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1 **Balancing effort and benefit – how taxonomic and**
2 **quantitative resolution affect the ecological assessment of**
3 **pesticide effects**

4 **Author list**

5 Liana Liebmann^{1,2}, Philipp Vormeier^{1,3}, Oliver Weisner¹, Matthias Liess^{1,3}

6 ***Affiliations***

7 ¹ Helmholtz Centre for Environmental Research – UFZ, Permoserstr. 15, 04318 Leipzig,
8 Germany

9 ² Department Evolutionary Ecology & Environmental Toxicology (E3T), Institute of Ecology,
10 Diversity and Evolution, Faculty of Biological Sciences, Goethe University Frankfurt, 60438
11 Frankfurt am Main, Germany

12 ³ Institute for Environmental Research (Biology V), RWTH Aachen University, 52062 Aachen,
13 Germany

14 E-mail: Liana.Liebmann@ufz.de

15

24 **Abstract**

25 Biological indices aim to reflect the ecological quality of streams based on the community's
26 species or trait composition. Accordingly, the capability to predict the ecological quality
27 depends on (i) the knowledge on the association of taxa or traits with stressors and (ii) the
28 taxonomic and quantitative resolution of taxa. Generally speaking, a higher resolution is
29 associated with a better linkage between environmental condition and biological response but
30 also with higher efforts and costs. So far it is unknown how the taxonomic and quantitative
31 resolution affect the ecological quality assessment of streams related to pesticide effects when
32 applying the invertebrate-based indicator $SPEAR_{pesticides}$. We investigated the ecological quality
33 of 101 streams considering four taxonomic levels (species, genus, family, order) and three
34 quantitative resolutions (abundance, three abundance classes, and presence-absence). In a
35 multiple linear regression analysis between 13 investigated stressors and $SPEAR_{pesticides}$, the
36 full models' explained variance remained fairly constant with decreasing taxonomic and
37 quantitative resolution. As expected, the highest association between pesticide pressure and
38 $SPEAR_{pesticides}$ was reached at a species/abundance resolution yielding an R^2 of 0.43. In
39 contrast, the lowest quantitative resolution of order level combined with presence-absence
40 information revealed an explained variance of 0.28 R^2 . We suggest the family/abundance class
41 resolution ($R^2 = 0.38$) as the best trade-off between effort and accuracy for large-scale
42 monitoring. Due to a comparable linear regression at family/abundance class resolution, the
43 assigned ecological quality classes were largely congruent (69%) to species/abundance
44 resolution. We conclude that the ecological quality assessment with $SPEAR_{pesticides}$ at
45 family/abundance class resolution can be used to link pesticide contamination and invertebrate
46 community structure with less taxonomic expertise and less quantification effort.

47 **Introduction**

48 The ecological status of streams can be determined by applying biological indices. These
49 indices use the community's taxa or trait composition to assess the ecological quality of
50 streams. Indices such as the %EPT (Lenat, 1988), Saprobic Index (Kolkwitz R. and Marsson
51 M., 1909) and $SPEAR_{pesticides}$ (Liess and Ohe, 2005) aim to reflect the general degradation,

52 oxygen deficiency, or pesticide contamination of streams. Species have specific ecological and
53 environmental requirements but also share ecological niches, increasingly so when comparing
54 taxonomic levels above the genus and family levels. Accordingly, with increasing taxonomic
55 resolution, the taxa composition can be linked more accurately to environmental conditions.
56 Similarly, increasing quantitative resolution of sampling, meaning the counting of individuals,
57 strengthens the association between community expression and environmental conditions. In
58 this trade-off between effort and accuracy, some countries use the taxonomic level of family
59 (e.g. Spain: Alba-Tercedor, J and Sánchez-Ortega, A (1988); USA: Barbour et al. (1999);
60 Australia: Smith et al. (1999); UK: Environment Agency (2008)). Instead, other countries apply
61 species level for determining aquatic indices (e.g. USA, Germany, the Netherlands, Slovakia,
62 and South Korea; Buss et al. (2015)). However, the strength of the relationship between
63 environmental factors and biological indicators also depends on the accuracy of the available
64 autecological information on the taxa investigated.

65 The Rhithron Feeding Type Index RETI (Schweder, 1992) and the SPEAR_{pesticides} index (Liess
66 and Ohe, 2005) , for example, contain taxonomic information for species, genus, family, and
67 order level, whereas the Saprobic Index is mainly based on species level information with less
68 information at genus level and almost no information at family or order level. Accordingly,
69 Schmidt-Kloiber et al. (2004) revealed that the AQUEM Assessment software (ASS, 2000) is
70 not applicable at genus and family level, since some indices like the Saprobic Index do not
71 have sufficient information available at lower taxonomic levels. They found that when lowering
72 taxonomic resolution to genus level, already half of the sites investigated were assigned a
73 divergent ecological quality class. Regarding the trade-off between effort and accuracy for a
74 lower quantitative resolution, Buchner et al. (2019) analyzed how presence-absence resolution
75 affects the accuracy of ecological class assignment for biological indices. The Saprobic Index
76 showed with a mean Spearman's ρ of 0.93 a strong significant correlation between abundance
77 and presence-absence data. For the General Degradation Metrics (GDM from ASS 2002),
78 75% of the ecological quality classes were identical with abundance resolution (Buchner et al.,
79 2019). Accordingly, a higher taxonomic and quantitative resolution of invertebrate monitoring

80 improves the ecological quality assessment with biological indices that involve both abundance
81 and taxonomy.

82 Our aim was to identify the optimum trade-off between effort and accuracy when lowering
83 taxonomic and quantitative resolution with SPEAR_{pesticides}. For this analysis, we evaluated an
84 extensive Germany-wide data set comprising relevant environmental factors including high
85 resolution sampling of pesticides and high resolution sampling and determination of
86 invertebrates. In detail, we (i) assessed the specificity of SPEAR_{pesticides} to indicate pesticide
87 pressure among multiple stressors when reducing taxonomic and quantitative resolution, (ii)
88 evaluated the indication power under reduced resolution, (iii) investigated how ecological
89 quality classification is impacted, and (iv) derived implications for pesticide effect monitoring.

90 **2 Material and methods**

91 **2.1 Sampling sites**

92 The data analyzed in this study was collected and presented by Liess et al. (2021). The
93 macroinvertebrates and pesticide concentrations were sampled in two field campaigns from
94 April to July in 2018 and 2019 in a Germany-wide monitoring study. This study provides
95 information on 101 sampling sites with a varying degree of agricultural use (0 - 100%) in the
96 hydrological catchment (n = 41 with less than 10 km²; n = 60 with 10 - 30 km²), covering 13
97 different stream types (EU Commission, 2000); see SI Liess et al. (2021)). These 101 sampling
98 sites contained 86 agricultural (more than 20% agricultural land use in the catchment area)
99 and 15 non-agricultural sites (less than 20% agricultural land use in the catchment area). For
100 the 11 sites monitored in both campaigns, the indicator values and environmental factors were
101 averaged.

102 **2.2 Pesticide monitoring and analyses**

103 Water samples (n = 320) were taken with event-driven automated samplers (MAXX TP5,
104 Rangendingen, Germany) to capture the peak concentrations induced by significant rise of
105 water level (5 cm, depending on stream bed, Liess et al. (1999)). After the device was
106 activated, a mixed sample of 500 mL with 40 subsamples (every 5 minutes) was collected over

107 a total period of 3 hours and 20 minutes (Liess et al., 2021). Additionally, streams were
108 sampled regularly with 250 ml grab samples (n = 520) every three weeks according to WFD
109 standards (EU Commission, 2000). 75 pesticides and 33 pesticide metabolites were selected
110 by prioritization according to active substance-related sale quantities, under consideration of
111 the current environmental quality standards (EQS) and the regulatory acceptable
112 concentrations (RAC) (Wick et al., 2018) (see list of substances in SI Liess et al. (2021)). All
113 830 water samples were analyzed for these 108 substances via target analysis with high
114 pressure liquid chromatograph coupled with a high-resolution tandem mass spectrometer LC-
115 HRMS/MS without enrichment by multiple-reaction-monitoring (MRM, (Halbach et al., 2021;
116 Reemtsma et al., 2013)).

117 **2.3 Transferring measured concentrations into invertebrate toxicity**

118 All pesticide concentrations from EDS and grab samples were converted into invertebrate
119 toxicity by calculating Toxic Units (TU; Sprague (1970)) with the substance-related acute LC₅₀
120 value of *Daphnia magna* or *Chironomus sp.* (most sensitive organism selected per substance)
121 (Münze et al., 2017). LC₅₀ values were collected from the Pesticide Property Data Base (PPDB
122 and the US EPA ECOTOXicology knowledgebase in case of lacking *Chironomus* data) (Lewis
123 et al., 2016). Peak exposure in streams to pesticides toxic to invertebrates was determined by
124 the local maximum single substance insecticidal toxicity measured according to Liess et al.
125 (2021).

126 **2.4 Macroinvertebrate sampling**

127 Macroinvertebrates were sampled after the main application of pesticides in June in 2018 and
128 2019 at each site (Liess et al., 2021). The macroinvertebrate sampling followed the Water
129 Framework Directive (WFD) guideline (Meier et al., 2006) commonly used by the German
130 federal states. At each site, a 50-meter section downstream of the monitoring site was
131 sampled. Regarding all microhabitat types, 20 subsamples were taken with a kicking net
132 according to their share in the stream section (at least 5%). Subsequently, macroinvertebrates
133 were filtered, sorted, and conserved in 90% ethanol until laboratory determination. Taxa were

134 identified to the highest possible taxonomic resolution using a binocular (Zeiss, Stereo
135 Discovery V.20, Carl Zeiss Microscopy GmbH; Jena, Germany).

136 **2.5 Calculation of the biological indicator - $SPEAR_{pesticides}$**

137 The effects of pesticides on macroinvertebrates were quantified using the bio-indicator
138 $SPEAR_{pesticides}$ by (Liess and Ohe, 2005). $SPEAR_{pesticides}$ provided the relative abundance of
139 vulnerable species within a community and was normalized to indicator values under pristine
140 conditions according to Liess et al. (2021).

141 Different taxonomic resolutions were considered by aggregating the taxa to the respective
142 taxonomic level and summing up the abundances before transforming the data. For
143 quantitative resolutions, the abundances were adjusted to presence-absence and three
144 different abundance classes (see 2.7.1). In conclusion, the same $SPEAR_{pesticides}$ formula were
145 applied to all taxonomic and quantitative resolutions.

146 $SPEAR_{pesticides}$ is calculated by using the following equation (Knillmann et al., 2018):

$$147 \quad SPEAR_{pesticides} = \frac{\sum_{i=1}^n \log_{10}(4x_i + 1) \cdot y_i}{\sum_{i=1}^n \log_{10}(4x_i + 1)}$$

148 where n is the total number of taxa in a sample, x_i is the abundance of taxon i (given as
149 individuals per m^2), and y is set to 1 if taxon i is classified as “at risk” (Liess and Ohe, 2005) –
150 i.e. vulnerable to pesticides under regular exposure events – and set to 0 otherwise.

151 Trait information on vulnerability is allotted to 1,581 species in the $SPEAR_{pesticides}$ database,
152 1,325 of which are sensitive and 256 insensitive. At the genus level, 255 taxa are assigned
153 sensitivity information, divided into 186 sensitive and 69 insensitive. Information for 53
154 sensitive and 111 insensitive families is available, while at the order level only four orders are
155 classified as sensitive (Ephemeroptera, Trichoptera, Plecoptera, Megaloptera) and 11 as
156 insensitive. According to the results of this study, the groups Diptera Gen. sp. and Crustacea
157 Gen. sp. were classified as not vulnerable and the order Megaloptera Gen. sp. as vulnerable.

158 **2.6 Biotic and abiotic parameters**

159 To reveal the relationships between environmental factors and ecological status, we collected
160 biotic and abiotic parameters as described in Liess et al. (2021). Oxygen content, water level,
161 conductivity, and water pressure were measured continuously from April to June (Liess et al.,
162 2021). Nutrients and metals were analyzed in EDS and grab samples. Flow velocity and water
163 level were measured every three weeks. Hydromorphological parameters were identified once
164 according to the guidelines of the WFD. For more in-depth information regarding each method
165 and parameter, please refer to Liess et al. (2021).

166 **2.7 Data analysis**

167 **2.7.1 Taxonomic and quantitative resolutions**

168 In a first step, the taxa numbers corresponding to each taxonomic level were summed. For
169 lower quantitative resolutions, the abundance was set to one or zero for presence-absence or
170 grouped into three classes for abundance class resolution. The smallest class contained
171 abundances from one to three, the middle class comprised abundances from four to 100, and
172 the third class included abundances greater than 100. Ultimately, the same $SPEAR_{pesticides}$
173 equation (see chapter 2.5) was applied to all taxonomic levels of abundance, presence-
174 absence, and abundance class resolution.

175 In a second step, we evaluated the differences between $SPEAR_{pesticides}$ values resulting from
176 four taxonomic (species, genus, family, and order) and three quantitative (abundance,
177 abundance class, and presence-absence) resolutions visually using linear correlation (see SI
178 figure 1) and with a paired Wilcoxon-test (see SI table 1). As for other German invertebrate
179 metrics under the WFD, the four $SPEAR_{pesticides}$ boundaries separating the five even quality
180 classes equal Ecological Quality Ratio values of 0.8, 0.6, 0.4, and 0.2 (Environment Agency,
181 2008; EU Commission, 2008; Liess et al., 2021).

182 **2.7.2 Linear regressions, visualization, and ecological class assignment**

183 The prediction quality of all resolutions was evaluated by a multiple linear regression analysis
184 with a two-way interaction of 13 environmental variables (for the environmental variables refer

185 to Liess et al. (2021)). The relative importance of each variable was calculated with the relaimp
186 package (version 2.2-3) in R and given in R^2 (Grömping, 2006).

187 The effects of pesticide pressure on macroinvertebrate communities evaluated by
188 $SPEAR_{pesticides}$ were modelled by single linear regressions for the different taxonomic and
189 quantitative resolutions. The indication power of different levels of taxonomic and quantitative
190 resolutions were assessed by comparing linear model R^2 values. The regression slopes + y-
191 intercepts of $SPEAR_{pesticides}$ were compared.

192 In addition, the percentage of congruence of the assigned ecological quality classes for all
193 lower taxonomic and quantitative resolutions compared to the species/abundance resolution
194 was evaluated. Data was processed using the software R (version 4.0.2, R Core Team, 2020).

195 All diagrams were generated with the “ggplot2” version 3.3.2 in R (Wickham, 2009).

196 **2.7.3 Compensation factor at order level**

197 At the order level, for each quantitative resolution, a compensation factor was empirically
198 derived. The compensation factor approximates the order-level $SPEAR_{pesticides}$ values to those
199 at species/abundance resolution. The validity of the approximation is represented by a similar
200 slope and y-intercept of the regression lines compared to species/abundance resolution as
201 well as by a higher congruence in ecological class assignment (Fig. 2, B and C).

202 **3 Results and discussion**

203 **3.1 Indication power (performance) of $SPEAR_{pesticides}$ using different taxonomic and** 204 **quantitative resolutions**

205 **3.1.1 Indicating pesticide pressure among other stressors**

206 $SPEAR_{pesticides}$ reflects the effect of pesticide pressure on the macroinvertebrate community
207 composition. To evaluate the specificity of the different resolutions of $SPEAR_{pesticides}$, we
208 assessed the response of $SPEAR_{pesticides}$ to 13 environmental and anthropogenic stressors
209 considered relevant for the benthic invertebrate community. Here, we used a multiple linear
210 regression analysis to identify the relative importance of each of the potential stressors:
211 pesticide pressure; deficient hydromorphology; deficient bed habitat structure; O_2 deficiency;

212 nutrient pollution by NH_4 ; NO_2 ; total phosphorus; flow velocity; metal toxicity; temperature; pH;
213 and stream morphology stressors (Liess et al., 2021). Two of the 13 stressors included were
214 associated with $\text{SPEAR}_{\text{pesticides}}$ at all taxonomic and quantitative resolutions investigated (Fig.
215 1). In the full models, the explained variance ranged from an R^2 of 0.55 to 0.61 for all taxonomic
216 and quantitative resolutions, while remaining relatively independent of the resolution of
217 $\text{SPEAR}_{\text{pesticides}}$. All full models reflected the specificity of $\text{SPEAR}_{\text{pesticides}}$ through the identified
218 main driver pesticide pressure (TU_{max}), which contributed an average explained variance of
219 $0.26 R^2$ up to the family level. Deficient hydromorphology added an average explained variance
220 of $0.16 R^2$ to the full model for all resolutions. NH_4 was only identified as relevant by abundance
221 class and presence-absence resolution – but at all taxonomic levels – and added an average
222 explained variance of $0.08 R^2$ to the full model. Deficient bed habitat structure contributed an
223 average R^2 of 0.07 and was found to be relevant mainly for abundance resolution but also for
224 the genus and family level at the two lower quantitative resolution. O_2 deficiency added an
225 average R^2 of 0.06 to the explained variance in the full models affecting only the order-level
226 full models' explained variance (Fig. 1). At the order level, the relevance of the main driver
227 pesticide pressure is underestimated. However, our results indicate that the specificity of
228 $\text{SPEAR}_{\text{pesticides}}$ toward toxic pressure decreased at lower quantitative resolution while NH_4 and
229 hydromorphology explain more and toxicity less of the variance. The specificity of
230 $\text{SPEAR}_{\text{pesticides}}$ is validated regarding the explanatory power of lower taxonomic resolutions up
231 to the family level in terms of multiple stressors.



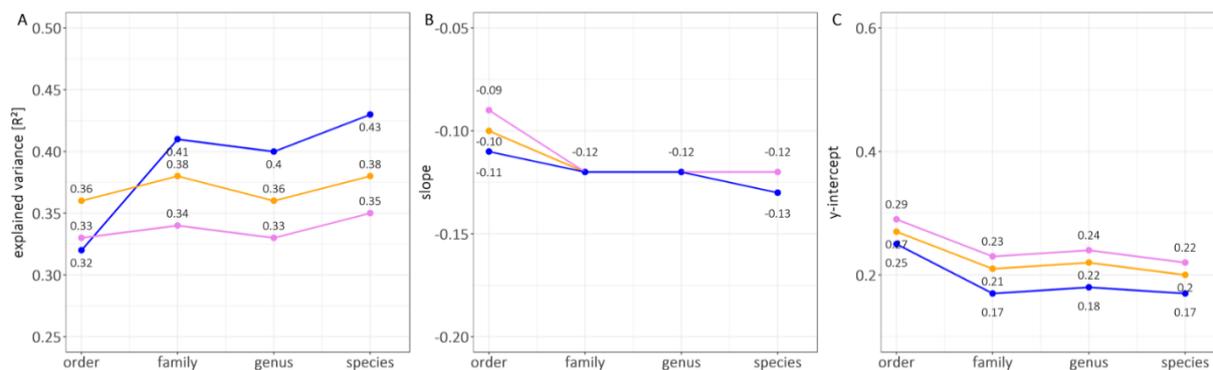
232

233 **Fig. 1 Overview of the multiple linear regression analysis. The relative importance of each of**
 234 **the five significant parameters for the different taxonomic and quantitative resolutions of**
 235 **SPEAR_{pesticides} is specified in explained variance by R² values. Black dots indicate the full models**
 236 **explained variance, whereas blue and red dots reflect the positive and negative influence of each**
 237 **stressor contributing to the full models' explained variance. Stars next to the R² indicate the level**
 238 **of significance (p < 0.05 = *, p < 0.01 = **, p < 0.001 = ***).**

239 With the multiple linear regression analysis, we confirmed the specificity of SPEAR_{pesticides}
 240 toward pesticide pressure at all quantitative resolutions up to the family level. In the following,
 241 we will focus on the indication power of SPEAR_{pesticides} at lower taxonomic and quantitative
 242 resolution. In a single linear regression analysis between SPEAR_{pesticides} and pesticide
 243 pressure, resulting R² values for all combinations of taxonomic and quantitative resolution
 244 ranged from 0.32 (order/presence-absence; Fig. 2 A) to 0.43 (species/abundance; Fig. 2 A).
 245 In general, R² values increased with increasing taxonomic and quantitative resolution, except

246 at the genus level (Fig. 2 A). We found only a slightly decreasing explanatory power (R^2) up to
247 the family level. With reducing the taxonomic resolution, the trait information basis changed.
248 At the genus level, genus sensitivity information was only available for 11% of genus taxa of
249 all sites, while 84% of genus taxa were linked at family level and 5% at order level. At species
250 level, only 17% of species taxa of all sites had sensitivity information. The majority of 80% had
251 sensitivity information at family level. Although the effect of pesticide pressure on the
252 invertebrate community could be linked up to the order level according to the R^2 values, in
253 some cases considerably deviating $SPEAR_{pesticides}$ values indicated a higher prediction
254 uncertainty at order level.

255 Order level lead to an increase of most $SPEAR_{pesticides}$ values. These higher $SPEAR_{pesticides}$
256 values at order level are not shifted in parallel (Fig. 2 A; SI). Streams with high ecological
257 quality deviated more than streams with lower ecological quality. Aggregating all higher
258 taxonomic trait information at order level is not feasible for every order, as many contradictory
259 higher trait information are combined. Some families classified as insensitive at the family level
260 (Hydropsychidae Gen. sp., Leptociridae Gen. sp., Sericostomidea Gen. sp., Ephemerellidae
261 Gen. sp., Leptophlebiidae Gen. sp., Nemouridae Gen. sp.) belong to the EPT group. The EPT
262 group is classified as sensitive at order level, thus increased the $SPEAR_{pesticides}$ values. In the
263 $SPEAR_{pesticides}$ database, 15 orders are implemented to which vulnerability information could
264 be assigned. Four orders are categorized as sensitive and 11 as insensitive towards
265 pesticides. To overcome the systematic shift of EPT rich streams at the order level, we derived
266 and applied a compensation factor of 1.2 for order/abundance and 1.3 for abundance class
267 and presence-absence resolution. This compensation factor approximated order level values
268 to that of species/abundance resolution to enable order level prediction (Fig. 2 B; SI). Despite
269 the compensation factor, many order-level $SPEAR_{pesticides}$ values still diverged from species-
270 level $SPEAR_{pesticides}$ values which can be explained by reduced numbers of taxa at order level.
271 Linearly adjusting the order-level $SPEAR_{pesticides}$ values only compensated the systematic shift
272 but did not allow to improve the prediction quality in terms of R^2 .



273
 274 **Fig. 2** SPEAR_{pesticides} as a function of maximum pesticide toxicity expressed (TU_{max}) for 101
 275 sampling sites. The three figures show the (A) explanatory power of the linear regression, (B)
 276 their respective slope, and (C) y-intercept. The points for each taxonomic level were connected
 277 as each level also contains some lower taxonomic resolutions.

278 3.1.2 Congruence and divergence between species and family SPEAR_{pesticides}

279 For most stream monitoring sites, species and family SPEAR_{pesticides} values were similar. With
 280 an explained variance of $R^2 = 0.41$, family/abundance resolution associate pesticide pressure
 281 and SPEAR_{pesticides} second accurately. Additionally, an identical slope and y-intercept of the
 282 linear regression at family level indicated a high congruence to species level regression
 283 SPEAR_{pesticides} values (Fig. 2 B, C). This congruence can be explained by three possible
 284 reasons. Firstly, trait information in the SPEAR_{pesticides} indicator at the species level is only
 285 available for one-third of our determined species, while most of the identified species are linked
 286 to family level trait information. The frequency of each taxa was not considered here. Secondly,
 287 if clear identification characteristics were absent due to damaging of individuals during the
 288 sampling process or too small to determine the taxa at species level, a lower taxonomic
 289 determination level was chosen to ensure the correct determination of the taxa. Thirdly, some
 290 trait information in the SPEAR_{pesticides} database was extrapolated from higher taxonomic levels
 291 to the species level due to missing information at the species level.
 292 These reasons are expected to explain the insignificant differences of the SPEAR_{pesticides} values
 293 between “species” and “family” level at the same quantitative resolution ($p = 0.44$, Wilcoxon
 294 test, Tab. 1; SI).

295 Still, some family $SPEAR_{pesticides}$ values deviated from species-level $SPEAR_{pesticides}$ values. For
296 instance, family values increased due to some insensitive taxa classified as sensitive at the
297 family level (Rhyacophilidae Gen. sp., Phryganeidae Gen. sp., Goeridae Gen. sp.,
298 Limnephelidae Gen. sp., Hydroptilidae Gen. sp.). Decreasing family $SPEAR_{pesticides}$ values can
299 be explained by sensitive taxa aggregated into insensitive families (Hydropsychidae Gen. sp.)
300 at lower taxonomic resolution. Additionally, if there are many species of the same family in one
301 stream, the $SPEAR_{pesticides}$ values at the family level decreased because the number of taxa
302 before log-transformation decreased.

303 In fact, family-level or even lower taxonomic levels are known to work with some
304 macroinvertebrate indices: the BMWP (Moolna et al., 2019), ASPT (Moolna et al., 2019), RETI
305 (Schmidt-Kloiber, A. & Nijboer, R. C., 2004), or Anglers' Riverfly Monitoring Initiative Index
306 (ARMI (Di Fiore and Fitch, 2016)). Taxonomic sufficiency (Ellis, 1985) is the taxonomic
307 resolution required for a specific endpoint to predict and judge impairment or change of an
308 ecosystem reliably. For macroinvertebrate indices, the level of adequate taxonomic resolution
309 depends on the determination depth of invertebrates but mainly on the available taxonomic
310 information in the index. Subsequently, depending on the invertebrate index and accordingly
311 the integrated taxonomic information, family level is sufficient to link multiple environmental
312 stressor to the macroinvertebrate community composition (SIGNAL, (Chessman, 1995)) or
313 show weak to no associations between environmental conditions and community composition
314 compared to the species level (SWAMPS, (Crowns et al., 1992)). Although species level has
315 the highest precision in reflecting the ecological status of streams (Norris and Hawkins, 2000;
316 Resh et al., 1995; Schmidt-Kloiber, A. & Nijboer, R. C., 2004) and is, thus, often treated as the
317 gold standard, similar to previous studies on other indicators, our analysis for $SPEAR_{pesticides}$
318 showed that family/abundance resolution can be used as a surrogate for species/abundance
319 resolution in terms of specificity and indication power.

320 **3.1.3 Effects of lower quantitative resolution on prediction quality**

321 The two lower quantitative resolutions analyzed – abundance class and presence-absence –
322 showed slightly less powerful associations between pesticide pressure and $SPEAR_{pesticides}$ of

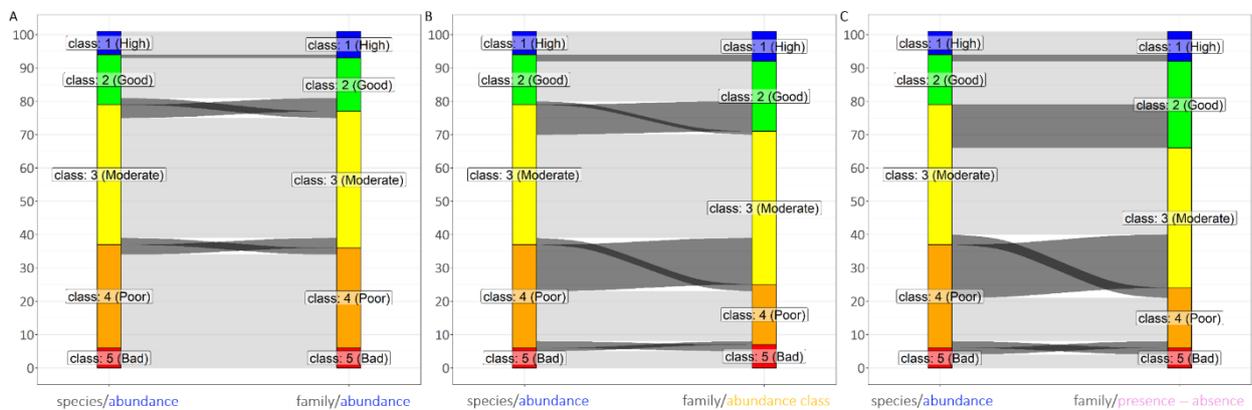
323 0.33 to 0.38 R^2 (Fig. 2 A) compared to abundance resolution. At lower quantitative resolutions,
324 the association between toxic pressure and $SPEAR_{\text{pesticides}}$ was observed to be relatively
325 independent from the taxonomic resolution. Within the same lower quantitative resolutions, the
326 explanatory power is similar for all taxonomic levels (Fig. 2 A). By converting abundances into
327 abundance classes or presence-absence data the relevance of the number of taxa increased
328 while the relevance of high abundances declined with a lower quantitative resolution. The
329 minor role of the abundance is reflected by the still high association (R^2) of $SPEAR_{\text{pesticides}}$ and
330 pesticide pressure at abundance class and presence-absence resolution, even at order level
331 (Fig. 2 A). Our correlation results at lower quantitative resolution showed for the first time that
332 abundance class and presence-absence resolution can also provide a powerful link between
333 pesticide exposure and $SPEAR_{\text{pesticides}}$.

334 **3.2 Indicating the ecological quality**

335 We analyzed how taxonomic and quantitative resolutions affect the assigned ecological quality
336 in terms of $SPEAR_{\text{pesticides}}$ classes. Lower taxonomic resolution only slightly affected the
337 ecological class assignments up to the family level. Family/abundance resolution yielded 88%
338 congruence in ecological class assignment compared to species/abundance resolution (Fig. 3
339 A). At the order level, $SPEAR_{\text{pesticides}}$ provided a distinctly reduced congruence of 56% in
340 ecological class assignment (Fig. 3; SI). In contrast, the AQUEM Assessment Software (ASS),
341 mainly focused on species taxonomic resolution, which only reached 50% congruence at the
342 genus level and 40% congruence at the family level in ecological classification (Schmidt-
343 Kloiber, A. & Nijboer, R. C., 2004).

344 Reducing the quantitative resolution affected ecological class assignment stronger compared
345 to a lower taxonomic resolution up to the family level. Species/abundance class resolution lead
346 to 74% congruence compared to species/abundance resolution. Species/presence-absence
347 resolution resulted in 65% congruence, the lowest accuracy at the species level in class
348 assignment with $SPEAR_{\text{pesticides}}$. Hence, at lower taxonomic and quantitative resolution, a class
349 shift of individual streams occurred. We found that a shift into a better ecological class was
350 more likely than into a worse class (see EXCEL workbook; SI). As indicated by the Aquem

351 Assessment Software applied at lower taxonomic resolution (ASS: Schmidt-Kloiber, A. &
 352 Nijboer, R. C. (2004)) and its General Degradation Metrics applied with presence-absence
 353 resolution (GDM: Buchner et al. (2019)), we also observed a class shift of up to two classes
 354 but in our case only at the order level (Fig. 3; SI). We observed that class shifts mainly occurred
 355 when the species/abundance resolution $SPEAR_{pesticides}$ value is close to an ecological class
 356 boundary. Therefore, we suggest that at lower taxonomic and quantitative resolution the
 357 assigned ecological class – but also the $SPEAR_{pesticides}$ value itself – should be considered for
 358 evaluation of individual streams.



359
 360

361 **Fig. 3 Comparison of ecological classification in terms of $SPEAR_{pesticides}$ under A)**
 362 **species/abundance and family/abundance resolution (88% congruence), B) species/abundance**
 363 **and family/abundance class resolution (69% congruence), and C) species abundance and**
 364 **family/presence-absence resolution (62% congruence). The ecological classes are displayed**
 365 **from “High” (blue) to “Bad” (red) and symbolize the ecological quality of invertebrate**
 366 **communities for each of the 101 streams (y-axis). The light grey horizontal bars show that the**
 367 **majority of streams are classified equally. The dark grey horizontal bars show streams shifting**
 368 **one class higher or lower due to lowering the taxonomic and quantitative resolution.**

369 **3.3 Implications for pesticide effect monitoring**

370 Large-scale monitoring implies working with multiple field personnel with different
 371 macroinvertebrate sampling skills, resulting in a higher likelihood of sampling errors that can
 372 occur when estimating macroinvertebrate abundances in the field (Metzeling et al., 2013). To
 373 avoid those sampling errors according to Giehl et al. (2014), many monitoring studies are
 374 based on presence-absence resolution, rather than abundance resolution. As an example,

375 Australian state-wide biomonitoring has adopted the use of family/presence-absence
376 resolution rather than species/abundance resolution (Marshall et al., 2002). Our study for
377 SPEAR_{pesticides} also revealed that abundance class resolution is a simplified approach for
378 differently skilled field personnel and less prone to sampling errors. This lower quantitative
379 resolution provides less accurate but still adequate indication of pesticide pressure and the
380 related ecological status. Whereas the trade-off for presence-absence resolution is higher as
381 the derived ecological status deviated in 35% of cases from the maximum resolution
382 assessment and should be restricted to monitoring where the scope is to only get an
383 impression of the analyzed invertebrate community composition. Reduced quantitative
384 resolution requires less material for the conservation of taxa and consequently the monetary
385 expenditure is lower (Bush et al., 2019; Marshall et al., 2006). Depending on the availability of
386 personnel, time, and material, the respective quantitative resolution should be chosen.
387 Accordingly, we suggest applying SPEAR_{pesticides} at family/abundance class resolution as a
388 good trade-off for large-scale monitoring. The deviating ecological classifications are minor
389 restrictions over the gained time efficiency and ensured determination validity in stream
390 monitoring. Classifying the sampled macroinvertebrates into abundance classes at the family
391 level in the field also leads to lower lethal effects in the community, as the proportion of
392 preserved sample for laboratory determination is reduced. Hence, already stressed but also
393 intact macroinvertebrate communities are protected, benefiting from lower taxonomic and
394 quantitative resolution. Taxonomically and quantitatively sufficient for SPEAR_{pesticides} is the
395 sampling and determination of macroinvertebrate data at family/abundance class resolution.

396 Conclusion

- 397 • Family/abundance class resolution provides the best trade-off for time-efficient large
398 scale pesticide effect monitoring and ecological stream assessment with
399 SPEAR_{pesticides}. It represents a simplified but reliable approach which is also applicable
400 for Citizen Science programs.

- 401 • The multiple stressor analysis has shown that at reduced quantitative resolution, the
402 $SPEAR_{pesticides}$ specificity decreased slightly while NH_4 also began to codetermine
403 $SPEAR_{pesticides}$.
- 404 • The derived compensation factors at order level enabled an approximation of the order
405 $SPEAR_{pesticides}$ values to the species level $SPEAR_{pesticides}$ values. Still, though, the
406 indication power is reduced at order/abundance resolution.
- 407 • Also order/presence-absence resolution can associate toxic pressure to the
408 invertebrate community composition, but reflecting the assigned ecological status only
409 half as accurate as species/abundance resolution.

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