

**This is the accepted manuscript version of the contribution published as:**

**Drechsler, M.** (2020):

Model-based integration of ecology and socio-economics for the management of biodiversity and ecosystem services: State of the art, diversity and current trends

*Environ. Modell. Softw.* **134** , art. 104892

**The publisher's version is available at:**

<http://dx.doi.org/10.1016/j.envsoft.2020.104892>

# Model-based integration of ecology and socio-economics for the management of biodiversity and ecosystem services: state of the art, diversity and current trends

Martin Drechsler, Helmholtz Centre for Environmental Research – UFZ, Department of Ecological Modelling, Permoserstr. 15, 04318 Leipzig; martin.drechsler@ufz.de.

## Abstract

Models are important tools to integrate ecological and socio-economic knowledge to better understand and manage social-ecological systems. Challenges include, among others, the adequate representation of feedback loops between the socio-economic and the ecological subsystems, uncertainties and human behaviour. To analyse how well models are able to address these challenges in the fields of biodiversity conservation and ecosystem services management the present paper systematically reviews recent mechanistic models of the field and analyses them with respect to a number of binary criteria. The reviewed models generally contain quite a few of the above-mentioned system features but still fall short when it comes to the adequate representation of the socio-economic dimension. Sorting the models by the labels given to them by their authors, such as “ecological-economic” or “system-dynamic”, allows assessing model variation within and among labels and indicates that there is a fruitful level of diversity in the models.

## Highlights

- Papers with models integrating ecological and socio-economic knowledge are reviewed
- Considered fields of application are biodiversity conservation and ecosystem services
- Models contain important features but generally miss important socio-economic issues
- There seems to be a fruitful level of diversity among the models in the field

## Key words

agent-based models, bio-economic models, ecological-economic models, social-ecological models, system-dynamic models.

## 37 **1 Introduction**

38 Despite various political and scientific initiatives (MA (2005), CBD (2007), IPBES: Díaz et al.  
39 (2018), TEEB (2010)), the loss of biodiversity and the degradation of ecosystem services is still  
40 ongoing worldwide (EEA 2019). Next to climate change, one of the main drivers of these losses is  
41 human land use (MA 2005), which in turn is driven by the socio-economic conditions, such as  
42 agricultural policy (e.g., Lakner et al 2019), under which the land users operate.

43

44 To understand the human-induced decline of biodiversity and ecosystem services, both the  
45 ecological and the socio-economic dimensions have to be considered in an integrated manner  
46 (Shogren et al. (1999), TEEB (2010), Wätzold et al. (2006)). A challenge here is that the coupled  
47 ecological and socio-economic processes are usually

- 48 • dynamic, partly involving feedbacks, so that, e.g., the ecological subsystem affects the  
49 socio-economic subsystem and vice versa (Polasky and Segerson 2009),
- 50 • spatially structured, so that, e.g., the distance between habitats affects the dispersal and the  
51 survival of a species (Hanski 1999), or the deforestation in the tropics leads to the  
52 fragmentation of rainforests (Taubert et al. 2018),
- 53 • subject to risk and uncertainty, since for instance the dynamics of species populations are  
54 often influenced by random factors (Beissinger et al. 2002), or systems may tip from one  
55 state to another where the location of the tipping point is fully or partly unknown (Bauch et  
56 al. (2016), and
- 57 • influenced by non-trivial behaviour of humans, where humans partly try to maximise their  
58 economic profits but in their decisions are also influenced by other motives (e.g.,  
59 Bartkowski and Bartke 2018).

60

61 Mathematical models have become important tools for the improvement and support of the  
62 understanding, prediction and management of complex systems, such as the world's climate system  
63 (e.g., Neelin 2010) or the dynamics of animal diseases (e.g., Vicente et al. 2019), which also holds  
64 for those cases in which several scientific disciplines, in particular natural sciences and social  
65 sciences, need to be considered in an integrated manner (Voinov (2008), Clark (2010), Flichman et  
66 al. (2011), Schlüter et al. (2012), Drechsler (2020)).

67

68 Purposes of models are manifold (Baumgärtner et al. 2008), probably the most important being a  
69 better understanding of systems and observed phenomena (e.g., Getzin et al. 2006, Taubert et al.  
70 2018), prediction of system dynamics (e.g., Bauch et al. 2016), and decision support (e.g., EFSA et  
71 al. 2017).

72

73 To increase their relevance for decision support, models should ideally consider that human land  
74 use is often influenced by agricultural, environmental and other policies (de Vries and Hanley 2016,  
75 Pe'er et al. 2017). A major obstacle to the development of policies is asymmetric information  
76 (Laffont and Mortimort 2002). Following the environmental-economics literature (e.g., Hanley et  
77 al. 2007), a case of asymmetric information is present when one actor has more information about  
78 some relevant facts than another (e.g., when landowners know their costs of carrying out a  
79 conservation measure, but a conservation agency that wishes to introduce a conservation policy  
80 does not).

81

82 The question arises, (Q1) to what extent current models used for the integration of ecology and  
83 socio-economics in the fields of biodiversity conservation and the management of ecosystem  
84 services, capture the above-mentioned issues of spatial, temporal and behavioural complexity and  
85 address the needs of deciders and policy makers.

86

87 To address this question, a quantitative analysis of research papers from the field is carried out,  
88 based on a systematic search from the past five years and focusing on mechanistic models. The  
89 papers are analysed with respect to a number of characteristics, including the purpose for which the  
90 models have been developed and analysed, structural features such as spatial structure or dynamics,  
91 issues of system complexity such as emergence, models of human behaviour, issues relevant in  
92 policy making, and the method by which the model is implemented and analysed.

93

94 With regard to agent-based models, a similar analysis has been carried out by Groeneveld et al.  
95 (2017) who focus on behaviour of land users, and Egli et al. (2019) who explore the consideration  
96 of resilience in agent-based models. However, while these two reviews address model features  
97 similar to those listed above, they consider different fields of model applications and, most  
98 importantly, consider only agent-based models.

99

100 Agent-based models, however, form only a limited subset of models integrating ecology and socio-  
101 economics in the fields of biodiversity conservation and the management of ecosystem services.  
102 Instead, next to agent-based, models in the field are known under very different labels, such as bio-  
103 economic or system-dynamic. Including these other labels on the one hand broadens the scope of  
104 the analysis and on the other hand raises additional research questions: (Q2) Do different labels  
105 reflect different model structures, possibly implying different abilities of the models to capture the  
106 features, and fulfil the purposes, described above? Further, how strongly are models of different  
107 labels related to each other and do the different labels indicate the presence of separate  
108 modelling communities with only limited communication among each other? To address these

109 questions, the reviewed models are sorted by the labels given to them, and the models found under  
110 each label are analysed separately and the results compared among the different labels.

111

112 The way models are built and labelled may change over time, and thus research question Q3 reads:  
113 What are the magnitudes and statistical significance of these changes? To address the three research  
114 questions Q1–Q3, relevant papers from the years 2015–2019 are identified and analysed as outlined  
115 above and described in detail below in the Methods. The Results section presents the outcomes of  
116 the analyses, and in the Discussion conclusions are drawn from these outcomes to formulate  
117 answers to the three research questions.

118

119

## 120 **2 Methods**

### 121 **2.1 Data collection**

122 The research context of the present analysis is mechanistic, or process-based, models that explicitly  
123 consider both the ecological and the socio-economic dimension of problems in the field of  
124 biodiversity, resource, and ecosystem services management. The basis of the quantitative literature  
125 review is thus a search with Scopus, performed in February 2020, for „TITLE-ABS-KEY

- 126 • ( ( model\* W/10 ( social\* OR socio\* OR economic\* ) )
- 127 • AND ( model\* W/10 ecological\* )
- 128 • AND ( biodiversity OR resource OR ( ecosystem W/5 service ) ) )
- 129 • AND ( EXCLUDE ( SUBJAREA , "MEDI" ) OR EXCLUDE ( SUBJAREA , "BIOC" )
- 130 OR EXCLUDE ( SUBJAREA , "ARTS" ) OR EXCLUDE ( SUBJAREA , "PSYC" ) OR
- 131 EXCLUDE ( SUBJAREA , "NURS" ) OR EXCLUDE ( SUBJAREA , "CENG" ) OR
- 132 EXCLUDE ( SUBJAREA , "NEUR" ) OR EXCLUDE ( SUBJAREA , "IMMU" ) OR
- 133 EXCLUDE ( SUBJAREA , "MATE" ) OR EXCLUDE ( SUBJAREA , "PHYS" ) OR
- 134 EXCLUDE ( SUBJAREA , "CHEM" ) OR EXCLUDE ( SUBJAREA , "HEAL" ) OR
- 135 EXCLUDE ( SUBJAREA , "PHAR" ) OR EXCLUDE ( SUBJAREA , "VETE" ) )

136

137 This search string was chosen to detect modelling papers considering (i) social, economic, or socio-  
138 economic dimension, (ii) an ecological dimension, and (iii) addressing issues of biodiversity,  
139 resources or ecosystem services (note that in Scopus the command W/n requires that the two  
140 connected keywords must not more than n words apart from each other). And to avoid an explosion  
141 of work load, thematic fields most likely to be irrelevant in the context of biodiversity were  
142 excluded. In total 1082 papers were identified, from which 73, 66, 85, 86 and 101 papers fell into  
143 the years 2015, 2016, 2017, 2018 and 2019, which form the basis of the present analysis.

144

145 Next, all the 411 papers were subjected to a rigorous manual screening process to avoid mis-  
146 classification of papers due to malfunctioning search filters (cf. Zare et al. 2017) and ensure that the  
147 remaining papers fall into the above-defined context. About twenty papers dropped out because they  
148 were not accessible (e.g., because they appeared in a journal of minor range), or not written in  
149 English language, and/or where the model description was too superficial (because, e.g., the paper  
150 is an extended abstract for a conference) to allow for the below-described detailed characterisation  
151 of the model. Then, a considerable proportion of papers dropped out whose models are of a  
152 qualitative/conceptual or statistical nature. Statistical models were excluded because these models  
153 are by their nature largely independent of the research context: a linear regression analysis, e.g.,  
154 simply correlates data and looks the same, regardless whether the data are from biological, social or  
155 astronomic observations (cf. Drechsler et al. 2007). Quite a few papers also dropped out, because  
156 other than suggested by their title, keywords or abstract (which were the targets of the Scopus  
157 search), they included only one of the two dimensions of ecology and socio-economics or even  
158 none of them. Lastly, models containing a mere aggregation of numbers into an index, such as  
159 ecological-footprint analysis or simple multi-criteria analysis (that did not include any further  
160 ecological and socio-economic modelling) were discarded.

161

## 162 **2.2 Selection of model characteristics and labels**

163 As a result, 122 papers (cf. Appendix A in the Supplementary Material) are retained which (or  
164 precisely, of course: the models in which) are characterised by the purpose for which the model was  
165 built, the structure of the model, particular issues addressed, and the formal implementation and  
166 analysis of the model. The selection of the criteria was influenced by Drechsler et al. (2007),  
167 Baumgärtner et al. (2008), Schlüter et al. (2012), Egli et al. (2019) and Drechsler (2020), and the  
168 criteria are chosen to explicitly address the issues raised in the Introduction:

169

170 Is the model

- 171 1. general?
- 172 2. specific?
- 173 3. built for positive analysis?
- 174 4. built for normative analysis?
- 175 5. used to improve the understanding of the modeled system or develop theory?
- 176 6. used to make predictions of the system dynamics?
- 177 7. used to include stakeholders and/or support management decisions?

178

179 Does the model explicitly consider

- 180 8. system dynamics?

181 9. spatial structure?  
182 10. randomness/stochasticity?  
183 11. feedback loops?  
184 12. individual agents?  
185 13. networks?  
186  
187 Regarding human actions, does the model address issues of  
188 14. asymmetric information?  
189 15. prediction and learning of actors?  
190 and is based on a  
191 16. utilitarian framework, in particular the human model of *homo oeconomicus*?  
192 17. or employing alternative, “non-classical” models of human behaviour?

193  
194 Does the model analysis address issues of  
195 18. non-linearity, such as resilience or discontinuous transitions/tipping points?  
196 19. emergence, in particular spatial pattern formation?  
197 20. policy interventions on the system?

198  
199 Is the model formulated through  
200 21. (differential or difference) equations?  
201 22. algorithms and rules?  
202 and is it solved or analysed  
203 23. analytically, or  
204 24. numerically/via computer simulation?

205  
206 And, how was the model labeled by its authors?

207  
208 Similar to Drechsler (2020), a general model is understood as a model that is not applied to a  
209 specific (geographically localised) case but formulated in a general manner by quantities like  
210 economic cost, population growth rate, etc. If a model is analysed numerically, it is counted as  
211 general if a systematic sensitivity analysis is carried out to explore the general behaviour of the  
212 model. A specific model, in contrast, is applied and parametrised to a specific case without any  
213 sensitivity analysis. Positive analysis is understood (cf. Drechsler 2020) as exploring how the  
214 system looks like or will develop, while normative analysis is understood as addressing how a  
215 system should be (managed) to maximise certain objectives like cost-effectiveness or social  
216 welfare.

217

218 Characteristic 7 is meant to capture participatory modelling approaches (e.g., Voinov et al. 2018),  
219 which includes research where stakeholders are not only sources of information (like in a survey) but  
220 actively and significantly involved in the research. Characteristic 12 contains multi-agent systems  
221 (Gilbert 2007) but also game theoretical approaches (Tadelis 2013; often with only two agents),  
222 while networks (characteristic 13) can, e.g., be ecological (Pascual and Dunne 2005) or social  
223 networks (Bruggeman 2008).

224

225 Asymmetric information (characteristic 14) was explained in the Introduction. Characteristic 16  
226 measures whether decisions are based on the maximisation of some utility function (as does, e.g. a  
227 welfare-maximising policy maker or the rational utility-maximising *homo oeconomicus*), while  
228 non-classical models (characteristic 17) may include altruism or inequity aversion in agents (e.g.,  
229 Fehr and Schmidt 1999), satisficing behaviour (Simon 1979) or social interaction such as copying  
230 opinions or decisions of other agents (Weisbuch 2005). Or the decisions may be modeled by  
231 statistical models (with the rest of the model being mechanistic as explained above) fitted to  
232 observed decisions (Lewis and Plantinga 2007), so that a utility function is, or cannot be, derived  
233 explicitly.

234

235 Most models in the sample include some non-linear elements. However, such non-linearity in the  
236 model structure (e.g., logistic growth of an animal population) does not necessarily lead to non-  
237 trivial model dynamics that can address issues of resilience or tipping points. Similarly, resource  
238 management models may contain a feedback loop in their formulation but may be analysed only to  
239 identify a monotonic optimal harvesting trajectory. Non-linearity in the context of the present paper  
240 (characteristic 18) does not refer to such non-linearities in the model *formulation* but only to the  
241 question of whether issues of non-linearity like resilience are addressed explicitly in the model  
242 *analysis*.

243

244 Characteristic 20 addresses, e.g., agricultural policies or the establishment of nature reserves. Some  
245 models contain policies as an exogenous driver but do not explicitly analyse the effects of these  
246 policies: these models are not classified as considering policy intervention. The last question above  
247 addresses the observation that models used in the present context have quite different labels, such as  
248 bio-economic or system-dynamic.

249

## 250 **2.3 Determination of the frequency distributions of the model characteristics (research** 251 **question Q1)**

252 Having analysed all 122 papers with regard to the 24 characteristics, the relative frequencies of the



models falling into each of these 24 categories are counted. Regarding the authors' label, many models have been given only an unspecific label like "simulation model" or "coupled model", and are classified as "not-specified". Although the label "network model" is specific, it is not particularly relevant to the issue of ecological and socio-economic integration, and so "network models" are also classified as "not-specified". For the five labels with the highest numbers of models the frequencies of the 24 characteristics are counted in the same way as for the set of all 122 models.

260

## 261 **2.4 Determination of relationships between model labels (question Q2)**

262 The counts of the model characteristics in the different model labels are used to identify  
263 characteristics typical for each model label to obtain a first understanding of the differences and  
264 similarities between the model labels. This is followed by a quantitative assessment of the  
265 relationships between the five model labels. For this, the frequency distributions are considered as  
266 vectors and Pearson correlation coefficients between these vectors are calculated. As an indicator of  
267 overlap and communication between different modelling communities, for each model label the  
268 journals are identified in which models of that label are published and the journals are identified  
269 that contain a particular number of model labels.

270

## 271 **2.5 Determination of trends in the consideration of model characteristics and model labels** 272 **(question Q3)**

273 Although five years of data is a rather short time frame to detect significant trends, two trend  
274 analyses are carried out using linear regression with time as the explaining variable. In the first  
275 analysis the explained variables are the frequencies of the model characteristics within the sample  
276 of all 122 models, while the second analysis considers the frequencies of the five model labels as  
277 explained variables.

278

## 279 **3 Results**

### 280 **3.1 Distribution of the model characteristics (research question Q1)**

281 The general distribution of the characteristics, based on all 122 models, is shown in Fig 1. About  
282 one third of the models is general while two thirds are specific (note that the sum of general and  
283 specific models is slightly above 100 %, since a few papers contain both a general and a specific  
284 analysis). Almost 80 % of the models were used for a positive analysis while about 30 % are  
285 normative (note that some papers contain both a positive and a normative analysis). About half of  
286 the models were analysed for system understanding and theory development, about two third were  
287 used for prediction, while only about five percent were developed and applied in a participatory  
288 approach.

290 Half of the models consider spatial structure, stochasticity and feedback loops; 80 % are dynamic,  
 291 but less than one third consider agents and less than ten percent consider networks. Asymmetric  
 292 information is considered only in less than five percent of all models, and the consideration of  
 293 prediction and learning is only slightly more abundant. About half of all models include some sort  
 294 of utilitarian approach of decision making, less than 20 % assume non-classical human behaviour,  
 295 while in the rest of the models modes of decision making or human behaviour are irrelevant or not  
 296 addressed. About one third of the models address issues of non-linearity, while only ten percent  
 297 consider emergence. Effects of policy intervention were analysed in about 30 % of the models.  
 298 Lastly, about one third of the models is equation-based while two thirds are algorithmic, and almost  
 299 all models were analysed numerically or by computer simulation.

300

### 301 **3.1 Distribution of the model characteristics for the dominant model labels (research question** 302 **Q2)**

303 As described in the Methods, an issue of interest is the label attached to the model by its author(s).  
 304 In the sample of 122 papers, 24 models were labeled “socio-ecological” or “social-ecological”  
 305 (SEM), 23 are “ecological-economic” (EEM), 14 “bio-economic” (BEM), 13 “agent-based” or  
 306 “multi-agent” (ABM), and eight are “system-dynamic” (SDM). Five models have two labels like  
 307 “agent-based social-ecological model” or “ecological-socio-economic model” and were sorted in  
 308 both respective classes. In addition, the labels “land use model” and “multi-objective model” are  
 309 observed three times each and two “game models” are in the sample. The labels “business model”,  
 310 “carbon cycle model”, “disease-economic model”, “growth model” and “socio-hydrological” occur  
 311 once each, and all other models are “not specified” as defined in the Methods.

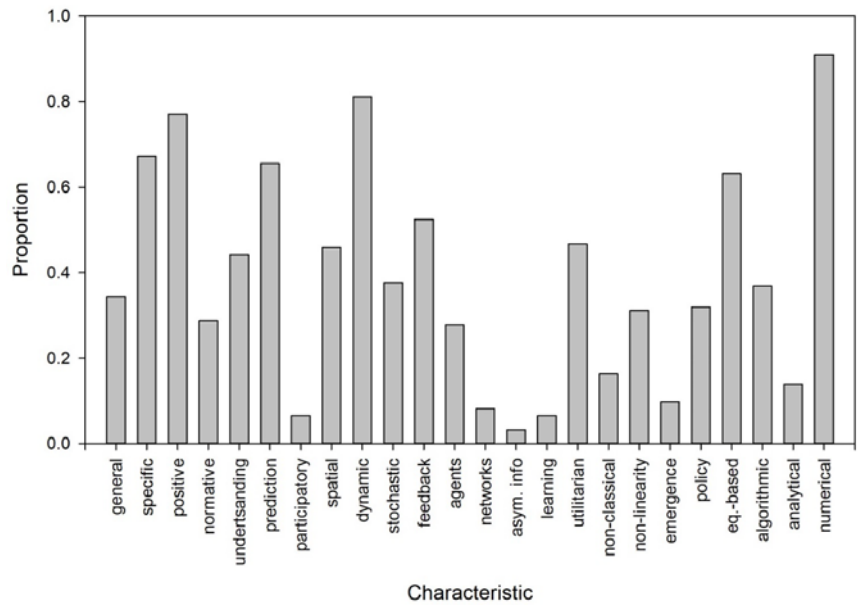
312

313 For the five most frequent labels, Fig. 2 shows the characteristics distributions, analogous to Fig. 1.  
 314 Frequencies between 0.25 and 0.75 (this choice is, if course, somewhat arbitrary) in a characteristic  
 315 indicate that both, presences and absences are quite abundant, so that this characteristic is neither  
 316 very typical nor very untypical for the considered model label. Larger frequencies  $\geq 0.75$ , in  
 317 contrast, indicate that the characteristic is typical for the model label while smaller frequencies  $\leq$   
 318 0.25 indicate that the characteristic is untypical. Table 1 summaries for each label the typical and  
 319 the untypical characteristics.

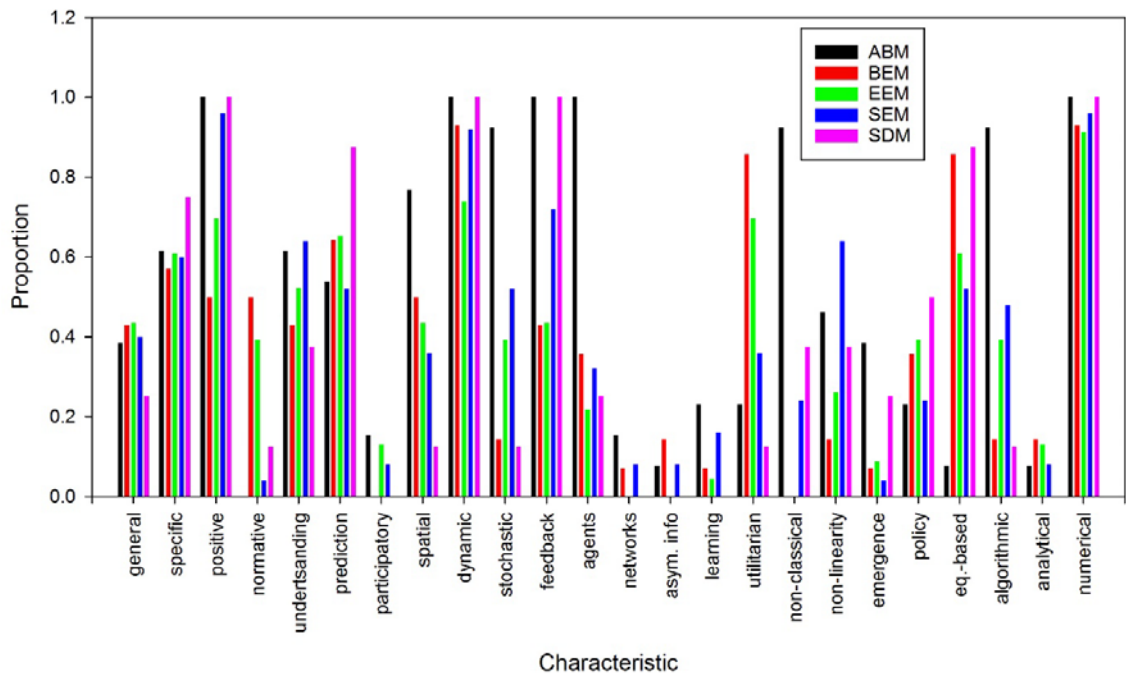
320

321 By these definitions of “typical” and “untypical”, participatory approaches or decision support are  
 322 untypical purposes for all model labels. Except for the EEM, all model labels are typically dynamic.  
 323 Typically, there is no consideration of networks, asymmetric information or learning. With the  
 324 exceptions of ABMs, consideration of emergence is untypical. And models of all labels are typically

325 analysed numerically.  
326



327  
328 Figure 1: Relative frequencies of the model characteristics over all 122 model studies (for the  
329 numerical values, see Appendix B in the Supplementary Material).  
330



331  
332 Figure 2: Relative frequencies of the model characteristics within the five most dominant model  
333 labels: agent-based models (ABM), bio-economic models (BEM), ecological-economic models  
334 (EEM), social-ecological models (SEM) and system-dynamic models (SDM) (for the numerical  
335 values, see Appendix B in the Supplementary Material).  
336

337 Beyond this, ABMs (recall: according to the authors' own labelling) are typically used for positive  
338 and not for normative analysis, contain many of the features like stochasticity and feedbacks, etc.,  
339 and consider non-classical human behaviour. Policy analysis is untypical, as well as an equation-  
340 based formulation. Instead the models are typically formulated (and solved) algorithmically.

341

342 BEMs are typically deterministic, employ a utilitarian framework like the human model of *homo*  
343 *oeconomicus* and are equation-based. They typically disregard non-linearities and are equation-  
344 based. Agent-based approaches, including the consideration of non-classical behaviour are untypical  
345 in EEMs. SEMs are very similar to the ABMs but in contrast – and like the EEMs – typically  
346 disregard non-classical behaviour and emergence.

347

348 SDMs are typically specific and not general, are used for positive rather than normative analysis  
349 and for prediction. They are typically non-spatial and deterministic but consider feedbacks. They do  
350 not consider agents, and since they are typically not used for normative analysis like cost-  
351 effectiveness analysis they also do not use a utilitarian framework. Like the BEMs they are typically  
352 equation-based.

353

354 If a model characteristic is found typical (or untypical) in only one model label then that model  
355 label may be regarded as distinctive in the considered characteristic (grey shaded cells in Table 1).  
356 By this definition, ABMs are distinctive by their consideration of spatial structure and stochasticity  
357 and, naturally, by their use of an agent-based approach. This allows them to distinctively consider  
358 non-classical behaviour and implies that they are formulated algorithmically rather than through  
359 mathematical equations. BEMs, in contrast, are distinctive by their use of a utilitarian framework  
360 and their formulation through equations. EEMs and SEMs are, by the present definition, not  
361 distinctive in any characteristic.

362

363 However, one should note that these outcomes, of course, depend on the choice of the thresholds  
364 (0.75 and 0.25) by which a characteristic is denoted as typical or untypical, respectively. The fact  
365 that SEMs, e.g., are not distinctive for feedback loops is because on the one hand the proportion of  
366 models with feedback loops in that label (0.72) is below the threshold of 0.75; and on the other  
367 hand, by the chosen definition, a model label is distinctive in a characteristic only if it is the *only*  
368 label in which that characteristic is typical. And because feedback loops are typical also for ABMs,  
369 if the threshold was lowered for instance to 0.7, feedback loops would be typical for both ABMs  
370 and SEMs – and none of both labels would be judged distinctive with respect to feedback loops.  
371 Therefore, if a label is not denoted as distinctive by a particular characteristic this does not imply

372 that this characteristic is rare in that label, but it means that the label does not stand out in this char-  
 373 acteristic relative to the other labels.

374

375 Table 1: Typical and untypical model characteristics for the five model labels: agent-based models  
 376 (ABM), bio-economic models (BEM), ecological-economic models (EEM), social-ecological  
 377 models (SEM) and system-dynamic models (SDM). The crosses mark the typical and untypical  
 378 characteristics in each model label and grey shaded areas mark characteristics that render the  
 379 respective model label distinctive (for further details, see text).

	ABM		BEM		EEM		SEM		SDM	
	typ.	untyp.	typ.	untyp.	typ.	untyp.	typ.	untyp.	typ.	untyp.
general										×
specific									×	
positive	×						×		×	
normative		×						×		×
understanding										
prediction									×	
participatory		×		×		×		×		×
spatial	×									×
dynamic	×		×				×		×	
stochastic	×			×						×
feedbacks	×								×	
agents	×					×				×
networks		×		×		×		×		×
asym. info		×		×		×		×		×
learning		×		×		×		×		×
utilitarian		×	×							×
non-classical	×			×		×		×		
non-linearity				×						
emergence				×		×		×		×
policy		×						×		
eq.-based		×	×						×	
algorithmic	×			×						×
analytical		×		×		×		×		×
numerical	×		×		×		×		×	

380

381

### 3.2 Relations between the models (question Q2 continued)

In the previous section the main similarities and differences between ABMs, BEMs, EEMs, SEMs and SDMs have been highlighted. According to Table 1, quite a lot of the 24 characteristics are either typical or untypical or neither of both in most or even all model labels. Only 12 instances were found where a model label is distinctive by a certain characteristic (grey-shaded cells in Table 1). The degree of overlap between the five model labels can be measured quantitatively by regarding the frequency distributions in Fig. 2 as vectors and calculating the Pearson's correlation coefficients between these vectors (Table 2).

The average correlation between a model label with the four respective others labels ranges from  $1.8/4 = 0.45$  (ABM) to 0.71 (EEM) and 0.74 (SEM). The highest pairwise correlation is between BEM and EEM (0.90), SEM and SDM (0.81) and SEM and EEM (0.80). These results identify quite a substantial overlap between the model labels, in particular between bio-economic, ecological-economic and social-ecological models. One may wonder whether this overlap also reflects in the journals in which the models were published. Comparing the diagonal cells of Table 3 with the off-diagonal cells reveals that for each model label the number of "unique" journals in which no other model label was observed (given in the diagonal cells of the Table) is about equal to the number of journals in which also other model labels were found (given in the off-diagonal cells).

Table 2: Correlations between the model labels agent-based (ABM), bio-economic (BEM), ecological-economic (EEM), social-ecological (SEM) and system-dynamics (SDM), based on the frequency distributions in Fig. 2.

	ABM	BEM	EEM	SEM	SDM	Row sum – 1
ABM	1.00	0.20	0.39	0.71	0.50	1.80
BEM	0.20	1.00	0.90	0.62	0.67	2.39
EEM	0.39	0.90	1.00	0.80	0.73	2.82
SEM	0.71	0.62	0.80	1.00	0.81	2.94
SDM	0.50	0.67	0.73	0.81	1.00	2.71

Although the smallness of the sample size demands some caution, this indicates that the authors of different model labels partly prefer different journals to publish their papers but that there is also a substantial overlap through "shared" journals. The broadest journals in that sense, that contain more than two model labels, are *Ecological Economics* (containing BEM, EEM, SEM and SDM), *Ecological Modelling* (containing ABM, BEM, EEM and SEM), *Journal of Environmental Management* (containing ABM, BEM, EEM, SEM), *Proceedings of the National Academy of Sciences of the USA*, PNAS (containing ABM, BEM, SEM and SDM) and *Land Use Policy* (containing BEM, EEM and SEM).

414

415 **3.3 Trends (question Q3)**

416 The first part of the trend analysis is to relate the number of papers represented by each model label  
 417 to the publication year (for the numerical values, see Appendix C in the Supplementary Material).  
 418 Table 4 shows the average annual changes and their statistical significance obtained by linear  
 419 regression. While BEMs, EEMs, and SDMs exhibit slight but highly insignificant declines, there is  
 420 a weakly significant increase of about one ABM paper per year and a strongly significant increase  
 421 of about two SEM papers per year. For comparison, for the years 2015–2019 the sum of the  
 422 numbers of ABMs, BEMs, EEMs, SEMs and SDMs equals 27, 20, 24, 20, 31, reflecting in a (highly  
 423 insignificant) average annual change of 0.8 (Table 4). Thus, SEMs increase significantly faster than  
 424 the average of all five model labels while ABMs seem to exhibit an average growth.

425

426 Table 3: Journals or conference proceedings in which the 122 model papers have been published,  
 427 sorted by the five model labels agent-based (ABM), bio-economic (BEM), ecological-economic  
 428 (EEM), social-ecological (SEM) and system-dynamic (SDM). The diagonal cells contain the  
 429 sources found only for the respective model label, while the off-diagonal cells contain the sources in  
 430 which papers of the two respective model labels were found. Journals in which four (three) model  
 431 labels were found are bold-faced (in italics).

	ABM	BEM	EEM	SEM	SDM
ABM	Ecol. Complex. Fishes Frontiers Ecol. Env. Frontiers Env. Sci. Proc. IEEE Conf.* Reg. Env. Change	<b>Ecol. Model.</b> <b>J. Env. Manag.</b> <b>PNAS</b>	<b>Ecol. Model.</b> <b>J. Env. Manag.</b>	Ecol. Appl. <b>Ecol. Model.</b> Env. Soft. <b>J. Env. Manag.</b> <b>PNAS</b>	<b>PNAS</b>
BEM		Aust. J. Agr. Res. Ec. Biol. Inv. J. Env. Econ. Manag. Nat. Res. Model.	Agricult. Syst. <b>Ecol. Econ.</b> <b>Ecol. Model.</b> <b>J. Env. Manag.</b> <i>Land Use Policy</i>	<b>Ecol. Econ.</b> <b>Ecol. Model.</b> <b>J. Env. Manag.</b> <i>Land Use Policy</i> <b>PNAS</b>	<b>Ecol. Econ.</b> IOP Earth & Env. Sc. <b>PNAS</b>
EEM			Adv. Syst. Sci. Appl. Aquaculture Biol. Cons. CEUR Works. Proc. Cons. Biol. Glob. Change Biol. IEEE Transact. Cyber. Syst. Res. Behav. Sci.	<b>Ecol. Econ.</b> <b>Ecol. Model.</b> Env. Res. Econ. Intl. J. Commons <b>J. Env. Manag.</b> <i>Land Use Policy</i>	<b>Ecol. Econ.</b>
SEM				Earth Syst. Dyn. Ecol. Appl. JASSS Marine Policy Ocean Coast. Manag. Sci. Total Env. Scientific Reports Sustainability (Switz.) Theor. Ecol. Urban Ecosystems	<b>Ecol. Econ.</b> <b>PNAS</b>
SDM					J. Cleaner Prod.

\*Proceedings of the IEEE Conference on Decision and Control

\*\*Proceedings – 2016 International Conference on Computational Intelligence and Applications, ICCIA 2016

Table 4: Average annual changes in the number of published papers within the time frame 2015–2019 for agent-based (ABM), bio-economic (BEM), ecological-economic (EEM), social-ecological (SEM) system-dynamic models (SDM), and all five model labels (All); and corresponding p-values from the underlying linear regression.

Model label	ABM	BEM	EEM	SEM	SDM	All
Annual change	0.8	−0.3	−0.1	1.8	−0.4	0.8
p-value	0.2	0.7	0.8	0.004	0.5	0.7

Analogously, Table 5 shows the trends in selected model characteristics (for the numerical values, see Appendix C in the Supplementary Material). Most characteristics exhibit insignificant trends with p-values above 0.2, while significant trends with p-values equal to or below 0.07 are observed only for the characteristics *general*, *stochastic* and *non-linearity*.

Table 5: Average annual changes in the number of published papers within the time frame 2015–2019 that contain a particular model characteristic, and corresponding p-values from the underlying linear regression. Only model characteristics with p-values  $\leq 0.2$  are shown.

Characteristic	Annual change	p-value
general	1.6	0.006
understanding	1.5	0.1
stochastic	1.5	0.07
agents	1.6	0.2
utilitarian	1.5	0.2
non-linearity	1.8	0.04
algorithmic	1.8	0.15
analytical	0.8	0.2

## 4 Discussion and Conclusions

### 4.1 Summary of the main results

122 papers dealing with the model-based integration of ecology and socio-economics in the fields of biodiversity conservation, ecosystem services and natural resource management from the years 2015–2019 were analysed with regard to 24 characteristics. The characteristics measure the purpose of the model study, structural features of the model, consideration of information, human behaviour, system complexity, environmental policy, as well as the formal implementation and analysis of the model.



459

460 In most model characteristics there is high variation among the models so that between 25 and 75  
461 percent of the models have that characteristic and the complementary percentage does not (cf. Fig.  
462 1). Rather frequently observed characteristics ( $> 75\%$ ) are: *positive analysis*, *dynamic*, and  
463 *numerical analysis*; while *networks*, *asymmetric information*, *learning in agents*, *non-classical*  
464 *agent behaviour* and *emergence* are considered only relatively rarely ( $< 25\%$ ); and only very few  
465 studies are *participatory*. These observations do not seem to change over time, as a trend analysis  
466 shows (cf. Table 5): significant increases are found only in the characteristics *general*, *stochastic*  
467 and *non-linearity* which are quite abundant anyway in the sample.

468

469 Some of the large variation in the first-mentioned characteristics with occurrences between 25 and  
470 75 percent is captured by the labels assigned to the models by the authors. The five most frequent  
471 labels found in the sample are “agent-based models” (ABM) (13), “bio-economic models” (BEM)  
472 (14), “ecological-economic models” (EEM) (23), “social-ecological models” (SEM) (25) and  
473 “system-dynamic models” (SDM) (8). In comparison to the other models, the ABMs have a high  
474 number of structural features and – next to the classical assumption of utility-maximising agents –  
475 also consider alternative models of human behaviour (although most of the ABMs do assume  
476 utility-maximising agents, confirming Groeneveld et al. (2017)) (cf. Table 1). Similar to  
477 observations made by Egli et al. (2019), emergence is also considered only in a minority of all  
478 ABMs in the sample. On the opposite end are the BEMs which contain only relatively few  
479 structural features and make use of a utilitarian framework such as the human model of *homo*  
480 *oeconomicus*. In between one can find the EEMs, the SEMs and the SDMs (which however also  
481 differ between each other, such that e.g., the SDMs rarely consider spatial structure, usually contain  
482 feedback loops and are often equation-based).

483

484 However, although there are differences between the models with different labels there are still  
485 many similarities, which can be seen in Fig. 2 and Table 1 – and is numerically confirmed by high  
486 pairwise correlation coefficients (taking the frequency distributions in Fig. 2 as vectors) (cf. Table  
487 2). While ABMs somewhat differ, the other four model labels are correlated quite strongly,  
488 especially BEM with EEM, SEM with SDM, and SEM with EEM, with correlation coefficients  $\geq$   
489 0.8.

490

491 That the ABMs are closely related to each other and somewhat different from the other model labels  
492 is also indicated by the results of a cluster analysis of the model characteristics. In Appendix D of  
493 the Supplementary Material, pairwise correlations between the model characteristics are calculated.  
494 Each characteristic is represented by a 122-element vector whose  $k$ -th element is one (zero) if the  $k$ -

th model in the sample of all models has (does not have) the characteristic. A high correlation between two characteristics  $i$  and  $j$  indicates that if characteristic  $i$  is present (absent) in a model then characteristic  $j$  is likely to be present (absent) in that model, as well. The two overlapping clusters D and E in Fig. D2 contain the characteristics *spatial*, *stochastic*, *algorithmic*, *agents* and *non-classical behaviour*, which are five of the typical characteristics of ABMs and exactly those five which render ABMs distinctive (Table 1). The other clusters in Fig. D2 cannot be mapped one-to-one to model labels. Clusters A and B contain many of those characteristics that are either typical or untypical for all 122 models (Fig. 1), while cluster C is difficult to relate and interpret.

The partial overlaps between the model labels indicated by Table 2 is also reflected in the journals in which the reviewed model papers have been published (Table 3). Although some journals were found to contain only one model label, a number of journals have published models of several labels – in particular *Ecological Economics*, *Ecological Modelling*, *Journal of Environmental Management* and PNAS which include four model labels each.

While there is a substantial overlap between the structure of the models within the different model labels, some differences exist with regard to temporal developments. While the number of published SEMs in the sample increased by about two per year, the numbers in the other model labels seem to remain rather constant (Table 4).

#### 4.2 Evaluation of the main results

Addressing research question Q1, there appears to be quite a good balance between general and specific models, between positive and normative analyses, and between the model purposes of system understanding and prediction; also features like dynamics, spatial structure, stochasticity, agents and feedback loops are quite often considered. On the downside, issues that are likely to be relevant and complicate policy making in the real, such as asymmetric information, non-rational behaviour of agents and social networks, are rarely addressed; and despite calls for more participation of stakeholders in research, models are still very rarely used in participatory research. As the trend analysis in Table 5 indicates, this does not seem to change over time (addressing research question Q3).

Addressing question Q2, the labels attached to the models by their author(s) partly indicate a typical structure of the model, so that agent-based models (ABMs) tend to differ from the models with other labels, in particular the system-dynamic (SDM) and bio-economic models (BEMs). Explaining these differences between the models with different labels is clearly beyond the scope of the present paper. One explanation could be the disciplinary background (even though the 122

models were selected for interdisciplinarity) of the model authors. Although with different characteristics, Drechsler et al. (2007) and Drechsler (2020) compared in a similar way ecological, economic and ecological-economic models from the field of biodiversity conservation and found that the ecological models tended to contain more structural features and were more complex than the economic models.

On the other, despite these differences there is quite some overlap between the model labels, indicated by the correlation coefficients in Table 2, which reflect that, e.g., the models have similar purposes, share some structural features like dynamics, and share deficiencies like the wide absence of participatory applications. The overlap between the models also reflects in the observation that various journals include models of several labels. These observation can altogether be regarded as a positive signal, such that there seems to be enough diversity in the modelling cultures to avoid “inbreeding” but enough similarities to allow for communication and cross-fertilisation among different modelling communities.

A problematic observation, though, is that the number of papers with the present type of mechanistic models, designed to integrate ecology and socio-economics for the management of biodiversity and ecosystem services, grows only slowly (addressing research question Q3; cf. Table 4). Qualitatively this is confirmed by another Scopus search (not shown), carried out in June 2020, with the same search string as in section 2.1 and adding in turn the terms “agent-based”, “bio-economic”, “ecological-economic”, “socio-ecological” OR “social-ecological” and “system-dynamic”. Although the tedious manual screening described in section 2 was excluded for simplicity (so the two searches can be compared only loosely), the qualitative result is the same such that, except for SEM, the number of papers seems to remain constant over the past decade; while the number of SEM papers has increased from about 10 in 2010 to about 50 in 2019.

#### **4.3 Limitations of the analysis and future research**

The present study involves a number of limitations. The literature search via Scopus (section 2.1) will certainly not have found all the papers of the considered field. This is indicated, e.g., by the fact that the same search a few months later led to slightly different results. Further, the manual screening of the obtained papers may have not been free of any errors. In addition, the choice of the search terms is subjective and may have led to the omission of relevant articles. For instances, it did not yield any papers with *integrated models* that also contain applications in the field of environmental management.

Related to this is the fact that the selection of the model characteristics (section 2.2) is subjective,

567 too. Smaller errors may have also occurred in the evaluation of the models with regard to the model  
568 characteristics (section 2.3). Although it is used only as a supplement to the correlation analyses  
569 described in section 2.4, the manual cluster analysis in Appendix D of the Supplementary Material  
570 involves subjective elements. Lastly, for time constraints only five years of data were considered  
571 which limits the significance of the results of the trend analysis (section 2.5).

572

573 While the errors in sections 2.2–2.4 are probably minor, the errors in the generation of the data base  
574 in section 2.1 and the trend analysis in section 2.5 call for further research. The use of alternative –  
575 possibly broader – search terms and a longer time frame could lead to interesting insights beyond  
576 those of the present study.

577

#### 578 **4.4 Conclusion**

579 To conclude, mechanistic models integrating ecology and socio-economics for the management of  
580 biodiversity and ecosystem services by now consider quite a lot of system features such as spatial  
581 structure, agents and feedback loops, but are still largely (to varying degrees, depending on the  
582 model label) disregarding quite a number of issues such as the distribution and acquisition of  
583 information among and by the human actors, as well as the way in which the human actors translate  
584 available information into decisions. The human factor is also under-represented in the sense that  
585 only very few models are used within participatory studies. Nevertheless, there seems to be a  
586 fruitful level of diversity in modelling cultures but also sufficient overlap for cross-fertilisation,  
587 which may help further improving the models in the field.

588

#### 589 **Acknowledgments**

590 I am very grateful for constructive comments of three anonymous reviewers and the journal editor,  
591 Sondoss Elsayah, that were of substantial help in the revision of this manuscript.

592

#### 593 **References**

- 594 Bartkowski, B., Bartke, S., 2018. Leverage points for governing agricultural soils: A review of  
595 empirical studies of European Farmers' decision-making. *Sustainability* 10 (9), 3179.
- 596 Bauch, C.T., Sigdel, R., Pharaon, J., Anand, M., 2016. Early warning signals of regime shifts in  
597 coupled human-environment systems. *Proceedings of the National Academy of Sciences of the*  
598 *United States of America* 131 (51), 14560–14567.
- 599 Beissinger, S.R., McCullough, D.R. (eds.), 2002. *Population Viability Analysis*. University of  
600 Chicago Press.
- 601 Bruggeman, J., 2008. *Social Networks: An Introduction*. Routledge.

602 CBD, 2007. *CBD Handbook. Convention on Biological Diversity*.  
 603 <https://www.cbd.int/convention/refrhandbook.shtml>.

604 Clark, C.W., 2010. *Mathematical Bioeconomics: The Mathematics of Conservation*, 3rd Edition.  
 605 Wiley.

606 de Vries, F.P., Hanley, N., 2016. Incentive-based policy design for pollution control and biodiversity  
 607 conservation: A review. *Environmental and Resource Economics* 63, 687–702.

608 Díaz, S., Pascual, U., Stenseke, M., et al., 2018. Assessing nature's contributions to people. *Science*  
 609 359(6373), 270–272.

610 Drechsler, M., 2020. *Ecological-economic Modelling for Biodiversity Conservation*. Cambridge  
 611 University Press.

612 Drechsler, M., Grimm, V., Mysiak, J., Wätzold, F., 2007. Differences and similarities between  
 613 ecological and economic models for biodiversity conservation. *Ecological Economics* 62(2), 232–  
 614 241.

615 EFSA (European Food Safety Authority), Depner K., Gortazar C., Guberti V., Masiulis M., More S.,  
 616 Oļševskis E., Thulke H.-H., Viltrop A., Woźniakowski G., Cortiñas Abrahantes J., Gogin A.,  
 617 Verdonck F., Dhollander S., 2017. Scientific Report on the epidemiological analyses of African  
 618 swine fever in the Baltic States and Poland. *EFSA Journal* 15 (11), 5068, 59 pp.

619 Egli, L., Weise, H., Radchuk, V., Seppelt, R., Grimm, V., 2019. Exploring resilience with agent-  
 620 based models: State of the art, knowledge gaps and recommendations for coping with  
 621 multidimensionality. *Ecological Complexity* 40, Part B, art. 100718.

622 European Environment Agency (EEA), 2019. *The European environment – state and outlook 2020*  
 623 *(SOER). Knowledge for transition to a sustainable Europe*. Luxembourg: Publications Office of the  
 624 European Union.

625 Fehr, E., Schmidt, K.M., 1999. A theory fairness, competition, and cooperation. *Quarterly Journal*  
 626 *of Economics* 114, 817–868.

627 Flichman, G. (Ed.), 2011. *Bio-Economic Models applied to Agricultural Systems*. Springer.

628 Getzin, S., Yizhaq, H., Bell, B., Erickson, T.E., Postle, A.C., Katra, I., Tzuk, O., Zelnik, Y.R.,  
 629 Wiegand, K., Wiegand, T., Meron, E., 2016. Discovery of fairy circles in Australia supports self-  
 630 organization theory. *Proceedings of the National Academy of Sciences of the USA* 113, 3551–3556

631 Gilbert, N., 2007. *Agent-based Models*. SAGE Publications.

632 Groeneveld, J., Müller, B., Buchmann, C.M., Dressler, G., Guo, C., Hase, N., Hoffman, F., John, F.,  
 633 Klassert, C., Lauf, T., Liebelt, V., Nolzen, H., Pannicke, N., Schulze, J., Weise, H., Schwarz, N.,  
 634 2017. Theoretical foundations of human decision-making in agent-based land-use models.  
 635 *Environmental Modelling and Software* 87, 39–48.

636 Hanley, N., Shogren, J.F., White, B., 2007. *Environmental Economics: In Theory and Practice*, 2<sup>nd</sup>  
 637 ed. Palgrave MacMillan

- 638 Hanski, I., 1999. *Metapopulation Ecology*. Oxford University Press.
- 639 Harvey, J.A., Heinen, R., Armbrrecht, I., et al., 2020. International scientists formulate a roadmap for  
640 insect conservation and recovery. *Nature Ecology & Evolution* 4, 174–176.
- 641 Laffont, J.-J., Mortimort, D., 2002. *The Theory of Incentives: The Principal-Agent Model*. Princeton  
642 University Press.
- 643 Lakner, S., Holst, C., Dittrich, A., Hoyer, C., Pe'er, G., 2019. Impacts of the EU's Common  
644 Agricultural Policy on Biodiversity and Ecosystem Services. In: Schröter, M., Bonn, A., Klotz, S.,  
645 Seppelt, R., Baessler, C. (eds.) *Atlas of Ecosystem Services*, pp. 383–389. Springer.
- 646 Lewis, D.J., Plantinga, A.J., 2007. Policies for habitat fragmentation: combining econometrics with  
647 GIS-based landscape simulations. *Land Economics* 83, 109–127.
- 648 Millennium Ecosystem Assessment, 2005. *Ecosystems and Human Well-being: Biodiversity  
649 Synthesis*. World Resources Institute.
- 650 Neelin, J.D., 2010. *Climate Change and Climate Modelling*. Cambridge University Press.
- 651 Pascual, M., Dunne, J.A., 2005. *Ecological Networks: Linking Structure to Dynamics in Food Webs*.  
652 Santa Fe Institute Studies on the Sciences of Complexity. Oxford University Press.
- 653 Pe'er, G., Zinngrebe, Y., Hauck, J., Schindler, S., Dittrich, A., Zingg, S., Tschardtke, T., Oppermann,  
654 R., Sutcliffe, L.M.E., Sirami, C., Schmidt, J., Hoyer, C., Schleyer, C., Lakner, S., 2017. Adding  
655 some green to the greening: improving the EU's ecological focus areas for biodiversity and farmers.  
656 *Conservation Letters* 10 (5), 517–530.
- 657 Polasky, S., Segerson, K., 2009. Integrating ecology and economics in the study of ecosystem  
658 services: some lessons learned. *Annual Review of Resource Economics* 1 (1), 409–434.
- 659 Railsback, S., Grimm, V., 2019. *Agent-Based and Individual-Based Modeling*, 2<sup>nd</sup> ed. Princeton  
660 University Press.
- 661 Schlüter, M., McAllister, R.R.J., Arlinghaus, R., 2012. New horizons for managing the  
662 environment: a review of coupled social-ecological systems modelling. *Natural Resource Modelling*  
663 25 (1), 219–272.
- 664 Shogren, J.F., Tschirhart, J., Anderson, T., 1999. Why economics matters for endangered species  
665 protection. *Conservation Biology* 13 (6), 1257–1261.
- 666 Simon, H.A., 1979. Rational thinking in business organizations. *American Economic Review* 69,  
667 493–513.
- 668 Tadelis, S., 2013. *Game Theory: An Introduction*. Princeton University Press.
- 669 Taubert, F., Fischer, R., Groeneveld J., et al., 2018. Global patterns of tropical forest fragmentation.  
670 *Nature* 554, 519–22.
- 671 TEEB 2010. *The Economics of Ecosystems and Biodiversity: Mainstreaming the Economics of  
672 Nature: A synthesis of the approach, conclusions and recommendations of TEEB*.

673 [www.teebweb.org/our-publications/teeb-study-reports/synthesis-report/#.Ujr2cX9mOG8](http://www.teebweb.org/our-publications/teeb-study-reports/synthesis-report/#.Ujr2cX9mOG8) (last  
674 accessed 10 October 2019).

675 Vicente J., Apollonio M., Blanco-Aguiar J.A., Borowik T., Brivio F., Casaer J., Croft S., Ericsson  
676 G., Ferroglio E., Gavier-Widen D., Gortázar C., Jansen P.A., Keuling O., Kowalczyk R., Petrovic  
677 K., Plhal R., Podgórski T., Sange M., Scandura M., Schmidt K., Smith G.C., Soriguer R., Thulke  
678 H.-H., Zanet S., Acevedo P., (2019). Science-based wildlife disease response. *Science* 364 (6444),  
679 943–944.

680 Voinov, A.A., 2008. *Systems Science and Modeling for Ecological Economics*. Academic Press.

681 Voinov, A., Jenni, K., Gray, S., et al., 2018. Tools and methods in participatory modeling: Selecting  
682 the right tool for the job. *Environmental Modelling and Software* 109, 232–255.

683 Wätzold, F., Drechsler, M., Armstrong, C.W., Baumgärtner, S., Grimm, V., Huth, A., Perrings, C.,  
684 Possingham, H.P., Shogren, J.F., Skonhøft, A., Verboom-Vasiljev, J., Wissel, C. 2006. Ecological-  
685 economic modeling for biodiversity management: Potential, pitfalls, and prospects. *Conservation*  
686 *Biology* 20, 1034–41.

687 Weisbuch, G., 2006. Social opinion dynamics. In: Chakraborti, B.K., Chakraborti, A., Chatterjee, A.,  
688 (eds.) *Econophysics and Sociophysics: Trends and Perspectives*. Wiley, pp. 339–336.

689 Zare, F., Elsawah, S., Iwanaga, T., Jakeman, A.J., Pierce, S.A., 2017. Integrated water assessment  
690 and modelling: A bibliometric analysis of trends in the water resource sector. *Journal of Hydrology*  
691 552, 765–778.