# This is the accepted manuscript version of the contribution published as:

Tara, N., **Arslan, M.**, Hussain, Z., Iqbal, M., Khan, Q.M., Afzal, M. (2019): On-site performance of floating treatment wetland macrocosms augmented with dyedegrading bacteria for the remediation of textile industry wastewater *J. Clean Prod.* **217**, 541 – 548

# The publisher's version is available at:

http://dx.doi.org/10.1016/j.jclepro.2019.01.258

# On-site performance of floating treatment wetland macrocosms augmented with dyedegrading bacteria for the remediation of textile industry wastewater

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• Rhizospheric bacteria





#### 1 ABSTRACT

2 Floating treatment wetlands are considered cost-effective remediation options for various 3 types of wastewater. Their effectiveness has been shown in several lab-scale and pilot-scale studies; however, there is a paucity of published data on pilot-scale systems treating genuine 4 wastewater. This study aims to assess the performance of a pilot-scale system, carrying 5 6 Phragmites australis in combination with three plant growth promoting and pollutant-7 degrading bacteria (Acinetobacter junii strain NT-15, Rhodococcus sp. strain NT-39, and Pseudomonas indoloxydans strain NT-38) for the treatment of textile industry wastewater 8 9 (Interloop Limited, Faisalabad, Pakistan). Fifteen floating treatment wetlands macrocosms were established employing plants and bacteria separately or in combination. Each unit was 10 capable to carry 1000-liter of wastewater and the system was operated in a batch-wise mode 11 for the period of 2 years. After a year of installation, performance of all FTWs units was 12 optimal. A high removal in organic and inorganic pollutants was observed in the vegetated 13 tanks, whereas combined application of plants and bacteria further enhanced the removal 14 performance, i.e., chemical oxygen demand was reduced to 92%, biochemical oxygen 15 demand to 91%, color to 86%, and heavy metals to approximately 87% in the wastewater. 16 17 The augmented bacteria displayed persistence in water as well as in the roots and shoots of P. australis suggesting a potential partnership with the host towards enhanced performance. 18 Treated wastewater met the National Environmental Quality Standards of Pakistan to be 19 20 discharged in the surface water without any potential risks. This pilot-scale study is a step forward towards sustainable remediation of the textile wastewater in the field. 21

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Keywords: Textile wastewater, wastewater treatment, floating treatment wetlands, plantbacteria partnership

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#### 26 INTRODUCTION

Textile industry plays a key role in the economy of many developing countries including 27 Pakistan. However, it also causes severe water pollution due to the direct wastewater 28 29 discharges in the surface water (Noreen et al. 2017). The textile industry of Faisalabad (Pakistan) represents a wide array of processes such as desizing, bleaching, mercerizing, 30 dyeing, printing and finishing of products. This results into production of a broad range of 31 32 organic and inorganic pollutants, which are discharged into the Ravi river and Chenab river through the Paharang and Madhuana drains without any pre-treatment (Azizullah et al. 2011; 33 Imtiazuddin et al. 2012). These discharges consequently deteriorate the water quality leading 34 35 to harmful effects on aquatic organisms, irrigating crops, and on human health (Ullah et al. 2017; Najam-us-Sahar et al. 2017; Daud et al. 2017). Since a few years, this condition has 36 become worse as the level of toxic pollutants in the river's body has surged dramatically. 37

Many physicochemical techniques are effective in treating textile wastewater but they 38 demand high operational costs, engineering skills, environmental incursions, labor 39 administration, and several other operational processes (Liu et al. 2009; Srinivasan et al. 40 2014). Contrarily, advances in ecological engineering have made us develop environmental-41 friendly methods for the effective treatment of contaminated water (Giannetti et al., 2002; 42 Ijaz et al., 2015; Saumya et al., 2105; Greenway, 2017). One such method is the installation 43 of floating treatment wetlands (FTWs) in which emergent aquatic macrophytes are planted 44 artificially on a floating raft that allows plants to grow hydroponically on the water body 45 46 (Nahlik and Mitsch 2006; Wu et al., 2017; Shahid et al. 2018). Plant roots hang down to the pelagic zones in the water column offer both mechanical and biological filtering. Mechanical 47 48 filtering is achieved through the physical effect of plant roots such as sedimentation, adsorption, filtration, etc. The biological filtering however is the result of bacterial 49

50 degradation and plant uptake (Zhang et al. 2106; Merkhali et al. 2015). In both situations, plant roots play several important roles such as: (1) their presence reduce the water 51 turbulence and avoid re-suspension of sediments (Smith and Kalin 2000); (2) they attach 52 53 suspended matter onto the root surface which is subsequently precipitated in the systems bottom or adsorbed on to the biofilm; and (3) when adsorbed, bacterial action results into the 54 degradation of organic pollutants (Prajapati et al. 2017). Macrophytes can hinder the algal 55 growth by competing for nutrition and sunlight (Li et al. 2010). Lastly, these systems also act 56 as habitate for fish harvesting (Faulwetter et al. 2011), livestock grazing (Ladislas et al. 57 2013), improve landscape aesthetics (Ijaz et al. 2015), and provide a significant amount of 58 biomass to be used for bio-energy purposes (Shahid et al. 2018). 59

Floating wetlands have been traditionally employed to treat variety of wastewater 60 such as stormwater (Headley et al. 2008), sewage effluents (Todd et al. 2003), municipal 61 wastewater (Arshad et al. 2017; Sirage et al. 2017), industrial wastewater (Li et al. 2012), 62 acid mine drainage and poultry processing wastewater (Smith and Kalin 2000; Todd et al. 63 2003). These systems, however, are limited in certain contexts such as poor stability and less 64 efficient purification performance. Among them, less potential of indigenous bacteria toward 65 the degradation of toxic organic compounds and weak metabolic capabilities of plants are 66 prominent ones (Arslan et al. 2017). To improve the treatment efficiency of FTWs, additional 67 approaches have been recommended among which inoculation of plant growth promoting 68 (PGP) and pollutant-degrading bacteria is an effective method (Shehzadi et al. 2014; Saleem 69 et al. 2018). To date, a few studies have explored the efficiency of FTWs in the presence of 70 bacterial partnership for the remediation of different wastewaters (Kabra et al. 2013; 71 Watharkar et al. 2015; Rehman et al. 2018). Many of them were conducted at microcosm 72 scale for which field-scale performance is speculated. In this study, *in situ* pilot-scale FTW 73 system was established in combination with dye degrading bacteria for the enhanced cleanup 74

of wastewater of a textile industry, i.e., Interloop Limited, Khurrianwala, Faisalabad,
Pakistan. There were enhanced pollutant removal and toxicity reduction in the bacterially
assisted FTWs as a result of effective plant-bacterial partnership towards remediation of
textile industry wastewater.

79

#### 80 MATERIAL AND METHODS

#### 81 Collection and characterization of textile wastewater

The untreated textile wastewater was obtained from an outlet of the wastewater equalization tank of Interloop Limited, Khurrianwala, Faisalabad, Pakistan. The wastewater was analyzed for several physicochemical parameters such as pH, color, electrical conductivity (EC), chemical oxygen demand (COD), 5-day biochemical oxygen demand (BOD<sub>5</sub>), total dissolved solids (TDS), total suspended solids (TSS), total nitrogen (TN), phosphate (PO<sub>4</sub><sup>-2</sup>), Phenols, Chlorides, Sulphates, heavy metals and toxicity by using standard methods (APHA, 2005).

#### 88 Bacterial strains

Three bacterial strains, i.e., Acinetobacter junii NT-15 (NCBI accession: MF478980), 89 Pseudomonas indoloxydans NT-38 (NCBI accession: MF478985), and Rhodococcus sp. NT-90 91 39 (NCBI accession: MF326802), were used in the present study. The strain A. junii NT-15 was isolated from the activated sludge, whereas strain P. indoloxydans NT-38 and 92 Rhodococcus sp. NT-39 were isolated from the root interior and rhizosphere of Polygonum 93 aviculare and Poa labillardierei, respectively (Tara et al. 2018). These strains were screened 94 because of their high dye resistance and decolorization potential. These bacterial strains were 95 cultivated as separate cultures at 30 °C for 24 h in Luria-Bertani (LB) broth. The cells were 96 harvested by centrifugation (×12,000 rpm) for 5 minutes at 4 °C and resuspended together at 97

ratio of 1:1:1 in 0.9% (w/v) sterile NaCl solution. The cell suspension of each pure culture
was adjusted as per the guidelines of the turbid metric method (Sutton, 2011). In the end, one
liter of bacterial consortium was used as an inoculum in the FTWs as per experimental
design.

#### 102 Construction and implementation of FTWs for textile wastewater treatment

Floating treatment wetlands were built in the vicinity of Interloop Limited Khurrianwala, 103 Faisalabad (Fig. 1A-D). Plastic tanks of 1000 liters of capacity (dimensions: 1.2 meter length, 104 1.2 meter width, and 1.2 meter height) were used to establish macrocosms. These tanks were 105 painted from all the sides to prevent algal growth. Accordingly, floating mats of equal 106 dimensions were prepared using polyethylene sheet as recommended earlier (Ijaz et al. 2015). 107 Aluminum foil was used to cover all sides of the mats in order to protect them from 108 ultraviolet radiations. Eight holes of equal diameter were drilled in each floating mat and 109 planted with 24 healthy seedlings of *Phragmites australis*: three plants, with the height of 110 approximately 60 cm and weight of 45 to 65 g, were inserted in each hole of the mat. P. 111 *australis* was selected based on its successful application in phytoremediation of industrial 112 wastewater as well as its ability to resist high concentrations of azo dyes (Todorovics et al. 113 2005; Stefanakis et al. 2014; Ha and Anh 2017; Rehman et al. 2018). These seedlings were 114 previously grown in the nursery developed in the vicinity of NIBGE, Faisalabad. Soil and 115 coconut shaving were used to support the plant seedlings. A total of 15 macrocosms were 116 established by placing floating mats on the tank surface. The plant seedlings were then 117 allowed to augment roots in tap water for a period of 1-month. When the plant growth was 118 119 optimum, tap water was replaced with the textile wastewater. The one-liter inoculum of previously isolated bacterial strains was introduced in the FTWs as per the experimental 120 design. Different treatments were established to study: (1) effect of textile effluent on plant 121

122 growth; (2) effect of vegetation on textile effluent degradation; (3) effect of bacterial augmentation on textile effluent degradation; (4) effect of bacterial augmentation on plant 123 growth and textile effluent degradation; and (5) plant growth in tap water (control). The 124 sequencing fill-and-draw batch mode method was used to withdraw decontaminated water. 125 Hydraulic retention time was set to 10 days based on lab-scale observations where a 126 continuous decrease in organic and inorganic load was observed in initial 10-days. The 127 system was operated for 2 years from March 2016 to March 2018. The water samples of 128 approximately 1.5 liters were collected every two days. However, in this article, the results 129 are reported when the system's performance was optimum/constant (after 6 months of 130 installation). The treatment performance was established by measuring wastewater 131 parameters such as pH, EC, TDS, TSS, COD, BOD<sub>5</sub>, TN, PO<sub>4</sub><sup>-2</sup>, color, phenols, chlorides, 132 sulphates, and heavy metals (Fe, Ni, Cr, and Cd) immediately after sample collection; 133 elsewise samples were stored until analysis according to standard methods (APHA, 2005). 134 The experiment was performed in the ambient conditions of the industry for the period of 2 135 years. In this period, precipitation on average was 29 mm (Faisalabad's climate conditions); 136 however, during rainy period, water level was manually maintained to 1000 litters by 137 covering the macrocosms with plastic sheets. 138

## 139 Determination of persistence of inoculated bacteria

140 The survival/persistence of inoculated bacteria in the treated wastewater and plant interior 141 (root and shoot interior) was enumerated by viable plate count method. Samples were 142 collected every 3 months. Shoots and roots were surface sterilized, homogenized, and the 143 slurry was plated onto LB agar plates having 100 mg  $l^{-1}$  of azo dyes as established previously 144 (Ijaz et al. 2015). Similarly, treated wastewater was also plated on LB agar plates and

incubated at 30 °C for 24 h. Restriction fragment length polymorphism (RFLP) analysis was
performed to identify the identity of inoculated strains (Afzal et al. 2012).

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#### 148 Plant biomass

The plant tissues were harvested every 3 months to measure shoot height, root length, shoot dry weight, and root dry weight. Briefly, shoots and roots were harvested one inch above and below the mats, respectively. The root length and shoot height were measured by using hand ruler. To measure dry biomass, plant samples were first transported to the laboratory, placed in an oven at 80 °C for 72 h, and then weighed using a digital balance. Samples were stored at -80 °C until further analysis.

#### 155 Fish toxicity bioassay

156 Fish bioassay was performed once, during the optimal period, to test the toxicity of untreated and treated wastewater. A freshwater dwelling fish species Rohu, Labeorohita rohita 157 (Family: Cyprinidae, Order: Cypriniformes) was selected for the bioassay due to the local 158 ecological significance. Briefly, the species is present in the local streams and lakes and 159 therefore direct discharges may lead to health issues (Khan et al. 2017). This species has 160 previously found to accumulate heavy metals above the natural/background levels (Hamid et 161 al. 2016). In this study, healthy specimens were obtained from Faisalabad Fish Hatchery, 162 treated with KMnO4 solution (0.05 %) for two minutes to remove dermal contamination, and 163 then shifted to glass aquaria containing uncontaminated water. Each fish specimen had an 164 average body length of ~8.5  $\pm$  0.8 cm and weight of ~4.0  $\pm$  0.1 g. The specimens were 165 acclimatized for 15 days prior to the effluent exposure. Subsequently, four fish groups each 166 containing 10 individuals were placed in the aquariums (dimensions: depth 30 cm, width 30 167

168 cm, and length 45 cm). These aquariums were filled with 30 L of treated and untreated
169 wastewater, as per the experimental design. Finally, the survival rate was determined by
170 counting the number of alive fish every 24 h for a period of four days.

#### 171 Long-term maintenance strategy

Maintenance strategy of the system involved harvesting of the plants and inoculation of the bacterial consortium in the wetland macrocosms after every 3 months during the experimental period. Accordingly, long-term system's performance was also monitored by measuring temporal response of basic water quality parameters, i.e., COD, BOD, Color, TSS, TDS, TN, Phosphates.

#### 177 Statistical analysis

R-statistical language was used to compare different treatments for water quality parameters
(pH, EC, COD, BOD<sub>5</sub>, TDS, TSS, color, Phenols, Chlorides, and Sulphates), nutrients (TN,
and PO4<sup>-2</sup>), heavy metals (Fe, Cr, Ni, and Cd), plant growth parameters (root length, shoot
length, fresh biomass, and dry biomass), and fish toxicity assay. The comparison was made
through one-way ANOVA and Duncan's test was used after testing homogeneity of variance.

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#### 184 RESULTS AND DISCUSSION

#### 185 Wastewater characteristics

In developing countries, unprecedented advances in textile industry have escalated the discharge of wastewater into the natural streams (Vikrant et al. 2018). In this study, physicochemical analysis of the textile industry wastewater displayed a high pollution due to COD (513 mg  $l^{-1}$ ), BOD (283 mg  $l^{-1}$ ), TDS (5251 mg  $l^{-1}$ ), TSS (324 mg  $l^{-1}$ ), Phenols (0.85 mg

190  $\Gamma^{-1}$ ), Chlorides (1383 mg  $\Gamma^{-1}$ ), Sulphates (311 mg  $\Gamma^{-1}$ ), Fe (14.4 mg  $\Gamma^{-1}$ ), Ni (7.57 mg  $\Gamma^{-1}$ ), Cr 191 (9.67 mg  $\Gamma^{-1}$ ), and Cd (0.88 mg  $\Gamma^{-1}$ ). Their concentrations were higher than the established 192 National Environmental Quality Standards (NEQS) of Pakistan (Table 1). This suggests an 193 initial treatment of the wastewater of textile industry of Pakistan before discharging into the 194 Ravi and Chenab rivers.

#### 195 Performance evaluation for physicochemical parameters

The performance of FTWs in the presence of vegetation, bacteria, and vegetation and bacteria 196 together, was evaluated by studying wastewater parameters at different time intervals (Table 197 2 - 4, Fig. 2). In the beginning of the second year, the system performance was relatively 198 optimal. Based on these observations, in-depth investigations were performed. Briefly, it was 199 found that FTWs vegetated with P. australis removed a high proportion of organic and 200 inorganic pollutants as compared to the un-vegetated FTWs. This removal was further 201 enhanced when *P. australis* and bacterial consortium were applied in combination. Briefly, in 202 the presence of both partners, COD was reduced to 91%, BOD to 92%, and color to 86% 203 (Fig. 2A-C). Earlier studies have also shown highest pollution reduction of textile wastewater 204 when plants and bacteria employed synergistically (Shehzadi et al. 2014; Watharkar et al. 205 2015; Hussain et al. 2018a); this could be attributed to the combined enzymatic activities of 206 both bacteria and plants to transform organic matter into simple metabolites (Kabra et al. 207 2013; Khandare et al. 2013). These transformed products can be taken up by plants as a part 208 of nutrient assimilation process or eliminated in the form of gases, e.g., CO<sub>2</sub> and N<sub>2</sub>. With the 209 passage of time, this reduction in organic load was increased and maximum remediation was 210 211 observed in the period of 10 days. Accordingly, a positive correlation for reduction of COD and color was seen in textile wastewater. It has been reported previously that COD 212 determines the extent of oxidizable contaminants and hence establish positive correlation 213

with the color of wastewater (Tara et al. 2018). The observation was consistent in this study
as well. Likewise, pH of the remediated water moved from basic to neutral conditions (Table
2). This could also be associated with the combined action of dyes transformation by bacteria
and plants (Watharkar et al. 2015). Accordingly, a significant amount of phenols, chlorides,
and sulphates were removed from the wastewater (Figure 2D). Earlier studies have reported
similar results in different types of CWs (Shehzadi et al. 2014; Hussain et al. 2018a,b).

Vegetation had a prominent effect on TDS and TSS removal as compared to the 220 unvegetated but inoculated treatments (Table 2). Briefly, vegetation reduced TDS from 5251 221 mg  $l^{-1}$  to 2265 mg  $l^{-1}$  (up to 57%) and TSS from 324 mg  $l^{-1}$  to 165 mg  $l^{-1}$  (up to 49%). The 222 bacterial inoculation in vegetated treatment further increased reduction in TDS and TSS up to 223 1399 mg  $l^{-1}$  (73%) and 148 mg  $l^{-1}$  (54%), respectively. These results are in accordance to the 224 previously published findings (Shehzadi et al. 2014; Tara et al. 2018; Hussain et al. 2018a,b). 225 As of these findings, vegetated treatments also displayed better removal of nutrients (TN and 226  $PO_4^{-2}$ ) than the un-vegetated treatments whereas highest removal efficiency was seen in the 227 presence of plant-bacteria synergism (Table 3). The removal efficiency for TN was recorded 228 up to 60% in the vegetated FTWs, which was increased to 87% on bacterial inoculation. In an 229 earlier study, Sun et al. (2009) reported similar results where TN removal was enhanced up to 230 72% by the addition of bacteria in FTWs whereas only 50% removal was achieved in the 231 treatments without bacteria. Similar results were obtained for  $PO_4^{-2}$  reduction as bacterial 232 augmentation improved the removal as compared to the vegetation only. Nevertheless, 233 percent removal for TN was higher than the  $PO_4^{-2}$ , which can be linked to the fact that 234 nitrogen, in addition to the plant uptake, is eliminated in the form of N<sub>2</sub> gas whereas  $PO_4^{-2}$ 235 stays in the system due to precipitation (Tao and Wang 2009; Vymazal, 2010). 236

Long-term system's performance was monitored to evaluate pollutant removal during the studied period, as shown in Figure 3. It was found that the system was relatively stable throughout the experimental period; nevertheless, performance was better in the warm months.

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#### 242 Reduction in heavy metals from wastewater and their bioaccumulation in plant tissues

The studied heavy metals (Fe, Ni, Cr, and Cd) were significantly removed by FTWs (Table 243 4). The removal efficiency was equally valid for both essential (Fe and Ni) and non-essential 244 (Cr and Cd) heavy metals. Nevertheless, bacterial inoculation improved the bioaccumulation 245 and phytoremediation potential of P. australis. The mean concentration values of heavy 246 metals in root, shoots, and leaves were observed in the following order: Fe > Cr > Ni > Cd. 247 Absorbed metals were partitioned among different parts of *P. australis*. Roots retained higher 248 concentration of heavy metals than shoots and leaves with the following order: 249 root>leaf>stem. These results are in accordance to the earlier findings demonstrating that 250 underground part of *P. australis* shows a higher storage capacity than the aboveground 251 organs (Štrbac et al. 2014; Kucaj and Abazi 2015; Amira and Leghouchi 2017). It has been 252 established that roots and rhizomes of P. australis possess large intercellular air spaces of 253 254 cortex parenchyma that allow accumulation of heavy metals far away from the metabolically active structures to avoid toxic effects (Mendonça et al. 2015). The present study confirms 255 such a potential of *P. australis* in phytoremediation for the studied heavy metals which was 256 further enhanced by bacterial inoculation, i.e., the system was able to remove Cr and Fe more 257 than 90%, Ni more than 80%, and Cd more than 60%. The role of bacteria in sorbing metallic 258 ions on their cell walls is prominent and has also been reported in terms of increasing 259 bioavailability followed by uptake by plants (Jilani and Khan 2013; Khan et al. 2014). 260

#### 261 Inoculated bacteria displayed persistence in FTWs

262 In phytoremediation, plant-associated microbial populations play a key role to mineralize 263 organic pollutants (Khan et al. 2014). Many studies have reported that the phytoremediation potential of plants can be correlated with the proportion of bacteria in the surrounding 264 environment (Yousaf et al. 2011; Shehzadi et al. 2014; Rehman et al. 2019). The artificial 265 266 augmentation of bacteria in such a system establishes intimacy with their host depending upon the environmental conditions such as nutrient supply and their ability to colonize the 267 host environment. Moreover, bacteria improve plant growth by reducing biotic and abiotic 268 269 stress, followed by N<sub>2</sub> fixation, production of phytohormones, and solubilization of various essential nutrients (Glick et al. 2010; Khan et al. 2014). In this study, in order to confirm the 270 improvement in performance of P. australis due to bacterial augmentation, persistence of 271 inoculated bacteria was evaluated within the plant tissues and wastewater at the beginning of 272 the experiment. Results showed a high persistence of inoculants in the plant interior, 273 especially in the roots, followed by the wastewater (Fig. 4). Furthermore, RFLP analysis 274 indicated that the relative proportion of inoculated bacterial with roots and shoots was 275 increased significantly in the initial 3-months period, which later became stable. Although the 276 trend was similar for both plant parts, their relative proportion was higher in the roots as 277 compared to the shoots. This might be due to the fact that The inoculated bacteria were 278 previously isolated from the roots and shoots of *P. australis* grown in the textile wastewater 279 and therefore they probably have adapted necessary mechanism for proliferation in these 280 hostile conditions (Afzal et al. 2014). By contrasts, a decline in bacterial counts was observed 281 in the water, which might be due to the fact that contaminated water is less favorable 282 environment for the inoculated bacteria. More specifically, rhizo- and endophytic bacteria 283 needs a symbiotic partner such as plant roots to survive and proliferate, however, the absence 284 can cause poor survival and growth (Arslan et al. 2014). Additionally, this might be attributed 285

to the natural selection pressure for the newly grown community that allows bacteria tomigrate from less favorable environment to the more favorable environment.

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#### 290 Plant biomass and growth

An important parameter in phytoremediation efficiency is the plant tolerance and survival in 291 the presence of pollutants. The growth of *P. australis* was increased sharply in the first year, 292 which was then sustained or remained unaffected during the remaining studied period (Figure 293 5). This could be attributed to the senescence effect. Nevertheless, when compared with the 294 contaminated water, it was found that wastewater inhibited the plant biomass significantly 295 (Table 5). This shows the toxic effects of textile wastewater on plant metabolism (Watharkar 296 et al. 2015; Shehzadi et al. 2016; Ramya et al. 2017). In an earlier study, Typha domingensis 297 exhibited similar behavior in which reduction in plant growth was seen in the presence of 298 299 textile wastewater (Shehzadi et al. 2014). Nevertheless, inoculation with bacteria rehabilitated the plant growth with a prominent increase in root length (46%), shoot length 300 (37%), root fresh weight (58%), shoot fresh weight (49%), root dry weight (55%), and shoot 301 dry weight (46%) (Table 5). This increase in plant growth parameters could be associated 302 with the pollutant degrading capabilities as well as ACC deaminase potential of the 303 inoculated bacteria. Fatima et al. (2016) reported similar results in which plant growth was 304 significantly improved by the augmented of endophytic bacteria due to their PGP activities 305 such as indole acetic acid production, siderophores formation, and ACC deaminase activity. 306

**307 Detoxification of wastewater** 

308 Fish toxicity bioassay suggested a successful reduction in wastewater's toxicity after treatment with FTWs. This reduction was partially effective when plant and bacteria were 309 applied separately, i.e., death of three and five fish out of 10 specimens. Whereas, bacterial 310 inoculation in vegetated FTWs resulted in complete detoxification as no fish died even after 311 72 h of exposure (Table 6). These results strengthen the earlier observations in which high 312 reduction was observed for COD, BOD<sub>5</sub>, TDS, TSS, color, and heavy metals in the presence 313 of vegetation and bacteria. Earlier studies also reported similar findings in which use of 314 plants in combination with bacteria was an efficient strategy to remove wastewater toxicity 315 (Tanner and Headley 2011; Ijaz et al. 2016; Rehman et al. 2018). 316

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#### 318 CONCLUSIONS

This study investigated the performance of FTWs in the presence of three bacterial strains to 319 remediate textile wastewater at pilot-scale. FTWs vegetated with P. australis and inoculated 320 bacteria displayed highest pollutants removal. Bacterial persistence was prominent in the root 321 and shoot interior suggesting optimal partnership in pollutant degradation. The system 322 attenuated both organic and inorganic contaminants and wastewater was found to be 323 completely detoxified. The system was operated for two years suggesting long-term potential 324 for the on-site remediation practices. It is concluded that bacterially assisted FTWs can be 325 exploited to treat industrial wastewaters in countries with more economic constraints like 326 Pakistan, where capital and operational cost are of prime importance. 327

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#### 329 **Funding statement**

Authors are thankful to Higher Education Commission (HEC), Pakistan for providing

funding. The grant number is 1-52/ILS-UITSP/HEC/2014

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#### Table 1

Parameter	Unit	Wastewater	NEQS
Temperature	°C	38 (2.63)	40
pH		8.8 (0.82)	6-10
Electrical conductivity (EC)	mS cm <sup>-1</sup>	8.2 (0.64)	NG
Color	$m^{-1}$	66 (4.4)	NG
Chemical oxygen demand (COD)	mg l <sup>-1</sup>	513 (37.6)	150
Total organic carbon (TOC)	$mg l^{-1}$	201 (15.9)	NG
Biochemical oxygen demand (BOD <sub>5</sub> )	mg l <sup>-1</sup>	283 (17.9)	80
Phenol	mg l <sup>-1</sup>	0.85 (0.06)	0.1
Chloride	mg l <sup>-1</sup>	1383 (61.8)	1000
Sulphate	mg 1 <sup>-1</sup>	311 (21.9)	600
Nitrogen	mg l <sup>-1</sup>	28.7 (6.0)	40
Phosphate	$mg l^{-1}$	16.4 (3.6)	NG
Total dissolved solids (TDS)	mg l <sup>-1</sup>	5251 (404)	3500
Total solids (TS)	mg l <sup>-1</sup>	5420 (395)	NG
Total settleable solids (TSeS)	mg l <sup>-1</sup>	19.9 (1.1)	NG
Total suspended solids (TSS)	mg l <sup>-1</sup>	324 (29.7)	150
Iron (Fe)	mg l <sup>-1</sup>	14.4 (0.64)	2.0
Nickel (Ni)	mg l <sup>-1</sup>	7.57 (0.38)	1.0
Chromium (Cr)	mg l <sup>-1</sup>	9.67 (0.26)	0.1
Cadmium (Cd)	mg l <sup>-1</sup>	0.88 (0.02)	0.1
Toxicity		Highly toxic	NG

Quality parameters of textile industrial wastewater collected from Interloop, Khurrianwala located in Faisalabad, compared to the National Environmental Quality Standards (NEQS), Pakistan (Hussain et al., 2018b)

Each value is the mean of three replicates; standard error among three replicates is presented in parenthesis. NG, not given in National Environmental Quality Standards (NEQS) list.

#### Table 2

Remediation of the textile industrial wastewater by floating treatment wetlands vegetated with Phragmites australis and inoculated with bacterial consortium

Days	s Control				T1				Τ2				Т3			
	pН	EC	TDS	TSS	pН	EC	TDS	TSS	pН	EC	TDS	TSS	pН	EC	TDS	TSS
0	8.73	8.22	5251	324	8.74	8.34	5252	325	8.75	8.27	5251	325	8.74	8.25	5252	324
	(0.20)	(0.64)	(404)	(30)	(0.16)	(0.66)	(404)	(31)	(0.19)	(0.61)	(404)	(29)	(0.20)	(0.61)	(404)	(29)
2	8.49	7.72	4934	319	8.43	7.27	4635	275	8.19	6.64	4250	256	8.55	7.34	4699	310
	(0.18)	(0.64)	(411)	(29)	(0.17)	(0.55)	(354)	(19)	(0.12)	(0.55)	(355)	(18)	(0.08)	(0.67)	(430)	(28)
4	8.38	7.36	4677	306	8.14	6.42	4103	241	7.83	5.73	3632	225	8.35	6.84	4356	296
	(0.17)	(0.69)	(442)	(28)	(0.13)	(0.43)	(279)	(18)	(0.13)	(0.43)	(277)	(12)	(0.13)	(0.61)	(389)	(29)
6	8.58	6.7	4295	300	7.67	5.35	3397	224	7.46	4.35	2735	206	8.58	6.37	4016	290
	(0.09)	(0.72)	(462)	(29)	(0.10)	(0.53)	(339)	(16)	(0.10)	(0.49)	(314)	(11)	(0.09)	(0.69)	(441)	(27)
8	8.88	6.2	3974	294	7.41	4.42	2843	199	7.14	3.49	2183	171	8.81	5.67	3632	286
	(0.08)	(0.70)	(447)	(29)	(0.10)	(0.43)	(278)	(16)	(0.10)	(0.46)	(295)	(12)	(0.11)	(0.62)	(394)	(26)
10	9.04	5.7	3655	285	7.35	3.57	2265	165	7.12	2.26	1399	148	8.87	5.13	3292	271
	(0.04)	(0.61)	(389)	(28)	(0.06)	(0.35)	(222)	(12)	(0.09)	(0.30)	(185)	(10)	(0.09)	(0.46)	(296)	(24)

Floating treatment wetlands vegetated with *P. australis* (T1), *P. australis* and bacteria (T2), and bacteria only (T3). Each value is the mean of three replicates; the standard error of three replicates is presented in parenthesis. TDS and TSS are mentioned in mg  $l^{-1}$ , whereas EC in mS cm  $^{-1}$ .

CEP.

#### Table 3

Nutrients (TN and  $PO_4^{2^-}$ ) removal from textile industrial wastewater by floating treatment wetlands vegetated with *Phragmites australis* and inoculated with bacterial consortium

Days		TN (n	ng l <sup>-1</sup> )			$PO_4^{2-}(mg l^{-1})$						
	Control	<b>T1</b>	T2	Т3	Control	T1	T2	Т3				
0	28.7 (6.0)	28.7 (6.0)	28.9 (5.6)	28.9 (5.7)	16.4 (3.6)	16.4 (3.6)	16.5 (3.6)	16.5 (3.6)				
2	27.0 (5.4)	22.6 (4.5)	19.3 (4.1)	25.9 (5.1)	16.1 (3.6)	15.7 (3.4)	15.1 (3.4)	15.9 (3.3)				
4	25.9 (5.1)	20.3 (3.9)	15.7 (2.9)	23.6 (4.7)	15.9 (3.4)	14.5 (3.1)	13.2 (2.8)	15.6 (3.5)				
6	24.5 (4.9)	17.4 (3.2)	12.0 (1.9)	22.2 (4.4)	15.8 (3.4)	13.1 (2.8)	11.8 (2.7)	15.2 (3.2)				
8	23.2 (4.6)	14.8 (2.7)	8.7 (1.8)	20.1 (3.7)	15.6 (3.4)	11.3 (2.3)	10.0 (2.3)	14.8 (3.0)				
10	22.3 (4.5)	11.3 (2.0)	3.7 (0.6)	19.0 (3.5)	15.3 (3.3)	10.1 (2.1)	8.2 (1.9)	14.2 (2.9)				

Floating treatment wetlands vegetated with *P. australis* (T1), *P. australis* and bacteria (T2), and bacteria only (T3). Each value is the mean of three replicates; the standard error of three replicates is presented in parenthesis.

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Metal	Control			T1		Τ2				
	Initial conc. (mg l <sup>-1</sup> )	Water $(mg l^{-1})$	Root (mg kg <sup>-1</sup> )	Shoot (mg kg <sup>-1</sup> )	Leaves (mg kg <sup>-1</sup> )	Water (mg l <sup>-1</sup> )	Root (mg kg <sup>-1</sup> )	Shoot (mg kg <sup>-1</sup> )	Leaves (mg kg <sup>-1</sup> )	Water (mg $l^{-1}$ )
Fe	14.4 (0.64)	2.63 (0.27)	5.90 (0.45)	1.97 (0.15)	3.94 (0.30)	1.05 (0.11)	6.69 (0.37)	2.23 (0.12)	4.46 (0.25)	9.0 (0.47)
Ni	7.57 (0.38)	1.72 (0.12)	2.92 (0.19)	0.97 (0.06)	1.95 (0.13)	1.08 (0.09)	3.25 (0.18)	1.08 (0.06)	2.16 (0.12)	5.43 (0.35)
Cr	9.67 (0.26)	1.50 (0.14)	4.08 (0.09)	1.36 (0.03)	2.72 (0.06)	0.40 (0.10)	4.63 (0.16)	1.54 (0.05)	3.09 (0.10)	6.53 (0.38)
Cd	0.88 (0.06)	0.29 (0.06)	0.24 (0.03)	0.08 (0.01)	0.16 (0.02)	0.26 (0.03)	0.31 (0.01)	0.10 (0.01)	0.21 (0.01)	0.72 (0.04)

.16 (v.. ralis and bacteria (. .esis. Floating treatment wetlands vegetated with P. australis (T1), P. australis and bacteria (T2), and bacteria only (T3). Each value is the mean of three replicates; the standard error of three replicates is presented in parenthesis.

#### Table 5

Effect of bacterial inoculation on biomass and root and shoot length of *Phragmites australis* vegetated in floating treatment wetlands

Treatment	Lengt	h (cm)	Fresh bio	omass (g)	Dry biomass (g)		
	Root	Shoot	Root	Shoot	Root	Shoot	
Tap water	67.8 (2.6)	407 (19.1)	1189 (52.1)	2194 (62.0)	152 (6.1)	510 (13.9)	
Wastewater	43.6 (2.5)	294 (14.8)	936 (64.2)	1803 (48.9)	84 (7.9)	370 (13.9)	
Wastewater and bacteria	54.8 (4.9)	336 (6.9)	1083 (48.9)	1994 (94.4)	122 (7.5)	434 (26.5)	

Each value is the mean of three replicates; the standard error of three replicates is presented in parenthesis.

#### Table 6

Fish toxicity assay of textile industrial wastewater detoxified by floating treatment wetlands vegetated with *Phragmites australis* and inoculated with bacterial consortium (BC)

Treatment	Death over time							Total deaths	Detoxification status			
	0 h	1 h	2 h	4 h	8 h	12 h	24 h	48 h	72 h	96 h		
P. australis (T1)	0	0	0	0	0	0	0	0	1	2	3	Partial
P. australis + BC (T2)	0	0	0	0	0	0	0	0	0	0	0	Complete
BC (T3)	0	0	0	0	0	0	2	1	1	1	5	Partial
Control	0	2	0	1	1	1	4	0	0	0	9	Negligible



Fig. 1. Development and implementation of floating treatment wetlands (FTWs) macrocosms for the remediation of wastewater at Interloop Limited Khurrianwala, Faisalabad. Wastewater equalization tank (A), floating mat for the plantation of *Phragmites australis* seedlings (B), nursery of P. australis in the vicinity of NIBGE, Faisalabad (C), and growth of *P. australis* in FTWs macrocosms at the end of experimentation (D).



**Fig. 2**. Reduction of COD (A), BOD (B), Color (C), and for Phenols, Chlorides, and Sulphates (D) in the textile wastewater by floating treatment wetlands vegetated with *P. australis* (T1), *P. australis* and bacteria (T2), and bacteria only (T3). Each value is the mean of three biological replicates. Error bars indicate standard error among three replicates.



Fig. 3: Performance evaluation of FTW in terms of individual parameters, and overall response.



**Fig. 4.** Bacterial survival and colonization in water and in root and shoot interior of *Phragmites australis* for the macrocosms containing wastewater with vegetation and bacterial inoculation (T2).



Fig. 5: Biomass of *Phragmites australis* per plant obtained after every 3-months of harvesting.

## Highlights

Bacterially assisted FTWs macrocosms were engineered to treat real textile effluent

Combined application of plants and bacteria had a prominent effect on pollution reduction

Inoculated bacteria displayed persistence in different components of the FTWs

Treated wastewater met the National Environmental Quality Standards (NEQS) of Pakistan