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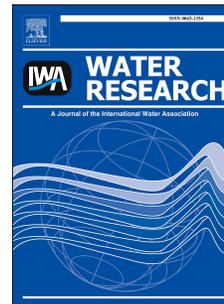
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Dissolved organic matter in a tropical saline-alkaline lake of the East African Rift Valley.

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Dissolved organic matter in a tropical saline-alkaline lake of the East African Rift Valley.**Authors and affiliations**

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27 **Abstract**

28 Saline-alkaline lakes of the East African Rift are known to have an extremely high primary
29 production supporting a potent carbon cycle. To date, a full description of carbon pools in these
30 lakes is still missing. More specifically, there is not detailed information on the quality of
31 dissolved organic matter (DOM), the main carbon energy source for heterotrophs prokaryotes.
32 We report the first exhaustive description of DOM molecular properties in the water column of
33 a meromictic saline-alkaline lake of the East African Rift. DOM availability, fate and origin
34 were studied either quantitatively, in terms of dissolved organic carbon (DOC) and nitrogen
35 (DON) or qualitatively, in terms of optical properties (absorbance) and molecular
36 characterization of solid-phase extracted DOM (SPE-DOM) through negative electrospray
37 ionization Fourier Transform Ion Cyclotron Resonance Mass Spectrometry (FT-ICR-MS).
38 DOM availability was high (DOC ~ 8.1 mM in surface waters) and meromixis imprinted a
39 severe quantitative and qualitative change on DOM pool. At the surface, DOM was rich in
40 aliphatic and moderately in aromatic molecules and thus mirroring autochthonous microbial
41 production together to photodegradation. At the bottom changes were extreme: DOC increased
42 up to 5 times (up to 50 mM) and, molecular signature drifted to saturated, reduced and non-
43 aromatic DOM suggesting intense microbial activity within organic sediments. At the
44 chemocline, DOC was retained indicating that this interface is a highly reactive layer in terms
45 of DOM processing. These findings underline that saline-alkaline lakes of the East African Rift
46 are carbon processing hot spots and their investigation may broaden our understanding of
47 carbon cycling in inland waters at large.

48

49 **Key words:** Dissolved organic carbon, African soda lake, Meromixis, FT-ICR-MS, CDOM

50

51

52 **1. Introduction**

53 Endorheic saline-alkaline lakes are known from all continents. In terms of water volume and
54 surface, they represent 40% and 20% of total lakes, respectively (Meybeck, 1995; Dodds and
55 Whiles, 2002; Wurtsbaugh et al., 2017). The saline-alkaline lakes of the East African Rift
56 (thereafter SLARs) are among the most distinctive in the world. SLARs contribute up to ~75%
57 of total lake area in this region and cover a total area of ~13000 km² with single lake areas
58 ranging from 8800 km² (Lake Turkana) down to tens of crater lakes smaller than 0.5 km².
59 Their pH and alkalinity are typically in excess of 9.5 and 50 meq L⁻¹, respectively (Hecky and
60 Kilham, 1973). A striking feature is a daily pelagic primary production of over than 10 g O₂ m⁻²
61 (Melack, 1982; Oduor and Schagerl, 2007), which is significantly higher than that recorded in
62 freshwater ecosystems (Lewis, 2011). This high productivity is sustained by planktonic
63 cyanobacteria (up to 10⁷-10⁸ cells ml⁻¹, Zinabu and Taylor, 1997) that proliferate owing to
64 constant high water temperatures (> 20 °C, the “endless summer” according to Kilham and
65 Kilham, 1989), intense and almost constant solar radiation throughout the year at Equator
66 (~173 W m⁻²; Odongo et al., 2016), and very high availability of dissolved inorganic carbon
67 (Talling et al., 1973).

68 Overall, SLARs showed a very dynamic carbon cycle, yet a clear picture of carbon availability
69 and fluxes is still lacking. In particular, information about dissolved organic matter (DOM) and
70 its qualitative properties is largely disregarded and incompletely understood (Jirsa et al., 2013).
71 DOM is a complex mixture of highly diverse molecules that differ in composition, structure,
72 origin, residence time and bioavailability (Dittmar and Stubbins, 2014). Besides being the
73 indispensable source of carbon and energy for the aquatic microbiota, DOM interacts with light
74 by attenuating and modifying the underwater radiation spectra (Thrane et al., 2014) and is
75 adsorbed to minerals or desorbed after reductive dissolution of iron minerals (Elkins and
76 Nelson, 2002).

77 The objective of this study is to provide a first exhaustive description of the DOM quantity and
78 molecular properties in a SLAR. Having in mind that, at the Equator, the seasonal changes are
79 reduced (Odongo et al., 2016), this study specifically focused on a meromictic lake with severe
80 biogeochemical gradients along the water column (Melack, Kilham and Fisher, 1982;
81 MacIntyre and Melack, 1982). In particular we aimed to provide a better understanding of the
82 impact of (i) water stratification, (ii) pelagic primary production and photodegradation. DOM
83 was analyzed to (a) quantify it in terms of dissolved organic carbon (DOC) and nitrogen (DON)
84 (b) describe its spectroscopic properties through absorbance, and (c) characterize its molecular
85 property through Fourier Transform- Ion Cyclotron Resonance Mass Spectroscopy (FT-ICR-
86 MS).

87 High primary production together with high potential evaporation is expected to determine
88 large DOC concentrations in the water column. High density of pelagic cyanobacteria and high
89 and constant incident solar radiation are likely to strongly affect DOM quality as well.
90 Cyanobacteria release aliphatic compounds (Bittar et al., 2015) as a consequence, saturated
91 ($H/C > 1.5$) and relative oxygen-poor ($O/C < 0.5$) molecules might be significant. Simultaneously,
92 aromatic compounds are expected to be low as consequence of photodegradation induced by
93 elevated incident radiation.

94

95 **2. Material and methods**

96 **2.1. Study site and sampling strategy**

97 Lake Sonachi (90 km NW of Nairobi at 1,884 m a.s.l.; $0^{\circ}46'57.68''S$; $36^{\circ}16''E$) is a small (0.18
98 km^2) meromictic saline-alkaline lake located in a volcanic crater (0.84 km^2) surrounded by
99 riparian vegetation (mainly *Vachellia xanthophloea*) (MacIntyre and Melack, 1982) and
100 adjacent to the larger freshwater Lake Naivasha (c.a. 150 km^2) (Figure SI 1). Climate is warm

101 and semiarid, with annual potential evaporation (ca. 1800 mm/year Olaka, 2011) exceeding
102 annual precipitation (ca. 640 mm/year www.climatedata.eu).

103 Groundwater seepage from Naivasha is likely to be the main water input, whereas evaporation
104 is the outflow (Gaudet and Melack, 1981). A steep crater rim protects the lake from winds and
105 the water column mixing is limited (MacIntyre and Melack, 1982).

106 Water column depth varied from 3 m to 18 m with generally stable meromictic conditions.
107 Holomixis episodes were reported only during low water level periods (MacIntyre and Melack,
108 1982; Verschuren, 1999). Primary production ranged from 2.2 to 56 mM O₂ m⁻³ h⁻¹ and
109 chlorophyll concentration in the euphotic zone reached peaks of 80 µg L⁻¹ (Melack, 1982;
110 Melack et al., 1982).

111 Pelagic biomass (wet weight) reached up to 3 g L⁻¹ during a bloom of *Arthrospira fusiformis*
112 (Ballot et al., 2005). Nutrient concentrations (SRP, NH₄⁺+NH₃, NO₂⁻ and NO₃⁻) are typically
113 lower than 0.4 µM in the mixolimnion, but can rise up at the sediment surface (up to 130 µM
114 and 4000 µM for SRP and NH₄⁺+NH₃ respectively) (Melack et al., 1982). Chemocline and
115 monimolimnion are anoxic and euxinic, with concentrations of hydrogen sulfide (H₂S) up to
116 1.2 mM (McIntyre and Melack, 1982).

117 Verschuren (1999) reported that the uppermost sediment layer, in the 30-50%, consists of
118 highly porous gelatinous organic aggregates of settled decomposing algae and cyanobacteria
119 and that organic carbon represents the 10% of total sediment dry weight (unpublished data from
120 authors).

121 Sampling was performed on June 18, 2017, in the middle of the lake, at the deepest site. One
122 litre of water was sampled at midday from seven depths: 0, -0.5, -1, -2, -3, -4 and -4.5 m with a
123 Rilsan[®] tube connected to a syringe with a three-way Teflon valve (Tassi et al., 2004). Sample
124 at 4.5 m depth was collected at the bottom lake at the interface between sediments and water.

125

126 **2.2. Lake characterization and underwater light attenuation**

127 Water temperature, electrical conductivity, oxidation/reduction potential (ORP), pH and
128 dissolved oxygen profiles were recorded between 3 and 4 p.m. at every 0.5 m, using a YSI
129 Multiparameter probe. ORP was converted into redox potential at 25°C ($E_{0(25^{\circ}\text{C})}$) according to
130 Wolkersdorfe (2008).

131 Adsorption spectra between 280 and 1000 nm of unfiltered samples were used to estimate
132 underwater light attenuation spectra at different depths. Underwater light spectra were obtained
133 according to the Beer-Lambert law:

$$134 \quad I_{\lambda} = I_{0(\lambda)} e^{-z \tau_{\lambda}} \quad (1)$$

135 Where $I_{0(\lambda)}$ is the input solar intensity at ground level at wavelength λ . τ_{λ} is optical depth at
136 wavelength λ ($\tau_{\lambda} = \ln(10) \text{ Abs}_{\lambda}$). Abs_{λ} is the absorbance values at wavelength λ . Spectra were
137 performed with a UV1700 Pharma Spec (Shimadzu) spectrophotometer using a 1-cm quartz
138 cell varying λ between 280 and 1000 nm. The ASTM G173-03 Reference was used as input
139 solar spectrum (<http://rredc.nrel.gov/solar/spectra>).

140

141 **2.3. Nutrients, bulk DOM and chromophoric DOM**

142 Water samples (100 mL) were filtered (0.45- μm pore-size nylon filters) and acidified with
143 ultrapure HCl for dissolved inorganic nitrogen ($\text{DIN} = \text{NH}_4^+ + \text{NH}_3 + \text{NO}_2^- + \text{NO}_3^-$). NO_2^- and NO_3^-
144 were analyzed with colorimetric method with an autoanalyzer Bran+Luebbe after nitrate
145 reduction with a copper-cadmium column. $\text{NH}_4^+ + \text{NH}_3$ concentrations were estimated with the
146 salicylate method of Reardon (1969). To remove silicate interference, the pH of solution was
147 set to 1. To reduce the interference of salts and sulfite, all samples were previously diluted to
148 1/10 -1/30.

149 Dissolved organic carbon (DOC) and total dissolved nitrogen (TDN) were measured by
150 oxidative combustion and infrared analysis using a Shimadzu 5000 A total organic carbon

151 analyzer coupled to a Total Nitrogen unit. Dissolved organic nitrogen (DON) was calculated as
152 the difference between TDN and DIN.

153 For DOM quality analysis, water samples (200 mL) were filtered (combusted GFF Whatman
154 filters), acidified with ultrapure HCl (pH =2), and stored in pre-combusted glass bottles.

155 Samples for spectroscopy were filtered just before their analysis with 0.22- μ m-pore nylon
156 membranes to remove particle interferences and, were analyzed at room temperature. Prior any
157 analysis, samples were diluted to 1/25-1/60 (to get a final DOC concentration of 2-4 mg L⁻¹) to
158 attenuate the interference of salinity. Two spectroscopic descriptors were assessed to provide
159 qualitative information about the chromophoric DOM (CDOM), according adsorption spectra
160 of filtered and diluted samples from 200 nm to 800 nm with a UV1700 Pharma Spec
161 (Shimadzu) spectrophotometer using a 1-cm quartz cell. The ultraviolet absorbance SUVA₂₅₄,
162 related to the DOM aromaticity level, was calculated as the absorbance at 254 nm (a_{254}) divided
163 by the DOC concentration and by the cuvette path length (Weishaar et al., 2003). Absorbance
164 spectral slope between 275 and 295 nm was estimated as an indicator of CDOM molecular size,
165 with high values indicating low molecular size and vice versa (Helms et al., 2008). Data were
166 collected in double beam mode using deionized water as the blank.

167

168 **2.5. Fourier Transform Ion Cyclotron Resonance Mass Spectrometry (FT-ICR-MS).**

169 Filtered acidified water samples (20 mL) were processed through 100 mg styrene-divinyl-
170 polymere type (PPL) solid-phase cartridges (Agilent, Waldbronn, Germany) to desalt the
171 extract (Dittmar et al., 2008) for subsequent electrospray ionization (ESI) mass spectrometry.

172 The solid-phase extracted DOM (SPE-DOM) was eluted twice with 2 mL of methanol. The
173 extracts were diluted with methanol to 5 mL and were kept frozen until analysis (for further
174 details see Raeke et al., 2016). Recovery rates ranged from 30 to 85% being the lowest
175 efficiencies correspond to samples at deepest layers (4 and 4.6 m depth, SI2).

176 SPE extracts were diluted 1:1 (v/v) with ultrapure water and analyzed with an FT-ICR-MS
177 equipped with a dynamically harmonized analyzer cell (Solari XR, Bruker Daltonics Inc.,
178 Billerica, MA) and a 12 T refrigerated actively shielded superconducting magnet (Bruker
179 Biospin, Wissembourg, France). The instrument is located at the ProVIS Centre for Chemical
180 Microscopy within the Helmholtz Centre for Environmental Research. Samples were infused
181 by an autosampler at a flow rate of $10 \mu\text{L min}^{-1}$. For each spectrum, 256 scans (ion
182 accumulation time=15 ms) were co-added in the mass range 150-3000 m/z with a 4 MWord
183 time domain. Mass spectra were internally calibrated with a list of peaks (m/z 250-550, $n >$
184 110) commonly present in natural organic matter, the mass accuracy after internal linear
185 calibration was better than 0.1 ppm. Peaks were considered if the signal/noise (S/N) ratio was
186 greater than four.

187 Formulae were assigned to peaks in the mass range 150-700 m/z allowing for elemental
188 compositions $\text{C}_{1-60}\text{H}_{1-120}\text{O}_{1-40}\text{N}_{0-2}\text{S}_{0-1}$ with an error range of ± 0.5 ppm according to Lechtenfeld
189 et al. (2014). Relative peak intensities were calculated based on the summed intensities of all
190 assigned peaks in each sample. As data set detection limit a relative signal intensity of 0.01%
191 was chosen to compensate for variability in sensitivity between different measurements. A
192 molecular formula of a mass peak is referred to as “molecule” in this article although one
193 molecular formula can represent several different structural isomers.

194 Only formulas with $-10 \leq \text{DBE} - \text{O} \leq +10$ (DBE: double bound equivalents) were considered
195 for further data evaluation (Herzprung et al. 2014).

196 The following descriptors were estimated after molecular formula identification: number of
197 detected molecules (richness) and, molecular mass, H/C, O/C, N/C, S/C, DBE, modified
198 aromaticity index (AImod, Koch and Dittmar, 2016) and nominal oxidation state of carbon
199 (NOSC, LaRowe and Van Cappellen, 2011) weighted averages. C, H, O, N and S refer to the
200 number of atoms of carbon, hydrogen, oxygen, nitrogen and sulfur in each assigned molecule.

201 The ratio H/C, DBE and AI_{mod} provide information about the degree of saturation of a molecule
202 (i.e. number of C-C double bonds) whereas O/C ratio and NOSC are descriptors of the
203 (average) carbon oxidation state. According to the values of the different DOM indicators, the
204 DOM molecular classes were split into, Carboxyl-rich alicyclic like molecules (CRAM),
205 Aliphatics, Aromatics and Condensed&Aromatic (CA; molecular categories constrains are
206 detailed at Table SI 1). Be aware that each formula can describe numerous isomers this
207 classification suggests the most likely structure (Rossel et al., 2017). CRAM, aliphatic and
208 aromatic molecule are subdivided into O-poor or O-rich molecules if O/C is <0.5 or ≥ 0.5
209 respectively.

210

211 **2.6. Statistics and data analysis**

212 Hierarchical cluster analysis (HCA) of FT-ICR-MS elemental formula data was performed to
213 test differences among samples. HCA included the 2338 components that were common in all
214 samples after ranking the SPE-DOM mass peak intensities for each of the seven samples
215 separately. Squared Euclidean distance with Ward linkage method was applied (Murtagh and
216 Legendre, 2014) by assigning first rank to the variable with the highest intensity.

217 Inter Sample Rankings Analysis (ISRA) allowed the visualization of DOM quality differences
218 in van Krevelen diagrammes (Herzsprung et al. 2017).

219 For each SPE-DOM molecular class and descriptor, the weighted average was estimated from
220 relative intensity of the peak of each assigned molecule.

221 At Sonachi, the water column showed a meromictic stratification with a chemocline that
222 separate two main water masses: mixolimnion and monimolimnion (details in the results
223 section).

224 A simple end-member mixing model was used to quantify the magnitude of mixing at
225 chemocline, according to the following mass balance:

$$y = \frac{C_{ch} - C_{mix}}{C_{mon} - C_{mix}} \quad (2)$$

$$y + x = 1$$

228

229 x and y are the relative contributions of mixolimnion and monimolimnion. C_{ch} , C_{mix} and C_{mon}
 230 are the measured tracer concentrations (conservative solute) at chemocline, mixolimnion and
 231 monimolimnion respectively. C_{mix} represented the average value of the tracer concentrations
 232 between surface (0 m) and 3 m depth ($n=5$). Specific electrical conductivity at 20°C was
 233 adopted as conservative tracer (Leibundgut et al., 2011).

234 Calculation of the magnitude of mixing (i.e. x and y in eq. 2) allowed the estimation of the
 235 expected concentration of a reactive solute at the chemocline ($RS_{(exp;ch)}$):

$$RS_{(exp;ch)} = x RS_{(mix)} + y RS_{(mon)} \quad (3)$$

237 where $RS_{(mix)}$ and $RS_{(mon)}$ are the measured concentrations of a reactive solute at mixolimnion
 238 and monimolimnion respectively.

239 Differences between expected and observed values provided a quantitative information (η)
 240 about net retention or release of a reactive solute at the chemocline:

241

$$\eta = \frac{(RS_{(obs;ch)} - RS_{(exp;ch)})}{RS_{(exp;ch)}} * 100 \quad (4)$$

243

244 If $\eta > 0$ (or $\eta < 0$) the reactive solute was released (or retained) at chemocline (Butturini et al.,
 245 2016). Mass balance was performed for DOC and DON.

246

247 3. Results

248 3.1. Limnologic characterization

249 Temperature and oxygen in the water column ranged between 20.5 and 22.4°C, and $<10^{-3}$ and
250 0.1 mM (0-43% of saturation), respectively. A sharp change denoted the presence of a
251 thermocline and an oxycline at 1.5 m depth. pH was almost steady at 9.6 from surface down to
252 3 m depth, but decreased slightly to 9.5 below 4 m. Specific electrical conductivity was
253 homogeneous (*ca.* 9.8 mS cm⁻¹) from the surface to 3 m depth and increased rapidly to 11.5 and
254 16.2 mS cm⁻¹ at 4 m and at the bottom (i.e. sediment-water interface). The vertical profile of
255 redox potential ($E_{0(25^{\circ}\text{C})}$) showed a complex pattern: its value was steady at 298 mV in the first
256 1.5 m, then it gradually decreased to 140 mV at 3 m depth, and finally decreased abruptly to
257 less than -64 mV below 4 m. Overall, vertical profiles of pH, specific electrical conductivity
258 and $E_{0(25^{\circ}\text{C})}$, revealed the presence of a chemocline at about 3.5-4 m depth that separated a
259 mixolimnion from a monimolimnion (Figures 1a-e). According to the mixing model, 74.2% of
260 the water at chemocline came from mixolimnion and the rest from the monimolimnion. DIN
261 concentration averaged 1.8 ± 1.2 $\mu\text{M N}$ and did not show any clear vertical trend (Table SI3).

262

263 **3.2. Light regime in water column.**

264 According to the adsorption spectra measured in unfiltered samples, 99% of the total radiation
265 was absorbed within the oxycline; at chemocline darkness was total (**Figure SI2**). UVA and
266 UVB radiation disappeared within the first 25 cm (Figure 1f).

267

268 **3.3. DOC, DON and CDOM**

269 Dissolved organic carbon (DOC) averaged 8.1 ± 0.8 mM in the mixolimnion and increased to
270 48.8 mM in the monimolimnion (Figure 1g). Total dissolved organic nitrogen (DON) averaged
271 800 ± 90 μM in the mixolimnion and increased in the monimolimnion (Figure 1h). DOC and
272 DOM showed similar vertical profiles. However DOC/DON ratio was not steady and showed
273 an abrupt decline from 8.8 at mixolimnion, to 1.8 at the chemocline. The mass balance revealed

274 that the chemocline was a highly reactive site for DOM; while DOC was retained ($\eta=-47\%$),
275 DON was released ($\eta=105\%$; Figures 1g-h). $SUVA_{254}$ was almost constant in mixolimnion
276 (4.5 ± 0.3 mgC L/m) but it dropped to <1 mgC L m⁻¹ at monimolimnion (Figure 1i). On the
277 other hand, spectral Slope $S_{275-295}$, oscillated widely, without a clear vertical gradient with a
278 subtle accumulation of relatively larger molecules at chemocline and smaller ones at 0,5 m
279 depth respectively.

280

281 3.4. SPE-DOM

282 FT-ICR-MS detected 6221 elemental formulae. A subtotal of 2338 formulae were shared
283 between all seven samples. Number of identified formulae in each sample oscillated between
284 3449 and 4344 (Table SII).

285 The most abundant compositional group (39-49% of relative peak intensity) was represented by
286 O-poor CRAM-like molecules, structurally similar to sterols and hopanoids (Hertkorn et al.,
287 2006). Aromatics and Condensed&Aromatics (CA, resulting from vascular plant
288 leaching/degradation, Kellerman et al., 2015) were much lower and comprised between 1.5 to
289 4.2% of the total signal. Aromatic O-poor molecules were slightly more abundant than the O-
290 rich fraction. Aliphatic molecules were also relevant, contributing 12.3-19.6% of the total peak
291 intensity with the O-poor fraction being four times more intense (Table 2).

292 HCA of SPE-DOM data highlighted the impact of water stratification on DOM; the sample at
293 chemocline resulted the most dissimilar one (Figure 2).

294 Van Krevelen space visually illustrated how SPE-DOM differed across samples (Figure 2). The
295 chemocline sample showed relatively high peak intensities (ranks 1 and 2) in the region $H/C>1$
296 and $O/C<0.6$. In contrast, the weakest peaks (ranks 6 and 7) were preferentially located at
297 $O/C>0.5$.

298 In the monimolimnion, the most intense peaks were mainly located at $H/C > 1.2$ and
299 $0.2 < O/C < 0.8$. Meanwhile the weakest ones were for molecules with $1 < H/C < 1.5$ and $0.2 <$
300 $O/C < 0.6$.

301 In the mixolimnion, at depths 0.5, 1, 2 and 3 m, the most intense signal was related to
302 molecules with $H/C < 1.4$ and $O/C > 0.4$. In surface water, the distribution of ranks in the van
303 Krevelen diagram differed a little from that at the mixolimnion and the most intense peaks shift
304 slightly to the region $H/C < 1.5$ and $0.2 < O/C < 0.5$.

305 At the chemocline and the bottom, the 30 and 38.7% of cases with rank 1 and 2 (i.e. the
306 relatively high peak intensities) were located in the region $H/C > 1.5$ and $0.2 < O/C < 0.53$ of the
307 van Krevelen diagram. This percentage dropped to less than 3.8% in the mixolimnion samples.
308 Within the mixolimnion, the 80-86% of cases with rank 1 and 2 were located in the region
309 $0.65 < H/C < 1.5$ and $0.25 < O/C < 0.67$. This percentage decreased to less than 63% at the
310 chemocline and within the monimolimnion.

311 At the lake bottom, under permanent darkness, anoxic and reduced conditions, organic rich
312 sediments released a large amount (DOC~50 mM) of DOM enriched in highly reduced
313 (NOSC<-0.38, Figure 3a) and saturated (DBE<6.8; Figure 3b), O-poor aliphatic and CRAM
314 molecules (Figure 3c-d), and strongly depleted in aromatic ones (Figure 3e).

315 HCA (Figure 2) also suggested that SPE-DOM at the surface (0 m) differed slightly from the
316 rest of samples of the mixolimnion. More specifically, at surface, the relative intensity of SPE-
317 DOM molecules that were slightly more aliphatic (higher H/C, Mann Whitney U test,
318 $p < 0.0001$) and more reduced (lower O/C and lower NOSC, Mann Whitney test, $p < 0.0001$;
319 Figure SI7) increased.

320 Non-oxygen heteroatom molecules (CHN_1O , CHN_2O , CHOS, and CHNOS) represented 51.3 -
321 53.4% of the total of the molecules identified (Table SI2 and Figure SI 3-6). N-bearing
322 molecules were especially relevant representing $40 \pm 0.9\%$ of SPE-DOM molecules with a

323 relative signal intensity averaging $18.5\pm 1.2\%$. On the other hand, S-bearing molecules were
324 less relevant ($7.8\pm 3.6\%$ of SPE-DOM molecules and $3.5\pm 0.7\%$ of relative intensity). CHNO
325 and CHOS were smaller (lower molecular mass), more oxidized (higher NOSC) and less
326 aromatic (lower AImod) than CHO molecules (Figure SI8).

327

328 **4. Discussion**

329 Vertical profiles of chemical and physical properties of lake water (Electrical conductivity,
330 Oxygen, pH and light extinction rate) reported in the present study were similar to those
331 reported in previous studies (Melack et al., 1982; Melack and MacIntyre, 2016) evidencing that
332 meromixis is common at Sonachi. The high radiation extinction rate could reflect the
333 abundance of pelagic algal biomass (e.g. cyanobacteria; Ballot et al., 2005) and the large
334 amount of DOM suggesting that its photo-degradation is effective only within the first few
335 centimeters of the water column.

336 DOC concentrations within the water column were 1-2 orders of magnitude higher than those
337 typically reported for freshwaters (Sobek et al., 2007; Chen et al., 2015) and even slight higher
338 than those reported for endorheic saline waters around the world (Song et al., 2019; Osburn et
339 al., 2011; Mariot et al., 2007).

340 SPE-DOM showed marked differences with in respect to what is usually reported in inland
341 waters. Contributions of CRAM and aromatics were remarkably lower respect to (Riedel et al.,
342 2016; Stubbins et al., 2006), while aliphatics were significantly higher (Figure 4). Higher and
343 comparable aliphatic relative contributions were reported only in a coastal Antarctic pond (i.e.
344 Pony lake, Kellerman et al., 2018) and in a tropical reservoir during a cyanobacterial bloom
345 (Bittar et al., 2015), respectively. Overall, SPE-DOM at Sonachi was more reduced, less
346 aromatic and more saturated than typical SPE-DOM from inland waters (Raeke et al., 2017;
347 Dadi et al., 2017; Kellerman et al., 2018).

348 The relative low contribution of aromatics and CA pointed to the presence of highly degraded
349 and processed DOM, altered by intense photodegradation (Stubbins et al., 2006; Bittar et al.,
350 2015). The relatively high abundance of aliphatics might reflect DOM release from pelagic
351 cyanobacteria (Welker and Von Dohren, 2006; Biller et al., 2015). Aliphatics are considered to
352 be unaffected by photodegradation (Stubbins et al., 2006; Timko et al., 2015) and their relative
353 relevance might have been additionally amplified by the photodegradation of aromatics.

354 At mixolimnion, SPE-DOM properties were homogeneous thus evidencing a rapid mixing
355 between the most superficial and oxic layer and the underneath anoxic mixolimnion. A subtle
356 change was detected at surface with a slight relative accumulation of O-poor CRAM and O-
357 poor aliphatic compounds together to a relative decrease of O-rich aromatics (Figure 3). These
358 small molecular shifts, in comparison to the rest of the mixolimnion, might be attributable to
359 photo decarboxylation of O-rich aromatic molecules (Lv et al., 2017), suggesting that
360 photodegradation might be limited to the first few centimeters of the water column as
361 suggested by underwater UVA and UVB vertical profiles (Figure 1e).

362 Interestingly, most SPE-DOM descriptors (i.e. NOSC_w , Mass_w , Aliphatic_w , Aromatic_w , O/C_w ,
363 H/C_w ; Table 1 and Figures 3) showed abrupt changes at chemocline. This finding, together with
364 DOM mass balance estimates, evidenced that the chemocline was a site of remarkable
365 readjustment of DOM concentration and of the SPE-DOM molecular signature.

366 At the bottom and at the chemocline, the additional increase in the relative contribution of
367 saturated O-poor aliphatics was interpreted to reflect the release of microbially derived DOM
368 (Lechtenfeld et al., 2015) and the degradation of cyanobacteria/algae detritus accumulated at
369 the sediments-water interface (Verschuren, 1999). The abundance of aliphatic substances at the
370 lake bottom might be even larger than that detected by FT-ICR-MS analyses. Raeke et al.
371 (2016) reported that PPL sorbents are less efficient in retaining O-poor ($\text{O/C} < 0.4$) and saturated
372 (i.e. $\text{H/C} > 1.5$) molecules. DOM recovery rates of SPE extractions in samples collected at the

373 bottom and at the chemocline (appendix Table SI1) were lower compared to the samples from
374 the mixolimnion. Consequently, a large fraction of aliphatic compounds might be missing in
375 the MS analysis of bottom waters. Additionally, highly reduced conditions at the bottom
376 together with high pH probably boosted DOM release because of the deprotonation of mineral
377 sediments with subsequent reduction of their DOM adsorption potential (Grybos et al., 2009).

378 Organic rich sediment of SLAR host a highly complex and unique consortium of alkaliphilic
379 prokaryotes (Grant and Jones, 2016). Almost nothing is known about the magnitude of their in-
380 situ activities but a rough idea can be deduced from Ceiling (1996) who reported values of
381 $\delta^{13}\text{C}_{\text{DIC}}$ signatures higher than +10‰ in interstitial water pointing to the occurrence of high
382 microbial activities. Therefore, the high contribution of O-poor saturated aliphatic substances in
383 out the bottom samples is likely a signature of high microbial activity at the sediment-water
384 interface.

385 FT-ICR-MS has been extensively used to explore SPE-DOM chemodiversity across lakes at
386 regional scale (Kellerman et al., 2015), but the impact of lake stratification on DOM molecular
387 signatures is almost overlooked. Dadi et al. (2017) reported that in a reservoir, with a DOC/Fe
388 molar ratio of around 40, the unsaturated and O-rich SPE-DOM components were adsorbed
389 onto ferric iron minerals under epilimnetic oxic conditions and were released under anoxic and
390 reduced conditions at the sediments following reduction of solid ferric iron to soluble ferrous
391 iron. At Lake Sonachi the DOM absorption/desorption processes related to iron redox cycling
392 appeared to be less relevant because the oxic conditions persisted only in the first meter of the
393 water column. Further, total dissolved iron concentration in water was below 0.7 μM
394 (unpublished data from authors) and the DOC/Fe molar ratio was higher than 10000. In
395 consequence the influence of adsorption/desorption processes of unsaturated and O-rich DOM
396 bound to ferric iron minerals can be considered negligible.

397 Regarding to non-oxygen heteroatomic molecules, the relative intensity of N-bearing molecules
398 was about 3-4 times higher than that reported in large rivers and it was in the same range of that
399 measured in ancient ocean waters (Riedel et al., 2016) and in submarine hydrothermal fluids
400 (Gomez-Saez et al., 2016). On the other side, the percentage of CHOS molecules was similar to
401 that reported in freshwaters (Riedel et al., 2016, Kellermann et al., 2018) but much lower than
402 in hydrothermal fluids with evidence of abiotic sulfurization by H₂S (Gomez-Saez et al., 2016)
403 or in sewage sludge waters (Lv et al., 2017). Further, CHOS molecules tend to be slightly
404 smaller and more oxidized than CHO molecules suggesting that abiotic sulfurisation of CHO
405 molecules seemed to be unimportant (Melendez-Perez et al., 2018).

406

407 **5. Conclusions and implications**

408 DOM is a fundamental component of the study of the carbon cycle in the hydrosphere (Dittmar
409 and Stubbins, 2014; McCallister et al., 2018). Apart from few remarkable exceptions (Waiser
410 and Robarts, 2000; Mariot et al., 2007, Osburn et al., 2011, Song et al., 2019), DOM in saline-
411 alkaline endorheic systems has attracted little attention and these ecosystems are missing in
412 reviews focused on DOM quantity and quality in inland waters worldwide (Sobek et al., 2007;
413 Alvarez-Cobelas et al., 2012; Chen et al., 2015; Kellerman et al., 2018). Thus, the SLARs, the
414 most productive waterbodies worldwide (Lewis, 2011), are not taken into account in these
415 global assessments.

416 In conclusion our study reveals that:

- 417 • SLARs provide unique reservoirs of microbially derived, photodegraded, little aromatic
418 and saturated O-poor DOM suggesting that a full understanding of the processes
419 governing DOM availability and properties in inland waters (Sobek et al., 2007; Chen et
420 al., 2015; Kellermann et al., 2015) needs to include a recognition of the processes taking
421 place in endorheic saline-alkaline lakes characterized by a high productivity.

422

- 423 • The DOM molecular signature in endorheic saline-alkaline Lake Sonachi reflected the
424 pressure exerted by photodegradation acting at the same time as microbial production
425 and degradation. Photodegradation might be at work within the first few centimeters of
426 the water surface, while the greatest changes in quantity, reactivity and molecular
427 signature were located at the chemocline under steep geochemical gradients.
428
- 429 • These findings emphasize that, next to regional-scale climatic and hydrological drivers,
430 small-scale internal discontinuities create striking changes in the DOM molecular
431 signatures of lentic waters.
432

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445

447 **References**

- 448 Alvarez-Cobelas, M., Angeler, D. G., Sánchez-Carrillo, S., & Almendros, G., 2012. A
449 worldwide view of organic carbon export from catchments. *Biogeochemistry*, 107, 275-
450 293.
- 451 Ballot, A., Krienitz, L., Kotut, K., Wiegand, C., & Pflugmacher, S., 2005. Cyanobacteria and
452 cyanobacterial toxins in the alkaline crater lakes Sonachi and Simbi, Kenya. *Harmful*
453 *algae*, 4, 139-150.
- 454 Biller, P., Ross, A. B., & Skill, S. C., 2015. Investigation of the presence of an aliphatic
455 biopolymer in cyanobacteria: Implications for kerogen formation. *Organic*
456 *Geochemistry*, 81, 64-69.
- 457 Bittar, T. B., Stubbins, A., Vieira, A. A., & Mopper, K., 2015. Characterization and
458 photodegradation of dissolved organic matter (DOM) from a tropical lake and its
459 dominant primary producer, the cyanobacteria *Microcystis aeruginosa*. *Marine*
460 *Chemistry*, 177, 205-217.
- 461 Butturini, A., Guarch, A., Romaní, A. M., Freixa, A., Amalfitano, S., Fazi, S., & Ejarque, E.,
462 2016. Hydrological conditions control in situ DOM retention and release along a
463 Mediterranean river. *Water Research*, 99, 33-45.
- 464 Cerling, T. E., 1996. Pore water chemistry of an alkaline lake: Lake Turkana,
465 Kenya. *Limnology, Climatology and Paleoclimatology of the East African Lakes*, 225,
466 240.
- 467 Chen, M., Zeng, G., Zhang, J., Xu, P., Chen, A., & Lu, L., 2015. Global landscape of total
468 organic carbon, nitrogen and phosphorus in lake water. *Scientific Reports*, 5, 15043.
- 469 Dadi, T.; Harir, M.; Hertkorn, N.; Koschorreck, M.; Schmitt-Kopplin, P.; Herzsprung, P., 2017.
470 Redox conditions affect DOC quality in stratified freshwaters. *Environ. Sci. Technol.*, 51,
471 13705-13713.

- 472 Dittmar, T., & Stubbins, A., 2014. 12.6—Dissolved organic matter in aquatic systems. Treatise
473 on Geochemistry, 2nd edn. Elsevier: Oxford, p. 125-156.
- 474 Dittmar, T., Koch, B., Hertkorn, N., & Kattner, G., 2008. A simple and efficient method for the
475 solid-phase extraction of dissolved organic matter (SPE-DOM) from seawater. *Limnol.*
476 *Ocean. Methods*, 6(6), 230-235.
- 477 Dodds, W. K. and Whiles, M.R., 2002. *Freshwater ecology: concepts and environmental*
478 *applications of limnology*. 2nd edn. Academic press. p. 829
- 479 Elkins, K. M., & Nelson, D. J., 2002. Spectroscopic approaches to the study of the interaction
480 of aluminum with humic substances. *Coord. Chem. Rev.*, 228, 205-225.
- 481 Gaudet, J. J., & Melack, J. M., 1981. Major ion chemistry in a tropical African lake basin.
482 *Freshwater Biology*, 11, 309-333.
- 483 Gomez-Saez, G. V., Niggemann, J., Dittmar, T., Pohlabein, A. M., Lang, S. Q., Noowong, A.,
484 ... & Bühring, S. I., 2016. Molecular evidence for abiotic sulfurization of dissolved
485 organic matter in marine shallow hydrothermal systems. *Geochim. Cosmochim.*
486 *Acta*, 190, 35-52.
- 487 Gonsior, M., Peake, B. M., Cooper, W. T., Podgorski, D., D'Andrilli, J., & Cooper, W. J.,
488 2009. Photochemically induced changes in dissolved organic matter identified by
489 ultrahigh resolution Fourier transform ion cyclotron resonance mass spectrometry.
490 *Environ. Sci. Technol.*, 43, 698-703.;
- 491 Grant, W. D., & Jones, B. E., 2016. Bacteria, archaea and viruses of soda lakes. In *Soda Lakes*
492 *of East Africa*. Springer, Cham. pp. 97-147
- 493 Grybos, M., Davranche, M., Gruau, G., Petitjean, P., & Pédrot, M., 2009. Increasing pH drives
494 organic matter solubilization from wetland soils under reducing
495 conditions. *Geoderma*, 154, 13-19.

- 496 Hecky, R. E., & Kilham, P., 1973. Diatoms in alkaline, saline lakes: ecology and geochemical
497 implications 1. *Limnol. Oceanogr.*, 18, 53-71.
- 498 Helms, J. R., Stubbins, A., Ritchie, J. D., Minor, E. C., Kieber, D. J., & Mopper, K., 2008.
499 Absorption spectral slopes and slope ratios as indicators of molecular weight, source, and
500 photobleaching of chromophoric dissolved organic matter. *Limnology and*
501 *Oceanography*, 53(3), 955-969.
- 502 Hertkorn, N., Benner, R., Frommberger, M., Schmitt-Kopplin, P., Witt, M., Kaiser, K., ... &
503 Hedges, J. I., 2006. Characterization of a major refractory component of marine dissolved
504 organic matter. *Geochim. Cosmochim. Acta*, 70, 2990-3010.
- 505 Herzsprung, P., Hertkorn, N., von Tümpling, W., Harir, M., Friese, K., & Schmitt-Kopplin, P.
506 (2014). Understanding molecular formula assignment of Fourier transform ion cyclotron
507 resonance mass spectrometry data of natural organic matter from a chemical point of
508 view. *Anal. Bioanal. Chem.*, 406(30), 7977-7987.
- 509 Herzsprung, P., Osterloh, K., von Tümpling, W., Harir, M., Hertkorn, N., Schmitt-Kopplin,
510 P., Meissner, R., Bernsdorf, S., Friese, K. (2017). Differences in DOM rewetted and
511 natural peatlands- results from FT-ICR-MS and bulk optical parameters. *STOTEN*, 586,
512 770-781.
- 513 Jirsa, F., Gruber, M., Stojanovic, A., Omondi, S. O., Mader, D., Körner, W., & Schagerl, M.
514 (2013). Major and trace element geochemistry of Lake Bogoria and Lake Nakuru, Kenya,
515 during extreme drought. *Chem. Erde-Geochem.*, 73, 275-282.
- 516 Kilham P, Kilham SS., 1989. Endless summer: internal loading processes dominate nutrient
517 cycling in tropical lakes. *Freshwater Biol.* 23:379-389.
- 518 Kellerman, A. M., Guillemette, F., Podgorski, D. C., Aiken, G. R., Butler, K. D., & Spencer, R.
519 G., 2018. Unifying concepts linking dissolved organic matter composition to persistence
520 in aquatic ecosystems. *Environ. Sci. Technol.*, 52, 2538-2548.

- 521 Kellerman, A. M., Kothawala, D. N., Dittmar, T., & Tranvik, L. J., 2015. Persistence of
522 dissolved organic matter in lakes related to its molecular characteristics. *Nat. Geo.*, 8,
523 454.
- 524 Koch, B. P., and Dittmar, T., 2016. Erratum: From mass to structure: an aromaticity index for
525 high-resolution mass data of natural organic matter. *Rapid Comm. Mass Spectrom.*, 30,
526 250. doi: 10.1002/rcm.7433
- 527 LaRowe, D. E., & Van Cappellen, P., 2011. Degradation of natural organic matter: a
528 thermodynamic analysis. *Geochim. Cosmochim. Acta*, 75, 2030-2042.
- 529 Lechtenfeld, O. J., Kattner, G., Flerus, R., McCallister, S. L., Schmitt-Kopplin, P., & Koch, B.
530 P., 2014. Molecular transformation and degradation of refractory dissolved organic
531 matter in the Atlantic and Southern Ocean. *Geochim. Cosmochim. Acta*, 126: 321-337.
- 532 Lechtenfeld, O.J.; Hertkorn, N.; Shen, Y.; Witt, M.; Benner, R., 2015. Marine sequestration of
533 carbon in bacterial metabolites. *Nat. Commun.* 6:6711.
- 534 Leibundgut, C., Maloszewski, P., & Külls, C., 2011. Tracers in hydrology. John Wiley & Sons.
535 P. 415.
- 536 Leinemann, T., Preusser, S., Mikutta, R., Kalbitz, K., Cerli, C., Höschen, C., ... &
537 Guggenberger, G., 2018. Multiple exchange processes on mineral surfaces control the
538 transport of dissolved organic matter through soil profiles. *Soil Biol. Biochem.*, 118, 79-
539 90.
- 540 Lewis Jr, W. M., 2011. Global primary production of lakes: 19th Baldi Memorial Lecture.
541 *Inland Waters*, 1, 1-28.
- 542 Lv, J., Li, D., Lou, L., wu, T. & Zhang, S., 2017. Molecular transformation of natural and
543 anthropogenic dissolved organic matter under photo-irradiation in the presence of nano
544 TiO₂. *Wat. Res.*, 125: 201-208.

- 545 MacIntyre, S., & Melack, J. M., 1982. Meromixis in an equatorial African soda lake 1. *Limnol.*
546 *Oceanogr.*, 27, 595-609.
- 547 Massicotte, P., Asmala, E., Stedmon, C., & Markager, S., 2017. Global distribution of
548 dissolved organic matter along the aquatic continuum: Across rivers, lakes and
549 oceans. *STOTEN*, 609, 180-191.
- 550 McCallister, S. L., Ishikawa, N. F., & Kothawala, D. N., 2018. Biogeochemical tools for
551 characterizing organic carbon in inland aquatic ecosystems. *Limnol. Oceanogr. Letters*, 3,
552 444-457.
- 553 Mariot, M., Dudal, Y., Furian, S., Sakamoto, A., Vallès, V., Fort, M., & Barbiero, L., 2007.
554 Dissolved organic matter fluorescence as a water-flow tracer in the tropical wetland of
555 Pantanal of Nhecolândia, Brazil. *STOTEN*, 388, 184-193.
- 556 Melack, J. M., & MacIntyre, S., 2016. Morphometry and Physical Processes of East African
557 Soda Lakes. In *Soda Lakes of East Africa* (pp. 61-76). Springer, Cham.
- 558 Melack, J. M., 1982. Photosynthetic activity and respiration in an equatorial African soda lake.
559 *Freshwat. Biol.*, 12, 381-400.
- 560 Melack, J. M., Kilham, P., & Fisher, T. R., 1982. Responses of phytoplankton to experimental
561 fertilization with ammonium and phosphate in an African soda lake. *Oecologia*, 52, 321-
562 326.
- 563 Melendez-Perez, J. J., Martinez-Mejia, M. J., Barcellos, R. L., Fetter-Filho, A. F., & Eberlin,
564 M. N., 2018. A potential formation route for CHOS compounds in dissolved organic
565 matter. *Mar. Chem.*, 202, 67-72.
- 566 Meybeck, M., 1995. Global distribution of lakes. In *Physics and chemistry of lakes*. Springer,
567 Berlin, Heidelberg. p. 1-35
- 568 Murtagh, F., & Legendre, P., 2014. Ward's hierarchical agglomerative clustering method:
569 which algorithms implement Ward's criterion?. *J. Class.*, 31(3), 274-295.

- 570 Odongo, V. O., van der Tol, C., Becht, R., Hoedjes, J. C., Ghimire, C. P., & Su, Z., 2016.
571 Energy partitioning and its controls over a heterogeneous semi-arid shrubland ecosystem
572 in the Lake Naivasha Basin, Kenya. *Ecohydrology*, 9, 1358-1375.
- 573 Oduor, S. O., & Schagerl, M., 2007. Phytoplankton primary productivity characteristics in
574 response to photosynthetically active radiation in three Kenyan Rift Valley saline–
575 alkaline lakes. *J. Plankton Res.*, 29, 1041-1050.
- 576 Olaka, L. A. (2011). Hydrology across scales: sensitivity of East African lakes to climate
577 changes. PhD thesis.
- 578 Osburn, C. L., Wigdahl, C. R., Fritz, S. C., & Saros, J. E., 2011. Dissolved organic matter
579 composition and photoreactivity in prairie lakes of the US Great Plains. *Limnol.*
580 *Oceanogr.*, 56, 2371-2390.
- 581 Riedel, T., Zark, M., Vähätalo, A. V., Niggemann, J., Spencer, R. G., Hernes, P. J., & Dittmar,
582 T., 2016. Molecular signatures of biogeochemical transformations in dissolved organic
583 matter from ten world rivers. *Front. Earth Sci.*, 4, 85.
- 584 Raeke, J., Lechtenfeld, O. J., Tittel, J., Oosterwoud, M. R., Bornmann K., and Reemtsma, T.,
585 2017. Linking the mobilization of dissolved organic matter in catchments and its removal
586 in drinking water treatment to its molecular characteristics. *Water Research* 113: 149-159
- 587 Raeke, J., Lechtenfeld, O.J., Wagner, M., Herzsprung, P., Reemtsma, T., 2016. Selectivity of
588 solid phase extraction of freshwater dissolved organic matter and its effect on ultrahigh
589 resolution mass spectra. *Environ. Sci. Processes & Impacts*, 18, 918 – 927.
- 590 Reardon, J., 1969. Salicylate method for the quantitative determination of ammonia nitrogen
591 U.S. Patent No. 3,432,395. Washington, DC: U.S. Patent and Trademark Office.
- 592 Rossel, P. E., Stubbins, A., Rebling, T., Koschinsky, A., Hawkes, J. A., & Dittmar, T., 2017.
593 Thermally altered marine dissolved organic matter in hydrothermal fluids. *Organic*
594 *geochemistry*, 110: 73-86.

- 595 Salm, C. R., Saros, J. E., Fritz, S. C., Osburn, C. L., & Reineke, D. M., 2009. Phytoplankton
596 productivity across prairie saline lakes of the Great Plains (USA): a step toward
597 deciphering patterns through lake classification models. *Can. J. Fish. Aquat. Sci.*, *66*(9),
598 1435-1448.
- 599 Schagerl, M. (Ed.), 2016. Soda Lakes of East Africa. Berlin: Springer.p. 408
- 600 Sobek, S., Tranvik, L. J., Prairie, Y. T., Kortelainen, P., & Cole, J. J., 2007. Patterns and
601 regulation of dissolved organic carbon: An analysis of 7,500 widely distributed lakes.
602 *Limnol. Oceanogr.*, *52*, 1208-1219.
- 603 Song, K., Shang, Y., Wen, Z., Jacinthe, P. A., Liu, G., Lyu, L., & Fang, C., 2019.
604 Characterization of CDOM in saline and freshwater lakes across China using
605 spectroscopic analysis. *Water research*, *150*, 403-417.
- 606 Stubbins, A., Spencer, R. G., Chen, H., Hatcher, P. G., Mopper, K., Hernes, P. J., ... & Six, J.,
607 2010. Illuminated darkness: molecular signatures of Congo River dissolved organic
608 matter and its photochemical alteration as revealed by ultrahigh precision mass
609 spectrometry. *Limnol. Oceanogr.*, *55*, 1467-1477.
- 610 Stubbins, A., Uher, G., Law, C. S., Mopper, K., Robinson, C., & Upstill-Goddard, R. C., 2006.
611 Open-ocean carbon monoxide photoproduction. *Deep Sea Research Part II: Tropical*
612 *Studies in Oceanography*, *53*(14-16), 1695-1705.
- 613 Talling, J. F., 2011. Some distinctive subject contributions from tropical Africa to fundamental
614 science of inland waters. *Inland Waters*, *1*, 61-73.
- 615 Talling, J. F., Wood, R. B., Prosser, M. V., & Baxter, R. M., 1973. The upper limit of
616 photosynthetic productivity by phytoplankton: evidence from Ethiopian soda lakes.
617 *Freshwater Biol.*, *3*, 53-76.

- 618 Tassi, F., Fazi, S., Rossetti, S., Pratesi, P., Ceccotti, M., Cabassi, J., ... & Vaselli, O. (2018).
619 The biogeochemical vertical structure renders a meromictic volcanic lake a trap for
620 geogenic CO₂ (Lake Averno, Italy). *PloS one*, 13, e0193914.
- 621 Tassi, F., Vaselli, O., Giannini, L., Tedesco, D., Nencetti, A., Montegrossi, G., & Yalire, M. M.
622 2004. A low-cost and effective method to collect water and gas samples from stratified
623 crater lakes: the 485 m deep lake Kivu (DRC). *Proc. IAVCEI Gen. Ass.*, Puchon, Chile,
624 p. 14-19.
- 625 Thrane, J. E., Hessen, D. O., & Andersen, T. 2014. The adsorption of light in lakes: negative
626 impact of dissolved organic carbon on primary productivity. *Ecosystems*, 17, 1040-1052.
- 627 Timko, S. A., Maydanov, A., Pittelli, S. L., Conte, M. H., Cooper, W. J., Koch, B. P., ... &
628 Gonsior, M., 2015. Depth-dependent photodegradation of marine dissolved organic
629 matter. *Front. Mar. Sci.*, 2, 66.
- 630 Verschuren, D. (1999). Influence of depth and mixing regime on sedimentation in a small,
631 fluctuating tropical soda lake. *Limnol. Oceanogr.*, 44, 1103-1113.
- 632 Waiser, M. J., & Robarts, R. D., 2000. Changes in composition and reactivity of allochthonous
633 DOM in a prairie saline lake. *Limnol. Oceanogr.*, 45, 763-774.
- 634 Weishaar, J. L., Aiken, G. R., Bergamaschi, B. A., Fram, M. S., Fujii, R., & Mopper, K., 2003.
635 Evaluation of specific ultraviolet absorbance as an indicator of the chemical composition
636 and reactivity of dissolved organic carbon. *Environ. Sci. Technol.*, 37, 4702-4708
- 637 Welker, M., & Von Döhren, H., 2006. Cyanobacterial peptides—nature's own combinatorial
638 biosynthesis. *FEMS microbiology reviews*, 30, 530-563.
- 639 Wolkersdorfer, C., 2008. *Water management at abandoned flooded underground mines:
640 fundamentals, tracer tests, modelling, water treatment*. Springer Science & Business
641 Media.

- 642 Wurtsbaugh, W. A., Miller, C., Null, S. E., DeRose, R. J., Wilcock, P., Hahnenberger, M., ... &
643 Moore, J., 2017. Decline of the world's saline lakes. *Nat. Geo.*, 10(11), 816.
- 644 Zinabu, G. M., & Taylor, W. D. (1997). Bacteria—chlorophyll relationships in Ethiopian lakes
645 of varying salinity: are soda lakes different?. *J. Plankton Res.*, 19, 647-654.
- 646

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648 Table 1. SPE-DOM molecular characterization at Sonachi. Molecular class criteria and SPE-DOM descriptors are defined at table SII. All values are signal
 649 intensity weighted. CA: Condensed and aromatic.

650

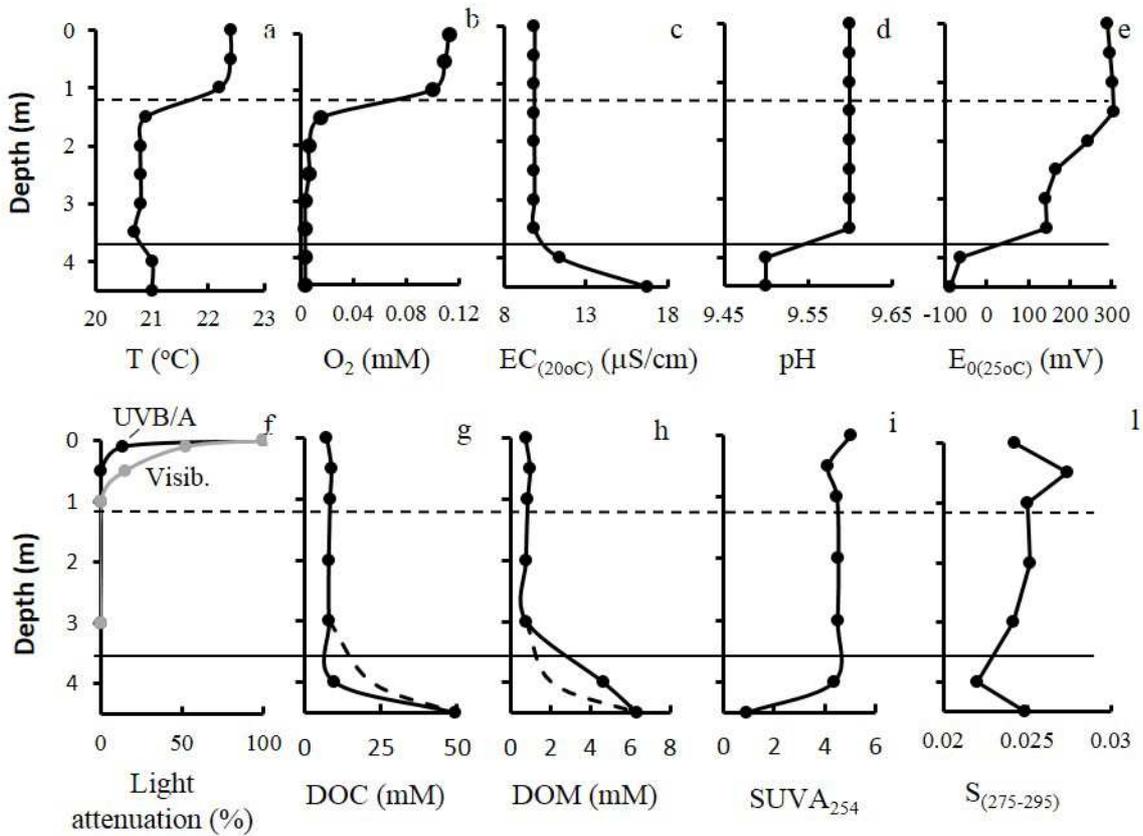
Depth (m)	CRAM		Aliphatic		Aromatic		CA
	O poor	O rich	O poor	O rich	O poor	O rich	
0	41.85	18.15	10.05	2.21	2.47	1.52	0.90
0.5	39.54	19.36	9.25	2.11	2.42	1.72	0.85
1	40.40	18.80	9.55	2.18	2.35	1.67	0.88
2	40.02	19.05	9.50	2.19	2.38	1.68	0.87
3	40.02	18.79	9.47	2.38	2.36	1.59	0.87
4	49.13	9.79	16.81	2.82	1.23	0.23	0.16
4.5	39.59	16.15	12.60	2.94	1.59	0.63	0.29

651

Depth (m)	# of molecules	Mass _w (Da)	NO _{SC} _w	AI _{mod} _w	DBE _w	C _w	N _w	S _w	O/C _w	H/C _w	S/C _w	N/C _w
0	4207	356.6	-0.336	0.227	7.092	17.16	0.24	0.03	0.46	1.31	0.0019	0.0155
0.5	3701	355.6	-0.310	0.225	7.117	16.98	0.26	0.04	0.47	1.30	0.0036	0.0172
1	3809	355.7	-0.318	0.227	7.105	17.03	0.25	0.04	0.46	1.30	0.0032	0.0163
2	4208	357.1	-0.314	0.225	7.122	17.07	0.25	0.04	0.47	1.30	0.0030	0.0165
3	4344	357.4	-0.317	0.225	7.114	17.11	0.25	0.03	0.47	1.30	0.0025	0.0165
4	3948	365.0	-0.499	0.194	6.758	18.02	0.21	0.04	0.42	1.38	0.0034	0.0132
4.5	3449	354.4	-0.379	0.200	6.724	17.03	0.25	0.03	0.46	1.35	0.0023	0.0163

652 **Figures**

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656 Figure 1. Vertical profiles of water temperature, dissolved oxygen, Electrical conductivity (EC),
 657 pH, E_0 , underwater light attenuation, DOC, DON, $SUVA_{254}$ and $S_{(275-295)}$. Dashed curves in the
 658 lower portion of panels *g* and *h* show expected trends assuming conservative solute mixing
 659 between mixolimnion and monimolimnion. Horizontal lines indicate depths of
 660 thermocline/oxycline (dashed) and chemocline (solid).

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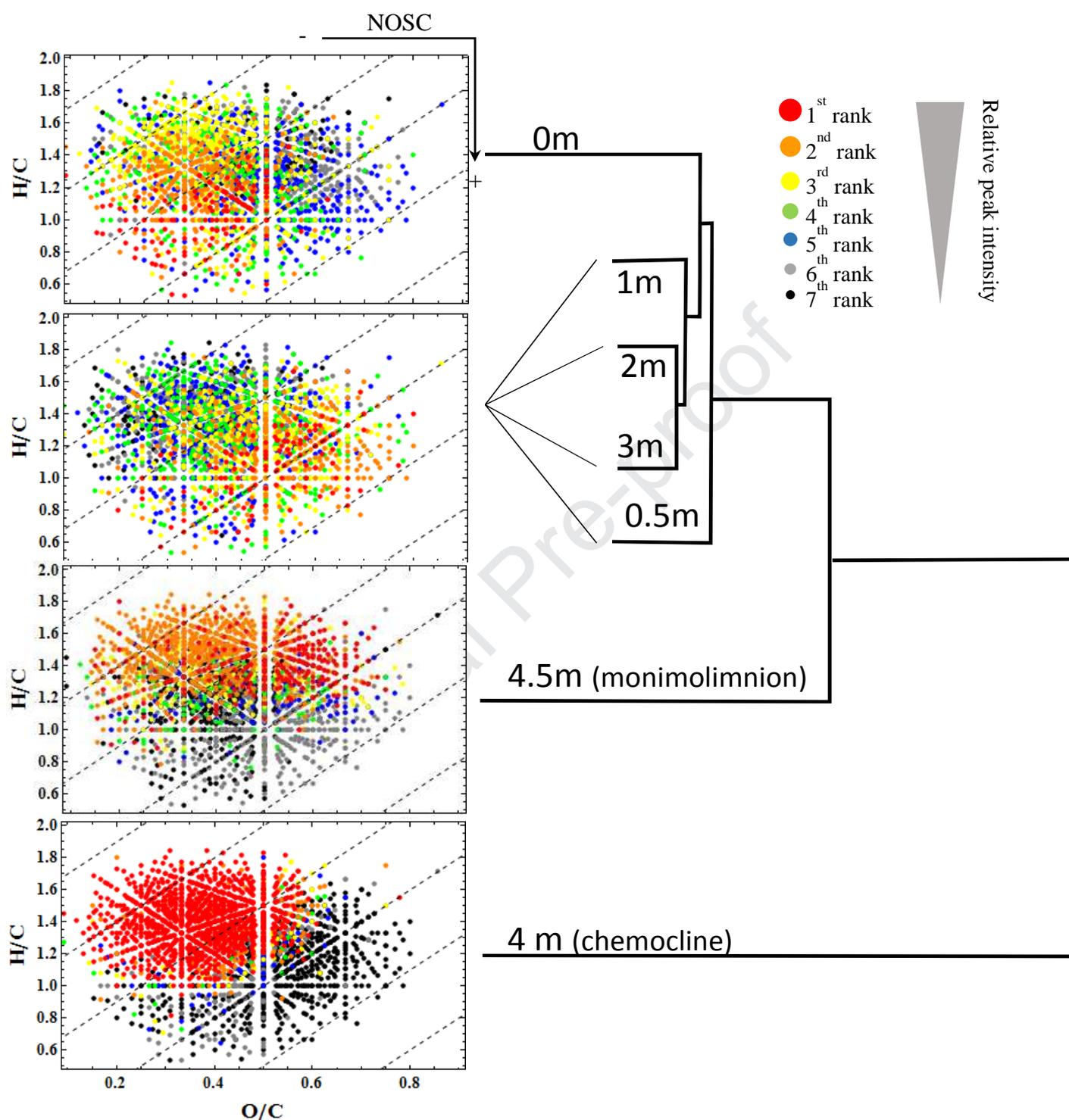
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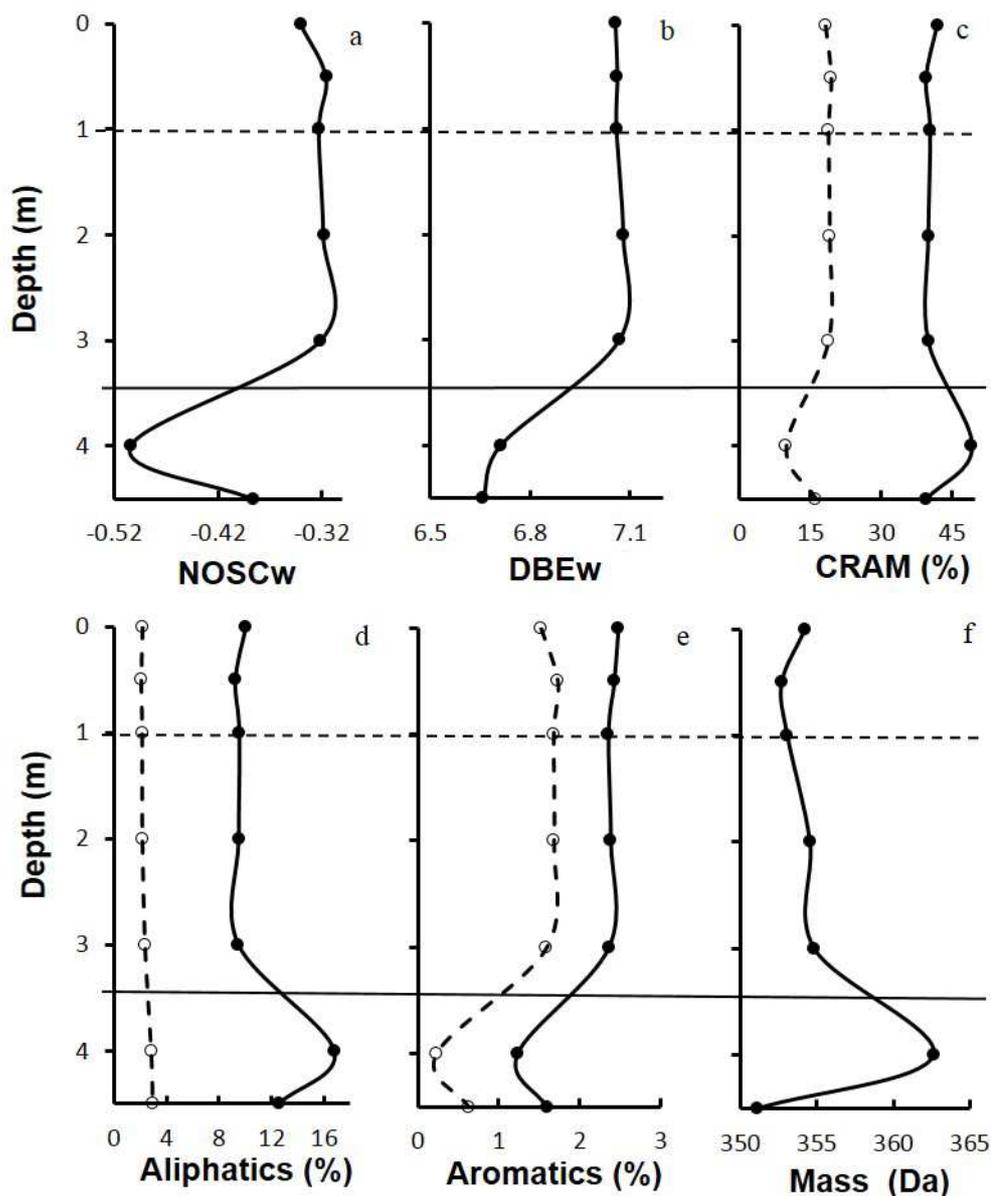
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Figure 2. HCA of all SPE-DOM components detected with FT-ICR-MS and van Krevelen diagrams showing the inter sample rankings and molecular composition of SPE-DOM in the water column. Samples at 0, 0.5, 1, 2 and 3 m depth are within the mixolimnion. See appendix SI 3-6 for specific Van Krevelen diagrams of CHO, CHN1S, CHN2O and CHOS molecules



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699 Figure 3. Vertical profiles of two SPE-DOM molecular descriptors (NOSC and DBE) and four
 700 molecular classes (CRAM, aliphatic, aromatic and peptides) and molecular size (Mass). All
 701 values are weighed averages of peak intensities. Dashed and solid lines in profiles of CRAM,
 702 Aliphatic and Aromatic indicate O-rich and O-poor molecules respectively. The two horizontal
 703 lines describe the approximate depth of oxycline/thermocline (dashed line) and chemocline
 704 (solid line).

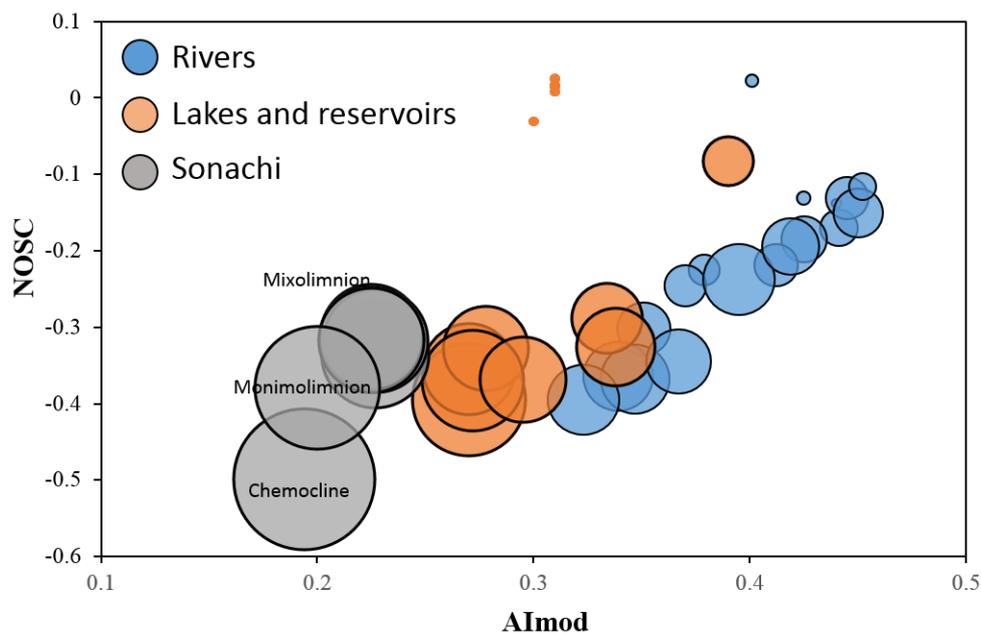
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712 Figure 4. Bubble plot describing the relationship between saturation degree (AI mod) and
713 Carbon oxidation state (NOSC) in rivers (blue bubbles) lakes and reservoirs (orange bubbles)
714 and Sonachi (gray bubbles). Bubbles size is proportional to the relative abundance of aliphatic
715 molecules. Literature data are from Kellerman, et al. (2018). Small orange dots are values
716 measured from Rappbode reservoir (Dadi et al., 2017)
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Highlights

- First detailed description of DOM in a tropical meromictic saline-alkaline lake.
- High amount of autochthonous DOM was found in water column.
- Solid-phase extracted DOM is microbially derived, photodegraded and aliphatic.
- Changes in DOM quantity and quality were located at the chemocline.

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