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1	Assessment of the key evolutionary traits that prevent extinctions in human-
2	altered habitats using a spatially explicit individual-based model
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20	ABSTRACT
21	Identifying key evolutionary strategies that support population persistence remains a
22	challenging task for biodiversity conservation. Here we assess if animal adaptations to
23	cope with low densities (i.e. that facilitate mate-findings or promote spatial aggregation
24	of individuals) can allow species to persist in human-altered habitats. A spatially
25	explicit and individual-based model was developed to assess if, and under what
26	circumstances, such adaptations maintain population viability. The model was
27	parameterised with data from the movement and demography of the spur-thighed
28	tortoise (Testudo graeca) and simulated scenarios with differences in adult
29	survivorships, initial population sizes and habitat alterations. Habitat alterations reduced
30	population viability, and extinction rates were dependent on population characteristics
31	and mate-finding distance. In contrast, philopatry around the birthplace did not prevent

extinctions. Our results highlight the importance of considering specific spatial traits of

33 species when assessing their vulnerability to human habitat alterations.

35	<b>KEYWORDS</b> : Allee effect; habitat fragmentation; habitat loss; low-movement species;
36	philopatry, spatial traits
37	HIGHLIGHTS
38	We developed STEPLAND, a spatially explicit IBM to assess population viability
39	We investigated if some spatial traits allow persistence in human-altered habitats
40	Mate-finding efficiency was essential for the viability of low density populations
41 42	We evidenced the importance of spatial traits when assessing species' vulnerability
43	

### 44 **1. INTRODUCTION**

45 With more than half the Earth's land surface currently altered by human activities (Barnosky et al., 2012), habitat loss and fragmentation has been identified as one of the 46 main drivers of biodiversity loss (Brooks et al. 2002). Therefore, assessing and 47 predicting how species cope with habitat alterations are relevant challenges (Sala 2000). 48 When assessing the vulnerability of animal species to habitat alterations, the 49 consideration of the spatial component inherent to animal movement is crucial (Fahrig, 50 2007; Villard & Metzger, 2014; Niebuhr et al., 2015). Habitat alterations have been 51 52 identified to cause animal mobility loss worldwide (Tucker et al., 2018), and loss of mobility can directly affect population persistence, but also ecosystem processes, such 53 54 as predator-prey interactions, nutrient cycling and disease transmission (Tucker et al., 2018). 55

56 The interaction between movement ability and the landscape structure determines the response of organisms to the various landscape elements, which is 57 known as species' functional connectivity (Tischendorf & Fahrig, 2000; Schooley & 58 Wiens, 2003). According to this concept, connectivity is highly dependent on the 59 organisms, population characteristics and landscapes being studied (Auffret et al. 2015). 60 Therefore, human habitat alterations can critically reduce functional connectivity, which 61 is especially detrimental for low-density populations of species with low movement 62 capacity that are prone to Allee effects (e.g., lower population densities lead to fewer 63 mating opportunities and diminished reproductive fitness) (Philip 1957; Wosniack et al. 64 2014). However, some of these species may have developed specific traits that allow 65 them to persist in a low-density context if they evolved under such circumstances. 66 67 Gascoigne et al. (2009) hypothesised that adaptations to mitigate mate-finding failures could also help to buffer the effects of human-altered habitats on population viability. 68 69 These authors identified two evolutionary routes for this: (i) evolution of mechanisms such as calls, pheromones and hermaphroditism that make mate-finding at low density 70 71 more efficient or less necessary; (ii) mechanisms that lead to the permanent or periodic aggregation of individuals and, therefore, weaken the strength of mate-finding Allee 72 73 effects (e.g., reproductive aggregations). When populations are reduced by habitat alternations, Gascoigne et al. (2009) predicted that the species with adaptations of type 74 75 (i) would be less likely to suffer from mate-finding Allee effects than species with 76 adaptations of type (ii) because they had already adapted to cope with low density.

To find out if, and under which circumstances, these mechanims can allow 77 populations to cope with human-altered habitats, it is necessary to embed them into a 78 demographic context (e.g., considering mortality, growth, and reproduction) and to 79 study the emerging spatial population dynamics. Correlated random walk models 80 (Kareiva & Shigesada, 1983) are powerful tools that reproduce the inherent stochasticity 81 of animal movements and behaviour (Kramer-Schadt et al. 2004, Börger et al. 2008, 82 Anadón et al. 2012). These mechanistic models simulate random walk trajectories that 83 emulate complex movement or behavioural processes, such as habitat selection or the 84 85 maintenance of home ranges (e.g., Moorcroft & Barnett, 2008). Individual-based 86 models are particularly suitable to study and predict the dynamics of complex systems, 87 which emerge from decision-making individuals that differ and interact locally (DeAngelis & Grimm, 2014). Hence integrating correlated random walk models into 88 89 individual-based models that consider demographic processes should allow us to assess if and under which circumstances the mechanisms proposed by Gascoigne et al. (2009) 90 91 allow maintenance of population viability in human-altered habitats (Kramer-Schadt et al. 2005). 92

This study assessed the hypothesis that biological traits developed to deal with 93 low densities could allow animal populations to cope with the negative effects of 94 human-altered habitats. Two traits were tested as being representative of the two 95 evolutionary routes identified by Gascoigne et al. (2009): for the first route we used 96 "mate-finding efficiency", modelled as the maximum distances among individuals that 97 allows for detection of conspecifics and subsequent mating, and for the second route we 98 99 used "philopatry", modelled as the tendency of individuals to stay close to their 100 birthplaces (keeping home ranges) that potentially avoids movement into low density 101 areas.

102 To test our hypotheses we developed STEPLAND, an individual-based 103 simulation model that integrates the spatially explicit movement of individuals with the 104 demographic processes of mortality, aging and recruitment. This model was 105 parameterised with extensive bibliographic and field data of the spur-thighed tortoise 106 (Testudo graeca). This species is ideally suited to address our overall objective because 107 it is threatened by habitat fragmentation and loss (IUCN 2016), and low population 108 densities and low movement capacity make it prone to mate-finding Allee effects. Thus, 109 understanding the mechanisms that may allow T. graeca to deal with low densities in

- human-altered habitats is a key for management of this species, and parameterization 110 111 with field data is essential for this. For example, movement patterns and home ranges proved to be very different among individuals of T. graeca that inhabited natural and 112 human-altered habitats (Anadón et al. 2012). As the effects of mate-finding efficiency 113 114 and philopatry on population viability are likely to be context-dependent and be strongly affected by low-density conditions, they were tested herein for different 115 landscapes, population densities and demographic conditions (i.e., adult survival). 116 STEPLAND was applied to 12 simulation scenarios that arose from combining three 117 118 initial population sizes, two adult survival rates and two distinct (natural and altered) landscapes (Table 1; Fig. 1). These scenarios well represent the range of real conditions 119 120 of low density populations of the selected study system. For each scenario, the sensitivity of the extinction rates of T. graeca populations over 200 years was tested in 121
- relation to changes in the tested biological traits.

123

124 Table 1. The simulation scenarios and variation in the spatial traits considered in the simulations. To
125 generate simulation scenarios, we combined the values of the initial population size, adult survival rates
126 and landscape type, which resulted in 12 scenarios. To test the impact of the mate-finding efficiency and
127 natal philopatry on population viability, we individually varied the values of parameters DIST in SITE for
128 each scenario. The values of these parameters, which were maintained across simulations ("standard
129 parameterisation"), are shown in bold. Table C1 in Appendix C describes the model's full

130 parameterisation.

Tested	Model	Values	Description	Justification/Reference
condition	variable			
or spatial				
trait				
<u></u>				
Scenarios				
Population	$N_0$	45/90/180	Initial number of tortoises	Anadón et al., (2009)
size			representing very low to	
			medium densities.	
Adult	S	0.95/0.98	Annual survival rates of	Sanz-Aguilar et al. (2011)
survival			adults	Rodríguez-Caro et al. (2013)
rate				
Habitat	LANDSCAPE	Natural/	Landscapes of 3 km $\times$ 3 km	Real landscapes from SE
		Human-	that represent the natural	

quality		altered	(Galera) and human-altered conditions (Bas)	Spain
Species traits				
Mate- finding efficiency	DIST	100/ 300/ <b>500</b> /700	Threshold of the maximal distance between females and males that allows mating (in metres).	The standard value matches the sum of the mean maximal annual displacement distances reported for males and females (Anadón et al. 2012)
Natal philopatry	SITE	YES/NO	Natal philopatry is simulated by maintaining the location of the focal- point attractor of tortoises for their lifetimes ("YES"). Alternatively, lack of such behaviour is simulated by reassigning this location to the last position reached each year ("NO")	The standard value represents strong site-fidelity as found by Sanz-Aguilar et al. (2011)



Figure 1. The simulation scenarios. Twelve plausible *T. graeca* populations were generated by

134 combining two (natural and human-altered) landscapes from SE Spain, three initial population sizes  $(N_0)$ 

that represent low-density populations, and two adult survival rates  $(S_a)$ . The grey areas on the map on the left show the approximate Western Mediterranean distribution range of *T. graeca* (triangles represent small populations). The two considered landscapes are located north of the species' distribution range in SE Spain.

139

## 140 **2. METHODS**

## 141 2.1 STEPLAND: a spatially individual-based model that integrates animal 142 movement and demography

143 The spatially explicit individual-based model STEPLAND was implemented in Python

144 2.7 and integrated demographic processes (i.e., mortality, aging and recruitment) with a

145 detailed model of animal movement. The purpose of our study is to investigate the

146 response of population viability to changes in biological traits, population

147 characteristics and/or habitat landscapes. The model description follows the Overview,

148 Design concepts and Details protocol (ODD) proposed by Grimm et al. (2006, 2010).

149 Figure 2 presents a general model overview and its description is summarised in the

150 following paragraphs. A full ODD version is presented in Appendix A. STEPLAND

151 was parameterised with extensive data on the spur-thighed tortoise's (*Testudo graeca*)

movement and demography. However, after suitable re-parameterization it can also be

applied to other study systems with similar movement profiles (species with movement

adjustable to correlated walk trajectories).

155



157

158 Figure 2. The main processes implemented into STEPLAND. Each simulation started with an initial 159 population that inhabited a given landscape and was governed by a specific model parameterisation. The 160 model output comprised the generated tortoise populations after T years (here T=200). After parameter 161 uploading, tortoises were subjected to the three basic demographic processes; movement, reproduction 162 and mortality. Landscape characteristics influenced tortoise movements, while spatial constraints 163 impacted reproduction (i.e.,: long distances between individuals, or barriers between males and females 164 hinder mating). The effects of the selected parameters on population dynamics were assessed by a 165 sensitivity analysis. See Table 1 and Fig. 1 for a description of the parameters and the simulation 166 scenarios, respectively.

167

### 168 **2.2 Studied species**

The spur-thighed tortoise Testudo graeca is a medium-sized tortoise whose main 169 Western European population is found in SE Spain (Fig. 1). This species inhabits 170 171 mainly semiarid shrublands and traditional agricultural landscapes, and is heavily threatened by habitat loss and fragmentation (IUCN 2016). T. graeca individuals are 172 173 non-territorial with overlapping home ranges that range from 1 to 5 ha in SE Spain (Anadón et al. 2006). They mature at the age of 8-10 years and use to live as long as 30-174 175 40 years (Díaz-Paniagua et al. 2001; Rodríguez-Caro et al. 2019). T. graeca females can store sperm for prolonged times after mating, a biological trait that allows clutches to be 176

177 fertilised for prolonged time periods after mating (Roques et al. 2004). Clutch sizes vary178 mostly between 1 and 7 (the authors, unpublished data).

179 The spur-thighed tortoise is ideally suited to study the interaction between habitat alteration and individual movement behaviour on population dynamics. Firstly, 180 181 radiotracking data, together with long-term capture-recapture studies, suggest that 182 individuals show a low-dispersal capacity and high site fidelity (Sanz-Aguilar et al. 183 2011, Anadón et al. 2012). These features make low-density populations especially 184 vulnerable to human habitat alterations (Sanz-Aguilar et al. 2011). Secondly, the basic 185 demographic aspects of T. graeca populations have been well studied (e.g., Díaz-Paniagua et al. 2001, Sanz-Aguilar et al. 2011, Rodríguez-Caro et al. 2013, 2019). 186 187 Additionally, an individual-based model that simulates the movement behaviour of adult T. graeca males and females in natural and human-altered landscapes has already 188 189 been parameterised and implemented (Anadón et al. 2012). STEPLAND extends this 190 movement model towards a dynamic population model by adding the demographic 191 processes of mortality, aging and recruitment.

192

### **2.3 Model use and parameterisation**

Entities, attributes and spatial scales. STEPLAND contains two types of 194 entities: landscape and individuals (tortoises in this study). The 3 km  $\times$  3 km landscape 195 (Fig. 1) was composed of a grid of  $10 \text{ m} \times 10 \text{ m}$  cells. Each grid cell was characterised 196 by its position (x and y coordinates), and also by its assignment to a habitat category. In 197 198 this study, the four habitat categories used by Anadón et al. (2012) were considered: intensive land use, traditional agriculture land, flat natural areas and natural areas on 199 200 slope, and an additional non-permeable infrastructure category. Landscapes were 201 enclosed by a non-permeable border that resulted in isolated populations.

The attributes of individuals were: sex, age, their locations over time and the coordinates of their focal point attractor to represent philopatry (i.e., individual's tendency to maintain a stable home range around the focal point). Each individual obtains a movement parameterisation from a pool of parameterisations that were compatible with observed radiotracking data (following Anadón et al. 2012; see Appendix A for further movement parameterisation information).

Input, initialisation and output. The model inputs included the initial population, 208 209 the landscape and model parameterisation (Fig. 2). In this study, the initial number of 210 individuals for the different age classes followed a stable age distribution predicted by 211 an age-stage structured deterministic matrix population model (Caswell 2001) with the 212 demographic parameters used in STEPLAND. The R package POPBIO (Stubben and Milligan 2007, R Core Team 2014; Appendix B) was used to derive the stable age 213 distribution that fitted age distributions observed in field (Rodríguez-Caro et al, 2019). 214 215 The individuals in the initial population were randomly distributed across the 216 landscapes and their initial locations were the same in all the simulation scenarios that 217 started with the same density. To avoid biases due to the excessive aggregation or 218 disaggregation of individuals we checked mean minimum Euclidean distances between 219 females and males in tentative initial populations before selecting suitable initial 220 populations (see section "Input, initialisation and output" in Appendix A for further information). Model parameterisations were based on Anadón et al. (2012), empirical 221 222 data and a literature review (see Table C1 in Appendix C). The data output comprised the surviving individuals of the population present at the end of the simulation. 223

224 Temporal scales, process overview and scheduling. Each time step represented one day. In this study, simulations were run for 200 years. The model consisted of 225 226 submodels for movement, reproduction, mortality and aging happened at the end of the year (Fig. 2). The scheduling of the different processes was adapted to the biology of 227 228 the species. Tortoise movement occurred on a daily basis, but followed a seasonal 229 activity pattern with no movement in winter and little movement in summer (i.e., 230 hibernation and aestivation periods), but peak movement during the mating season 231 (April to May) (Fig. A4 in Appendix A).

232 I) Movement submodel: STEPLAND incorporates the movement model and 233 movement parameterisations of Anadón et al. (2012) based on the analysis of radiotracking data using inverse and individual-based modelling approaches. 234 235 These authors found that tortoises showed high site fidelity in natural landscapes, and only slight variability in their movement patterns. However, 236 237 movement behaviour varied strongly among individuals in altered landscapes. A 238 wide range of movement patterns was found particularly for females, which 239 ranged from strong homing behaviour to unbounded movements. For males, Anadón et al. (2012) determined 129 and 51 different parameterisations of nine 240

parameters governing individual movement that were compatible with the
empirical movement data in natural and altered landscapes, respectively. For
females, there were 155 and 5,282 parameterisations of the nine parameters,
respectively (see Appendix A for further details). To reflect the individual
variability observed in the movement patterns, we randomly assigned one of
these movement parameterisations to a tortoise at birth (corresponding to its sex
and landscape type).

Individual movement consisted in a sequence of steps to the 248 249 neighboured 10 m  $\times$  10 m cells (Fig. 3A). The stochastic rules that determined the number of these "cell-to-cell" steps were governed by the two probability 250 distributions, PMOV and DMOV, which were directly taken from the detailed 251 radiotracking data (that divided the active period of 1 day into four 2- to 3-hour 252 253 periods). Firstly, on a given day, the probability of moving during a given 2-3 hour period was given by the corresponding probability of PMOV, and 254 255 depended on the month and sex of the tortoise in accordance with its activity. hibernation and aestivation periods (Fig. A4 in Appendix A). Secondly, if the 256 257 PMOV trial indicated movement, DMOV decided how many cell-to-cell steps 258 the individual actually moved during this period (Fig. A4 in Appendix A). While 259 PMOV and DMOV governed the number of cell-to-cell steps in which an 260 individual moved, the actual movement decisions for the cell-to-cell step were stochastic, but depended on three sets of weights that described autocorrelation 261 in movement, habitat use and homing behaviour towards a focal point attractor. 262 263 The three normalised weights were multiplied to calculate the final probability of a cell being selected (Fig. 3B). See Appendix A for a complete description of 264 the movement submodel. 265

266

267 II) *Demographic submodel*: demography in STEPLAND included ageing,
268 reproduction and mortality. Reproduction included the three processes: mating,
269 sperm storage and clutching (Fig. 2).

Mating occurred once during the spring. The STEPLAND model
assumed that mature females and males mated during this period if: (i) the
distance between female and male was shorter than the value of the parameter

DIST and (ii) females and males were not separated by barriers such as main
roads, dense human infrastructures or intensive agriculture.

Sperm-storage and multiple paternity, which are specific traits of spur-275 thighed tortoises (Roques et al. 2004), were modelled as a temporal dissociation 276 between mating events and clutching. Some studies have indicated that sperm 277 can be viable in female oviducts for as long as 3-4 years (e.g., Cutuli et al. 278 2013). Therefore, it was assumed that females could reproduce up to 3 years 279 after their last mating. Adult females are able to lay up to three clutches a year as 280 281 observed in real populations (Díaz-Paniagua et al. 1996, 1997; unpublished 282 data).

Mortality was a single stochastic event at the end of each simulated year. Annual survival rates varied among age classes, including eggs, immature individuals, subadults and adults. Finally, surviving individuals aged one year. Appendix C provides the parameters values and an extended description of the criteria followed to select them.

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289

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Figure 3. The movement submodel. (A) Movement of a simulated tortoise on a given day: starting from 291 292 the arrow, the tortoise moves during three of the four active periods (represented by different colours), 293 with six, five and three cell-to-cell steps each (movement directions are represented by numbers). (B) 294 Stochastic movement decision for one step: dependent on three sets of weights that described 295 autocorrelation in movement, habitat use and homing behaviour towards to a focal point attractor. The 296 three normalised weights (broken arrows) were multiplied to calculate the final probability of a 297 neighbouring cell being selected. The asterisk indicates the individual's current position. The black dot 298 represents the focal point attractor governing homing behaviour. The continuous arrow represents the 299 previous movement. The cell colours indicate different habitat categories. See Appendix A for a complete 300 description of the movement submodel.

301

## 302 **2.4 Experimental design**

303 Simulation scenarios. Twelve simulation scenarios (Table 1, Fig. 1) were assembled that resulted from combining three initial population sizes, two annual adult survival 304 rates and two distinct landscapes. Simulations started with 45, 90 and 180 tortoises  $(N_0)$ , 305 which covered very low to medium population densities, as found in natural T. graeca 306 307 populations in SE Spain (0.05, 0.1 and 0.2 tortoises/ha; Anadón et al., 2009). As the growth rate of long-living species is very sensitive to the annual survival rates (S) of 308 309 adults, independent simulations were conducted for two S values that represented high and medium survival for T. graeca populations (Sanz-Aguilar et al. 2011, Rodríguez-310 Caro et al. 2013;  $S_{\text{LOW}} = 0.95$  and  $S_{\text{HIGH}} = 0.98$ ; Table C1 in Appendix C). The 311 combinations of demographic parameters resulted in theoretical growth rates of  $\lambda_{LOW}$  = 312 1.008 (for  $S_{\text{LOW}}$ ) and  $\lambda_{\text{HIGH}} = 1.02$  (for  $S_{\text{HIGH}}$ ) (see Table C1 in Appendix C). These 313 theoretical growth rates were calculated with the R package POPBIO, and represent 314 theoretical population growth (i.e., without effects of demographic or environmental 315 stochasticity or spatial Allee effects; Appendix B). However, in the populations 316 317 simulated with our spatially explicit individual-based model, spatial constraints could impede the reproduction of some females in fragmented landscapes (see below) and 318 319 result, together with demographic stochasticity, in lower growth rates, or even in 320 population extinctions.

To represent natural and human-altered landscapes (the LANDSCAPE variable 321 322 in the subsequent analyses), two typical  $3 \text{ km} \times 3 \text{ km}$  landscapes from SE Spain that 323 currently host T. graeca populations were used (Fig. 1). These two landscapes were also 324 used to parameterise the movement submodel for individual tortoises (Anadón et al. 2012). The "Galera" landscape represented the natural landscapes with very low cover 325 326 of intensive land uses and non-permeable infrastructures (1%), whereas the "Bas" landscape represented altered landscapes with a much higher proportion of intensive 327 land uses or non-permeable infrastructures (26%). As the "Galera" landscape lacked 328 329 internal barriers, we assumed the population to be potentially panmictic (i.e., in theory, 330 all individuals can breed if located close enough to one another). In contrast, and as mentioned above, we assumed that intensive land uses (i.e., intensive crops) and non-331 permeable infrastructures (i.e., a highway) imposed barriers for reproduction (Fig. 1). 332

Note that the movement of tortoises across barriers was not totally impeded, but became 333 334 little likely because movement in human-altered landscapes was influenced by habitat 335 type (Fig. 3B; Fig. A7 in Appendix A). Anadón et al. (2012) determined individual 336 movement parameterisations separately for males and females, and also for natural and 337 the altered landscapes (see Appendix A for further movement parameterisation 338 information). Therefore, for the Galera simulations, the movement parameterisations of individuals were randomly selected from the pool of natural landscape 339 parameterisations, and for the Bas simulations from the pool of altered landscape 340 341 parameterisations.

Simulation experiments. For all 12 scenarios, we assessed the effects of mate-342 343 finding efficiency and philopatry on population viability (Fig.1 and Table 1). To test the influence of mate searching efficiency, the parameter DIST was varied over a wide 344 345 range of values (100 m, 300 m, 500 m and 700 m). Note that the mean maximal displacement distances observed during one-year period in the radiotracked T. graeca 346 males and females were 189-275 m and 148-271 m, respectively, as reported by Anadón 347 et al. (2012). In the standard simulations, tortoises' birthplaces were maintained as a 348 focal point attractor for their whole lifetime to represent philopatry (SITE = Yes). To 349 simulate lack of philopatry, the standard parameterisation was used (with DIST = 500), 350 but the focal point attractor was reassigned to the individual's last position at the end of 351 each year (SITE = No) (Table 1). This resulted in "wandering" behaviour. 352

353 **2.5 Data analysis** 

354 The total number of independent model simulations was 4,608 (12 simulation scenarios 355  $\times$  5 parameterisations of spatial traits  $\times$  64 replicates; Table 1). Scripts developed in R 356 (R Core Team 2014) and Python 2.7 processed model outputs to calculate extinction rates. Multivariate binomial GLM models were used to assess the sensitivity of each 357 358 evaluated parameter for population extinction. The variables related to the scenarios were treated as factors ( $N_0$ , S, LANDSCAPE), while model parameters were treated as 359 360 numerical (DIST) or factor (SITE) variables. The selection of the best model for the parameterisations that represented the variations in each trait was based on Akaike's 361 362 information criterion (AIC; Burnham and Anderson 2004) using a stepwise backward algorithm. Model structures with  $\Delta AIC < 2$  were assumed to receive similar support 363

364 (Hastie and Pregiborn 1992). *Post hoc* analyses (ANOVA function) served to calculate
365 the explained deviance of each parameter to the selected model.

**366 3. RESULTS** 

## 367 3.1 Effects of landscape type and population characteristics on population 368 viability

In the first set of simulation experiments, the standard model parameterisation (Table 1) 369 was used to test the influence of the 12 scenarios on population viability. The results 370 showed that variation in population and landscape characteristics in the scenarios 371 372 affected the population dynamics of *Testudo graeca* (Table 2). The extinction rates (EX) over a 200-year time horizon differed strongly among the simulation scenarios. 373 Lower initial density increased the extinction risk (mean EX = 0.37 for  $N_0 = 45$ , mean 374 EX = 0.08 for  $N_0$  = 90, and mean EX = 0 for  $N_0$  = 180). Landscape type also influenced 375 population viability as the extinction rate was higher in the altered landscape (mean EX 376 = 0.18) than in the natural one (mean EX = 0.11). Finally, as expected, the lower adult 377 survival (0.95 vs. 0.98) strongly enhanced the extinction rate (mean EX = 0.29 for  $S_{LOW}$ , 378 379 and mean EX = 0.01 for  $S_{\text{HIGH}}$ ).

**Table 2.** Extinction rates (i.e., proportion of the 64 replicate simulations that became extinct) of the *T*. *graeca* populations over 200 years. Simulations were replicated 64 times using standard parameterisation based on both empirical data and a literature review (see Table C1 in Appendix C). The simulation scenarios combined different initial population sizes ( $N_0$ ), adult survival rates ( $S_{LOW} = 0.95$ ;  $S_{HIGH} = 0.98$ ) and habitat quality (natural and human-altered landscapes). Note that the deterministic growth rate, estimated from the demographic parameters, yielded  $\lambda_{LOW} = 1.008$  (for  $S_{LOW}$ ) and  $\lambda_{HIGH} = 1.02$  (for

 $S_{HIGH}$ ). Population extinction can arise from spatial constraints and demographic stochasticity, especially

**387** for those scenarios with  $N_0 = 45$  and  $S_{\text{LOW}}$ .

3	8	8

	Very low density ( $N_0 = 45$ tortoises)		Low density $(N_0 = 90 \text{ tortoises})$		Medium density ( $N_0 = 180$ tortoises)		
Landscape	$S_{\rm LOW}$	$S_{ m HIGH}$	$S_{ m LOW}$	$S_{ m HIGH}$	$S_{ m LOW}$	$S_{ m HIGH}$	
Natural	0.66	0.02	0.02	0	0	0	
Altered	0.77	0.02	0.30	0	0	0	

389

## **390 3.2 Effect of mate-finding efficiency on population viability**

Mate-finding efficiency, represented by the maximal distance DIST between males and females where mating occurred (within the 100-700 m range), strongly influenced the

population dynamics and extinction probability of T. graeca. For low adult survival 393  $(S_{LOW})$ , all populations went extinct for the lowest mate searching efficiency (i.e., DIST 394 = 100 m), and most of them for the low mate searching efficiency (i.e., DIST = 300 m; 395 Fig. 4). However, for DIST = 700 m, all simulations were viable for high adult survival 396  $(S_{\text{HIGH}})$ , and also for the majority of simulations run with  $S_{\text{LOW}}$ . Only the low density 397 simulations that started with  $N_0 = 45$  individuals showed a moderate percentage of 398 extinctions for DIST = 700 (Fig. 4). Finally, intermediate DIST values (i.e., DIST = 300 399 m for  $S_{\text{HIGH}}$  and DIST = 500 m for  $S_{\text{LOW}}$ ) produced a wide range of population 400 dynamics (Fig. 4). In these cases, extinction rates for  $S_{LOW}$  were related with landscape 401 type and the initial population size; but for  $S_{\text{HIGH}}$  the mainly driver of variation was the 402 403 initial population size (Fig. 4). Table D1 in Appendix D shows the extinction rates of 404 the simulations.

Multivariate binomial GLM models, used to assess the sensitivity of population extinction to the four variables DIST, S,  $N_0$ , and LANDSCAPE, quantitatively revealed the relevance of the mate-finding efficiency for population viability. Most of the variation was explained by the DIST parameter (i.e. DIST explained 54.28% of the deviance, while the total deviance explained by the model was 65.23%; Table 3).

411**Table 3.** The GLM results that describe the sensitivity of the three tested spatial traits for population412extinction over 200 years. Parameters govern processes in relation to movement and behaviour (mate-413finding efficiency and natal philopatry). The best model only includes those variables or interactions with414 $\Delta AIC < 2$ . See Table 1 for the description of the parameters and variables.

Spatial trait	Parameter	Variables included in the GLM model	$\Delta$ % of explained deviance
Mate-finding efficiency	DIST	DIST	54.28
		S	6.25
		$N_{0}$	4.51
		LANDSCAPE	0.19
		Total model	65.23
Natal philopatry	SITE	S	30.78
		$N_0$	25.44
		LANDSCAPE	0.92
		Total model	57.14



**418** Figure 4. Extinction rates predicted by GLMs in dependence of DIST, for different survival rates (*S*), and 419 initial population densities ( $N_0$ ). This parameter imposes a maximal distance threshold between females 420 and males to allow mating (in metres). The simulation scenarios included two different adult survival 421 rates ( $S_{LOW} = 0.95$ ;  $S_{HIGH} = 0.98$ ), habitat types and the initial population sizes. See Table 1 and Table C1 422 in Appendix C for further model parameterisation information.

423

## 424 **3.3 Sensitivity of philopatry**

In our simulations, the maintenance, or not, of stable home ranges around tortoises'
birthplaces had no effect on the extinction rates of the *T. graeca* populations (Table 3).
The SITE parameter, which governs the maintenance of the stable home range of
tortoises among years and emulates philopatry, displayed only weak explanatory power.
Table D1 in Appendix D shows the extinction rates.

430

## 431 **4. DISCUSSION**

The evolutionary history of species may result in traits that allow them to cope with human-altered habitats. For example, Gascoigne et al. (2009) hypothesised that evolutionary adaptations to persist under low-density conditions could allow species also to deal with human-altered habitats. This study employed an individual-based simulation model to assess the effects of such traits, mate-finding efficiency and philopatry, on the viability of a low-movement species within a range of contexts given
by variations in initial population size, natural *vs*. human-altered landscapes and adult
survival.

Our results showed that increasing mate-finding efficiency had a high potential 440 441 to avoid population extinctions if initial population sizes are low, whereas philopatry had no effect. The typical human-altered landscape negatively affected the viability of 442 443 *Testudo graeca* populations compared to the natural landscape, but this depended on 444 population characteristics (the initial population size and survival rates of adults). 445 Landscape type did not affected viability in the scenarios with high adult survival rates (resulting in 2% maximal possible population growth), but extinction rate was 446 447 especially sensitive to the landscape type in the scenarios with lower adult survival rates (0.8% maximal population growth) and a low and very low initial density of tortoises 448 449 (Table 2). These results agree with those of Pearson et al. (2014) who, in the broader climate change context, found that species' vulnerability depends on a mixture of spatial 450 and demographic factors. These authors also identified population size and population 451 452 growth among the most important demographic predictors.

453

## 454 **4.1 Role of spatial traits in mitigating mate-finding Allee effects**

The consideration of species traits in ecological modelling allows from the study of the 455 population dynamics of one or several species (e.g. discrete traits distribution models 456 457 reproducing predator-prey dynamics; Tirok & Gaedke, 2010), to the description of the 458 temporal dynamics of a community (e.g. biomass produced by phytoplankton considering distribution-based aggregate models under different selection regimes; 459 Klauschies et al. 2018). To produce robust predictions under global change scenarios, 460 461 recent studies claim the necessity to consider relevant species traits when projecting species range distributions (Regos et al. 2019) or species vulnerability (Pearson et al. 462 2014). In this work, two adaptations with the potential to increase mating encounters in 463 human-altered habitats were assessed: mate-finding efficiency and philopatry to the 464 birth site. The population dynamics of spur-thighed tortoises was especially sensitive to 465 the mate-finding efficiency, but not to philopatry. These results confirmed the 466 expectations of Gascoigne et al. (2009), who predicted that the adaptations which 467

468 evolved to function well at low densities would have a very high potential to increase469 the resilience of populations that declined in human-altered habitats.

470 Mate-finding efficiency. The maximum distance of male and female individuals to be allowed for mating encounters, simulated by the DIST parameter in STEPLAND, 471 472 turned out to be a key factor for population dynamics and extinction. Despite being a 473 simplification of a complex biological process, our results agree well with our 474 knowledge on *T. graeca*. In our model simulations, a reduction in extinction risk 475 occurred for the DIST values of 300 m and 500 m, distances that match the sum of the 476 mean maximal annual displacement distances reported for T. graeca males and females (189-275 m and 148-271 m, respectively; Anadón et al. 2012). Clearly unrealistically 477 478 high parameter values (e.g. > 500 m) would lead to a further reduction in the extinction 479 risk of small populations because they counteract the Allee effect. How conspecific 480 attraction underlies mating search strategies remains largely unknown (Galeotti et al. 481 2007), but our results support the notion that such a biological mechanism is key for 482 maintaining the viability of low-density populations (Reed and Dobson 1993, Berec and 483 Boukal 2004). Some evolutionary strategies, like pheromones, song or calls, prevent mate-finding failures; alternatively, hermaphroditism or the ability to maintain pair 484 bonds, can reduce the frequency with which mate-finding needs to occur (Gascoigne et 485 al. 2009). Mate-finding is driven in tortoises by the production, detection and 486 recognition of sex pheromones (Norris and López, 2011). However, when considering 487 488 their limited dispersal capabilities, we hypothesise that female sperm-storage may also 489 play a major role in supporting population viability under low-density conditions 490 (Roques et al. 2004; Gascoigne et al. 2009). This evolutionary strategy has a high 491 potential for mitigating the effects of human-altered habitats in limiting the availability of potential mates. However, further simulation experiments should elucidate the trade-492 493 offs for the evolutionary potential that an increase in the use of this trait may cause (like 494 endogamy from using the sperm of a few parents over years).

*Philopatry.* The tendency to return to the birthplace is a mechanism that allows
individuals to obtain breeding sites of higher quality and in closer proximity of other
individuals, than settling at random locations. However, because trade-offs may arise in
relation to inbreeding, settling patterns have proved to be variable among individuals,
depending on the quality among breeding sites and on individuals' characteristics (e.g.,
age and sex; Schjørring, 2001). Hence, philopatry acts as a plastic biological trait with

variable relevance for population persistence depending on the context. In our 501 502 simulation experiments, the stable home ranges that remained during the lifetimes of 503 tortoises around their birthplace did not affect the population dynamics compared to the 504 home ranges that could shift every year (by displacements of the focal point attractors to 505 the last location in the year). Stable home ranges may offer tortoises high quality 506 feeding patches, refuge areas and places for laving eggs (e.g., Kaddour et al. 2006; 507 Anadón et al. 2012; Lecq et al. 2014), and may prevent tortoises from wandering into low-density areas where mating encounters are rare. The fact that our simulation 508 509 experiments did not show effects of natal philopatry could be caused by the moderate annual displacements of home ranges, which did not cause populations to drift apart in 510 511 the 3x3 km simulated landscapes. However, omission of homing behaviour resulted in 512 immediate population extinction when simulating colonisation processes at a larger 513 geographic scale (10x10 km landscapes) (results not shown). Indeed, philopatry has proved be relevant for other chelonians (Caretta caretta) to prevent Allee effects during 514 515 range expansions (Stiebens et al., 2013).

516

## 517 4.2 Is *T. graeca* able to counterbalance Allee effects in human-altered 518 landscapes?

Habitat alterations generally reduce the quantity of high-quality habitats and enhance 519 520 border effects. Animal movement patterns in fragmented landscapes may play an important role in mitigating the detrimental effects of human habitat alterations, and in 521 522 maintaining populations' functional connectivity (Mimet et al. 2013, Niebuhr et al. 523 2015). The present study was carried out under low-density conditions, far away from 524 the carrying capacity possible in our two landscapes, and the same demographic parameters were assumed for the tortoises that inhabited both human-altered and natural 525 526 landscapes (except for movement parameterisations). Therefore, the differences in population dynamics between the corresponding scenarios in the natural and altered 527 528 landscapes could only arise from the spatial mechanisms linked to mate-finding Allee 529 effects (i.e., lower mating encounter rates due to barriers). Our simulation experiments 530 suggested that the human-altered landscape increased the extinction risk of T. graeca compared to the natural landscape by generating Allee effects in low-density 531 populations (i.e., isolation between sexes). As exemplified for T. graeca, some 532

populations that inhabited human-altered mosaic landscapes could disappear within
some decades (Alharbi and Petrovskii 2016). The fact that these tortoises are currently
present in highly disturbed areas is probably a result of their longevity and not due to
population viability (Henle et al. 2004).

537

## 538 **5. CONCLUSIONS**

Habitat changes have been identified among the main drivers of biodiversity loss 539 (Brooks et al. 2002). Therefore, assessing and predicting how species face human 540 541 habitat alterations are relevant challenges (Sala 2000). Our study exemplifies that considering the spatial component inherent to animal movement (spatial traits) is 542 relevant when assessing animals' vulnerability to human-altered habitats. We 543 demonstrate that integrating individual-based correlated random walk models and 544 demographic population models allows for the identification of the key biological traits 545 governing populations' viability. As suggested by Gascoigne et al. (2009), we found 546 547 that increasing mate-finding efficiency has a higher potential to avoid population 548 extinctions in human-altered habitats than adaptations that lead to the aggregation of 549 individuals. This information points to vulnerability of species that base their 550 reproductive strategies on philopatric behaviours. Moreover, our modelling approach could inform managing strategies of endangered species to, for example, assess and 551 increase the permeability of linear barriers that may cause isolation between sexes. 552 Further research is needed to explore the consequences of landscape changes not just for 553 554 population demography, but also for the evolutionary potential. Impacts on population 555 connectivity reduce gene flow on a landscape scale and promote local endogamy 556 (Méndez et al. 2014). Similarly, evolutionary strategies such as sperm-storage, that facilitate reproduction when mate encounters are scarce, may help to support 557 558 populations persistence but also promote local endogamy. Then, the study of trade-offs derived of the maintenance of reproduction in altered habitats turns out a relevant task 559 for individual-based models that consider specific traits of species in their spatial 560 561 context.

562

## 6. SOFTWARE AVAILABILITY

563 Model name: STEPLAND.

- 564 Developed by: Eva Graciá (egracia@umh.es), Ángel Luis García-García
- 565 (angelluis78@gmail.com).

566 Year available: 2019.

- 567 Availability: The full code of STEPLAND, its parameterisation and the main results
- 568 files are available at XXX.
- 569 Cost: Free and open source.
- 570 Language: Python.
- 571
- 572 **7. ACKNOWLEDGEMENTS**

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- 587
- 588 **8. REFERENCES**

Alharbi, W. G., & Petrovskii, S. V. (2016). The Impact of Fragmented Habitat's
 Size and Shape on Populations with Allee Effect. Mathematical Modelling of
 Natural Phenomena, 11(4), 5-15. DOI: 10.1051/mmnp/201611402

- Anadón, J. D., Giménez, A., Perez, I., Martinez, M., & Esteve, M. A. (2006).
  Habitat selection by the spur-thighed tortoise *Testudo graeca* in a
  multisuccessional landscape: implications for habitat management. Biodiversity
  and Conservation, 15(7), 2287-2299. DOI: https://doi.org/10.1007/s10531-0048226-4
- 3. Anadón, J. D., Wiegand, T., & Giménez, A. (2012). Individual based
  movement models reveals sex biased effects of landscape fragmentation on
  animal movement. Ecosphere, 3(7), 1-32. DOI: <u>https://doi.org/10.1890/ES11-</u>
  00237.1
- 4. Auffret, A. G., Plue, J., & Cousins, S. A. (2015). The spatial and temporal
  components of functional connectivity in fragmented landscapes. Ambio, 44
  Suppl 1(Suppl 1), S51–S59. DOI:10.1007/s13280-014-0588-6
- 5. Barnosky, A.D., Hadly, E.A., Bascompte, J., Berlow, E.L., Brown, J.H.,
- 605 Fortelius, M., Getz, W.M., Harte, J., Hastings, A., Marquet, P.A., Martinez,
- 606 N.D., Mooers, A., Roopnarine, P., Vermeij, G., Williams, J.W., Gillespie, R.,
- 607 Kitzes, J., Marshall, C., Matzke, N., Mindell, D.P., Revilla, E., & Smith, A.B.,
- 608 (2012). Approaching a state shift in Earth's biosphere. Nature 486, 52–58. DOI:
  609 https://doi.org/10.1038/nature11018
- 6. Berec, L., & Boukal, D.S. (2004). Implications of mate search, mate choice and
  divorce rate for population dynamics of sexually reproducing species. Oikos
  104, 122–132. DOI: 10.1111/j.0030-1299.2004.12753.x
- 7. Börger, L., Dalziel, B. D., & Fryxell, J. M. (2008). Are there general
  mechanisms of animal home range behaviour? A review and prospects for future
  research. Ecology letters, 11(6), 637-650. DOI: https://doi.org/10.1111/j.14610248.2008.01182.x
- 8. Brooks, T. M., Mittermeier, R. A., Mittermeier, C. G., Da Fonseca, G. A.,
  Rylands, A. B., Konstant, W. R., ... & Hilton Taylor, C. (2002). Habitat loss
  and extinction in the hotspots of biodiversity. Conservation biology, 16(4), 909923. DOI: https://doi.org/10.1046/j.1523-1739.2002.00530.x
- 9. Burnham, K. P., & Anderson, D. R. (2004). Multimodel inference:
  understanding AIC and BIC in model selection. Sociological Methods &
  Research. 33, 261–304. DOI: https://doi.org/10.1177/0049124104268644
- 624 10. Caswell H. (2001). Matrix population models. 2nd edn. Sinauer Associates.

- 625 11. Cutuli, G., Cannicci, S., Vannini, M., & Fratini, S. (2013). Influence of mating
  626 order on courtship displays and stored sperm utilization in Hermann's tortoises
  627 (*Testudo hermanni hermanni*). Behavioral Ecology and Sociobiology, 67(2),
  628 273-281. DOI: 10.1007/s00265-012-1447-8
- 12. DeAngelis, D.L., & Grimm, V., (2014). Individual-based models in ecology
  after four decades. F1000Prime Reports 6. DOI: https://doi.org/10.12703/P6-39
- 13. Díaz-Paniagua, C., Keller, C., & Andreu, A. C. (1995). Annual variation of
  activity and daily distances moved in adult Spur-thighed tortoises, *Testudo graeca*, in southwestern Spain. Herpetologica, 51: 225-233. DOI:
  http://www.jstor.org/stable/3892590
- 14. Díaz-Paniagua, C., Keller, C., & Andreu, A. C. (1996). Clutch frequency, egg
  and clutch characteristics, and nesting activity of spur-thighed tortoises, *Testudo graeca*, in southwestern Spain. Canadian Journal of Zoology, 74(3), 560-564.
  DOI: 10.1139/z96-061
- biaz-Paniagua, C., Keller, C., & Andreu, A. C. (1997). Hatching success, delay
  of emergence and hatching biometry of *Testudo graeca* in southwestern Spain.
  Journal of Zoology, London, 243: 543-553. DOI: https://doi.org/10.1111/j.14697998.1997.tb02800.x
- 643 16. Díaz-Paniagua, C., Keller, C., & Andreu, A. C. (2001). Long-term demographic
  644 fluctuations of the spur-thighed tortoise *Testudo graeca* in SW
  645 Spain. Ecography, 24(6), 707-721. DOI: 10.1111/j.1600-0587.2001.tb00532.x
- 17. Fahrig, L. (2007). Non-optimal animal movement in human-altered
  landscapes. Functional Ecology, 21(6), 1003-1015. DOI:
  https://doi.org/10.1111/j.1365-2435.2007.01326.x
- 649 18. Galeotti, P., Sacchi, R., Rosa, D. P., & Fasola, M. (2007). Olfactory
  650 discrimination of species, sex, and sexual maturity by the Hermann's tortoise
  651 *Testudo hermanni*. Copeia, 2007(4), 980-985. DOI:
  652 https://www.jstor.org/stable/25140714
- 653 19. Gascoigne, J., Berec, L., Gregory, S., & Courchamp, F. (2009). Dangerously few
  654 liaisons: a review of mate-finding Allee effects. Population Ecology, 51:355–
  655 372. DOI: 10.1007/s10144-009-0146-4
- 656 20. Grimm, V., Berger, U., Bastiansen, F., Eliassen, S., Ginot, V., Giske, J., ... &
  657 Huth, A. (2006). A standard protocol for describing individual-based and agent-

658	based models. Ecological modelling, 198(1), 115-126. DOI:
659	https://doi.org/10.1016/j.ecolmodel.2006.04.023
660	21. Grimm, V., Berger, U., DeAngelis, D. L., Polhill, J. G., Giske, J., & Railsback,
661	S. F. (2010). The ODD protocol: a review and first update. Ecological
662	modelling, 221(23), 2760-2768. DOI: 10.1016/j.ecolmodel.2010.08.019
663	22. Hastie, T. J. & Pregibon, D. (1992) Generalized linear models. Chapter 6
664	of Statistical Models in S eds J. M. Chambers and T. J. Hastie, Wadsworth &
665	Brooks/Cole.
666	23. Henle, K., Davies, K. F., Kleyer, M., Margules, C., & Settele, J. (2004).
667	Predictors of species sensitivity to fragmentation. Biodiversity &
668	Conservation, 13(1), 207-251. DOI:10.1023/B:BIOC.0000004319.91643.9e
669	24. IUCN. (2016). IUCN Red List of Threatened Species. International Union for
670	Conservation of Nature, Gland, Switzerland.
671	25. Kaddour, K. B., Slimani, T., El Mouden, E. H., Lagarde, F., & Bonnet, X.
672	(2006). Population structure, population density and individual catchability of
673	Testudo graeca in the central Jbilets (Morocco). Vie et Milieu, 56(1), 49-54.
674	26. Kanagaraj, R., Wiegand, T., Kramer-Schadt, S., Anwar, M., & Goyal, S.P.
675	(2011). Assessing habitat suitability for tiger in the fragmented Terai Arc
676	Landscape of India and Nepal. Ecography 34, 970–981. DOI:
677	https://doi.org/10.1111/j.1600-0587.2010.06482.x
678	27. Kareiva, P. M., & Shigesada, N. (1983). Analyzing insect movement as a
679	correlated random walk. Oecologia, 56(2), 234-238. DOI:10.1007/BF00379695
680	28. Klauschies, T., Coutinho, R.M., Gaedke, U. (2018). A beta distribution-based
681	moment closure enhances the reliability of trait-based aggregate models for
682	natural populations and communities. Ecological Modelling, 38146-77. DOI:
683	10.1016/j.ecolmodel.2018.02.001
684	1. Kramer-Schadt, S., Revilla, E., & Wiegand, T. (2005). Lynx reintroductions in
685	fragmented landscapes of Germany: Projects with a future or misunderstood
686	wildlife conservation?. Biological Conservation, 125(2), 169-182. DOI:
687	https://doi.org/10.1016/j.biocon.2005.02.015
688	2. Kramer Schadt, S., Revilla, E., Wiegand, T., & Breitenmoser, U. R. S. (2004).

- 690 influences on the dispersal of Eurasian lynx. Journal of Applied Ecology, 41(4),
  691 711-723. DOI: https://doi.org/10.1111/j.0021-8901.2004.00933.x
- Lecq, S., Ballouard, J. M., Caron, S., Livoreil, B., Seynaeve, V., Matthieu, L. A.,
  & Bonnet, X. (2014). Body condition and habitat use by Hermann's tortoises in
  burnt and intact habitats. Conservation physiology, 2(1), cou019. DOI:
  10.1093/conphys/cou019
- 4. Méndez, M., Vögeli, M., Tella, J.L., & Godoy, J.A. (2014). Joint effects of
  population size and isolation on genetic erosion in fragmented populations:
  finding fragmentation thresholds for management. Evolutionary Applications, 7,
  506–518. DOI: https://doi.org/10.1111/eva.12154
- 5. Mimet, A., Houet, T., Julliard, R., & Simon, L. (2013). Assessing functional connectivity: a landscape approach for handling multiple ecological requirements. Methods in Ecology and Evolution, 4(5), 453-463. DOI: https://doi.org/10.1111/2041-210x.12024
- Moorcroft, P.R., & Barnett, A. (2008). Mechanistic home range models and
  resource selection analysis: a reconciliation and unification. Ecology 89, 1112–
  1119. DOI: https://doi.org/10.1890/06-1985.1
- 707 7. Niebuhr, B. B., Wosniack, M. E., Santos, M. C., Raposo, E. P., Viswanathan, G.
  708 M., da Luz, M. G., & Pie, M. R. (2015). Survival in patchy landscapes: the
  709 interplay between dispersal, habitat loss and fragmentation. Scientific reports, 5.
  710 DOI: 10.1038/srep11898
- 8. Norris, D.O. & Lopez, K.H., 2011. Hormones and reproduction of vertebrates.
  Vol. 3. Academic Press/Elsevier, Amsterdam; Boston.
- 9. Pearson, R.G., Stanton, J.C., Shoemaker, K.T., Aiello-Lammens, M.E., Ersts,
  P.J., Horning, N., ... & Akçakaya, H.R., (2014). Life history and spatial traits
  predict extinction risk due to climate change. Nature Climate Change 4, 217–
  221. DOI: 10.1038/nclimate2113
- 717 10. Pérez, I., Giménez, A., Anadón, J. D., Martínez, M., & Esteve, M. A. (2002).
  718 Patrones de actividad estacional y diaria de la tortuga mora (*Testudo graeca*)
  719 *graeca*) en el sureste de la Península Ibércia. *Anales de Biología*, 24: 55-65.
- 11. Philip, J.R. (1957). Sociality and sparse populations. Ecology, 38:107–111. DOI:
   https://doi.org/10.2307/1932132

- 722 12. R Core Team. (2014). R: a language and environment for statistical computing.
  723 R Foundation for Statistical Computing, Vienna
- 13. Regos A, Gagne, L, Alcaraz-Segura, D, Honrado, J.P., Domínguez, J. (2019)
  Effects of species traits and environmental predictors on performance and
  transferability of ecological niche models. Scientific Reports, 4221. DOI:
  10.1038/s41598-019-40766-5.
- 14. Reed, J.M., & Dobson, A.P. (1993). Behavioural constraints and conservation
  biology: Conspecific attraction and recruitment. Trends in Ecology & Evolution,
  8, 253–256. DOI: 10.1016/0169-5347(93)90201-Y.
- 15. Rodríguez-Caro, R. C., Graciá, E., Anadón, J. D., & Gimenez, A. (2013).
  Maintained effects of fire on individual growth and survival rates in a spurthighed tortoise population. European journal of wildlife research, 59(6), 911913. DOI:10.1007/s10344-013-0764-1.
- 16. Rodríguez-Caro, R., Wiegand, T., White, E., Sanz-Aguilar, A., Giménez, A.,
  Graciá, E., van Benthem, K.J., Anadón, J.D. (2019) A low cost approach to
  estimate demographic rates using inverse modelling. Biological Conservation,
  237(358):365. DOI: 10.1016/j.biocon.2019.07.011.
- 739 17. Roques, S., Díaz-Paniagua, C., & Andreu, A. C. (2004). Microsatellite markers
  reveal multiple paternity and sperm storage in the Mediterranean spur-thighed
  tortoise, *Testudo graeca*. Canadian Journal of Zoology, 82(1), 153-159. DOI:
  https://doi.org/10.1139/z03-228
- 18. Stiebens, V. A., Merino, S. E., Roder, C., Chain, F. J., Lee, P. L., & Eizaguirre,
  C. (2013). Living on the edge: how philopatry maintains adaptive potential.
  Proceedings. Biological sciences, 280(1763), 20130305.
  DOI:10.1098/rspb.2013.0305
- 19. Sala, O. E., Chapin, F. S., Armesto, J. J., Berlow, E., Bloomfield, J., Dirzo, R., ...
  & Leemans, R. (2000). Global biodiversity scenarios for the year
  2100. Science, 287(5459), 1770-1774. DOI: 10.1126/science.287.5459.1770
- 20. Sanz-Aguilar, A., Anadón, J. D., Giménez, A., Ballestar, R., Graciá, E., & Oro,
  D. (2011). Coexisting with fire: the case of the terrestrial tortoise *Testudo graeca* in Mediterranean shrublands. Biological Conservation 144: 1040-1049.
  DOI: 10.1016/j.biocon.2010.12.023

- 21. Schjorring, S. (2001). Ecologically determined natal philopatry within a colony
  of great cormorants. *Behavioral Ecology* 12, 287–294. DOI:
  10.1093/beheco/12.3.287
- 22. Schooley, R. L., & Wiens, J. A. (2003). Finding habitat patches and directional
  connectivity. Oikos, 102(3), 559-570. DOI: 10.1034/j.1600-0706.2003.12490.x
- 759 23. Stubben, C., & Milligan, B. (2007). Estimating and analyzing demographic
  760 models using the popbio package in R. Journal of Statistical Software, 22(11), 1761 23. DOI: 10.18637/jss.v022.i11
- 762 24. Tirok, K. & Gaedke, U. (2010) Internally driven alternation of functional traits
  763 in a multispecies predator-prey system. Ecology, 91, 1748-1762, DOI:
  764 10.1890/09-1052.1
- 765 25. Tischendorf, L., & Fahrig, L. (2000). On the usage and measurement of
  766 landscape connectivity. Oikos, 90(1), 7-19. DOI: https://doi.org/10.1034/j.1600767 0706.2000.900102.x
- 768 26. Tucker, M.A., Böhning-Gaese, K., Fagan, W.F., Fryxell, J.M., Van Moorter, B.,
  769 Alberts, S.C., ... & Mueller, T. (2018). Moving in the Anthropocene: Global
  770 reductions in terrestrial mammalian movements. Science 359, 466–469. DOI:
  771 https://doi.org/10.1126/science.aam9712
- 772 27. Villard, M.-A., & Metzger, J.P. (2014). REVIEW: Beyond the fragmentation
  773 debate: a conceptual model to predict when habitat configuration really matters.
  774 Journal of Applied Ecology 51, 309–318. DOI: https://doi.org/10.1111/1365775 2664.12190
- 28. Wosniack, M. E., Santos, M. C., Pie, M. R., Marques, M. C. M., Raposo, E. P.,
  Viswanathan, G. M., & da Luz, M. G. E. (2014). Unveiling a mechanism for
  species decline in fragmented habitats: fragmentation induced reduction in
  encounter rates. Journal of The Royal Society Interface, 11(91), 20130887. DOI:
  https://doi.org/10.1098/rsif.2013.0887

**Appendix A.** Overview, Design concepts, and Details protocol of the individual-based model used in this study

## STEPLAND: a spatially explicit individual-based model to assess population viability

#### Purpose

The main objective of STEPLAND is to investigate the effects of biological traits, population characteristics (e.g., adult survival) and the habitat landscape, on population viability over a timespan of some hundreds of years. The model was implemented into Python 2.7 and its code will be available in XX repository.

#### Entities, attributes and spatial scales

STEPLAND comprises two different entities: landscapes and individuals. In this study we focus on individuals of the tortoise species *Testudo graeca*, and landscapes comprise an area of 3 km  $\times$  3 km composed of a 10 m  $\times$  10 m grid of cells (although other areas and grids are also possible). Each cell is characterised by its position (x and y coordinates) and its habitat category:

- H1: intensive land use
- H2: traditional agriculture land
- H3: natural flat areas
- H4: natural areas on slopes
- H5: non-permeable infrastructures

These categories were taken from a previous study that investigated the effects of habitat loss and fragmentation on tortoise movements (Anadón et al. 2012). Landscapes are enclosed by a non-permeable border and host closed and isolated populations (Fig. A1).

Tortoises are characterised by the following attributes: gender, age, time for stored sperm caducity, their actual location and their past locations. Additionally, the nine parameters AU1, AU2, dHB, rHB, H1W, H2W, H3W, H4W and H5W govern their movement and the location of their focal-point attractor. Except for H5W, these parameters were taken from the movement analysis of Anadón et al. (2012). See the description of the movement process for further information.



**Figure A1.** Landscapes used in STEPLAND. Habitat categories include: non-permeable infrastructures (black), intensive land use (grey), traditional agriculture lands (yellow), natural flat areas (light green) and natural areas on slopes (dark green). Grey landscapes show the barrier effects of landscape features for the mating submodel. The natural landscape lacks internal barriers, but roads and intensive agriculture divide the human-altered landscape.

### Input, initialisation and output

Inputs include the landscape, the initial population and model parameterisation; they are all included in a SQL database that is directly read by the model. Initially,  $N_0 = 45$ , 90 and 180 tortoises are randomly distributed within the landscape, which represents the medium and low densities that are typically found in the natural species populations in SE Spain (Anadón et al. 2009). Note that the initial populations (individuals and their locations) were the same in all the simulation scenarios that started with the same  $N_0$ . To avoid biases caused by extreme spacing of the initial locations of individuals, we tested 15 different sets of locations and selected for our study a set with medium mean minimum Euclidean distances between females and males (Table A1). Each parameterisation was then simulated 64 times in a given landscape using the same initial population.

**Table A1.** Test for biases in extinction probabilities after 200 years due to individuals' initial locations. We tested 15 sets of locations (5 per initial density) in the natural landscape with a standard parameterisation (Appendix C). Extinction rates were more variable for  $N_0$ =45 sets. For our study, three sets of locations with mean minimum Euclidean distances between males and females close to the obtained means from the tested simulations were selected.

Initial density	Mean extinction probability ± SD	Mean min. fem-male distance ± SD (m) in the tested sets of the initial locations	Mean min. fem-male distance ± SD in the selected set of initial locations
N <sub>0</sub> =45	0.51±0.18	673.5±303.4	627.5±174.35
<i>N</i> <sub>0</sub> =90	0.06±0.07	422.5±142.2	417.4±150.62
<i>N</i> <sub>0</sub> =180	0±0	295.9±38.8	338.6±169.19

The different age classes of the initial population follow a stable age distribution predicted by an age-stage structured deterministic matrix population model (Caswell, 2001) that was derived using the POPBIO package in the R software (Stubben & Milligan, 2007; R Core Team, 2014; R scripts are shown in Appendix B). The demographic parameters of the matrix model are the same as the demographic parameters of the simulation model, but the growth rate predicted by the deterministic matrix model does not consider demographic stochasticity and mate-finding Allee effects (Appendix B). The complete model parameterisation is shown in Appendix C.

Data output consists of CSV files that comprise the surviving tortoises of the population and their attributes (location in the landscape, gender and age) for every 10 years during the simulated years. Own scripts, developed in R (R Core Team 2016) and Python 2.7, allow these massive outputs to be processed. For operational reasons, the simulation stops if the total number of tortoises in the landscape exceeds 4,500.

#### Temporal scales, process overview and scheduling

In STEPLAND, each time step represents 1 day, while outputs are summarised for 10year periods. All simulations were run for 200 years (corresponding to approximately eight tortoise generations).

Tortoises are subject to the basic processes movement, reproduction, mortality and aging. They can move up to 4 times a day according to their attributes and movement parameters, the time of the year and the habitat category of surrounding cells (as modelled and parameterised by Anadón et al. 2012). Adult females can reproduce up to 3 times/year, as evidenced by own data and previous studies that have addressed the reproduction of this species (Díaz-Paniagua et al. 1996, 1997). In our model, the nutritional status of tortoises does not affect reproduction. However, spatial effects may hinder mating, including a too large distances between females and their nearest adult males, or presence of non-permeable barriers between them. Annual mortality rates are implemented through a single event of mortality at the end of the year. The age of surviving tortoises is then updated, and their movement attributes may also be updated (i.e., in some scenarios the last location of the year may be used as their focal point attractor; see movement submodel for further information). Figure A2 provides a general overview of the model and the following paragraphs describe the main processes in more detail.



**Figure A2.** Main processes implemented in STEPLAND. Inputs are the landscape and the initial tortoise population. During the simulation, the individuals are subject to the basic processes: movement, reproduction, mortality and aging; and may be subject to parameter updates (e.g., variations in the location of their focal point attractor). In this study, we simulated population dynamics for 200 years and evaluated the sensitivity of extinction rates to variations in the SITE and DIST parameters.

#### Movement

STEPLAND includes the movement model developed and parameterised by Anadón et al. (2012). They used individual-based modelling and pattern-oriented modelling to parameterise the movement model based on radio-tracking data. One main result of Anadon et al. (2012) is that land-use intensification in SE Spain had a strong impact on spur-thighed tortoises' movement and behaviour. While tortoises showed high site fidelity and slight individual variability in natural landscapes, movement and behaviour varied strongly among individuals in human-altered landscapes. Particularly for females, a wide range of movement patterns was found, which ranged from strong home behaviour to unbounded movements. Anadón et al. (2012) determined a total of 5,627 individual parameterisations of annual movements that were compatible with the radio-tracking data (Table A2). Females in human-altered landscapes formed the largest set of individual movement parameterisations (Table A3).

**Table A2.** Subset as an example of the parameterisations assigned to individuals for their movements as determined by Anadón et al. (2012). In our simulations, these parameterisations were randomly assigned to the initial individuals and to the newborns (depending on sex and landscape type).

Sex and habitat					Habit	at-deper	ndent we	ights for	cells
type	AU1	AU2	dHB	rHB	H1W	H2W	H3W	H4W	H5W
male_natural	0.05	0.7	50	25	0.1	5	5	8	0
male_natural	0.05	0.65	75	3	0.1	20	11	3	0
male_natural	0.75	0.75	25	10	0.1	15	2	13	0
female_natural	0.1	0.25	50	7	0.1	11	17	10	0
female_natural	0.05	0.3	50	10	0.1	5	2	7	0
female_natural	0.15	0.9	25	10	0.1	17	5	13	0
male_alterned	0.7	0.85	75	3	0.1	5	8	18	0
male_alterned	0.65	0.65	100	10	0.1	5	15	6	0
male_alterned	0.25	0.5	75	3	0.1	4	13	10	0
female_alterned	0.55	0.85	175	60	0.1	9	12	2	0
female_alterned	0.25	0.65	10000	30	0.1	8	16	1	0
female_alterned	0.25	0.85	500	4	0.1	15	10	3	0

AU1: autocorrelation between consecutive movement steps; AU2: autocorrelation between cell steps; dHB; distance threshold below which no homing behaviour occurs rHB: temporal delay (in days) of homing behaviour as dHB was exceeded; H1W: weight for cells with intensive land use: H2W: weight for cells with traditional agriculture; weight for cells with H3W: natural flat areas; weight for cells with H4W: natural areas on slopes; weight for cells with H5W: non-permeable infrastructures.

**Table A2.** Number of individual parameterisations for the movement process, as determined by Anadón et al. (2012).

	Natural landscape	Human-altered landscape
Males	139	51
Females	155	5,282

Tortoises older than the age threshold given by the parameter age\_first\_mov, can change their location up to 4 times/day (i.e., four movement steps; Fig. A3). The active period of one day is divided into four 2- to 3-hour periods, and each of these periods a random trial is made with the PMOV parameter being the probability of moving during a period. PMOV is dependent of the month and tortoise gender, for example, tortoises are inactive during their hibernation and aestivation periods (Fig. A4a). If the PMOV trial indicated movement, DMOV is the discrete probability of moving *s* cell steps in each movement step ("cell-to-cell movements"; Fig. A3). DMOV varies between adult males and females (Fig. A4b). It was also assumed that subadults have a reduced movement capability than adults. To implement this, the number of cell-to-cell steps in DMOV was assumed to increase gradually with age until the reproductive age:

DMOV <sub>subadult</sub> =round	(DMOV <sub>adult</sub> ×(subadult's	s age)/(reproductive a	ge)) (1)	J
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						10 m ◀──▶	
							10 m
6	1				3		
5	2			1	2		
4		3	4	5			
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2	1						
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**Figure A3.** Exemple of the simulated movement of a tortoise during a given day. Starting from the arrow, the tortoise takes three movement steps during the day (represented by different colours), with six, five and three cell-to-cell steps each (movement directions are represented by numbers).



**Figure A4.** Discrete probability distributions of the parameters PMOV (a); DMOV of females and males (b) and clutch\_size (c).

In each cell-to-cell movement step, three submodels interact in the selection of one of the eight neighbouring cells surrounding the tortoise by determining their associated weights, which are then multiplied to calculate its final probability to be occupied:

i) Autocorrelation: describes the tendency of keeping the previous direction. The eight cells surrounding the present location are numbered from i = 1, ..., 8, where the index i = 1 indicates the cell-by-cell step following exactly the previous direction. Then i = 2, ..., 8 are numbered clockwise. The weights  $d_i$  of the neighbouring cells with indices i = 1, ..., 5 (movement to the right) are calculated as:



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(2)

For the remaining weights that represent movement to the left, isotropic movement (i.e.,  $d_6 = d_4$ ,  $d_7 = d_3$ ,  $d_8 = d_2$ ) is assumed. Finally, weights  $d_i$  are normalised to add up to 1. As cell-to-cell steps are nested inside movement steps, two parameters were used to describe autocorrelation in movement. The AU1 parameter represents the autocorrelation between consecutive movement steps and affects only the first cell-to-cell step of each movement step. The AU2 parameter is the autocorrelation between the following cell-to-cell steps. Thus for the first cell-to-cell step, AU1, and AU2 were used for the other cell steps (Fig. A5).



**Figure A5.** Schematic representation of the autocorrelation submodel. Arrows differentiate steps or cell step transitions using parameters AU1 (in black) or AU2 (in white).

ii) *Homing behaviour:* modelled as the tendency of an individual to maintain a stable home range around a focal point attractor. Weights  $b_i$  are numbered in the same way as the weights for autocorrelation. We used the first location of the individual as the focal-point attractor. To model homing behaviour, two critical distances were assumed around the focal point. *D* is the individual's distance to its focal-point attractor and the *dHB* parameter is then the distance threshold below which no homing behaviour occurs. Firstly, if the individual is located close to the focal point, equal weights for the different directions are given (i.e.,  $b_i = 1/8$  for all i = 1,...8, when D < dHB). Thus, if the individual is close to its focal point attractor, no homing behaviours occurs. Secondly, if the individual is located farther away than distance 2\*dHB from the focal point attractor, the bias towards the focal point is maximal and the cell *j* closest to the focal-point

attractor has a probability of  $b_j = 1$  of being occupied, while the remaining seven cells have probability of  $b_i = 0$ . In intermediate situations in which dHB < D < 2dHB, the bias of  $b_j$  increases from 1/8 proportionally with distance D up to 2\*dHB with the formula:

$$b_j = (D - dHB)/(dHB)$$
 and  $b_i = (1 - b_j)/7$  for  $i \neq j$  (3)

In order to enhance the model's biological realism, a temporal delay was considered so that homing behaviour came into force only after exceeding the distance threshold *dHB* from the focal point. This allows tortoises to temporally leave their most regularly used areas to meet specific needs, such as egg-laying in females or mate-searching in males (Díaz-Paniagua et al. 1995, Pérez et al. 2002). The *rHB* parameter is the temporal delay (in days) of homing behaviour as *dHB* was exceeded (R, in days) (Fig. A6).



**Figure A6.** Schematic representation of the homing behaviour submodel. Asterisks indicate the location of tortoises, with distances  $D_1$  and  $D_2$  to the focal point attractor. NB:  $dHB < D_1 < 2dHB$  and  $D_2 > 2dHB$ .

STEPLAND allows to leave the focal point attractors of the tortoises during their entire lifespan in the same location (representing philopatry; parameter SITE = Yes); or can, alternatively, be reassigned to the individuals' last position at the end of each year (allowing individuals 'wandering; parameter SITE = No) (Table 1; Appendix C). *Habitat dependence:* in relation to habitat quality, the eight neighbouring cells receive habitat-dependent weights to be occupied through parameters H1W (intensive land use), H2W (traditional agriculture), H3W (natural flat areas), H4W (natural areas on slopes) and H5W (non-permeable infrastructures). The weights of the eight cells are subsequently normalised to add up to 1 (Fig. A7).



Figure A7. Schematic representation of the habitat dependence submodel. Cell colours represent habitat categories.

#### Reproduction

Only adult individuals (aged  $\geq$  7 years for males and  $\geq$  10 years for females) reproduce (Rodriguez-Caro et al. 2013). Sperm storage, a vital trait of spur-thighed tortoises (Roques et al. 2004), was modelled as a temporal dissociation between mating and clutching.

iv) Mating: once a year at the beginning of spring, mature females localise mature males in their surroundings (both with ages over female\_repro\_age and male\_repro\_age in Appendix C). Given the spatial constraints to find mates, mating was assumed to occur only if the Euclidean distance between the female and the closest male was (at the time scheduled for mating) below a minimal distance given by the DIST parameter. Moreover, mating occurred only if there was no barrier between males and females (e.g. main roads or dense human infrastructures) (Fig. A1). Sperm can be viable in the oviducts of females of the

species *T. graeca* for as long as 3-4 years (Cutuli et al. 2013). Therefore, we modelled sperm-caducity in a simple way: females that mated can be, for the same probability, reproductive for the following 3 years (represented by the sperm\_cad parameter). If females find mates, then the stored sperm-caducity timer is reset to zero; otherwise, 1 is added to the sperm-caducity timer. Females do not reproduce if their sperm-caducity timer reaches 3 and if they do not find a male during the next breeding season.

v) *Clutching*: the repro\_female parameter is the probability that females with available sperm will be gravid when clutching can take place in spring (Díaz-Paniagua et al. 1996, 1997; unpublished data). The number of eggs that gravid females lay is determined from the discrete probability distribution clutch\_size (Fig. A4c). Clutch viability is later simulated by applying mortality rates for each egg (Díaz-Paniagua et al. 1997, Díaz-Paniagua et al. 2006; see the next paragraph and Appendix C for further details regarding model parameterisation).

### Mortality

A single mortality event occurs at the end of each simulated year. Annual survival rates vary among age classes, including eggs ( $S_0$ ), immature individuals ( $S_{immatures}$ ; aged 1-3), subadults ( $S_{subadults}$ ; aged 4-6), and adults ( $S_{adults}$ ; aged  $\geq$ 7) (Díaz-Paniagua et al. 2001, 2002, Sanz-Aguilar et al. 2011). Eggs are identified as individuals whose age equals 0, while age cuttings apply (subadult\_age, female\_adut\_age and male\_adult\_age) for the transition among the remaining classes (Table C1 in Appendix C1). The individuals (and their attributes) that died disappear from the population.

#### Parameters uploading

- vi) *Ageing:* the age of each living individual is increased at the end of the year by one.
- vii) *Temporal changes in movement characteristics:* in our model, the individuals may maintain the same focal point attractor for their whole lifetime (which emulates philopatry around birthplaces) or they may change it annually to their last location in every year (which allows wandering of individuals). At the beginning of the simulation, the parameterisations of the tortoises related with

movement (i.e., the parameter set AU1, AU2, dHB, rHB, H1W, H2W, H3W, H4W and H5W) are randomly assigned to the newborns from parameterisations determined in Anadon et al. (2012), according to the sex of the individual and landscape. The movement parameterisation is maintained for the whole life of the tortoise, thus simulating "individual personalities" (Golubović et al. 2014, 2017; see Appendix C for further details regarding model parameterisation).

#### **Design concepts**

*Basic principles*: the model assumes that: i) tortoises maintain age-class transitions and survival rates in human-altered landscapes; ii) habitat loss and fragmentation cause changes in their movement and behaviour (as reported by Anadón et al. 2012); iii) non-permeable barriers hinder tortoise mating.

*Emergence:* simulations may result in different population dynamics due to variations in tortoises' reproductive success.

*Adaptation:* tortoises face fragmentation by modifying their movement decisions. They avoid intensive land uses and non-permeable infrastructures.

*Objective:* tortoises do not have any particular objective implemented into simulations but, as previously mentioned, they modify their movement decisions according to habitat characteristics.

*Sensing:* tortoises are able to detect habitat loss and fragmentation in landscapes of 3 km  $\times$  3 km (i.e., natural vs. human-altered landscapes). They also recognise their surrounding habitat in 30 m  $\times$  30 m (the eight adjacent grid cells).

*Interaction:* tortoises are able to mate if they are separated by a distance shorter than that given by the DIST parameter and if there are no geographical barriers between them.

Learning: tortoises do not change their behaviour according to their experience.

Prediction: tortoises do not have the capacity to predict future conditions.

*Stochasticity:* tortoises' initial location, and assigning movement parameters to them, are purely stochastic processes. The remaining "reproduction", "movement",

"mortality" processes become semi-stochastic by implementing discrete probability distributions.

Collectives: there are no defined collectives in the model.

Observation: output serves to calculate the number of living individuals of a

given population at a given time and, consequently, population viability over time.

#### References

Anadón, J. D., Wiegand, T., & Giménez, A. (2012). Individual-based movement models reveals sex-biased effects of landscape fragmentation on animal movement. *Ecosphere* 3: art64.

Anadón, J.D., Giménez, A., Ballestar, R., & Perez, I. (2009). Evaluation of Local Ecological Knowledge as a method for collecting extensive animal abundance data. Conservation Biology 23: 617-625.

Caswell, H., (2001). Matrix population models: construction, analysis, and interpretation, Second Edition. ed. Sinauer Associates, Sunderland, Massachusetts, USA.

Cutuli, G., Cannicci, S., Vannini, M., & Fratini, S. (2013). Influence of mating order on courtship displays and stored sperm utilization in Hermann's tortoises (*Testudo hermanni hermanni*), Behav. Ecol. Sociobiol. 67 (2013a) 273e281.

Díaz-Paniagua, C. Keller, C., Andreu, A.C. (2001). Long-term demographic fluctuations of the spur-thighed tortoise, *Testudo graeca*, in SW Spain. Ecography, 24: 707-721.

Díaz-Paniagua, C., Andreu, A. C., Keller, C. (2006). Effects of temperature on hatching success in field incubating nests of spur-thighed tortoises, *Testudo graeca*. Herpetological Journal, 16 (3): 249-257.

Díaz-Paniagua, C., Keller, C., & Andreu, A. C. (1995). Annual variation of activity and daily distances moved in adult Spur-thighed tortoises, *Testudo graeca*, in southwestern Spain. Herpetologica, 51: 225-233.

Díaz-Paniagua, C., Keller, C., & Andreu, A. C. (1996). Clutch frequency, egg and clutch characteristics, and nesting activity of spur-thigheed tortoises, *Testudo graeca*, in southwestern Spain. Canadian Journal of Zoology, 74: 560-564.

Díaz-Paniagua, C., Keller, C., & Andreu, A. C. (1997). Hatching success, delay of emergence and hatching biometry of *Testudo graeca* in southwestern Spain. Journal of Zoology, London, 243: 543-553.

**Formatiert:** Spanisch (Spanien-Modern)

Díaz-Paniagua, C., Keller, C., & Andreu, A. C. (2002). Life history and demography of *Testudo greca* in Southwestern Spain. Chelonii, 3: 214-222.

Golubović, A., Anđelković, M., Arsovski, D., Bonnet, X., & Tomović, L. (2017). Locomotor performances reflect habitat constraints in an armoured species. Behavioral Ecology and Sociobiology, 71(6), 93.

Golubović, A., Andjelković, M., Arsovski, D., Vujović, A., Iković, V. Djordjević, S., & Tomović, L. (2014). Skills or strength—how tortoises cope with dense vegetation?. Acta Ethologica, 17(3), 141-147.

Pérez, I., Giménez, A., Anadón, J. D., Martínez, M., & Esteve, M. A. (2002). Patrones de actividad estacional y diaria de la tortuga mora (*Testudo graeca graeca*) en el sureste de la Península Ibércia. *Anales de Biología*, 24: 55-65.

R Core Team. (2014). R: A language and environment for statistical computing. R Foundation for Statistical Computing, Vienna, Austria. URL <u>http://www.R-project.org/.</u>

Rodríguez-Caro, R. C., Graciá, E., Anadón, J. D., & Gimenez, A. (2013). Maintained effects of fire on individual growth and survival rates in a spur-thighed tortoise population. *European journal of wildlife research*, 59(6), 911-913.

Roques, S., Diaz-Paniagua, C., & Andreu, A.C. (2004): Microsatellite markers reveal multi-ple paternity and sperm storage in the Mediterranean spurthighed tortoise, *Testudo graeca*. Can. J. Zool. 82: 153-159.

Sanz-Aguilar, A., Anadón, J. D., Giménez, A., Ballestar, R., Graciá, E., & Oro, D. (2011). Coexisting with fire: the case of the terrestrial tortoise *Testudo graeca* in Mediterranean shrublands. *Biological Conservation* 144: 1040-1049.

Stubben, C. & Milligan, B. (2007). Estimating and analyzing demographic models using the popbio package in R. J. Stat. Softw. 22, 1–23.

**Formatiert:** Spanisch (Spanien-Modern)

Formatiert: Englisch (Großbritannien)

Formatiert: Englisch (Großbritannien)

Formatiert: Englisch (Großbritannien)