

380 evident in the definition and calculation of AH, which was reflected in the spatial and temporal

381	distribution of AH and its relationship with UHI. The method based on energy consumption
382	inventories follows the assumption that heat generated from anthropogenic energy consumption
383	is immediately released into the atmosphere as sensible heat (Kotthaus and Grimmond, 2012;
384	Smith et al., 2009). However, in reality, this heat is absorbed, stored, or dissipated into the
385	atmosphere as latent heat. Thus, AH_{inv} should be considered as the maximum sensible heat
386	generated by anthropogenic activities. In contrast, RS-SEB ignores the perturbation of AH on
387	terms other than the sensible heat term; thus, AH_{seb} could be considered as an increase in near-
388	surface atmospheric sensible heat caused by AH (Firozjaei et al., 2020; Kato and Yamaguchi,
389	2005). This explains why AH_{seb} had a stronger linear correlation and explanatory rate with
390	the surface UHI. In urban centers with a high density of high-rise buildings, AH_{seb} was
391	generally lower than AH _{inv} due to non-negligible heat storage perturbation and impacts of
392	building shading, and stronger heat storage efficiency in winter further reduced its value (Kato
393	and Yamaguchi, 2007; Wong et al., 2015; Yu et al., 2021). Because AH_{inv} is derived from
394	urban energy consumption and does not have a direct effect on the urban thermal environment,
395	its interpretation of UHI was lower. However, its interaction with other factors could be
396	explained as the efficiency of AH_{inv} as sensible heat release, causing an increase in LST. The
397	interaction of AH_{seb} with other factors represents the warming effect of anthropogenic
398	sensible heat influenced by surface properties and is influenced by the urban background
399	climate, building materials, and meteorological conditions. At night, the main energy source
400	besides AH is the heat stored in urban surface materials that dissipates in the form of long-wave
401	radiation (Zheng et al., 2021; Zhou et al., 2014). Therefore, the ΔS in this study reflected the

402 UHI distribution at night and confirmed the determining effect of the combined contribution of403 artificial surfaces and AH sources to the nighttime UHI.

404 In general, the differences in physical meanings and numerical values determine the 405 differences in the distribution characteristics and environmental effects of AH_{inv} and AH_{seb} 406 under various conditions. When examining the characteristics of human activities and their environmental impact, AH_{inv} is more appropriate, as it comes directly from energy 407 consumption and social economy, whereas AH_{seb} is a better choice for urban energy budgets 408 409 and direct impacts on the thermal environment. However, the suitability of these methods is not 410 absolute as the definitions of the AH input required by analytical or simulation methods must 411 be considered. Clarifying the implications of AH input is critical to obtain more accurate and 412 robust conclusions when investigating the driving role of human activities on UHI.

413

5.2. Implications for UHI mitigation

The inclusion of AH in climate simulations can considerably improve model performance 414 and reliability for meteorological parameters and air pollutants, whereas urban numerical 415 416 simulation studies that do not consider AH are physically incomplete (Chen et al., 2016; Molnar 417 et al., 2020). Because of the differences in the implications of AH estimated via different 418 methods, there are limitations in using only a broad definition of AH to study the urban thermal 419 environment and climate change (Liu et al., 2022a). It is critical to select the most appropriate 420 AH input under different AH coupling schemes for various climate and urban canopy numerical 421 models (Molnar et al., 2020; Narumi et al., 2009). Currently, the dominant coupling scheme in 422 studies is to use the AH obtained through inventory-based methods as sensible heat, but latent and storage heat should not be disregarded as they are important energy dissipation pathways
(Vahmani et al., 2022; Yu et al., 2021). Therefore, the scarcity of comparative simulation studies
using multiple AH inputs hinders a thorough understanding of the urban thermal environment,
thus limiting the development of AH-based UHI mitigation measures.

427 In contrast, despite extensive research into UHI driving mechanisms, it is difficult to 428 propose feasible thermal environment mitigation measures for developed land, whereas AH is 429 controllable as an important heat source within cities, and such control has been effective in 430 UHI mitigation (Liu et al., 2022b). Therefore, future research should focus on a deeper analysis and simulation of AH to achieve a reduction in energy consumption and an increase in urban 431 432 thermal comfort through the reasonable control of AH. However, the differences in AH effects 433 on the urban thermal environment are heavily influenced by the background climate, resulting 434 in some synchronization with UHII variations as well as notable regional differences. As a result of the spatial and temporal heterogeneity of UHI and driving mechanisms, thermal 435 436 mitigation measures must be tailored to local and temporal conditions, particularly in regions where the background environment differs markedly from the general situation, as in the 437 438 particular characteristics of Lanzhou and Guangzhou in this study. The continental arid climate, 439 urban basin topography, and large amount of bare soil around urban areas contribute to the cold 440 island effect during daytime in Lanzhou, whereas in Guangzhou, the warm and rainy climate 441 results in dense and moist land around the urban area, implying higher thermal inertia (Hafner and Kidder, 1999; Pandey et al., 2014), which may be an important reason for the considerably 442 443 nighttime UHI. Other regions have unique characteristics that contribute to the variability of

AH driving effects. However, the strong UHI and AH effects in the summer are an urgent 444 concern. Even if the AH input cannot be reduced, its warming effect can be mitigated to some 445 446 extent by adjusting other potential drivers, such as increasing urban vegetation cover and 447 building material albedo, or altering the urban landscape structure. However, although the UHI 448 is high at night in winter, this additional heat helps to improve thermal comfort in cold climates 449 and reduces heating energy consumption, demonstrating how capitalizing on the warming 450 impact of AH and ΔS works in cold scenarios. This study serves as an important theoretical 451 foundation for analyzing UHI from the perspective of AH and provides new ideas for more 452 precise urban climate simulations to serve as more effective urban thermal environment 453 optimization measures.

454 **5.3.** Limitations and uncertainties

455 The underlying mechanism of urban heat islands is a complex integration of multiple factors (Hu et al., 2020), particularly in multi-regional and multi-temporal analyses, making it 456 457 challenging to arrive at a uniform and systematic conclusion. Furthermore, this study 458 considered the impact of AH using two models, producing some reasonable results overall, but 459 it was difficult to gain a thorough understanding of the physical mechanism in all regions and 460 at all times due to the limitation of the scale of this study (Peng et al., 2012). More rigorous 461 analyses can be performed for specific regions and time phases. Furthermore, due to cumulative 462 parametric errors in the equation residuals, RS-SEB may produce larger errors in non-urban areas with less precise parameter settings than in urban areas, and the uncertainty assumptions 463 464 raise some issues (Park et al., 2016; Sailor, 2011). Therefore, improving the accuracy of the

465 estimations of the SEB terms and their parameters is a priority for subsequent studies. Possible
466 areas of improvement in the future could include more reasonable heat storage models, finer
467 meteorological observation data, and high-resolution multispectral remote sensing images for
468 improved AH estimations based on RS-SEB.

469

470 **6.** Conclusion

This study investigated the differences in the spatial and temporal distributions and impacts of different AH models and their relationships with UHI. First, we estimated AH using an RS-SEB model, an improved energy consumption-machine learning model, and nighttime Δ S in seven Chinese cities, extracted the UHI based on remote-sensing LST, and analyzed the driving effects of AH on UHI and the interactions with other factors using a linear mixed-effects model and geographic detectors.

477 Although AH_{inv} and AH_{seb} were similar in distribution, there were considerable 478 differences in temporal and spatial details, with AH_{inv} showing more aggregation in urban 479 centers and radiating outward, whereas AH_{seb} was relatively weaker. AH_{seb} was notably 480 weaker than AH_{inv} during the daytime in winter. UHI was strong during summer daytime and 481 winter nighttime and has a stronger positive relationship with AH_{inv} at this time, whereas 482 AH_{seb} and ΔS had a stronger linear correlation with UHI than AH_{inv} . The mixed-effects 483 model confirmed the inherent warming effect of AH on the cities, but also showed the significance of background climatic factors in affecting the differences between AH_{inv} and 484 485 AH_{seb} and their relationship with UHI. According to the geographic detectors, AH_{inv} had a

stronger impact on UHI during the daytime and nighttime in summer and winter than at other times, and the driving effects of AH_{seb} and ΔS were stronger. The interaction of other potential driving factors with AH enhanced the interpretation rate of UHI, and the general trend of the interaction effects was similar to that of separate AH effects, but varied considerably between different factors. The surface albedo had a more salient impact on the AH_{inv} driving effects, whereas the interaction of AH_{seb} with other factors had a high interpretation rate (close to 100%).

493 AH obtained based on different methods has different implications in various spatial and temporal scenarios, and it is necessary to clarify the definition of AH in urban thermal 494 495 environment research to reach more scientific and rigorous conclusions. In addition, the 496 selection of AH inputs that match the coupling scheme of numerical and urban canopy models 497 is an important topic that cannot be ignored in future urban climate simulations. More crucially, 498 AH is the most important controllable driving factor of UHI and the key to a feasible urban 499 thermal comfort optimization scheme, although the situational heterogeneity in the role of AH requires attention. 500

501

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507 **Reference**

- Alhazmi, M., Sailor, D.J., Anand, J., 2022. A new perspective for understanding actual
 anthropogenic heat emissions from buildings. Energy Build. 258, 111860.
 <u>https://doi.org/10.1016/j.enbuild.2022.111860</u>
- Bahi, H., Mastouri, H., Radoine, H., 2019. Review of methods for retrieving urban heat islands.
 3rd International Conference on Materials and environmental Science (ICMES). 27,
 3004-3009. <u>https://doi.org/10.1016/j.matpr.2020.03.272</u>
- 514Bollen, K.A., Brand, J.E., 2010. A General Panel Model with Random and Fixed Effects A515Structural Equations Approach. Soc. Forces 89, 1-34.516https://doi.org/10.1353/sof.2010.0072
- 517 Cao, Q., Yu, D.Y., Georgescu, M., Wu, J.G., 2016. Impacts of urbanization on summer climate
 518 in China: An assessment with coupled land-atmospheric modeling. J. Geophys. Res.519 Atmos. 121, 10505-10521. <u>https://doi.org/10.1002/2016jd025210</u>
- Chapman, S., Watson, J.E.M., Salazar, A., Thatcher, M., McAlpine, C.A., 2017. The impact of
 urbanization and climate change on urban temperatures: a systematic review. Landscape
 Ecol. 32, 1921-1935. <u>https://doi.org/10.1007/s10980-017-0561-4</u>
- 523 Chen, F., Yang, X., Wu, J., 2016. Simulation of the urban climate in a Chinese megacity with
 524 spatially heterogeneous anthropogenic heat data. J. Geophys. Res.-Atmos. 121, 5193 525 5212. <u>https://doi.org/10.1002/2015jd024642</u>
- 526 Chen, Q., Yang, X., Ouyang, Z., Zhao, N., Jiang, Q., Ye, T., Qi, J., Yue, W., 2020. Estimation
 527 of anthropogenic heat emissions in China using Cubist with points-of-interest and
 528 multisource remote sensing data. Environ. Pollut. 266, 115183.
 529 <u>https://doi.org/10.1016/j.envpol.2020.115183</u>
- Chow, W.T.L., Salamanca, F., Georgescu, M., Mahalov, A., Milne, J.M., Ruddell, B.L., 2014.
 A multi-method and multi-scale approach for estimating city-wide anthropogenic heat fluxes. Atmos. Environ. 99, 64-76. <u>https://doi.org/10.1016/j.atmosenv.2014.09.053</u>
- 533 Clinton, N., Gong, P., 2013. MODIS detected surface urban heat islands and sinks: Global
 534 locations and controls. Remote Sens. Environ. 134, 294-304.
 535 <u>https://doi.org/10.1016/j.rse.2013.03.008</u>
- Cong, J.P., Wang, L.B., Liu, F.J., Qian, Z.M., McMillin, S.E., Vaughn, M.G., Song, Y.M., Wang,
 S.S., Chen, S.S., Xiong, S.M., Shen, X.B., Sun, X., Zhou, Y.Z., Ho, H.C., Dong, G.H.,
 2022. Associations between metabolic syndrome and anthropogenic heat emissions in
 northeastern China. Environ. Res. 204, 111974.
 https://doi.org/10.1016/j.envres.2021.111974
- 541 Dewan, A., Kiselev, G., Botje, D., Mahmud, G.I., Bhuian, M.H., Hassan, Q.K., 2021. Surface
 542 urban heat island intensity in five major cities of Bangladesh: Patterns, drivers and
 543 trends. Sust. Cities Soc. 71, 102926. <u>https://doi.org/10.1016/j.scs.2021.102926</u>
- 544Dong, Y., Varquez, A.C.G., Kanda, M., 2017. Global anthropogenic heat flux database with545high spatial resolution. Atmos. Environ. 150, 276-294.546<u>https://doi.org/10.1016/j.atmosenv.2016.11.040</u>
- 547 Firozjaei, M.K., Weng, Q.H., Zhao, C.H., Kiavarz, M., Lu, L.L., Alavipanah, S.K., 2020.
 548 Surface anthropogenic heat islands in six megacities: An assessment based on a triple-

- 549source surface energy balance model. Remote Sens. Environ. 242, 111751.550https://doi.org/10.1016/j.rse.2020.111751
- Fu, X., Yao, L., Xu, W., Wang, Y., Sun, S., 2022. Exploring the multitemporal surface urban
 heat island effect and its driving relation in the Beijing-Tianjin-Hebei urban
 agglomeration. Appl. Geogr. 144, 102714.
 https://doi.org/10.1016/j.apgeog.2022.102714
- Gao, S., Janowicz, K., Couclelis, H., 2017. Extracting urban functional regions from points of
 interest and human activities on location-based social networks. T. gis 21, 446-467.
 <u>https://doi.org/10.1111/tgis.12289</u>
- Gong, P., Li, X., Wang, J., Bai, Y., Cheng, B., Hu, T., Liu, X., Xu, B., Yang, J., Zhang, W., Zhou,
 Y., 2020. Annual maps of global artificial impervious area (GAIA) between 1985 and
 2018. Remote Sens. Environ. 236, 111510. https://doi.org/10.1016/j.rse.2019.111510
- Gong, P., Liu, H., Zhang, M., Li, C., Wang, J., Huang, H., Clinton, N., Ji, L., Li, W., Bai, Y.,
 Chen, B., Xu, B., Zhu, Z., Yuan, C., Suen, H.P., Guo, J., Xu, N., Li, W., Zhao, Y., Yang,
 J., Yu, C., Wang, X., Fu, H., Yu, L., Dronova, I., Hui, F., Cheng, X., Shi, X., Xiao, F.,
 Liu, Q., Song, L., 2019. Stable classification with limited sample: transferring a 30-m
 resolution sample set collected in 2015 to mapping 10-m resolution global land cover
 in 2017. Sci. Bull. 64, 370-373. https://doi.org/10.1016/j.scib.2019.03.002
- Gu, S.H., Huang, C.R., Bai, L., Chu, C., Liu, Q.Y., 2016. Heat-related illness in China, summer
 of 2013. Int. J. Biometeorol. 60, 131-137. <u>https://doi.org/10.1007/s00484-015-1011-0</u>
- Hafner, J., Kidder, S.Q., 1999. Urban heat island modeling in conjunction with satellite-derived
 surface/soil parameters. J. Appl. Meteorol. 38, 448-465. <u>https://doi.org/10.1175/1520-</u>
 0450(1999)038<0448:Uhimic>2.0.Co;2
- 572 Hu, D., Meng, Q., Zhang, L., Zhang, Y., 2020. Spatial quantitative analysis of the potential 573 driving factors of land surface temperature in different "Centers" of polycentric cities: 574 A case study in Tianjin, China. Sci. Total Environ. 706, 135244. 575 https://doi.org/10.1016/j.scitotenv.2019.135244
- Hu, D., Meng, Q.Y., Schlink, U., Hertel, D., Liu, W.X., Zhao, M.F., Guo, F.X., 2022. How do
 urban morphological blocks shape spatial patterns of land surface temperature over
 different seasons? A multifactorial driving analysis of Beijing, China. Int. J. Appl. Earth
 Obs. Geoinf. 106, 102648. https://doi.org/10.1016/j.jag.2021.102648
- 580 Kato, S., Yamaguchi, Y., 2005. Analysis of urban heat-island effect using ASTER and ETM+
 581 Data: Separation of anthropogenic heat discharge and natural heat radiation from
 582 sensible heat flux. Remote Sens. Environ. 99, 44-54.
 583 https://doi.org/10.1016/j.rse.2005.04.026
- Kato, S., Yamaguchi, Y., 2007. Estimation of storage heat flux in an urban area using ASTER
 data. Remote Sens. Environ. 110, 1-17. <u>https://doi.org/10.1016/j.rse.2007.02.011</u>
- Kotthaus, S., Grimmond, C.S.B., 2012. Identification of Micro-scale Anthropogenic CO2, heat
 and moisture sources Processing eddy covariance fluxes for a dense urban
 environment. Atmos. Environ. 57, 301-316.
 https://doi.org/10.1016/j.atmosenv.2012.04.024
- Liang, Z., Wu, S.Y., Wang, Y.Y., Wei, F.L., Huang, J., Shen, J.S., Li, S.C., 2020. The
 relationship between urban form and heat island intensity along the urban development

- 592
 gradients.
 Sci.
 Total
 Environ.
 708,
 135011.

 593
 https://doi.org/10.1016/j.scitotenv.2019.135011
- Liu, K., Li, X., Wang, S., Li, Y., 2020. Investigating the impacts of driving factors on urban
 heat islands in southern China from 2003 to 2015. J. Clean Prod. 254, 120141.
 <u>https://doi.org/10.1016/j.jclepro.2020.120141</u>
- 597 Liu, X., Yue, W.Z., Zhou, Y.Y., Liu, Y., Xiong, C.S., Li, Q., 2021. Estimating multi-temporal 598 anthropogenic heat flux based on the top-down method and temporal downscaling 599 China. Resour. Conserv. Recycl. methods in Beijing. 172. 105682. 600 https://doi.org/10.1016/j.resconrec.2021.105682
- Liu, Y., Luo, Z., Grimmond, S., 2022a. Revising the definition of anthropogenic heat flux from
 buildings: role of human activities and building storage heat flux. Atmos. Chem. Phys.
 22, 4721-4735. <u>https://doi.org/10.5194/acp-22-4721-2022</u>
- Liu, Z.H., Lai, J.M., Zhan, W.F., Bechtel, B., Voogt, J., Quan, J.L., Hu, L.Q., Fu, P., Huang, F.,
 Li, L., Guo, Z., Li, J.F., 2022b. Urban Heat Islands Significantly Reduced by COVIDLockdown. Geophys. Res. Lett. 49, e2021GL096842.
 https://doi.org/10.1029/2021gl096842
- Meng, Q., Liu, W., Zhang, L., Allam, M., Bi, Y., Hu, X., Gao, J., Hu, D., Jancso, T., 2022.
 Relationships between Land Surface Temperatures and Neighboring Environment in
 Highly Urbanized Areas: Seasonal and Scale Effects Analyses of Beijing, China.
 Remote Sens. 14, 4340. <u>https://doi.org/10.3390/rs14174340</u>
- Meng, Q., Zhang, L., Sun, Z., Meng, F., Wang, L., Sun, Y., 2018. Characterizing spatial and temporal trends of surface urban heat island effect in an urban main built-up area: A 12year case study in Beijing, China. Remote Sens. Environ. 204, 826-837.
 <u>https://doi.org/10.1016/j.rse.2017.09.019</u>
- Ming, Y.J., Liu, Y., Liu, X., 2022. Spatial pattern of anthropogenic heat flux in monocentric and
 polycentric cities: The case of Chengdu and Chongqing. Sust. Cities Soc. 78, 103628.
 <u>https://doi.org/10.1016/j.scs.2021.103628</u>
- 619Mirzaei, P.A., Haghighat, F., 2010. Approaches to study Urban Heat Island Abilities and620limitations.Build.Environ.45,2192-2201.621https://doi.org/10.1016/j.buildenv.2010.04.001
- Mohajerani, A., Bakaric, J., Jeffrey-Bailey, T., 2017. The urban heat island effect, its causes,
 and mitigation, with reference to the thermal properties of asphalt concrete. J. Environ.
 Manage. 197, 522-538. <u>https://doi.org/10.1016/j.jenvman.2017.03.095</u>
- Molnar, G., Kovacs, A., Gal, T., 2020. How does anthropogenic heating affect the thermal
 environment in a medium-sized Central European city? A case study in Szeged,
 Hungary. Urban CLim. 34, 100673. <u>https://doi.org/10.1016/j.uclim.2020.100673</u>
- [dataset] Muñoz Sabater, J., 2019. ERA5-Land monthly averaged data from 1981 to present.
 <u>https://doi.org/10.24381/cds.68d2bb3</u>
- Narumi, D., Kondo, A., Shimoda, Y., 2009. Effects of anthropogenic heat release upon the
 urban climate in a Japanese megacity. Environ. Res. 109, 421-431.
 <u>https://doi.org/10.1016/j.envres.2009.02.013</u>
- 633 [dataset] NCEI GIS Team, 2021. Hourly/Sub-Hourly Observational Data. V3.0.0.
 634 <u>https://www.ncei.noaa.gov/maps/hourly/</u>

- Nie, W.-S., Sun, T., Ni, G.-H., 2014. Spatiotemporal characteristics of anthropogenic heat in an
 urban environment: A case study of Tsinghua Campus. Build. Environ. 82, 675-686.
 <u>https://doi.org/10.1016/j.buildenv.2014.10.011</u>
- Pandey, A.K., Singh, S., Berwal, S., Kumar, D., Pandey, P., Prakash, A., Lodhi, N., Maithani,
 S., Jain, V.K., Kumar, K., 2014. Spatio temporal variations of urban heat island over
 Delhi. Urban CLim. 10, 119-133. <u>https://doi.org/10.1016/j.uclim.2014.10.005</u>
- Park, C., Schade, G.W., Werner, N.D., Sailor, D.J., Kim, C.-H., 2016. Comparative estimates
 of anthropogenic heat emission in relation to surface energy balance of a subtropical
 urban neighborhood. Atmos. Environ. 126, 182-191.
 https://doi.org/10.1016/j.atmosenv.2015.11.038
- Peng, S.S., Piao, S.L., Ciais, P., Friedlingstein, P., Ottle, C., Breon, F.M., Nan, H.J., Zhou, L.M.,
 Myneni, R.B., 2012. Surface Urban Heat Island Across 419 Global Big Cities. Environ.
 Sci. Technol. 46, 696-703. <u>https://doi.org/10.1021/es2030438</u>
- Peng, T., Sun, C., Feng, S., Zhang, Y., Fan, F., 2021. Temporal and Spatial Variation of
 Anthropogenic Heat in the Central Urban Area: A Case Study of Guangzhou, China.
 ISPRS Int. Geo-Inf. 10, 160. <u>https://doi.org/10.3390/ijgi10030160</u>
- Pigeon, G., Legain, D., Durand, P., Masson, V., 2007. Anthropogenic heat release in an old
 European agglomeration (Toulouse, France). Int. J. Climatol. 27, 1969-1981.
 <u>https://doi.org/10.1002/joc.1530</u>
- Qian, J., Meng, Q., Zhang, L., Hu, D., Hu, X., Liu, W., 2022. Improved anthropogenic heat flux
 model for fine spatiotemporal information in Southeast China. Environ. Pollut. 299,
 118917. https://doi.org/10.1016/j.envpol.2022.118917
- Quah, A.K.L., Roth, M., 2012. Diurnal and weekly variation of anthropogenic heat emissions
 in a tropical city, Singapore. Atmos. Environ. 46, 92-103.
 <u>https://doi.org/10.1016/j.atmosenv.2011.10.015</u>
- Raj, S., Paul, S.K., Chakraborty, A., Kuttippurath, J., 2020. Anthropogenic forcing exacerbating
 the urban heat islands in India. J. Environ. Manage. 257, 110006.
 https://doi.org/10.1016/j.jenvman.2019.110006
- Ramirez-Aguilar, E.A., Lucas Souza, L.C., 2019. Urban form and population density:
 Influences on Urban Heat Island intensities in Bogota, Colombia. Urban CLim. 29,
 100497. <u>https://doi.org/10.1016/j.uclim.2019.100497</u>
- Rizwan, A.M., Dennis, Y.C.L., Liu, C.H., 2008. A review on the generation, determination and
 mitigation of Urban Heat Island. J. Environ. Sci. 20, 120-128.
 <u>https://doi.org/10.1016/s1001-0742(08)60019-4</u>
- Rozenfeld, H.D., Rybski, D., Andrade Jr, J.S., Batty, M., Stanley, H.E., Makse, H.A., 2008.
 Laws of population growth. P. Natl. Acad. Sci. 105, 18702-18707.
 <u>https://doi.org/10.1073/pnas.0807435105</u>
- Sailor, D.J., 2011. A review of methods for estimating anthropogenic heat and moisture
 emissions in the urban environment. Int. J. Climatol. 31, 189-199.
 <u>https://doi.org/10.1002/joc.2106</u>
- 675 Schneider, A., Mertes, C.M., 2014. Expansion and growth in Chinese cities, 1978-2010.
 676 Environ. Res. Lett. 9, 024008. <u>https://doi.org/10.1088/1748-9326/9/2/024008</u>
- 677 Sheiner, L.B., Grasela, T.H., 1991. An introduction to mixed effect modeling: concepts,

- definitions, and justification[J]. Journal of pharmacokinetics and biopharmaceutics. J.
 Pharmacokinet. Biop. 19, S11-S24. <u>https://doi.org/10.1007/bf01371005</u>
- Singh, V.K., Mughal, M.O., Martilli, A., Acero, J.A., Ivanchev, J., Norford, L.K., 2022.
 Numerical analysis of the impact of anthropogenic emissions on the urban environment
 of Singapore. Sci. Total Environ. 806, 150534.
 https://doi.org/10.1016/j.scitotenv.2021.150534
- Smith, C., Lindley, S., Levermore, G., 2009. Estimating spatial and temporal patterns of urban
 anthropogenic heat fluxes for UK cities: the case of Manchester. Theor. Appl. Climatol.
 98, 19-35. <u>https://doi.org/10.1007/s00704-008-0086-5</u>
- Sun, R., Wang, Y., Chen, L., 2018. A distributed model for quantifying temporal-spatial patterns
 of anthropogenic heat based on energy consumption. J. Clean Prod. 170, 601-609.
 https://doi.org/10.1016/j.jclepro.2017.09.153
- Tadono, T., Nagai, H., Ishida, H., Oda, F., Naito, S., Minakawa, K., Iwamoto, H., 2016.
 GENERATION OF THE 30 M-MESH GLOBAL DIGITAL SURFACE MODEL BY
 ALOS PRISM. 23rd Congress of the International-Society-for-Photogrammetry-andRemote-Sensing (ISPRS). 41, 157-162. <u>https://doi.org/10.5194/isprsarchives-XLI-B4-</u>
 157-2016
- Tao, H.R., Xing, J., Pan, G.F., Pleim, J., Ran, L.M., Wang, S.X., Chang, X., Li, G.J., Chen, F.,
 Li, J.H., 2022. Impact of anthropogenic heat emissions on meteorological parameters
 and air quality in Beijing using a high-resolution model simulation. Front. Environ. Sci.
 Eng. 16, 44. <u>https://doi.org/10.1007/s11783-021-1478-3</u>
- Ulpiani, G., 2021. On the linkage between urban heat island and urban pollution island: Threedecade literature review towards a conceptual framework. Sci. Total Environ. 751,
 141727. <u>https://doi.org/10.1016/j.scitotenv.2020.141727</u>
- Vahmani, P., Luo, X., Jones, A., Hong, T., 2022. Anthropogenic heating of the urban environment: An investigation of feedback dynamics between urban micro-climate and decomposed anthropogenic heating from buildings. Build. Environ. 213, 108841.
 https://doi.org/10.1016/j.buildenv.2022.108841
- Varquez, A.C.G., Kiyomoto, S., Khanh, D.N., Kanda, M., 2021. Global 1-km present and future
 hourly anthropogenic heat flux. Sci. Data 8, 1-14. <u>https://doi.org/10.1038/s41597-021-</u>
 <u>00850-w</u>
- Voogt, J.A., Oke, T.R., 2003. Thermal remote sensing of urban climates. Remote Sens. Environ.
 86, 370-384. <u>https://doi.org/10.1016/s0034-4257(03)00079-8</u>
- Wang, J.-F., Hu, Y., 2012. Environmental health risk detection with GeogDetector. Environ.
 Modell. Softw. 33, 114-115. <u>https://doi.org/10.1016/j.envsoft.2012.01.015</u>
- Wang, J.-F., Li, X.-H., Christakos, G., Liao, Y.-L., Zhang, T., Gu, X., Zheng, X.-Y., 2010.
 Geographical Detectors-Based Health Risk Assessment and its Application in the
 Neural Tube Defects Study of the Heshun Region, China. Int. J. Geogr. Inf. Sci. 24, 107127. <u>https://doi.org/10.1080/13658810802443457</u>
- 717 Wang, K., Aktas, Y.D., Malki-Epshtein, L., Wu, D., Bin Abdullah, M.F.A., 2022a. Mapping the 718 city scale anthropogenic heat emissions from buildings in Kuala Lumpur through a top-719 down and bottom-up approach. Sust. Cities Soc. 76, 103443. а 720 https://doi.org/10.1016/j.scs.2021.103443

- Wang, X.M., Meng, Q.Y., Zhang, L.L., Hu, D., 2021a. Evaluation of urban green space in terms
 of thermal environmental benefits using geographical detector analysis. Int. J. Appl.
 Earth Obs. Geoinf. 105, 102610. <u>https://doi.org/10.1016/j.jag.2021.102610</u>
- Wang, Y., Yi, G., Zhou, X., Zhang, T., Bie, X., Li, J., Ji, B., 2021b. Spatial distribution and influencing factors on urban land surface temperature of twelve megacities in China from 2000 to 2017. Ecol. Indic. 125, 107533.
 https://doi.org/10.1016/j.ecolind.2021.107533
- Wang, Y.C., Hu, D.Y., Yu, C., Di, Y.F., Wang, S.S., Liu, M.Q., 2022b. Appraising regional anthropogenic heat flux using high spatial resolution NTL and POI data: A case study in the Beijing-Tianjin-Hebei region, China. Environ. Pollut. 292, 118359.
 https://doi.org/10.1016/j.envpol.2021.118359
- Wang, Z., Meng, Q., Allam, M., Hu, D., Zhang, L., Menenti, M., 2021c. Environmental and anthropogenic drivers of surface urban heat island intensity: A case-study in the Yangtze River Delta, China. Ecol. Indic. 128, 107845.
 https://doi.org/10.1016/j.ecolind.2021.107845
- Ward, K., Lauf, S., Kleinschmit, B., Endlicher, W., 2016. Heat waves and urban heat islands in
 Europe: A review of relevant drivers. Sci. Total Environ. 569, 527-539.
 <u>https://doi.org/10.1016/j.scitotenv.2016.06.119</u>
- 739
 Wolpert, D.H., 1992.
 Stacked generalization.
 Neural
 Networks
 5, 241-259.

 740
 https://doi.org/10.1016/S0893-6080(05)80023-1
- Wong, M.S., Yang, J.X., Nichol, J., Weng, Q.H., Menenti, M., Chan, P.W., 2015. Modeling of
 Anthropogenic Heat Flux Using HJ-1B Chinese Small Satellite Image: A Study of
 Heterogeneous Urbanized Areas in Hong Kong. IEEE Geosci. Remote Sens. Lett. 12,
 1466-1470. <u>https://doi.org/10.1109/lgrs.2015.2409111</u>
- Wu, Y., Hou, H., Wang, R., Murayama, Y., Wang, L., Hu, T., 2022. Effects of landscape patterns
 on the morphological evolution of surface urban heat island in Hangzhou during 20002020. Sust. Cities Soc. 79, 103717. <u>https://doi.org/10.1016/j.scs.2022.103717</u>
- Xu, D., Zhou, D., Wang, Y.P., Meng, X.Z., Gu, Z.L., Yang, Y.J., 2021. Temporal and spatial
 heterogeneity research of urban anthropogenic heat emissions based on multi-source
 spatial big data fusion for Xi'an, China. Energy Build. 240, 110884.
 https://doi.org/10.1016/j.enbuild.2021.110884
- Yang, B., Liu, H., Kang, E.L., Hawthorne, T.L., Tong, S.T.Y., Shu, S., Xu, M., 2022a. Traffic
 restrictions during the 2008 Olympic Games reduced urban heat intensity and extent in
 Beijing. Commun. Earth Environ. 3, 105. https://doi.org/10.1038/s43247-022-00427-4
- Yang, B., Yang, X., Leung, L.R., Zhong, S., Qian, Y., Zhao, C., Chen, F., Zhang, Y., Qi, J., 2019.
 Modeling the Impacts of Urbanization on Summer Thermal Comfort: The Role of
 Urban Land Use and Anthropogenic Heat. J. Geophys. Res.-Atmos. 124, 6681-6697.
 https://doi.org/10.1029/2018jd029829
- Yang, M., Wang, H., Yu, C.W., Cao, S.-J., 2022b. A global challenge of accurately predicting
 building energy consumption under urban heat island effect. Indoor Built Environ.,
 1420326x221123222. <u>https://doi.org/10.1177/1420326x221123222</u>
- Yang, Q., Huang, X., Li, J., 2017. Assessing the relationship between surface urban heat islands
 and landscape patterns across climatic zones in China. Sci. Rep. 7, 9337.

764 <u>https://doi.org/10.1038/s41598-017-09628-w</u>

- Yu, Z., Hu, L.Q., Sun, T., Albertson, J., Li, Q., 2021. Impact of heat storage on remote-sensing
 based quantification of anthropogenic heat in urban environments. Remote Sens.
 Environ. 262, 112520. <u>https://doi.org/10.1016/j.rse.2021.112520</u>
- Zhan, C., Xie, M., 2022. Land use and anthropogenic heat modulate ozone by meteorology: a
 perspective from the Yangtze River Delta region. Atmos. Chem. Phys. 22, 1351-1371.
 <u>https://doi.org/10.5194/acp-22-1351-2022</u>
- Zhao, G.S., Liu, J.Y., Kuang, W.H., Ouyang, Z.Y., Xie, Z.L., 2015. Disturbance impacts of land
 use change on biodiversity conservation priority areas across China: 1990-2010. J.
 Geogr. Sci. 25, 515-529. https://doi.org/10.1007/s11442-015-1184-9
- Zhao, L., Lee, X., Smith, R.B., Oleson, K., 2014. Strong contributions of local background
 climate to urban heat islands. Nature 511, 216-219. <u>https://doi.org/10.1038/nature13462</u>
- Zheng, Y., Huang, L., Zhai, J., 2021. Divergent trends of urban thermal environmental characteristics in China. J. Clean Prod. 287, 125053.
 <u>https://doi.org/10.1016/j.jclepro.2020.125053</u>
- Zheng, Y., Weng, Q., 2018. High spatial- and temporal-resolution anthropogenic heat discharge
 estimation in Los Angeles County, California. J. Environ. Manage. 206, 1274-1286.
 <u>https://doi.org/10.1016/j.jenvman.2017.07.047</u>
- Zhou, D.C., Zhao, S.Q., Liu, S.G., Zhang, L.X., Zhu, C., 2014. Surface urban heat island in
 China's 32 major cities: Spatial patterns and drivers. Remote Sens. Environ. 152, 51-61.
 <u>https://doi.org/10.1016/j.rse.2014.05.017</u>
- Zhou, Y., Weng, Q., Gurney, K.R., Shuai, Y., Hu, X., 2012. Estimation of the relationship
 between remotely sensed anthropogenic heat discharge and building energy use. ISPRSJ. Photogramm. Remote Sens. 67, 65-72.
 https://doi.org/10.1016/j.isprsjprs.2011.10.007
- Zhou, Y.Y., Li, X.C., Asrar, G.R., Smith, S.J., Imhoff, M., 2018. A global record of annual urban dynamics (1992-2013) from nighttime lights. Remote Sens. Environ. 219, 206-220.
 https://doi.org/10.1016/j.rse.2018.10.015

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