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1 Research Paper	1	Research Paper
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Front and back yard green analysis with subpixel vegetation fractions from earth observation data in a city

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18 19 20	 Highlights This paper introduces a novel approach to the backyard green space in cities. 	us-far mostly neglected urban front and
21 22	• We calculated subpixel vegetation fractions f the entire city with a spectral unmixing techn	from RapidEye remote-sensing data for ique.
23 24	• We applied a novel 'house-attached front and the city of Leipzig with accuracy of 96%.	d backyard green derivation algorithm' to
25 26	• Key findings include that the total amount of 2000 ha or 40% of the amount of public gree	front and backyard green space is almost n.
27 28	• In 15 out of the 63 total districts, we found m than public green space.	nore house-attached front and backyard
29		

31 Abstract

This paper introduces a novel approach to green space availability in cities that includes the 32 thus-far mostly neglected urban front and back yard green space around residential buildings 33 on privately owned ground. To quantify the full spatial scope of urban green space, we 34 calculated subpixel vegetation fractions from RapidEye remote-sensing data for the entire city 35 with a spectral unmixing technique that enabled us to model the extent of urban vegetation with 36 a high degree of confidence (MAE 7%, R² 0.92). We then applied a novel 'urban front and back 37 yard green space derivation algorithm', namely, a masking of the fractional vegetation data 38 using GIS vector data of land cover, in order to delineate the front and backyard greenspace of 39 residential houses in a city with an accuracy of 96%. Combining these two approaches, we can 40 calculate the area of urban front and back yard green space for the entire city (including different 41 42 residential structure types) and compare this data to the area of public (parks, urban forests) and semi-public (allotment gardens) green spaces that have been used for prevailing per capita green 43 44 space availability analyses. The new method is exemplified at the city of Leipzig, Germany, which provides different residential structures concerning house types and the surrounding 45 46 green that are characteristic of many European cities. Key findings include that the total amount of urban front and back yard green space is almost 2000 ha, which is ~40% of the amount of 47 public green space (4768 ha). In 15 out of the 63 total districts, there is more private house than 48 public green space, which highlights the importance of these urban front and back yard green 49 50 space for the analysis of urban livelihoods and a tool for detailed ecosystem services-oriented urban planning. 51

52 Keywords

Front and back yard green, remote sensing, RapidEye, delineation, urban green space
availability, spectral unmixing, sub-pixel mapping, urban planning

55 **1 Introduction**

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Until now, the literature on green space availability in cities has mainly referred to public
(parks, forests; Voigt et al., 2014) and different types of gardens when either mapping (Dennis
et al., 2018) or calculating accessibility and availability (Comber et al., 2008; Fuller & Gaston,
2009; Kabisch & Haase, 2012; Kabisch et al., 2014; Kabisch et al., 2016; Larondelle & Haase,
2017). Urban front and back yard green space is the only major type of urban green space that

has rarely been analysed in its full and differentiated quantitative scope, even more sporadically 62 on a citywide level (for a promising approach see Dennis et al., 2018). Residential green space 63 refers to the vegetation that surrounds residential buildings, directly attached to it in form of 64 flower beds, lawns with or without trees in the front and backyard, or between houses in the 65 case of a block location. Residential green space is also related to different forms of property 66 rights and ownership constellations in the city and, therefore, this type of green space can hardly 67 be derived from official land-use data. This is the research gap that we intend to fill with this 68 69 study.

In general, urban green spaces can be subdivided into three groups of ownership: firstly, public green space including parks, cemeteries and urban forests; secondly, semi-public allotment garden areas and sport facilities, and thirdly, private-house and backyard green with limited (physical or jurisdictional) access for the tenants (Fischer et al., 2016). We define urban front and back yard green space as that which directly surrounds the residential houses, including front and backyard green space and what is most probably private and not public.

Numerous qualitative studies have shown the importance of house and backyard green space 76 77 on public health (Bosch and Sang, 2017), happiness, life satisfaction and for urban biodiversity 78 (Taylor and Taylor Lovell, 2014; Strohbach et al., 2009). There are areas in a city in which the 79 major allotment of green space can be found around privately owned houses and backyards. 80 Even though only a limited share of urban dwellers, namely the residents of the respective houses, have direct access to these areas, they are important for climate regulation, air cooling, 81 local recreation, and human well-being to list a few important ecosystem services. With regard 82 to an increasing number of elderly and aged residents in our European cities, local recreational 83 amenities have gained in importance in enabling healthier lifestyles among older people 84 85 (Brookfield et al., 2015).

Regardless of its importance for the city, the literature widely overlooks front and back yard
green space front and back yard green as an important form of urban vegetation (Coolen and
Meesters, 2012), mainly due to limited data availability and data access (Breuste et al., 2016).
To overcome this lack of assessment of urban front and back yard green space, we provide a
straightforward blueprint for how to generate such a dataset using remote sensing imagery and
a limited number of vector layers.

With the use of remote sensing data, we achieve a spatially continuous analysis and open up the potential for long-term trend analysis. In order to quantify the amount of green space, we use a subpixel mapping approach with Random Forest Regression, a relatively new approach that is regularly applied to different fields in generating classifications of urban land cover.
Random Forest Regression is widely used across disciplines and, as in the case for multispectral
remote sensing data, is effective in handling multidimensional data with a high degree of
collinearity. This is achieved in training the model with a multitude of decision trees. The entire
methodology used in this study will soon be available in the open-source EnMapBos QGis
Plugin (van der Linden, 2015) – an in-house development that facilitates techniques of analysis
based on advanced remote sensing.

With this research paper we therefore introduce a GIS- and regression-tree-based scheme for the creation of a city-wide urban front and back yard green space delineation in the form of a spatial data set. In addition, we perform an analysis of this urban front and back yard green space type that includes a quantitative comparison with public and semi-public green spaces in order to search for complementarities and synergies.

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108 **2 Study Area**

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Leipzig is a fast growing, compact city (580,000 inhabitants) located in the eastern part of 110 111 Germany and featuring a wide variety of building types and a wide variety of public, private and semi-public (with opening times) green spaces across its 63 districts. The provision with 112 green space in the city is below mid-European and also German average. While the distribution 113 of public and semi-public green spaces has already been analysed in previous studies (e.g. 114 Haase and Nuissl, 2010; Haase and Gläser, 2009), front and back yard green space is, as in the 115 case of all other cities, beyond the scope of the current research. The largest body of public 116 green space in Leipzig – the Weiße-Elster-Pleiße floodplains – is arranged along a north south 117 gradient to the west of the city centre (Haase & Gläser, 2009). Semi-public allotment gardens 118 are relatively equally distributed across the city, particularly within the wetlands of the 119 floodplains and along major roads and railways. The major residential types in the city are 120 Wilhelminian-period perimeter development blocks, alignment blocks built in the 1950s and 121 1960s, prefabricated high-rise buildings from the socialist period (dating back to the 1970ies 122 and 1980ies of the 20th century), late 19th and early 20th century villa districts, perimeter block 123 estates of the 1950s and 1960s, single-house colonies in the peripheral districts and, after 1990, 124 townhouses built to fill vacant lots in the inner city (Haase and Nuissl, 2010). Between these 125 major urban built structures, privately owned residential green space is distributed very 126 unevenly. Its share ranges from almost zero to considerable green coverage around the houses, 127

with larger absolute amounts of front and back yard green space in the townhouse and villaareas in comparison to the denser inner-city Wilhelminian perimeter-block quarters.

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131 **3 Data & Methods**

With the exception of sealed surfaces, urban front and back yard green space - including the 133 front and backyard green space of residential houses that are predominantly on private 134 135 individual property or that of private housing companies – are all green spaces as defined above. We thereby calculate the fractional amount of vegetation of each of the pixels situated in the 136 privately owned areas around the houses in order to calculate the actual amount of green space. 137 We exclude bare soil since, in the moment of analysis, it is not vegetated. However, this can 138 rapidly change and is therefore often incorporated in green space definitions. Our analysis only 139 incorporates ground-based front and back yard green space as defined above, rooftop green 140 space as well as wall green remains excluded in order to keep the delineation algorithm simple 141 and to minimize miss-classifications. 142

143 From our dataset we can derive multiple indicators, comparable in different (spatial) reference units (e.g. land-use type, urban districts, per capita etc.) and between entire cities. Thereby, both 144 the total area values (m²) and the relative values (%) can be obtained. To derive sub-pixel 145 fractions of vegetation, we used a Random Forest Regression combined with ensemble learning 146 147 and synthetically mixed training data, an approach successfully applied to map urban land cover 148 (Okujeni et al. 2013, Okujeni et al. 2017, Rosentreter et al. 2017). The underlying assumption of this method is that spectral reflection changes with land-cover composition. Pure spectra of 149 150 trees show different spectral characteristics than pure spectra of soils or impervious surfaces such as streets or pavements. Due to the high degree of heterogeneity in the urban environment 151 152 however, most pixels do not contain pure spectral information but are rather a mixture between different land-cover classes. The spectral signature depends on the environmental heterogeneity 153 resulting in different fractional abundance of each land-cover class in different pixels. 154

The relationship between the covered area and the signal change can be described by a linear function (Keshava & Mustard, 2002). Important in regard to the high degree of urban heterogeneity is that the training data needs to cover as many as possible combinations of different land cover compositions and spectral signals to train the model. In this study we overcome this challenge by generating training data that features pure spectral information in order to calculate synthetic mixtures between the spectra of pure land-cover classes. This approach simplifies the training data collection greatly because only pure spectra need to besampled.

164 **3.1 Data**

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Our quantitative assessment of house and backyard green space makes use of the capability of 165 RapidEye satellite data sensor, which acquires data in five spectral bands (R,G,B, red-edge & 166 167 near infra-red) with a ground resolution of 6.5 meters resampled to 5 m. For our study, we acquired two cloud-free RapidEye images of Leipzig and its surroundings from the year 2012 168 (14.05.2012) and corrected for atmospheric influences with ATCOR (Richter, 2011). For the 169 validation of our regression model we compiled a validation dataset based on very high-170 resolution (VHR) satellite imagery from Google Earth. The dataset encompasses 141 validation 171 sites and spatially explicit mapping of vegetated areas, thus enabling us to calculate the fraction 172 of green space in each of the polygons. For the purpose of deriving the exact location of three 173 regarded forms of urban green space, a set of spatial vector data is used (Table 1). Most 174 prominently in this regard are the biotope map and ATKIS (Amtliche Topographisch-175 Kartographische Informationssystem Deutschland) land-use database, which allow the 176 detection of a variety of urban infrastructure. 177

179	Table 1.	Vector datasets	used in th	is study	with date,	source and a	a detailed	description
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Datasets	Masked out areas	Date	Source
City border	Rural surrounding	2014	Free State of
-	_		Saxony
Biotope map	Public green and blue infrastructure	2014	Free State of
			Saxony
	Semi-public green infrastructure	2014	Free State of
			Saxony
	Agricultural areas	2014	Free State of
			Saxony
	Railway corridors	2012	Free State of
	-		Saxony
	Industrial and commercial sites	2014	Free State of
			Saxony
Roads	Road surfaces	2014	City of Leipzig
ATKIS	Building footprints	2014	Free State of
			Saxony
Public trees	Canopies of public trees	2012	City of Leipzig
			and UFZ

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182 **3.2 Methods**

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184 3.2.1 Estimation of subpixel vegetation fractions For the estimation of subpixel vegetation fractions, we used spectral unmixing with Random 185 Forest Regression trained with synthetic mixtures of pure land cover signals. Vegetation 186 fractions are derived from a RapidEye scene. The Random Forest Regression models are 187 188 embedded in an ensemble learning system to reduce the calculation time and to improve the subsequent generalization of the unmixing results (Figure 1; Breiman, 1996). Resulting maps 189 are validated with a reference dataset created by visual interpretation of high-resolution imagery 190 as described above (see section 3.1). 191



Figure 1. Flowchart of the applied subpixel vegetation modelling approach using spectral unmixingwith a RapidEye imagery library.

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196 3.2.1.1 Training data generation

We collected pure spectra of different urban land-covers present in the remote sensing data using Google Earth imagery. They were either belonging to the vegetation class or the nonvegetation class. With the help of the higher-resolution imagery from Google Earth, larger areas of homogenous land cover could be identified more accurately. Then, the spectral signal was obtained from the RapidEye imagery. We captured the original spectral variability in the scene

by including as many different land-surface materials as possible - e.g. different rooftops, 203 204 ground layers and tree, shrub and grass species. Overall, our library of pure spectra contains 39 spectra of the vegetation class and 62 spectra of the non-vegetation class. 205





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Figure 2. The concept of synthetic spectral mixing shown in four iterative steps. Firstly, a mixing complexity needs to be determined; secondly, pure spectra are drawn and thirdly, their respective share of their contribution is determined. Thereof synthetic mixtures are finally calculated.

211 The library of pure spectra was used to calculate the synthetic mixtures, i.e. the number of 212 213 training data sets that were later used to train the Random Forest Regression. In order to generate a single synthetic mixture, several parameters and components need to be set including 214 215 the mixing complexity, the actual spectra contributing to the mixture and the respective

fractions of the contributing spectra. Subsequent parameterization was carried out in a randomized and automatic manner. The entire process consists of four steps (see Figure 2):

- a. First, the mixing complexity was determined. It specified how many different spectra
 were contributing to the synthetic mixture (Figure 2, part 1). We considered binary,
 ternary and quaternary complexities that varied randomly per mixture.
- b. In a second step, the actual spectra contributing to the mixture were randomly drawn
 (Figure 2, part 2). For binary complexities, two pure spectra were used and for ternary
 and quaternary complexities, three and four pure spectra were used. We allowed for
 intra and interclass mixtures.
- c. After the pure spectra were determined, the fractions that each spectrum contributed
 needed to be specified (Figure 2, part 3). The fractions defined the extent to which each
 spectrum contributed to the mixture. We therefore used a randomized method under the
 condition that all fractions sum up to 100%.
- d. Finally, the synthetic mixing could be applied (Figure 2, part 4). In our analysis, we
 calculated that there were 1000 spectral mixtures for each model in the ensemble
 learning (see section 3.2.1.2).
- 232
- 233 234

3.2.1.2 Random forest regression ensemble

Random Forest Regression is a classification procedure based on the process of ensemble 235 236 learning. In this study, the Random Forest Regression was itself embedded in an ensemble learning system (Breiman, 1996). An ensemble learning approach consists of a specified 237 238 number of Random Forest Regression models that are trained with a subset of the training data. Each of the resulting models is then applied separately to the RapidEye image. The resulting 239 240 fraction-maps of the different model runs are subsequently combined by averaging. The approach of using training sample subsets for the development of models is called bootstrap-241 aggregation (bootstrapping, bagging) and was first introduced by Breiman (1996). In our 242 approach, we trained 20 Random Forest Regression models with sample subsets consisting of 243 1000 random synthetic mixtures (see section 3.2.1.1). 244

245

246 3.2.1.3 Validation

We validated the final subpixel vegetation fraction map with reference to the aforementioned dataset (section 3.2.1.2). As a performance measure, we used the Mean Absolute Error (MAE)

$$\text{MAE} = \frac{\sum_{i=1}^{n} |y_i - x_i|}{n} \tag{1}$$

with y_i being the fraction of the reference polygons, x_i the average estimated fraction inside the reference polygons, and *n* the number of reference polygons. The model fit was assessed by the coefficient of determination R².

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256 *3.2.2 Calculation of the extent of parks and forests, gardens and front and back yard green*

Since the full extent of urban front and back yard green space is unknown, we iteratively 257 masked out every urban structure that did not represent urban-front and back yard green space 258 259 from our vegetation fraction map. Public and semi-public blue and green infrastructure, agricultural areas, railway corridors, and industrial and commercial sites were masked out using 260 a biotope map. Thereof, and according to the mean statistics of the vector data we have for the 261 city, we defined buffers around roads (16m) and public trees (r = 5m) and clipped these from 262 263 our imagery. Public trees were excluded, because in this paragraph the creation of the private front and back yard green layer is explained. Finally, buildings whose location was derived 264 from the ATKIS dataset were masked out. The remainder of the satellite image is front and 265 backyard green space. 266

For the delineation of public green space, the land-use classes of forests, parks, graveyards and 267 268 grasslands, that are located in the forest or show early signs of shrub/tree encroachment as indicators of not entirely pure agricultural usage, were extracted from the biotope map. As semi-269 public green space, we defined allotment gardens and sport facilities as in the biotope map. By 270 masking the remote sensing data iteratively using these datasets, we were able to determine the 271 272 amount of total house-attached front and back yard green, allotment and community gardens, as well as public parks and forests at city level. Thereof, we calculated nine relative indicators: 273 front and back yard green, gardens, parks and forests respectively per urban district, per capita 274 and per land-use type (cf. Figure 3 in the Results section). 275

We validated the urban front and back yard green space dataset by randomly distributing 50 sampling points across the regarded dataset. Through visual interpretation, we were able to validate that 96% of the sampling points were correctly classified as urban front and back yard green space. The visual interpretation was carried out by a visiting of the randomly selected sites and an assessment of the authors whether the site belongs to most probably private houseattached or to public green. Of great help were fences and other visible delineations which indicate private properties or that a green lot belongs to a house. Since the other datasets are
directly derived from a single, well-established GIS source, there is no need for validation.

285 **4 Results**

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287 *4.1 Subpixel vegetation fractions*

The final vegetation map portrays the fractions of vegetation for the entire city with a high 288 degree of certainty and allows us to distinguish between house-attached, public and semi-public 289 green space. The map shows a clustering of public green space along a north-south gradient to 290 the west of the city centre, which is a large remnant of alluvial forest situated on the floodplains. 291 292 Allotment and community gardens are distributed relatively equally across Leipzig's districts with a lower abundance in the central districts. Urban front and back yard green space can be 293 294 found in all of the districts and is especially abundant in the outer districts where residential 295 quarters were constructed after 1990.



Figure 3. The extent of front and back yard green space, allotment and community gardens and urban
parks and forests as important green space types in the city of Leipzig

An R² of 0.92 and an MAE of 7.4 % highlight a very robust overall model fit (figure 4). This is reinforced by the fact that the model data closely follows the diagonal, representing a hypothetical perfect fit. It can be derived from the formula that we are overestimating vegetation fractions in pixels featuring sparse vegetation as well as underestimating the amount of vegetation in areas of dense plant cover.



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Figure 4. Validation plot showing the R², the mean absolute error (MAE) and the model formula of a fitted linear least squares regression between reference fraction and estimated fraction.

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311 *4.2 Extent of the different urban green space types*

Analysing the new dataset of urban -front and back yard green space, we find a clear trade-off 313 between public, semi-public and most probably private green spaces (Figure 5). Districts with 314 high shares of public park and forest green space, namely the Leipzig floodplain forest or large 315 parks, exhibit a lower share of house green space and vice versa. While the former group 316 consists mostly of peripheral districts, including former village cores, are included, the latter 317 group is mainly comprised of inner-city districts. Therefore, we can conclude that inner-city 318 districts offer considerable availability of house green space compared to more peripheral 319 districts which, surprisingly, have less backyard green space. Central parts of the so-called 320 "Inner East" (centre-located reddish patches in Figure 5B), an under-privileged area with a high 321 share of migrants and low-income families with many children, report both a comparatively 322 low supply of house-attached and public green space, meaning that we also find clear 323 inequalities in green space availability in the city and, indicating 'injustice' in terms of access 324 to green space in Leipzig (Low, 2013; Kabisch et al., 2016; cf. again Figure 5). 325

The prefabricated socialist housing estates in the west of Leipzig, predominantly in the settlement of Grünau, show the highest supply of urban house green space by far, and thus, housing quality has potentially been underestimated if only analysed with previously existing

public green space datasets (west-located reddish patches in Figure 5B). Houses in such 329 prefabricated estates are often surrounded by larger lawns around the single blocks that are 330 richly equipped with flower beds and some bushes or short trees. This type of green space is 331 different from that in the old built-up perimeter blocks as the lawns are publicly accessible and, 332 in most of the cases, not fenced in. However, the flower beds are under the care of the local 333 residents of the blocks and thus this 'lawn-green' is dedicated to those families living in the 334 335 large housing estates. Small signs also provide information as to which housing company the 336 block and the lawns around belong to.

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Figure 5. Ternary plot showing the relative amount of each type of green space in relation to the total
amount (front and back yard green, gardens and parks and forests) for Leipzig's 63 districts (A),
whose locations are shown in (B).

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Overall, in terms of absolute area (Table 2), public green spaces in the form of forest or parks make up almost half of Leipzig's total green space. However, in terms of vegetated area, front and back yard green spaces cover an area of almost 2000 hectares compared to 4800 hectares of public green space. Due to the higher share of lawn vegetation, the vegetation density of urban front and back yard green spaces in Leipzig is lower compared to parks, forests and gardens (Table 2).

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Table 2. Spatial extent of the three regarded green space types, their mean vegetation density and theirshare in the total city area

Green Space Type	Absolute area (ha)	Vegetated area (ha)	Vegetation density (%)	Share of the absolute total area in the city (%)	Share of the vegetated total area in the city (%)
Front and back yard green space	3518	1990	56	12	7
Allotment and community gardens	2166	1364	62	7	5
Parks and forests	5662	4768	84	19	16

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In terms of the per capita supply, we find that all densely populated districts in Leipzig suffer 354 from a significant undersupply of house-attached as well as parks, forests and gardens. Only a 355 small number of districts (7) benefit from above-average house-attached and publicly accessible 356 357 green space, while only 8 districts with dense population (according to Leipzig census data) contain above-mean front and back yard green space. For the garden-type green space, the 358 359 picture is less segregated, with more people benefitting either from private or allotment garden green space; the number of people suffering from a double undersupply is lower compared to 360 361 the chart showing the population number per public green space (Figure 6).





Figure 6. Share of parks and forests, gardens, and front and back yard green space relative to district area (A&B) and in m² per capita (C&D), for each of the 63 districts of Leipzig and per land-use type (LUT) (E&F) with indication for the mean value of the x- and y-axis, respectively.

We only find a small number of districts benefitting from both semi-public green space and front and back yard green space, as the supply with allotments is more limited than that with public parks and urban forests. Districts with high share of allotments are found along the floodplains in Leipzig, from south to north, and thus cover inner-city and peripheral locations (Figures 5 and 6). Interestingly, here we find almost all settings of residential areas that Leipzig has: privileged inner-city, poorer inner-city, prefabricated socialist inner-city and peripheral single/detached house areas (Figure 6).

In terms of per capita supply, the majority of the districts suffer from below average front and backyard green and park- and forest-type green space. There are only three districts with exceptionally high amounts of public and front and back yard green space.

Land use	Fraction map	VHR resolution imagery	Street level impressions		
type		from Google Earth			
Terraced houses					
Villas					
Prefabricated housing estates					
Residential park after 1990					
Single and semi- detached houses					
Multi story tenement blocks					

Figure 7. Overview of the most prominent urban land-use types in Leipzig, consisting of the fraction
map, a very high-resolution (VHR) Google Earth satellite image and a characteristic image from the
sites.

Concerning the urban residential structure types, the dense old built-up multi-storey perimeter block and tenement fabric shows in the city centre and the inner-city residential ring exhibit the highest undersupply with all types of green space: front and back yard green, gardens as well as parks and forests, although it is the most wanted in Leipzig at the moment. Exclusively the old residential cores including the city centre and the former village cores show both above average public and front and back yard green space. A hotspot with high supply of houseattached front and back yard green space are clearly the single house areas and the terraced houses spread across the whole outer city area as well as the urban villa areas close to the floodplains and allocated more in the inner parts of the town (cf. again Figure 7).

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392 5 Discussion

The mixed method of RapidEye imagery and GIS-data combined with random forest 394 models appears to be an efficient, straightforward and 'safe' way to delineate different 395 types of urban green spaces including planning-relevant public green spaces such as large 396 parks and forests, partly fenced and opening-time restricted allotment gardens, but also, 397 house-attached green space as a category which has not yet been quantified for a city such 398 as Leipzig and, what is more, are so far neglected in a systematic and all-encompassing 399 green space assessment by the city planners. The results are in excellent agreement with 400 401 an estimate by Breuste (2016) that states that front and back yard green space has often been neglected in urban green space assessments "although private green can often 402 constitute up to 10% of the total greenspace in a city (page 213)". The spectral unmixing of 403 404 these different types of green space also allows a city wide and district specific quantification and assessment of distribution over- and undersupplies with both public 405 406 and – most probably – private front and back yard green, which is essential for a modern urban green planning in large cities (Buijs et al., 2016). In particular for ageing cities such 407 408 as Leipzig, a good estimate and a spatial allocation of front and back yard green spaces 409 provides a more complete picture for planners where people can relax from heat when 410 discussing adaptation to hot days and tropical night temperatures, one of the major health 411 concerns of the future (Schinasi et al., 2018).

In addition, and of major interest for both urban planners and urban ecologists, green space type-specific analyses between different urban structural types, and using housing areas from different eras, becomes possible at a very detailed spatial level. The validation of the extent of private green space could be validated by randomly distributing 50 sampling points across the dataset. That the resulting 96% of the sampling points were correctly classified is an excellent result. Since the other datasets, namely of parks, forests and allotment garden 418 colonies, are directly derived from a single well-established GIS data source, there is no need419 for validation (Haase and Nuissl, 2007).

420 We are convinced that the method and the models we have developed and presented in this article can be of great help to (a) identify different types of urban green spaces and 421 422 (b) quantify the amount of green space using spectral unmixing with trained random forest regression models. This method can also: (c) identify urban front and back yard 423 green space for the first time, which is rarely to be found in any other urban green space 424 classification created from remote-sensing data but that is of enormous importance for 425 less mobile population groups in our cities (elderly, young children, disabled or 426 427 chronically ill persons to list the most important affected groups; Brookfield et al., 2015). And, finally, this method can (d) bring the different types of public, semi-public and house 428 green space existing in a large city at the level of the local districts in mutual size and 429 availability relation which is key for a sustainable and just urban planning (Haase et al., 430 2017; Kabisch et al., 2016). However, it is important to state that our novel method 431 432 characterises the spatiality of the quantity of green space types and not the quality.

433 Such a classification which we presented in the paper is not doable using the classical Biotope map as it mixes and partly aggregates green space types which we want to have 434 435 separated. Moreover, the regional as well as the urban Biotope maps for many cities – among them Leipzig – are far too old for a present-day analysis and assessment. But, this 436 437 novel approach is also of interest from a purely methodological perspective using remote sensing data for urban planning purposes. Even though we combined RapidEye with 438 comparatively high-resolution satellite data, the calculation of vegetation fractions also 439 allows for the integration of sensors with lower spatial resolution like Sentinel and 440 Landsat. Fractions greatly enhance the accuracy of estimation in heterogeneous 441 landscapes by calculating the amount of green in every pixel and thus delivering the actual 442 extent of vegetation. Remote sensing indicators like the NDVI (Normalized Difference 443 Vegetation Index; Tucker, 1979) have the potential to lead to valuable insights about 444 urban greenspaces, however, they do not allow for exact spatial quantification because 445 the state of the vegetation is assessed by the chlorophyll content, in the case of the NDVI. 446 The entire procedure described in the methods section of this paper will be implemented 447 in the upcoming QGIS Plugin of the EnMap box (van der Linden, 2015), an in-house 448 development of the Geographical Institute of the Humboldt University of Berlin, 449 450 Germany. This will open up the potential of this fraction-mapping procedure to the wider 451 urban ecologists and environmental planning community who are aiming for more realworld, evidence-based implementation and monitoring of green infrastructure at the city
and urban landscape levels (Buijs et al., 2016). A potential coupling with threedimensional indicators, such as those recently introduced by Alavipanah et al. (2017),
would lead to very valuable indicators, for example, in the ability to determine the volume
of green space next to trees which is crucial to know for urban planners to determine the
shade and air pollution fixation potential of the green they implement.

Moreover, basing green space classification on remote sensing data has considerable 458 potential for using long-term series of imagery and respective long-term change detection 459 (e.g. ~40 years with Landsat images; Wulder et al., 2012). Remote sensing and the 460 functional indicators that can be derived from it as spatio-temporal traits and trait 461 variations (Wellmann et al., 2018; Lausch et al., 2016a, b) also permit a gradual distinction 462 and assessment of urban green space functionality, which is key to more accurately 463 monitoring and developing a better understanding of ecological (bio-geo-chemical) 464 processes in the vegetation layer of cities. Traits can be also used for a more qualitative 465 466 assessment (gains and losses) of taxonomic and structural biodiversity in cities (Lausch et 467 al., 2015, 2016a).

Uncertainties: As in every remote sensing study, our study is accompanied by certain 468 469 margins of error. Transferability: Since our study is based on several GIS layers, the transferability of our study in a 1:1 scheme is only given if these data sources are present. 470 471 Another limiting factor in terms of data availability might be the use of RapidEye remote 472 sensing data since it is not normally provided free of charge; however, with the new Planet Labs academic data policy the download of RapidEye scenes got feasible. Trade-offs 473 accompanied by the use of coarser satellite sensors, such as Landsat or Sentinel, and also 474 benefits of finer sensors (Worldview) will be reviewed in upcoming studies in order to 475 476 justify and enrich the results and statements achieved in this study.

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478 **6** Conclusions

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In this study, we applied a novel 'urban front and back yard green derivation algorithm,' namely
a masking of the fractional vegetation data using GIS vector data of land cover to delineate the
front and backyard green space of residential houses in a city with an accuracy of 96%.
Combining these two approaches, we can calculate the area of urban front and back yard green

space for the entire city (including different types of residential structures) and compare this 484 data to the area of public (parks, urban forests) and semi-public (allotment gardens) green 485 spaces that have been used for the analysis of prevailing per capita green space availability. In 486 combination with the remote sensing based spectral unmixing techniques, we are able to 487 delineate house-attached, semi-public and public green spaces in a mid-density Central 488 European city with a high degree of certainty (MAE 7%). For the city Leipzig – a case in point 489 in terms of structural characteristics and growth-shrinkage dynamics in Europe – key findings 490 include that the total amount of house green space is almost 2000 ha, which is ~40% of the 491 492 amount of public green (4768ha) and about 10% of the total urban area. In a quarter of all 63 total districts, front and back yard green space is the major source of green, a finding that 493 494 highlights the importance of this structure for urban green space analysis of livelihood but also for urban planning and ecosystem service assessments. 495

496 The delineation of property-attached green that is not part of green space classes in classical land-use maps is therefore of pronounced importance. Using the novel and more complete 497 498 dataset of urban green space, we can add this data to that of public (parks, urban forests) and semi-public (allotment gardens) green spaces to create a more complete picture of green space 499 (availability when compared to the population data) in cities and to classify 'hot' (well-500 supplied) and 'cold' (undersupplied) spots. As this new green space class relates to private 501 ground, it cannot be 'simply' part of the overall per capita green space availability equation 502 (Kabisch et al., 2016) but is also important as complementary proxy in approaching a real 503 assessment of green space availability for urban residents, almost all of whom have access to 504 the urban front and back yard green space of the property they live in. In addition, agricultural 505 land – be it used or unused – has, so far, not been included in the public green space data as it 506 507 is not clear how this land will develop in the sample city of Leipzig.

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