### This is the accepted manuscript version of the contribution published as:

Jax, K., Furman, E., Saarikoski, H., Barton, D.N., Delbaere, B., Dick, J., Duke, G., Görg, C., Gómez-Baggethun, E., Harrison, P.A., Maes, J., Pérez-Soba, M., Saarela, S.-R., Turkelboom, F., van Dijk, J., Watt, A.D. (2018): Handling a messy world: Lessons learned when trying to make the ecosystem services concept operational *Ecosyst. Serv.* **29** (Part C), 415 – 427

### The publisher's version is available at:

http://dx.doi.org/10.1016/j.ecoser.2017.08.001

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# 38 Abstract

- 39 The concept of ecosystem services is widely used in the scientific literature and increasingly also in
- 40 policy and practice. Nevertheless, operationalising the concept, i.e. putting it into practice, is still a

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41 challenge. We describe the approach of the EU-project OpenNESS (Operationalisation of Ecosystem 42 Services and Natural Capital), which was created in response to this challenge to critically evaluate 43 the concept when applied to real world problems at different scales and in different policy sectors. 44 General requirements for operationalization, the relevance of conceptual frameworks and lessons 45 learnt from 27 case study applications are synthesized in a set of guiding principles. We also briefly 46 describe some integrative tools as developed in OpenNESS which support the implementation of the 47 principles. The guiding principles are grouped under three major headlines: "Defining the problem 48 and opening up the problem space", "Considering ethical issues" and "Assessing alternative methods, 49 tools and actions". Real world problems are often "wicked" problems, which at first are seldom clear-50 cut and well-defined, but often rather complex and subject to differing interpretations and interests. 51 We take account of that complexity and emphasise that there is not one simple and straightforward 52 way to approach real world problems involving ecosystem services. The principles and tools 53 presented are meant to provide some guidance for tackling this complexity by means of a 54 transdisciplinary methodology that facilitates the operationalisation of the ecosystem services 55 concept. 56 57 58 Highlights 59 A set of guiding principles for applying the ecosystem service concept is proposed 60 • Tackling real world problems using the ecosystem services concept requires integrative tools 61 • There is not only one approach or tool; guidance for choosing between alternatives is needed 62 Involving knowledge brokers which are already familiar with the concept is often desirable 63 64 Keywords: ecosystem services, conceptual frameworks, integrative tools, guiding principles, 65 **OpenNESS** project 66 67 68 1. Introduction 69 The concept of ecosystem services (ES) is widely used in the scientific literature and increasingly also 70 in policy and practice documents (notably, MA 2005, TEEB 2010, European Commission 2006, 2011, 71 UNEP 2015). While the general idea of ES as the contribution of nature to human well-being is 72 intuitively appealing and easily understandable, putting the concept into practice is still a challenge 73 (Daily et al. 2008, Primmer & Furman 2012, Bouma and Van Beukering 2015). In this paper, we 74 provide an overview of an approach to the operationalisation of ecosystem services and natural 75 capital taken in the EU-project OpenNESS (Operationalisation of Ecosystem Services and Natural 76 Capital), which was created in response to this challenge. OpenNESS focused on testing how the ES 77 concept could be operationalised and applied to real world problems at different scales and in 78 different policy sectors, involving a wide range of stakeholders (see Furman et al. 2017a, Dick et al. 79 2017a). We describe here some major lessons learned when trying to make the ES concept 80 operational, expressed as ten guiding principles, and briefly sketch some integrative tools to 81 implement these principles as developed in OpenNESS. Detailed descriptions of the principles and

82 methodologies introduced here are elaborated in other papers of this special issue (see overview in
83 Furman et al. 2017a).

84 The goal of operationalisation is to put theoretical concepts into practice, by finding rules or guiding 85 principles for their application. Rules of application are most often not included in definitions of 86 concepts despite the argument that one basic requirement of sound and useful scientific concepts is 87 what van der Steen (1990) calls the "principle of operationality", which "concerns our ability to 88 decide whether some item does belong to the empirical reference of a concept" (van der Steen 1990, 89 p. 385). In other words, which phenomena can be subsumed under a concept and which cannot. This 90 is, however, only part of what "operationalisation" entails, as there is another, broader and more 91 practical dimension to operationalisation of concepts than the one defined by van der Steen (1990). 92 That is linking conceptual work with empirical work on real world conditions and situations in order 93 to find ways to express and use concepts in practice.

94 Operationalisation in this sense includes conceptual, procedural, and methodological aspects. We 95 therefore set out to consider relevant conceptual frameworks in relation to problems encountered in 96 a series of 27 real-life case studies in a range of ecological and socio-economic contexts in Europe 97 and worldwide (see Furman et al. 2017a, Dick et al. 2017a). A major purpose of conceptual 98 frameworks for ES, such as in the Millennium Ecosystem Assessment (MA 2005) framework, the 99 Cascade Model (Potschin and Haines-Young, 2011), the framework of the Inter-Governmental 100 Science-Policy Platform on Biodiversity and Ecosystem Services (IPBES; see UNEP 2015), or even the 101 interlinkages of the SDGs (Agenda2030, 2015), is to visualise a particular set of complex relationships 102 between humans and nature (namely those that contribute to human well-being) as an aid to 103 understanding. These frameworks show how ES may relate to ecosystem structure processes and 104 functions, and to the various benefits and values that promote human well-being. Such conceptual 105 frameworks provide important support in operationalising the concept of ES (see e.g. Saarikoski et al. 106 2015). Specifically, the usefulness of the Cascade Model as a framework was explored through a 107 focus group discussion in the Cairngorms case study (Dick et al. 2017b) and by a number of 108 questionnaires with all case studies (Potschin et al. 2017). Stakeholders highlighted three themes in 109 particular in how the cascade model supported their work: (i) strengthening communication, (ii) 110 developing understanding and (iii) coordinating actions. Conceptual frameworks allow for a general 111 orientation and understanding of the ES idea and help to distinguish and delimit different 112 phenomena, such as biodiversity, ecosystem functions, ES, and benefits that flow from ES. OpenNESS 113 stakeholders also expressed their need for common understanding of terms used as part of the ES 114 concept (Carmen et al. 2017). Clear definition of these related terms is a necessary but not sufficient 115 step to be able to measure them, to compare results between different studies, and to derive 116 generalisations from empirical data that should allow predictions such as what happens to ES and 117 human well-being when there is a particular change in biodiversity? However, when terms such as 118 "ecosystem services", "benefits", and "values" are applied in the field, and as a basis for action in 119 real-world contexts, we need to make them operational. That is, we need rules (or guidelines) stating 120 how they should be applied, including rules for measurement and implementation (Daily et al. 2008).

- 121 This does not mean that there must be one unique definition of each term for all purposes, but when
- 122 it comes to applying the ES concept, their meaning must be clear. In the context of OpenNESS
- 123 'measurement' is understood in a wide sense of both biophysical and monetary metrics, as well as
- 124 qualitative, but consistent descriptions of socio-cultural phenomena. Similarly, 'implementation' has
- 125 a broad interpretation of application 'beyond science' in terms of both changing public perspectives,
- 126 supporting action, and assessing outcomes (Ruckelshaus et al. 2015, and see below).
- 127 Bringing a conceptual framework into the field takes us far from the ordered world represented in
- 128 such frameworks, to a world where things are more complex and 'messy'. Conceptual frameworks
- 129 can, however, help to provide structure to this complex real-world, highlighting important inter-
- 130 linkages and avenues for measurement and assessment (Jann et al. 2007).
- 131 In OpenNESS we aimed to provide systematic guidance for operationalising the ES concept. We also 132 consider the concept of natural capital  $(NC)^2$  to a more limited extent. When looking at the state-of-133 the-art in the literature on operationalisation of ES much has been written on concepts and methods 134 (Potschin et al., 2016). Only recently, however, have scientists started to publish analyses that 135 elaborate on practices in various contexts (e.g. Hauck et al. 2014, Primmer et al. 2015, Spangenberg 136 et al. 2015, Grêt-Regamey et al., in press). The challenge with putting the concept of ES into practice 137 is that real-world problems, as already mentioned above, are seldom clear-cut and well-defined, but 138 often rather complex and 'messy', including both indirect and unexpected linkages, both ecologically 139 and socially (Norton and Nooan 2007, Langemeyer et al. 2016). Furthermore, they involve multiple 140 knowledge producers, interests and values, as well as shifting institutional, economic and political 141 environments (Balint et al. 2011, Salomaa et al, 2016). Likewise, the ways to solve such problems and 142 find the proper place for the application of the ES concept are not easy or straightforward, but may 143 require iteration and take unexpected turns before they materialise. In fact, there may often be 144 multiple paths and methodologies for tackling a problem, depending on their specific ecological, 145 social and political contexts. This was a major assumption before starting the project, which was 146 corroborated by the variety of case studies. Moreover, the ES concept may give rise to (alternative) 147 solutions that may compete with more conventional ways of dealing with problems, such as 148 engineered or technological solutions. Therefore, operationalising the concept of ES cannot be 149 captured as a simple one-size-fits-all solution and one simple scheme of application but needs to 150 take account of the variety of questions, contexts and purposes, both to avoid either overly 151 complicating or simplifying the issues at hand. In addition, it is not only necessary to describe the 152 potential of the ES concept but to be aware of the limitations of applying it. In fact, the concepts of 153 ES and NC, with their economic connotations of 'services' and 'capital', will always be only two 154 among many possible metaphors that capture the importance of nature to humans and cannot be 155 taken as a panacea for solving environmental problems (Larson 2011, Raymond et al. 2013, 156 Spangenberg et al. 2014).

<sup>&</sup>lt;sup>2</sup> See Furman et al. 2017a for a more detailed discussion of the natural capital concept.

157 OpenNESS allowed us, especially through its many case studies, to obtain insight into the multitude

- 158 and complexity of real-world problems and solutions. In this paper we describe this complexity and
- 159 elaborate a set of guiding principles and tools which may be appropriate for solving complex
- 160 environmental or socio-ecological problems. By this we do not mean to cover all aspects of
- 161 operationalisation in all steps of implementation, but to focus on procedural and methodological
- aspects which we see neglected in many studies and/or for which OpenNESS provided important
- 163 ideas and tools. In spite of the complexity and even messiness of real world problems, which we take
- 164 into account here, we acknowledge that some form of simplification is necessary in order to be able
- 165 to consider the broad range of contexts where operationalisation takes place, including the
- 166 limitations on time available, on personnel or institutional ambitions and skills, as well as on budgets
- 167 available. The resulting guidance tries to balance these concerns by providing a means to synthesise,
- 168 focus and make the procedural choices needed for the various settings in which people operate (see
- 169 Potschin et al. 2017, Harrison et al. 2017, Pérez-Soba et al. 2017, Turkelboom et al. 2017).
- 170 171

# 172 2. Methodological approach: a procedure to link research, real-world problems and societal173 challenges

- 174 The process of the OpenNESS project proved to be valuable in itself. An anonymous survey with 246
- 175 practitioners in OpenNESS case studies found out that "to a large extent the impact [of
- 176 operationalising the ES concept in OpenNESS] was attributed to a well conducted science-practice
- 177 interaction process (>70%)." (Dick et al. 2017a, p. xyz). In this section, we therefore briefly describe
- 178 the crucial procedural features of the project, as a potential guide for future projects aiming at
- 179 applying the ES concept.
- 180 In the project, we used a transdisciplinary approach to guide problem solving (Furman et al. 2017b).
- 181 There was an inter-play between researchers from various disciplines including sociology, political
- 182 sciences, environmental sciences, geography, economics, philosophy, biology and ecology, experts
- 183 from communication, policy, and business, as well as various local and EU level stakeholders (Carmen
- 184 et al. 2017; Dick et al. 2017a; Turkelboom et al. 2017).
- 185 The research design was based on an iterative application of ES assessment methods and tools in 27
- 186 place-based case studies in thirteen European and four non-European countries (see Wijnia et al.
- 187 2016). Out of the 27 case studies, 25 case study sites had local study teams which included both
- 188 researchers and local, non-scientific experts in the implementation of the ES concept, as well as Case
- 189 Study Advisory Boards (CABs) in which the various local stakeholder groups were represented (see
- 190 Dick et al. 2017a). The case study teams not only provided local knowledge but also were often
- 191 involved in refining the research questions to be explored and selecting the tools to be applied.
- 192 Together with the CABs, these place-based research teams tested the conceptual, methodological
- 193 and governance-oriented tools and approaches developed by the project. This included challenging
- 194 the tools and approaches with respect to the needs and requirements of local practitioners in putting
- 195 the ES concept into practice.

- 196 The case studies allowed us to test the applicability of our conceptual and methodological work on
- 197 ES by identifying potential solutions to specific problems in the case studies, as well as at the EU and
- 198 national levels. It also allowed us to further develop our conceptual and methodological
- 199 understanding by generalising from specific case studies, thereby guiding future users of the ES
- 200 concept. The interaction between those involved in conceptual and methodological development
- 201 and those testing concepts and methods in the case studies promoted common interests and social
- 202 learning through tackling problems from different angles in an open collaboration. Thus a
- 203 continuous, iterative dialogue led to research outcomes that were tested in practice and suitable for
- 204 implementation in the real world (Dick et al. 2017a).
- 205 The cases were also analysed according to the societal challenges where operationalisation of ES
- 206 could play a role. We selected four major challenges: human wellbeing, sustainable ecosystem
- 207 management, governance, and competitiveness. These are discussed in more detail in Potschin et al.
- 208 (2017).
- 209 In the following section, we first elaborate ten guiding principles for operationalising the ES concept
- $210 \qquad \text{and then in section 4 provide examples of important tools that were produced in OpenNESS and \\$
- 211 which support the implementation of these principles.
- 212

# 3. From problems to solutions: guiding principles for operationalising the ecosystem servicesconcept

- 215 Sometimes it is assumed that solving an environmental problem should work as a simple linear or 216 circular process roughly along the lines of the ideal policy cycle. Regarding ecosystem services, this 217 process starts with defining the problem, defining the relevant ES, assessing and valuing the ES, 218 suggesting solutions to decision makers, adopting and implementing the solution, monitoring and 219 evaluating the effect of the solution, and then recommencing the cycle once again to assess whether 220 any further adjustment is required. Such approaches are important and much progress has been 221 made in elaborating them further and adapting them (e.g. Chan et al. 2012, Förster et al. 2015); we 222 also make use of them as a simplified form of guidance to ecosystem service assessments (see 223 section 4, ESAST). Experiences from OpenNESS suggest, however, that this may often be an 224 oversimplification, which does not fit the way many real-world problems are tackled (see 225 Langemeyer et al. 2016).
- 226 In this section, we describe ten guiding principles which we deem are necessary to apply the ES
- 227 concept to a variety of problems. These principles were developed on the basis of several sources. In
- 228 part they are based on empirical experiences drawn from the case studies in OpenNESS, in part they
- 229 are reflections taken from the literature and/or from our previous work and then "tested" in the case
- 230 studies. We describe the principles as propositions and characterise the potential or actual problems
- 231 or obstacles to which they respond as well as the evidence that support them, both from the
- 232 literature, from the OpenNESS case studies and from other conceptual and methodological work
- 233 undertaken in the project.

- 234 The order of the principles we describe here does not imply a fixed sequence of the pathway from
- 235 problems to solutions, neither their importance; specific problems and situations require different
- entry points and sometimes quite convoluted pathways, and perhaps also with iteration (already
- emphasised by Chan et al. 2012, and see Mouchet et al. 2014, Langemeyer et al. 2016; also Potschin
- et al. 2017). This is why we deliberately use the term "principles" instead of "steps". Depending on
- the specific problem, some principles may not be pertinent at all. For instance, it is not always
- 240 necessary to make a complex choice of methods, in order to map, quantitatively assess, or formally
- value ES. Sometimes ES may simply be used as metaphors or heuristics for explaining and/or
- structuring a problem, without any need for quantification. In the following, we list and briefly
- characterise the principles we consider as crucial for operationalising the ES concept in various
- problem contexts. We do not claim, however, that this is all that can be said about operationalisation
- of ES, especially concerning methodological approaches; instead we focus on some *fundamental*
- 246 *issues*, which are often underestimated when applying ES approaches to real-world problems. In the
- 247 following section, we then briefly sketch some selected tools that OpenNESS assessed for supporting
- 248 the implementation of these principles.
- 249

#### 250 **3.1.** Defining a problem and opening up the problem space

251 A problem space consists of all the different aspects or components of a problem as well as the

- 252 (often various) possible pathways to its solution. Not all these components are evident to everyone
- 253 from the outset. Defining and framing (Bardwell 1991, Hajer 2006) the problem is easily
- underestimated, but crucial to an effective and efficient way of applying the ES concept in a useful
- 255 way (Wittmer & Gundimedia 2012). Therefore, adequate time should be devoted to it. It comprises
- 256 five of our guiding principles, which we now describe (see table 1 for summary).
- A) As real-world problems involve and affect people, it will in most cases be necessary to involve
   stakeholders from the very beginning, when the problem is defined and the entire problem space
   is laid out.
- 260 Stakeholders in our context are any individual or group of people who can affect the use, or is 261 affected by the use, of ecosystem services (Hauck et al. 2013). Relevant stakeholders can be defined 262 by answering questions, such as: Who is affected? Who derives benefits? Who manages the delivery 263 of ES? Who decides? Who can influence the policy or management rules (Lovens et al. 2015)? It is 264 important to identify together with stakeholders the relevant ES and the potential benefits that 265 different groups of people derive from these services (see also Potschin and Haines-Young 2011). 266 Other objects of value (e.g. built infrastructure, culture etc.), which may be linked to the problem, 267 should also be identified and included at the same time (see also principle F). While often done alone 268 by scientists providing a list of ES, it is often better to adopt a transdisciplinary approach (Cash et al. 269 2003), involving stakeholders with their specific local knowledge and interests in the selection of 270 relevant ES. Involving stakeholders in this way can also prevent potential biases from pre-elaborated 271 ES classifications that may be at odds with stakeholders' ways of perceiving these services. For

- example, Chan et al. (2012) illustrate the importance of this in the context of salmon fishing in British
- 273 Columbia. "Wild salmon fishing" would be perceived by most scientists normally as a provisioning
- service, however it also has a crucial cultural value for the local people, related to their cultural
- identity. (Fishing of) wild salmon thus was not replaceable by farmed salmon, as it would have been
- if salmon had been considered only as a provisioning service.
- 277 Missing or unsatisfactory stakeholder involvement can be a major impediment to using the ES
- 278 concept in real-world situations, both in terms of legitimacy, as well as in terms of missing crucial
- information on the respective social and ecological context (thus saliency) (Cash et al. 2003). On the
- 280 other hand, haphazard participation can be costly and ineffective in representing social interests
- 281 (Paloniemi et al. 2015).
- 282 Stakeholder involvement was a crucial element in the OpenNESS case studies, most of which
- 283 involved a Case Study Advisory Board (CAB) consisting of practitioners, policy-makers and place-
- based experts (Dick et al. 2017a; Saarikoski et al. 2017). Dick et al. (2017a) found that stakeholder
- 285 perspectives were involved in framing the issue in 40% of the OpenNESS case studies. Saarikoski et al
- 286 (2017) also found that a transdisciplinary research approach increased mutual understanding
- 287 between planners and researchers in several OpenNESS case studies, especially in cases where
- 288 stakeholders were involved in joint problem formulation. For example, in one Belgian case study (De
- 289 Cirkel), the research topics were defined based on strategic knowledge gaps of the project managers
- and on expertise available within the research team. This resulted in a research focus on the
- 291 landscape needs of local inhabitants and their perception of the functions of traditional orchards.
- 292 This approach enabled a direct uptake of some of the findings of the case study.
- 293

# 294 B) The role of scientists in approaching the problem should be clarified

- 295 Scientists can have various tasks and roles in contributing to solving real-world problems, from "pure 296 scientist" to issues advocate, but also as an "honest broker" towards finding (policy) alternatives 297 (Pielke 2007). The more deeply they are involved, the more important it will be for them to gain trust 298 and acceptance to be involved within the discourse (Chan et al. 2012). To increase legitimacy and 299 effectiveness (Cash et al. 2003, Heink et al. 2015), scientists should see themselves as having a 300 designated, but not dominant, role in a group of people who are collectively identifying and solving 301 the problems at stake. Too close a relationship may lead to a dependence on the researcher which 302 was not planned beforehand and which is out of scope of his or her project (Stone 2006).
- 303 In OpenNESS the main interaction with stakeholders was provided by creating CABs. Most cases
- 304 studies were initiated by researchers, identifying potential problems and being partly based on
- 305 previous research in the area (e.g. for landscape-ecological planning in urban and peri-urban areas in
- 306 Slovakia, Bezák et al. 2017), or for farmland management in Kiskunság in Hungary, Kelemen et al.
- 307 2015a). However, as a spin-off, e.g., of the Belgian case studies, researchers were contacted by the
- 308 city of Genk and the Provincial administration of Oost-Vlaanderen to start similar research with their

- 309 case studies (i.e. Stiemerbeek and Maarkebeek) to support solving already identified problems. More
- 310 than 80% of OpenNESS stakeholders responding to a questionnaire stated that the people involved
- 311 were trusted and that the researchers provided good facilitation (Dick et al., 2017a).
- 312

C) The complexity and often "wickedness" of the problem should be acknowledged and the nature of
the problem, including its social, ecological, administrative and economic spheres, should be
charted. Simplistic understandings of problems should be avoided to ensure that the problem and
how an ES approach might contribute to its solution is clearly expressed. Hidden links between
the different spheres need to be exposed.

318 Scientists as well as decision makers strive for clear-cut questions and problem descriptions. Many 319 real-world problems, especially those involving social-ecological systems with multiple stakeholders 320 and interests, are by their very nature "messy", often only vaguely captured; they are also complex, 321 uncertain and urgent (Funtowicz and Ravetz, 1994). Norton and Noonan (2007) call environmental 322 problems "wicked problems", by which they mean, following Rittel and Webber (1973), that "they do 323 not emerge as well-defined problems that are formulated similarly by different participants in the 324 discussion. There will be, on the contrary, varied complaints and varied explanations of what the 325 problem is, often associated with varied value positions and perspectives of the participants." 326 (Norton and Noonan 2007, p. 672). Defining the problem itself is a result of a social process (an 327 interaction of actors allowed to take part in problem definition). How a problem is defined (whether 328 structured, moderately structured or unstructured) affects significantly how it is handled (Hoppe 329 2011).

330 The variety of case studies in OpenNESS showed how different the entry points to a complex

problem can be. It may be, e.g., a social conflict (such as in the Cairngorms case study on woodland

332 creation, Dick et al 2017b) or the demand deriving from some formal regulatory requirement (such

- as in the Loch Leven case study, where the overarching aim was assessing the consequences of the
- 334 EU Water Framework-Directive for the delivery of ES). Nevertheless, the social, ecological and

335 political aspects will generally be linked in a complex manner.

336 García-Llorente et al. (2016) described this complexity for two protected areas which were also 337 OpenNESS case study sites (Doñana and Sierra Nevada), emphasising the differing perceptions and 338 priorities of environmental managers and researchers compared to that of users of ES. To account 339 for the multiple complex problem constellations experienced in OpenNESS case studies and 340 elsewhere, we developed and tested the so called OpenNESS Conceptual Nexus (ONEX), in order to 341 tailor the different entry points according to the specific situation. It is designed to find the most 342 appropriate pathway for approaching the problem at hand (see Potschin et al. 2017, Haines-Young et 343 al. 2017, and below). ONEX was tested in the case study in the Kiskunság region (Hungary) and 344 according to the key informant from that case study "enabled a 'comprehensive picture' of the case 345 study to be built up." Also, "the experience was found to 'shed light on non-trivial relationships'

- 346 between different aspects of the problems and issues that were the focus of the case study, that
- 347 were previously not so well articulated." (Haines-Young et al. 2017, p. 81)
- 348
- 349
- D) Political space and influence spheres should be defined. The manoeuvring space (what is possible 350 within the boundaries of e.g. a legal or societal situation) should be defined.

351 Defining political space and manoeuvring space is a matter of clarifying governance conditions as

- 352 well as power relations (Gómez-Baggethun et al. 2013, Berbés-Blázquez et al. 2016), for finding
- 353 consent between different stakeholders for potential solutions and an appropriate problem
- 354 delimitation and simplification. Manoeuvring space refers also to the space of possible solutions, as
- 355 they may be restricted by e.g. property rights, budgetary restrictions, or policy regulations (e.g. the
- 356 EU Water Framework Directive or the EU Habitats Directive). Being clear and transparent about such
- 357 limitations is important both to focus research as well as avoid unrealistic expectations among
- 358 stakeholders in terms of implementation (Reed et al. 2014, Spangenberg et al. 2015, Görg et al.
- 359 2014), which may undermine trust between researchers and stakeholders (Cash et al. 2003).
- 360 In their analysis on the possibilities of mainstreaming the ES concept in EU policy making, conducted 361 as part of the OpenNESS project, Schleyer et al. (2015) warn of raising wrong and unfulfillable 362 expectations. According to their study, the ES concept is only partly incorporated in EU policy making 363 and currently restricted to the environmental sector. Thus, its potential to address trade-offs with 364 other policy sectors (such as agriculture or regional development) and to identify possible synergies 365 is still limited due to the silo mentality of policy-making and other administrative challenges 366 (including power imbalances across sectors). To move forward, a deeper understanding of the 367 factors affecting the uptake is required, including communication barriers, stakeholder attitudes to 368 the ES concept, and tensions between policy sectors.
- 369 At a national to local scale, Bezak et al. (2017) in their OpenNESS case study in Slovakia identified,
- 370 e.g., the existence of partly contradictory legislation and regulations for spatial land use planning and
- 371 assessment. At a local scale, they found perceived obstacles for ES-based management in the
- 372 "complex land ownership structures in Slovakia where many owners are unknown and some private
- 373 properties have many owners" (p. 129) and also in the resistance of many politicians on the local
- 374 level, who perceived environmental legislation as an obstacle to rural development. To avoid
- 375 frustration of local stakeholders and to aid discussion of synergies and trade-offs, local networks
- 376 could play an important role, and 'local ownership' of these 'integration' frameworks should be 377 encouraged.
- 378

#### 379 *E)* Concepts and language should be adapted to the specific situations and stakeholders.

- 380 The required precision of concepts depends on the specific problem and situation. Vague concepts
- 381 may be sufficient or even better at some stages in the research process (theory formation) and also
- 382 for some application purposes (general communication about the value of nature for humans,

- 383 "didactic purpose", see e.g. Jax & Heink 2015). Vague concepts are also important in bringing people
- into the process, as no interpretation has yet been left out or closed and they feel that they still have
- a say in what the discussion is about (boundary concepts; see Abson et al. 2014), as also expressed
- by some OpenNESS stakeholders (Carmen et al. 2017). Vagueness, however, can become
- 387 problematic when decision-support is required, as argued above (section 1). It can impede
- 388 operationalisation of the ES concept in real world situations and implementation of results. The ES
- 389 terminology is replete with concepts that are either vague or for which multiple definitions exist (see
- 390 e.g. Jax 2016).
- 391 This also relates to the language used (Carmen et al. 2017). It is important to use terms and words
- 392 that are understood well by all in the process. Opening up the meaning of technical words is
- 393 essential for transferring the ES concept into practice. With its economic connotations, the language
- 394 used may not be familiar for many stakeholders and decision makers (e.g. Lamarque et al. 2011, Böck
- 395 et al. 2015), or it may appear to be less suitable to a stakeholders way of approaching their relations
- 396 with nature (e.g. Turnhout et al. 2013, see also the recent discussion on the IPBES framework, e.g.
- 397 Borie & Hulme 2015). ES language may thus often require "translation" when communicating it
- 398 (Gómez-Baggethun and de Groot 2010). In many previous studies, the questions put to elicit relevant
- 399 ES from stakeholders were (at least initially) not phrased in the ES terminology, especially at a local
- 400 or regional scale (e.g. Chan et al. 2012, Koschke et al. 2014).
- 401 Metaphors to describe nature should be tailored to specific audiences and decision-making contexts.
- 402 For example, the term 'natural capital' works well in economic discussions on environmental
- 403 accounting, 'green infrastructure' can work well in discussions with urban planners, 'ecosystems' fits
- 404 discussions with ecologists, 'Mother Earth' is suitable in discussions with indigenous peoples on
- 405 nature's sacredness, whereas 'nature' may still be the best term to communicate with general
- 406 audiences (Gómez-Baggethun and de Groot 2010). We should, however, also be aware that the
- 407 different metaphors are not simple translations, but often also carry decisive differences ("framings")
- 408 in terms of the values connected to them (Bardwell 1991, Larson 2011, Raymond et al. 2013), which
- 409 must be considered in their own right (see Section 3.2). Discussions about the exact framing are
- 410 always boundary negotiations, shaped by power relations, that define the precise meaning and
- 411 relevance of a problem across a variety of stakeholder perceptions and scientific disciplines and thus
- 412 require a truly inter- and transdisciplinary approach (Schleyer et al. 2017).
- 413 In some OpenNESS case studies, the appropriate terms were discussed with the stakeholders. It was 414 attempted to find 'context-relevant' and 'self-explanatory' terms in an interactive process (Ulenaers 415 et al. 2014). In some case studies the ES terminology was simplified as a response to interactions 416 with the stakeholders. In the case study in Sibbesborg, Finland, for example, the five steps of the 417 cascade model were pooled and reduced to three because some distinctions (in this case between 418 structure and function, and between service and benefit, respectively) were not clear to the planners 419 and eventually not required for the purpose, namely to "structure thinking and communicate with 420 planners and residents" (Jari Niemelä, personal communication, June 2016). Likewise, interviews on

- 421 important ES used in the Hungarian case study did not use the ES terminology since "previous
- 422 experience had shown that locals were not familiar with the term and had difficulties relating to the
- 423 scientific categories" (Kovács et al. 2015, p. 121).
- 424

#### 425 **3.2. Considering ethical issues**

426 Ethical issues in the use of the ES concept arise in various ways and are often unrecognized. Making427 them visible and considering them in research and application is necessary.

- F) When applying the ES concept in a specific situation, hidden and neglected issues, hidden values,
  and hidden links between issues should be revealed and made transparent. Hidden and
  suppressed issues and values may be the most important in terms of conflicts and conflict
  resolution between stakeholders.
- 432 Due to the complexity of real-world problems, researchers, decision makers and other stakeholders 433 striving to apply the ES concept may easily overlook some issues, as well as underlying values or links 434 between issues. Such things are often of ethical relevance (see Luck et al. 2012, Jax et al. 2013). They 435 refer to questions of justice such as who benefits, who carries the burdens of ES production or 436 impairment (Daw et al. 2011; 2015, Pascual et al. 2014; Phelps et al. 2015)? Thus, for example, after 437 the occurrence of mad cow disease in the 1990s and the resulting strong restrictions for feeding 438 meat and bone meal to cattle, Europe's import of soybeans from South America for feed strongly 439 increased. Importing ES (feed for more healthy animals and food) from South America has further 440 increased deforestation and the transformation of other natural ecosystems there and - via the use 441 of high loads of pesticides – partly led to diseases among the local population (WWF 2014). 442 Relevant items and values may not be captured by the ES concept, e.g. items which have no obvious 443 value to most people or which some people consider to have intrinsic value (Davidson 2013, Jax and 444 Heink 2015). But there are also links between issues that are easily overlooked. These include links 445 between different ES categories. Reyes García et al. (2015), for example, found that edible wild 446 plants for many people are not primarily a provisioning service, but their use is mainly continued 447 because they have a high cultural and recreational value and thus also represent cultural services 448 (see also Schnegg et al. 2014, Chan et al. 2012, 2016). Also, there are often crucial and complex links 449 between ecological, social and economic issues, which are decisive for understanding and solving a 450 problem, but may be missed if only one type of expertise is available (Abson et al. 2014).<sup>3</sup> These can 451 be severe problems for operationalising the ES concept in a way that is appropriate for the problem 452 at hand, not the least in the sense of reaching compromise solutions that are comprehensive and 453 acceptable to all or most stakeholder groups.
- In one OpenNESS case study, for example, García-Llorente et al. (2016) found that priorities given to
   specific ES in two Spanish protected areas (Sierra Nevada and Doñana Natural and National Parks)

<sup>&</sup>lt;sup>3</sup> Of course, also less obvious relationships in the biophysical system (e.g. groundwater recharge) must be visualised.

- 456 were different between managers and researchers on the one hand and ES users on the other; in
- 457 consequence only some of the ES considered as vulnerable and important by stakeholders were part
- 458 of the management plan of the protected areas, providing potential for conflict and loss of valued ES.
- In the Hungarian case study, winners and losers from conservation-related land use changes becameapparent through application of the ES concept and trade-off-analysis for different stakeholders. This
- 461 then provided a clearer view on the specific potentials for conflict (Kovacs et al. 2015).
- 462 Tools developed in OpenNESS (such as The ONEX; see below) also provide support in identifying
- 463 hidden values through raising questions and opening up the problem space to raise awareness of
- 464 issues (such as justice and value plurality) that otherwise might easily be overlooked. We also have
- 465 promoted methodologies for an integrated valuation of ES that is designed to cover a plurality of
- 466 values embraced by people (see Jacobs et al. 2016, and below). For example, one of the Scottish case
- 467 studies, showed that established methods such as QUICKscan can support elicitation of different
- 468 values and viewpoints, and aid communication (Dick et al. 2017).
- 469
- G) To avoid socially unacceptable results or decisions arising from application of the ES concept the
  social and political compatibility of outcomes should be assessed and potential winners and losers
  identified during the implementation process.
- 473 The outcomes of applying the ES concept may not always be welcomed by every stakeholder.
- 474 Unexpected issues (materially, economically or socially) may also arise which may compromise
- 475 particular interests and the desired problem solution. Payments for ES schemes may, for example,
- 476 lead to locally undesirable injustices (e.g. Muradian et al. 2013) or to a loss of previous intrinsic
- 477 motivation to protect nature without any further payments ("crowding out", see. Rode et al. 2015).
- 478 In some cases, monetary valuation may even violate stakeholder perceptions as they feel alienated
- 479 and consider their cultural, social and other non-monetary values as being ignored (Spangenberg et
- 480 al. 2015). This principle thus includes visualising implications of potential and alternative actions on
- 481 ES and biodiversity. It also calls for considering issues of justice and environmental values (Daw et al.
- 482 2011, Sikor et al. 2013).
- 483 In one of our Belgian case studies (De Cerkel), we observed the negative and initially "unexpected"
- 484 side-effects of promoting rural tourism: garbage, apple theft along paths, parking problems, and
- 485 damage to erosion grass strips. Thus one solution brought about unexpected problems. In another
- 486 case study (Cairngorms, Scotland), participatory recreational mapping was used to determine where
- 487 woodland could be located. The map showed roads as non-use areas as the survey focused on
- 488 recreational use. However, one stakeholder commented she really valued the 'view while driving to
- 489 work on a daily basis' along the roads and 'would not like the view blocked by trees' (Jan Dick,
- 490 personal communication, June 2017).
- 491 In OpenNESS we addressed this issue by developing participatory scenarios (Priess et al. 2017) and an
- 492 integrative valuation tool able to include biophysical, monetary and non-monetary valuation

- 493 methods and thus, in principle, to respect different cultural perceptions (Jacobs et al. 2016, Martín
  494 Lopez et al. 2017 and see below).
- 495

### 496 **3.3** Assessing alternative methods, tools and actions

497 H) ES tools and methods may not always be the only or even the best choice. To determine what

498 kind of assessment is needed for decision-making, deliberative tools for scoping the problem
499 space and the most appropriate tools for problem solving can be useful.

- 500 The ES concept may not be able to address all types of problems, maybe not even fundamentally 501 environmental ones. This may be because the ES concept and its terminology is not accepted due to 502 its specific framing, which some people think does not reflect their relationships with nature, e.g. 503 because they feel that this relationship cannot be expressed as a "service" but more as a "gift" (Borie 504 and Hulme 2015), or because they care about nature without having to receive a benefit from it 505 ("benefits to nature": Davidson 2013). Also, many types of problem have traditionally been dealt 506 with using other tools (e.g. in forest and water management or in traditional landscape planning; see 507 e.g. von Haaren & Albert 2011). Earlier experiences and ways of thinking may have led to a mindset 508 which expects that it is easier and more appropriate to handle the problem using conventional tools, 509 such as multifunctional landscapes, sustainable development, or identifying the need for a protected 510 area (see Raymond et al. 2013, Rozzi 2015, Norton 2015 for alternative approaches).
- 511 The ES concept and deliberative methods which involve stakeholders may not be needed for many
- 512 environmental issues. Existing environmental regulations have in most cases already been
- 513 deliberated by a legislature, and often define the relevant scope of the problem. They may
- 514 nevertheless be supported by an ES approach, which is discussed in several places for the WFD (see
- e.g. Vlachopoulou et al. 2014, Carvalho et al. 2017). At times, however, an ES approach may also be
- 516 used to challenge shortcomings of existing regulations.
- 517 In OpenNESS we documented a situation, where regulating services (removal of air pollutants) were
- 518 assessed to make a limited contribution to climate change mitigation and pollution removal in cities,
- 519 thus concluding that for this specific situation it was more effective to limit pollution sources than
- 520 using the assessed regulating services on green infrastructure sites as ecological sinks (Baró et al.
- 521 2014). In the Finnish bioenergy case study, local stakeholders considered sustainability assessment
- 522 criteria as performing better in describing and handling their situation than the ES approach. The
- 523 latter was seen as being insufficient to express crucial aspects of human well-being related to the
- 524 services (here: bioenergy and timber) like forest owner income, employment, and regional economy
- 525 (Saarikoski et al. 2017).
- 526
- 527

It is important to connect various data pools with each other. This requires including all relevant
 sectors (such as research disciplines, policy fields). However, not all ES need to be assessed, which
 leaves space to focus on the most relevant, assuming all relevant stakeholders are consulted.

531 When ecosystem services are assessed, it often happens that the knowledge used leans on existing 532 databases and institutions. However, the need is often beyond that. This requires modifications or 533 total revision in data gathering, management and sharing on various levels and various institutions, 534 both public and private. Given the costs involved in these data collection processes, it is also 535 important to fit the scale of data resolution to the nature of the problem at hand. Hauck et al. (2013), 536 for example, found in their study that synergies and trade-offs between different ES at a large scale 537 did sometimes not match that at a regional or local scale. Different applications of the ES concept 538 require different degrees of data accuracy, scale and reliability (Gómez-Baggethun and Barton 2013); 539 sometimes collecting too much data may not be effective or even helpful ('optimal ignorance').

540Assessments of ES may require diverse information. In the case of forest management, data is541needed on timber, berries, their pricing etc., often in a form that allows integrated analyses between

542 them, including traditional/local environmental data (Primmer & Furman 2012, Saarikoski et al.

- 543 2015).
- 544

545 J) Different methods, or combinations of methods, can be useful for answering different questions
 546 and therefore it is important to identify as precisely as possible what is wanted (in iterations with
 547 knowledge users) before starting to search for appropriate methods.

548 In terms of methods selection, it is necessary to tailor the methodology to the specific problem at 549 hand. This also should, wherever possible, be done in collaboration with knowledge users (Opdam et 550 al. 2015, Rodela et al. 2017, Harrison et al., 2017). Not every application of the ES concept requires 551 detailed mapping of ES, nor a formal valuation process. If - and at what stage of a process of 552 problem description and solution – particular methods are needed again depends on the specific 553 situation. This may not be clear at the outset as some methodological needs and the pertinence of 554 some approaches may surface only in the course of the process. Alternatively, assessments which are 555 too complex and detailed may lack focus and be ineffective. Very comprehensive methods may take 556 too long to produce results relative to the often short windows of opportunity involved with both 557 policy and practice. Flexibility is necessary here to adapt to the complexity and time-pressures 558 involved in addressing real-world problems (Potschin et al., 2017).

559 There are several methods and tools available to identify, quantify, map and value ES (see e.g.

560 contributions in Potschin et al. 2016). These include methods for quantifying and qualifying ES,

- 561 valuing them, stakeholder analysis, conceptual analysis, and a variety of social science methods (see
- also section 4 below). As said above, which methods need to be used depend on the specific problem
- 563 at hand, and even within a particular field (e.g. valuation) different methods may be possible.

For example, to assess and communicate the importance of urban gardens to policy makers in our
Barcelona case study, it was not necessary to map each ES and value it on a monetary basis. Instead,
social science methods, in particular interviews, were mainly used (Camps-Calvet et al. 2015).

567 In OpenNESS an iterative approach, determining needs, demands and feasibility was practiced. On 568 that basis guidance was developed for selecting methods appropriate to the respective problem(s) 569 and to the situation, e.g. decision-making context, expertise and data available, budgets of time and 570 money (see below and Barton et al., 2017, Jacobs et al. 2017; Dunford et al. 2017; Harrison et al. 571 2017; Perez-Soba et al. 2017, Turkelboom et al. 2017).

572

# 573 [Insert table 1 here]

574

#### 575 **4.** Selected tools from OpenNESS for supporting the implementation of our guiding principles

How can we implement the guiding principles described above? In this section we briefly introduce
some integrative tools, as developed in OpenNESS, and describe how they can support the
implementation of the principles presented above. More detailed accounts of the tools can be found
in other papers of this special issue, as indicated below.

580

#### 581 Ecosystem Service Assessment Support Tool (ESAST)

582 As an overarching guidance for ES assessments, the Ecosystem Service Assessment Support Tool 583 (ESAST) was developed to involve a broad range of stakeholders from the beginning (principle A) (Fig. 584 1). It is hosted in the web-platform Oppla (www.oppla.eu and see below) and offers practical, step-585 by-step guidance on how to carry out an ES assessment process and to integrate the results into 586 management and decision-making. The tool follows loosely the form of the classical "policy cycle" 587 and links to other tools (such as those described below). It starts, following principle A, with an 588 interactive problem formulation process to jointly with key stakeholders define the objectives as well 589 as biophysical and socio-cultural dimensions that are relevant for the management or decision-590 making situation (step 1). The next step (step 2) in ESAST is the identification of the ES and the 591 associated benefits and beneficiaries that are likely to be influenced by the management or policy 592 decisions at hand (principles D and G) followed by an analysis of the ways in which direct and indirect 593 drivers of change influence ecosystems and their capacity to provide services (step 3). To understand 594 the values that people assign to ES and the benefits that they derive from them, ESAST then provides 595 guidance for selecting biophysical assessment methods as well as monetary and non-monetary 596 valuation methods that are fit for purpose (step 4, principle J). In the next step, it then uses the 597 knowledge about ecosystem services to inform actual decision making (see Saarikoski et al. in this 598 issue) through a multitude of decision support tools that are available (see Barton et al. 2017) (step

- 599 5). These tools structure information on management and policy options and their consequences,
- 600 and highlight trade-offs between ES (see Turkelboom et al. 2017). Facilitation, mediation and dispute
- 601 resolution methods are helpful in highly conflictual trade-off situations (supporting principles F and
- 602 G). Overall, knowledge of ES is most effective when decision-makers and key stakeholders have been
- 603 closely involved in the assessment process to ensure that they find the information relevant and
- 604 reliable, and are ready act upon it.
- 605

#### 606 [Insert figure 1 here]

607

# 608 Tool for problem specification: OpenNESS Conceptual Nexus (ONEX)

609 To account for the messiness of real-world problems and the multitude of issues to which the ES

- 610 approach may be applied, the OpenNESS Conceptual Nexus (ONEX) tool was developed. It supports
- 611 the application of several of the above principles.
- 612 The application of the ES concept generally has to be approached iteratively because it usually
- 613 involves diverse groups of stakeholders (principle A), who need to develop a shared understanding of
- 614 issues and potential solutions. To support this deliberative approach, we have explored how
- 615 guidelines can be created to help people understand, discuss and apply key ideas in ways that are
- 616 relevant to their situation (principle E). These guidelines take the form of a 'conceptual nexus', or
- 617 network of concepts, termed ONEX (Potschin et al. 2017). The ONEX tool, starting from the general
- 618 question "what is the issue at hand", guides people via different potential entry points, namely the
- 619 types of ecosystems being considered, the stakeholders involved and the dominant social and
- 620 political processes within the study area (principle D). Via a number of questions, it enables people to
- 621 look at the relationships between ideas and gives them access to resources that allow them to build
- 622 deeper insights into issues (principle C, also supporting principle G). ONEX is implemented using the
- 623 freely available internet tool, Trello, which is widely used for project management. In an operational
- 624 context, researchers often have to work as 'knowledge brokers' (Pielke 2007, Reed et al. 2014) with
- 625 diverse groups of people. ONEX can help them by facilitating deliberative work involving ES, which
- 626 requires the co-production of knowledge and social learning. The focus is on how it can be used by
- 627 case studies to build a richer picture of their problem situation (principle C) by looking at, e.g., the
- 628 four OpenNESS Challenges of human well-being, sustainable ecosystem management, governance,
- and competitiveness (Potschin et al. 2017). The ONEX tool can be downloaded and used via
- 630 <u>https://trello.com/b/sm1lX0S0/the-onex-lab.</u>
- 631

#### 632 **Toolbox for Integrated Assessment and Valuation**

633 In support of principle I and J, OpenNESS guidelines on method selection for conducting integrated 634 assessment and valuation (Braat et al. 2014; Gómez-Baggethun et al. 2014; Kelemen et al. 2015b; 635 Barton and Harrison 2017) have been developed. The toolbox featuring a range of new and existing 636 methods for ES assessment includes different biophysical, socio-cultural and monetary techniques 637 (Dunford et al 2017; Harrison et al. 2017; Jacobs et al. 2017; Smith et al. 2017, Zulian et al. 2017). 638 Characteristics of each method in terms of their requirements (e.g. data, resources, expertise) and 639 purpose (e.g. mapping, deliberation, valuation) are identified, recognising that several methods may 640 suit a specific purpose or that combinations of methods may be useful for addressing certain 641 problems (principle I and J). Documentation describing the steps required to implement each 642 method is supplied. Several approaches which aim to provide guidance on selecting methods for 643 biophysical, socio-cultural and monetary valuation, including a set of interlinked decision trees 644 (Harrison et al. 2017), matrices of method considerations (Dunford et al. 2017), plural value 645 dimensions covered by different methods (Jacobs et al. 2017), and an online method selection tool 646 (Barton and Harrison 2017) are provided.

647 As an important lesson learned from the OpenNESS project we recommend that decision trees are 648 not used literally to make decisions on method choice, but as an organised way of asking questions 649 that aid method selection. Recognising that decision trees are limited by their fixed structures, they 650 are supplemented by other approaches such as the online method selection tool<sup>4</sup>. The tool uses 651 Bayesian Belief Network software to address method selection as a multi-criteria classification 652 problem. In contrast to the decision trees which lead to single method recommendations through a 653 series of binary choices, the method selection tool recommends portfolios of methods, which are a 654 narrower set of options for further consideration 'off-line'.

655

#### 656 Scenarios for regulatory frameworks

657 In order to provide better insights into the political and social manoeuvring space (principle D) and to 658 elucidate the social and political compatibility of outcomes from decisions involving ES (principle G), 659 we used a participatory scenario approach. One important aspect for the usefulness of the ES 660 concept for policy making is its potential relevance as a cross-cutting issue that goes beyond 661 biodiversity and nature conservation and integrates other dimensions of environmental policy (such 662 as the WFD in Europe) and other societal sectors and the policy fields they regulate (such as 663 agriculture and the Common Agricultural Policy or regional development and infrastructure policy). 664 Even though the uptake of the ES concept in policy is currently limited (see above), scenario

<sup>&</sup>lt;sup>4</sup> <u>http://openness.hugin.com/oppla/ValuationSelection</u>

665 approaches are useful for capturing the interplay between different policies and other drivers of 666 change (e.g. social, technological and environmental) and how this effects future changes in ES. 667 Based on participatory scenario development at the European level and in the case studies, policy 668 scenarios were developed to analyse policy options and their impact on future ecosystem change 669 (García-Llorente et al. 2016, Priess et al. 2017). Policy analyses were combined with scenario 670 approaches and modelling to better understand policy options (principle D) and their impacts on 671 future changes in land use and ES (supporting principle G). Based on this analysis, we identify and 672 explore alternative policy options that may have triggered (or at least fostered) certain changes 673 (Hauck et al. 2017).

674

#### 675 Digital interaction via Oppla

676 Oppla (www.oppla.eu) is the web platform jointly developed by the OpenNESS and OPERAs projects, 677 to facilitate knowledge exchange on ecosystem services, natural capital and nature-based solutions 678 gathered from around Europe and beyond. It not only contains the main outcomes of both projects 679 such as case studies, online tools and methods and guidance to select them, but it also supports 680 communication and dissemination activities such as information on events, and organisation of 681 consultancy and training (via webinars and MOOCS). Very importantly, it also hosts a community that 682 will exchange and transfer the new knowledge acquired, developing individual capacities to address 683 challenges associated with ecosystem services and natural capital (see Pérez-Soba et al. 2017). The 684 content of Oppla initially came from OPERAs and OpenNESS, but other members of the community 685 have contributed as well.<sup>5</sup>

- 686 The following key features describe Oppla's characteristics to make knowledge exchange
- 687 operational. Oppla is open to a multidisciplinary and cross-sectoral Community of Practice, involving
- 688 academic/research organisations, policy makers, NGOs, private companies, etc.
- 689 In line with Principle A, from the very start Oppla was designed with the input of a range of
- 690 potential users, and its development has been adjusted based on feedback on functionality, user
- 691 friendliness and content, collected through several approaches including direct dialogue with
- 692 representatives from the European Commission, European institutions, intergovernmental
- 693 bodies, private sector and other target groups.

<sup>&</sup>lt;sup>5</sup> The Oppla ownership by the OpenNESS and OPERAs consortia was transferred in 2017 to a private company to ensure Oppla use and development beyond the lifetime of both projects.

694 Supporting Principle J, Oppla is developed to facilitate the connection of various data sources. 695 For example, over 40 case studies are currently described in the Case Study Finder, a tool that 696 gathers and spatially displays all the case studies from both projects and some others. It covers a 697 wide range of ecosystem services at multiple scales, areas and management schemes. These case 698 studies provide fresh insights into the needs of those applying the ES concept in the field, as well 699 as an empirical resource for testing ES instruments, tools and methods. Each case study refers to 700 a contact person for further information and to a location. The Oppla community can hence learn 701 from each other by this way of knowledge sharing.

702

# 703 **5. Conclusions and outlook**

704 The approach towards operationalising the ES concept as developed in the OpenNESS project, whose 705 process and outcomes are described in Furman et al. (2017 a and b) and in various articles in this 706 special issue, was not constructed as a single formal framework or scheme. Rather it consists of a 707 number of guiding principles and a set of integrative tools for operationalising the ES concept. We 708 see the guiding principles as crucial to consider when applying the ES concept to real world 709 situations. They were derived from an iterative interplay between experiences on the ground – from 710 a broad array of case studies – and theoretical work. This procedure turned out to be extremely 711 useful, as evidenced by the feedback given by stakeholders of the project (Dick et al. 2017a). The 712 guiding principles also include a number of important caveats in order to avoid an over-simplistic and 713 potentially counterproductive use of the concept, always a danger whenever a new concept starts to 714 become "fashionable". Based on the practical and theoretical work and the guiding principles, we 715 developed integrative tools, in order to promote an inter- and transdisciplinary approach to 716 operationalising the ES concept, making full use of the potential both of the concept (where it has 717 often not been played out hitherto; Abson et al. 2014) and of a large-scale research project with (in 718 our case) over 150 project participants. We did not develop tools for all of the guiding principles 719 described, partly because such tools were already available in the rich literature on ES and beyond, 720 and partly because the needs for additional tools surfaced only during the project and thus could not 721 be implemented during the restricted duration of a project. Also, there is still room for integrating 722 the different tools developed even further, e.g. more closely linking ONEX with the Toolbox for 723 Integrated Assessment and Valuation. In any case we hope that our approach, presented in this 724 paper and in more detail in the other papers in this Special Issue, will be of use to other projects on 725 operationalising the ES concept, but even more for users on the ground. Concerning the latter, let us 726 emphasise that, given the complexity of the ES concept, our approach (and the ES concept in 727 general) should not be used by completely inexperienced non-scientific users. Instead, in most cases

- it will require a "knowledge broker" (scientist or other experienced person) already familiar with the
- concept, to make the best use of it, for the benefit of all stakeholders and of nature.

730

#### 731 Acknowledgements

- 732 This work was supported by the OpenNESS project (Operationalisation of Natural Capital and
- 733 Ecosystem Services: From Concepts to Real-world Applications), funded from the European Union
- 734 Seventh Framework Programme (FP7-ENV.2012.6.2-1) under grant agreement n1 308428. We also
- thank Leon Braat, Ulrich Heink, Marion Potschin-Young, Heidi Wittmer and an anonymous reviewer
- for their critical comments, as well as Suvi Vikström for kindly supporting our work on this paper.

737

# 738 Literature cited

- Abson, D. J., Von Wehrden, H., Baumgärtner, S., Fischer, J., Hanspach, J., Härdtle, W., Heinrichs, H.,
- 740 Klein, A. M., Lang, D. J., Martens, P. and Walmsley, D., 2014. Ecosystem services as a boundary

741 object for sustainability. Ecological Economics, 103, 29-37. doi:

742 <u>http://dx.doi.org/10.1016/j.ecolecon.2014.04.012</u>

- Balint, P.J., Stewart, R.E. and Desai, A., 2011. Wicked environmental problems: managing uncertainty
  and conflict. Island Press. In: J.A. Bouma, and P.J.H. Van Beukering (Eds.), Ecosystem services:
  from concept to practice. Cambridge University Press, Cambridge.
- 746 Baró, F., Chaparro, L., Gomez-Baggethun, E., Langemeyer, J., Nowak, D. J. and Terradas, J. 2014.

Contribution of ecosystem services to air quality and climate change mitigation policies: The case
 of urban forests in Barcelona, Spain. Ambio, 43, 466–479. doi: 10.1007/s13280-014-0507-x

- 749 Barton, D. N. and P. A. Harrison, (Eds.) 2017. Integrated valuation of ecosystem services. Guidelines
- and experiences. European Commission FP7, 2017. EU FP7 OpenNESS Project Deliverable 3.3-4.4.
   www.openness-project.eu/library
- 752 Barton, D. et al. 2017 Ecosystem Services [this issue]
- 753 Bezák, P., Mederly, P., Izakovičová, Z., Špulerová, J., Schleyer, C., 2017. Divergence and conflicts in
- 754 landscape planning across spatial scales in Slovakia: An opportunity for an ecosystem services-
- based approach? International Journal of Biodiversity Science, Ecosystem Services & Management13, 119-135.
- Böck, K., Muhar, S., Muhar, A., Polt, R., 2015. The ecosystem services concept: Gaps between science
   and practice in river landscape management. GAIA Ecological Perspectives for Science and
   Society 24, 32-40.
- Borie, M., Hulme, M., 2015. Framing global biodiversity: IPBES between mother earth and ecosystem
   services. Environmental Science & Policy 54, 487–496.

- 762 Braat, L.C., E. Gómez-Baggethun, B. Martín-López, D.N. Barton, M. García-Llorente, E. Kelemen, and
- 763 H. Saarikoski. 2014. Framework for integration of valuation methods to assess ecosystem service
- 764 policies. European Commission EU FP7 OpenNESS Project Deliverable 4.2. <u>www.openness-</u>

765 project.eu/library

- Bardwell, L.V., 1991. Problem-Framing: A perspective on environmental problem-solving.
   Environmental Management 15, 603-612.
- Berbés-Blázquez, M., González, J.A., Pascual, U., 2016. Towards an ecosystem services approach that
   addresses social power relations. Current Opinion in Environmental Sustainability 19, 134-143.
- 770 Bouma, J.A., van Beukering, P.J.H. (Eds). 2015. Ecosystem Services: From Concept to Practice.
- 771 Cambridge University Press, Cambridge.
- Cash, D. et al. 2003: Knowledge systems for sustainable development. Proceedings of the National
   Academy of Sciences 100(14): 8086-8091.

Camps-Calvet, M., Langemeyer, J., Calvet-Mir, L., Gómez-Baggethun, E., 2015. Ecosystem services
 provided by urban gardens in Barcelona, Spain: Insights for policy and planning. Environmental
 Science and Policy 62, 14-23.

- 777 Carmen, E. et al. 2017 Ecosystem Services [this issue]
- 778 Carvalho, L. et al. (2017) OpenNESS policy brief on WFD and ES<sup>6</sup>
- 779 Chan, K. M. A., Guerry, A. D., Balvanera, P., Klain, S., Satterfield, T., Basurto, X., Bostrom, A.,
- 780 Chuenpagdee, R., Gould, R., Halpern, B. S., Hannahs, N., Levine, J., Norton, B., Ruckelshaus, M.,
- 781 Russell, R., Tam, J. and Woodside, U., 2012. Where are Cultural and Social in Ecosystem Services?
- 782 A Framework for Constructive Engagement. BioScience, 62(8), 744-756. doi:
- 783 10.1525/bio.2012.62.8.7
- Chan, K.M.A, Balvanera, P., Benessaiah, K., Chapman, M., Díaz, S., Gómez-Baggethun, E., Gould, R.K.,
  Hannahs, N., Jax, K., Klain, S.C., Luck, G., Martín-López, M., Muraca, B., Norton, B., Ott, K., Pascual,
- 786 U., Satterfield, T., Tadaki, M., Taggart, J., Turner, N.J., 2016. Why Protect Nature? Rethinking
- 787 Values and the Environment. Proceedings of the National Academy of Sciences, 113(6), 1462-
- 788 1465. doi: 10.1073/pnas.1525002113
- 789 Daily, G. C., Matson, P. A., 2008. Ecosystem services: From theory to implementation. Proceedings of
- 790 the National Academy of Sciences, 105(28), 9455-9456. doi: 10.1073/pnas.0804960105
- Davidson, M. D., 2013. On the relation between ecosystem services, intrinsic value, existence value
   and economic valuation. Ecological Economics, 95, 171-177. doi:
- 793 http://dx.doi.org/10.1016/j.ecolecon.2013.09.002

<sup>&</sup>lt;sup>6</sup> Note to the editor: Not published yet, but should be online on the OpenNESS website by end of July. Details will be supplemented then otherwise quote will be deleted.

- Daw, T., Brown, K., Rosendo, S., Pomeroy, R., 2011. Applying the ecosystem services concept to
- poverty alleviation: the need to disaggregate human well-being. Environ Conserv 38, 370-379.
- Dick, J. et al., 2017a: Stakeholders' perspectives on the operationalisation of the ecosystem service
   concept: results from 27 case studies. Ecosystem Services [this issue]
- 798 Dick, J, Verweij, P, Carmen, E, Rodela R, Andrews. C. 2017b. Testing the ecosystem service cascade
- 799 framework and QUICKScan software tool in the context of land use planning in Glenlivet Estate
- 800 Scotland. International Journal of Biodiversity Science, Ecosystem Services and Management (in
- 801 press). doi: <u>http://dx.doi.org/10.1080/21513732.2016.1268648</u>
- 802 Dunford, R. et al. (2017) Ecosystem Services [this issue]
- 803 European Commission, 2006. Halting the Loss of Biodiversity by 2010 and Beyond. Sustaining
- 804 Ecosystem Services for Human Well-being. Communication from the Commission. Document
- 805 COM(2006)216 final, issued 22 May 2006. <u>http://eur-</u>
- 806 <u>lex.europa.eu/LexUriServ/LexUriServ.do?uri=COM:2006:0216:FIN:EN:PDF</u>
- 807 European Commission, 2011. Our Life Insurance, Our Natural Capital: an EU Biodiversity Strategy to
- 808 2020. Communication from the Commission to the European Parliament, the Council, the
- 809 Economic and Social Committee and the Committee of the Regions. Document COM(2011) 244
- 810 final, issued May 3, 2011. <u>http://eur-lex.europa.eu/legal-</u>
- 811 <u>content/EN/TXT/?uri=celex%3A52011DC0244</u>
- 812 Förster, J., Barkmann, J., Fricke, R., Hotes, S., Kleyer, M., Kobbe, S., Kübler, D., Rumbaur, C.,
- 813 Siegmund-Schultze, M., Seppelt, R., Settele, J., Spangenberg, J.H., Tekken, V., Václavik, T.,
- 814 Wittmer, H., 2015. Assessing ecosystem services for informing land-use decisions: a problem-
- 815 oriented approach. Ecology and Society 20 (3). doi: 10.5751/ES-07804-200331
- 816 Funtowicz, S.O., & Ravetz, J.R., 1994. The worth of a songbird: ecological economics as a post-normal
- 817 science. Ecological Economics, 10(3), 197-207. doi: <u>http://dx.doi.org/10.1016/0921-</u>
- 818 <u>8009(94)90108-2</u>
- 819 Furman, E. et al. 2017a [Editorial] Ecosystem Services [this issue]
- 820 Furman, E., Barton, D.N., Harrison, P., Kopperoinen, L., Mederly, P., Pérez-Soba, M., Potschin, M.,
- 821 Saarikoski, H., Schleyer, C., Smith, A., Watt, A., Young, J., Kelemen, E., Gómez-Baggethun, G.,
- 822 2017b. Operationalizing ecosystem services: Advancing knowledge on natural and cultural capital.
- 823 In: Paracchini M.L., Zingari P.C., Blasi C. (Eds.), Re-connecting Natural and Cultural Capital -
- 824 Contributions from Science and Policy. Luxemburg: Office of Publications of the European Union,
- 825 Luxembourg (in press)
- 826 García-Llorente, M., Harrison, P.A., Berry, P., Palomo, I., Gómez-Baggethun, E., Iniesta-Arandia, I.,
- 827 Montes, C., García del Amo, D., Martín-López, B., 2016. What can conservation strategies learn
- 828 from the ecosystem services approach? Insights from ecosystem assessments in two Spanish
- 829 protected areas. Biodivers. Conserv. 1-23. http://dx.doi.org/10.1007/s10531-016-1152-4

- 830 Gómez-Baggethun, E., de Groot, R., 2010. Natural capital and ecosystem services: The ecological
- foundation of human society. In: R. E. Hester and R. M. Harrison (Ed.), Ecosystem services: Issues
- in Environmental Science and Technology 30, Royal Society of Chemistry, Cambridge, pp. 118-145.
- 833 Gómez-Baggethun, E., Barton, D.N., 2013. Classifying and valuing ecosystem services for urban
- 834 planning. Ecological Economics, 86, 235-245. doi:
- 835 <u>http://dx.doi.org/10.1016/j.ecolecon.2012.08.019</u>
- 836 Gómez-Baggethun, E., Kelemen, E., Martín, B., Palomo, I., Montes, C., 2013. Scale misfit in ecosystem
- 837 service governance as a source of environmental conflict. Society & Natural Resources, 26(10),
- 838 1202-1216. doi: 10.1080/08941920.2013.820817
- 839 Gómez-Baggethun, E., B. Martín-López, D.N. Barton, L. Braat, H. Saarikoski, E. Kelemen, M. García-
- 840 Llorente, J. van den Bergh E., P. Arias, P. Berry, M. Potschin, H. Keene, R. Dunford, C. Schröter-
- 841 Schlaack, and P. Harrison. 2014. "EU FP7 OpenNESS Project Deliverable 4.1. . State-of-the-art
- report on integrated valuation of ecosystem services. European Commission FP7, 2014."
- 843 <u>www.openness-project.eu/library</u>
- 844 Görg, C., J. H. Spangenberg, V. Tekken, B. Burkhard, D. Thanh Truong, M. Escalada, K. Luen Heong, G.
- 845 Arida, L. V. Marquez, J. Victor Bustamante, H. Van Chien, T. Klotzbücher, A. Marxen, N. Hung
- 846 Manh, N. Van Sinh, S. Villareal & J. Settele, 2014. Engaging Local Knowledge in Biodiversity
- Research: Experiences from Large Inter- and Transdisciplinary Projects. Interdisciplinary Science
   Reviews, 39(4), 323-341. doi: 10.1179/0308018814Z.0000000095
- 849 Grêt-Regamey A, Sirén E, Brunner SH, Weibel B. 2017. Review of decision support tools to
- 850 operationalize the ecosystem services concept. Ecosystem services, (in press). Doi:
- 851 http://dx.doi.org/10.1016/j.ecoser.2016.10.012
- Haines-Young, R., Potschin, M.; Jax, K.; Görg, C.; Heink, U.; Kelemen, K., Schleyer, C. 2017. OpenNESS
  Conceptual Nexus (ONEX). Guidelines for testing the conceptual frameworks in case study areas
  using methods and data resources developed in WPs 2, 3 and 4. EU FP7 OpenNESS Project
  Deliverable 1.4. 83 pp. www.openness-project.eu/library.
- Hajer, M.A., 2006. Doing discourse analysis: coalitions, practices, meaning, in: van den Brink, M.,
- 857 Metze, T. (Eds.), Words matter in policy and planning. Disocurse theory and method in the social
- sciences. Netherlands Graduate Schhol of Urban and Regional Research, Utrecht, pp. 65-74.
- 859 Harrison PA, Dunford R, Barton DN, Kelemen E, Martin-Lopez B, Norton L, Termansen M, Saarikoski
- 860 H, Hendriks K, Gomez-Baggethun E, Czúcz B, Garcia-Llorente M, Howard D, Jacobs S, Karlsen M,
- 861 Kopperoinen L, Madsen A, Rusch G, van Eupen M, Verweij P, Smith R, Tuomasjukka D, Zulian G.
- 862 (2017, in review). Selecting methods for ecosystem service assessment: A decision tree approach.
- 863 Ecosystem Services.

- 864 Hauck, J., Görg, C., Varjopuro, R., Ratamäki, O., Jax, K., 2013. Benefits and limitations of the 865 ecosystem services concept in environmental policy and decision making: Some stakeholder 866
- perspectives. Environmental Science & Policy 25, 13-21.
- 867 Hauck, J., Görg, C., Werner, A., Jax, K., Bidoglio, G., Maes, J., Furman, E., Ratamäki O., 2014.
- 868 Supporting the development and implementation of the European Biodiversity Strategy to 2020: 869 Insights from a European project. Interdisciplinary Science Reviews, 39(4), 376–391. Doi:
- 870 http://dx.doi.org/10.1179/0308018814Z.0000000098
- 871 Hauck, J., Schleyer, C. Priess, J., Haines-Young, R., Harrison, P., Dunford, R., Kok, M. Young, J., Berry, 872 P., Primmer, E., Veerkamp, C., Bela, G., Vadineanu, A., Dick, J., Alkemade R., Görg, C., 2017. Policy 873 Scenarios of future change, OpenNESS Deliverable 2.5. www.openness-project.eu/library
- 874 Heink, U., Heubach, K., Jax, K., Kugel, C., Marquard, E., Neßhöver, C., Neumann, R., Paulsch, A., Tilch,
- 875
- S., Timaeus, J., Vandewalle, M., 2015. Conceptualizing credibility, relevance and legitimacy for the 876 evaluation of science-policy interfaces – challenges and opportunities. Science and Public Policy 877 42,676-689.
- 878 Hoppe, R., 2011. The governance of problems. Puzzling, powering and participation. Bristol: Policy 879 Press.
- 880 Jacobs, S., Dendoncker, N., Martín-López, B., Barton, D.N., Gomez-Baggethun, E., Boeraeve, F.,
- 881 McGrath, F.L., Vierikko, K., Geneletti, D., Sevecke, Katharina J., Pipart, N., Primmer, E., Mederly,
- 882 P., Schmidt, S., Aragão, A., Baral, H., Bark, Rosalind H., Briceno, T., Brogna, D., Cabral, P., De
- 883 Vreese, R., Liquete, C., Mueller, H., Peh, K.S.H., Phelan, A., Rincón, Alexander R., Rogers, S.H.,
- 884 Turkelboom, F., Van Reeth, W., van Zanten, B.T., Wam, H.K., Washbourne, C.-L., 2016. A new
- 885 valuation school: Integrating diverse values of nature in resource and land use decisions.
- 886 Ecosystem Services 22, Part B, 213-220.
- 887 Jacobs, S. et al. (2017) Ecosystem Services [this issue]
- 888 Jann, W., Wegrich, K., et al., 2007. Theories of the policy cycle. In: Fischer, F., Miller, G.J., Sidney, M.S.
- 889 (Eds.), Handbook of Public Policy Analysis. CRC Press, pp. 44-62.
- 890 Jax, K., Barton, D. N., Chan, K. M. A., De Groot, R., Doyle, U., Eser, U., Görg, C., Gomez-Baggethun, E., 891 Griewald, Y., Haber, W., Haines-Young, R., Heink, U., Jahn, T., Joosten, H., Kerschbaumer, L., Korn,
- 892 H., Luck, G. W., Matzdorf, B., Muraca, B., Neßhöver, C., Norton, B., Ott, K., Potschin, M.,
- 893 Rauschmayer, F., Von Haaren, C. and Wichmann, S., 2013. Ecosystem services and ethics.
- 894 Ecological Economics, 93, 260-268. doi: <u>http://dx.doi.org/10.1016/j.ecolecon.2013.06.008</u>
- 895 Jax, K. and Heink, U., 2015. Searching for the place of biodiversity in the ecosystem services

896 discourse. Biological Conservation, 191, 198-205. doi:

897 http://dx.doi.org/10.1016/j.biocon.2015.06.032

- Jax, K. 2016. Ecosystem functions: a critical perspective. In: Potschin, M., Haines-Young, R., Fish, R.
  and Turner, R. K. (Eds.): Routledge Handbook of Ecosystem Services. London: Routledge, pp. 2830.
- 901 Kelemen, E., Lazányi, O., Arany, I., Aszalós, R., Bela, Gy., Czúcz, B., Kalóczkai, Á., Kertész, M., Megyesi,
- 902 B., Pataki, G. 2015a. Sociocultural valuation of ecosystem services provided by the Kiskunság sand
- 903 ridge region (Ökoszisztéma szolgáltatásokról a kiskunsági Homokhátság társadalmának
- 904 szemszögéből), Természetvédelmi Közlemények, 21, 116-129. (in Hungarian) online available at:
   905 <u>http://www.mbt-biologia.hu/gen/pro/mod/let/let\_fajl\_megnyitas.php?i\_faj\_azo=1134</u>
- Kelemen, E., D. N. Barton, S. Jacobs, B. Martín-López, H. Saarikoski, M. Termasen, G. Bela, L. Braat, R.
  Demeyer, M. García-Llorente, E. Gómez-Baggethun, J. Hauck, H. Keune, S. Luque, I. Palomo, G.
- 908 Pataki, P. Tenerilli, and F. Turkelboom. 2015b. 'Preliminary guidelines for integrated assessment
- 909 and valuation of ecosystem services in specific policy contexts. EU FP7 OpenNESS Project
- 910 Deliverable 4.3., European Commission FP7'. <u>www.openness-project.eu/library</u>
- 911 Koschke, L., Van der Meulen, S., Frank, S., Schneidergruber, A., Kruse, M., Fürst, C., Neubert, E.,
- 912 Ohnesorge, B., Schröder, C., Müller, F., Bastian, O., 2014. Do you have 5 minutes to spare? -The
- 913 challenges of stakeholder processes in ecosystem services studies. Landscape Online 37, 1-25.
- Kovács, E., Kelemen, E., Kalóczkai, Á., Margóczi, K., Pataki, G., Gébert, J., Málovics, G., Balázs, B.,
  Roboz, Á., Krasznai Kovács, E., Mihók, B., 2015. Understanding the links between ecosystem
  service trade-offs and conflicts in protected areas. Ecosystem Services 12, 117-127.
- Lamarque, P., Quétier, F., Lavorel, S., 2011. The diversity of the ecosystem services concept and its
   implications for their assessment and management. Comptes Rendus Biologies 334, 441-449.
- Langemeyer, J., Gómez-Baggethun, E., Haase, D., Scheuer, S., Elmqvist, T., 2016. Bridging the gap
   between ecosystem services and land-use policy and planning: An exploration of multi-criteria
   decision analysis. Environmental Science & Policy (62), 45–56.
- Larson BMH., 2011. Metaphors for environmental sustainability: Redefining our relationship withnature. Yale University Press.
- Uvens, A.; Turkelboom, F.; Demeyer, R.; Garcia-Llorente, M.; Hauck, J.; Kelemen, E.; Teng, C.;
- 925 Tersteeg, J.; Lazányi, O.; Martin-Lopez, B.; Pataki, G. & Schiffer, E. (2015). Openness manual:
- 926 stakeholder analysis for environmental decision-making at local level. Publication developed in
- 927 the framework of OpenNESS (FP7 Project). INBO, Brussels: Belgium.
- 928 Luck, G. W., Chan, K. M. A., Eser, U., Gómez-Baggethun, E., Matzdorf, B., Norton, B. and Potschin, M.,
- 929 2012. Ethical considerations in on-ground applications of the ecosystem services concept.
- 930 BioScience, 62(12), 1020-1029. doi: 10.1525/bio.2012.62.12.4
- 931 Martín-Lopez et al. 2017 (this issue)
- 932 Millennium Ecosystem Assessment, 2005. Ecosystems and Human Well-being: Current State and
- 933 Trends, Volume 1. Island Press, Washington D.C.

- 934 Mouchet, M. A., Lamarque, P., Martín-López, B., Crouzat, E., Gos, P., Byczek, C. and Lavorel, S. 2014.
- 935 An interdisciplinary methodological guide for quantifying associations between ecosystem
- 936 services. Global Environmental Change, 28, 298- 308. doi:
- 937 http://dx.doi.org/10.1016/j.gloenvcha.2014.07.012
- 938 Muradian, R., Arsel, M., Pellegrini, L., Adaman, F., Aguilar, B., Agarwal, B., Corbera, E., Ezzine de Blas,
- 939 D., Farley, J., Froger, G., Garcia-Frapolli, E., Gómez-Baggethun, E., Gowdy, J., Kosoy, N., Le Coq,
- 940 J.F., Leroy, P., May, P., Méral, P., Mibielli, P., Norgaard, R., Ozkaynak, B., Pascual, U., Pengue, W.,
- 941 Perez, M., Pesche, D., Pirard, R., Ramos-Martin, J., Rival, L., Saenz, F., Van Hecken, G., Vatn, A.,
- Vira, B., Urama, K., 2013. Payments for ecosystem services and the fatal attraction of win-win
  solutions. Conservation Letters 6, 274-279.
- Norton, B. G. and Noonan, D., 2007. Ecology and valuation: Big changes needed. Ecological
- 945 Economics, 63(4), 664-675. doi: <u>http://dx.doi.org/10.1016/j.ecolecon.2007.02.013</u>
- Norton, B., 2015. Sustainable values, sustainable change. A guide to environmental decision making.
   The University of Chicago Press, Chicago.
- Opdam, P., Albert, C., Fürst, C., Grêt-Regamey, A., Kleemann, J., Parker, D., La Rosa, D., Schmidt, K.,
  Villamor, G.B., Walz, A., 2015. Ecosystem services for connecting actors lessons from a
  symposium. Change and Adaptation in Socio-Ecological Systems, 2: 1-7
- 951 Paloniemi, R., Apostolopoulou, E., Cent, J., Bormpoudakis, D., Scott, A., Grodzińska-Jurczak, M.,
- 952 Tzanopoulos, J., Koivulehto, M., Pietrzyk-Kaszyńska, A. and Pantis, J. D. 2015. Public participation
- 953 and environmental justice in biodiversity governance in Finland, Greece, Poland and the UK.
- 954 Environmental Policy and Governance, 25, 330-342. doi: 10.1002/eet.1672
- 955 Pascual, U., Phelps, J., Garmendia, E., Brown, K., Corbera, E., Martin, A., Gómez-Baggethun, E.,
- 956 Muradian, R., 2014. Social equity matters in Payments for Ecosystem Services. BioScience 64,
  957 1027-1036. doi: 10.1093/biosci/biu146
- 958 Perez-Soba, M. et al. (2017) Ecosystem Services [this issue]
- Phelps, J., Jones, C.A., Pendergrass, J.A., Gómez-Baggethun, E., 2015. Environmental liability: A
  missing use for ecosystem services valuation. Proceedings of the National Academy of Sciences,
  112(39), E5379. doi: 10.1073/pnas.1514893112
- 962 Pielke Jr., R., 2007. The Honest Broker. Making sense of science in policy and politics. Cambridge963 University Press, Cambridge.
- Potschin, M. and R. Haines-Young, 2011. Ecosystem Services: Exploring a geographical perspective.
   Progress in Physical Geography, 35(5), 575-594.
- Potschin, M.; R. Haines-Young, R. Fish and K. Turner (Ed.), 2016. Routledge Handbook of Ecosystem
   Services. Routledge, London and New York, 630 pp

- 968 Potschin, M. et al. (2017): Understanding the role of Conceptual Frameworks: Reading the
- 969 Ecosystem Service Cascade. Ecosystem Services [this issue]
- 970 Priess, J. et al. (2017).... Ecosystem Services [this issue]
- 971 Primmer, E. & Furman, E., 2012. Operationalising ecosystem service approaches for governance: Do
- 972 measuring, mapping and valuing integrate sector-specific knowledge systems? Ecosystem
  973 Services, 1(1), 85-92. doi: http://dx.doi.org/10.1016/j.ecoser.2012.07.008
- 974 Primmer, E., Jokinen, P., Blicharska, M., Barton, D.N., Bugter, R., & Potschin, M., 2015. Governance of
- 975 Ecosystem Services: A framework for empirical analysis. Ecosystem Services, 16, 158-166. doi:
   976 <u>http://dx.doi.org/10.1016/j.ecoser.2015.05.002</u>
- Raymond CM, Singh GG, Benessaiah K, Bernhardt JR, Levine J, Nelson H, Turner NJ, Norton B, Tam J,
  Chan KMA., 2013. Ecosystem services and beyond: Using multiple metaphors to understand

979 human-environment relationships. BioScience, 63(7), 536-546. doi: 10.1525/bio.2013.63.7.7

- 980 Reed, M.S., Stringer, L.C., Fazey, I., Evely, A.C. and Kruijsen, J.H.J., 2014. Five principles for the
- 981 practice of knowledge exchange in environmental management. Journal of Environmental982 Management 146: 337-345.
- Reyes-García, V., G. Menéndez-Baceta, L. Aceituno-Mata, L. Calvet-Mir, T. Garnatje, E. GómezBaggethun, R. Acosta, R. Rodríguez Franco, P. Domínguez, J. Vallès, M. Pardo-de-Santayana M.,
  2015. 'From famine foods to delicatessen: Interpreting trends in the use of wild edible plants
  through cultural ecosystem services'. Ecological Economics 120: 303–311.
- 987 Rittel, H.W.J.; Webber, M.M., 1973. Dilemmas in a General Theory of Planning. Policy Sciences, 4,
  988 155–169.
- Rode, J., Gómez-Baggethun, E., Krause, T., 2015. Motivation crowding by economic incentives in
   conservation policy: A review of the empirical evidence. Ecological Economics 117, 270-282.
- Rodela, R., Bregt, A.K., Perez-Soba, M., Verweij, P., 2017. When having a stake matters: pilot-testing
  spatial decision support systems with students vs. real stakeholders. Environmental Modelling &
  Software, (submitted December 2016; to be deleted if not accepted).
- 994 Rozzi, R., 2015. Earth Stewardship and the biocultural ethic: Latin American perspectives, in: Rozzi,
- 995 R., Chapin, F.S.I., Callicott, J.B., Pickett, S.T.A., Power, M.E., Armesto, J.J., May, R.H. (Eds.), Earth
- 996 stewardship. Linking ecology and ethics in theory and practice. Springer, Heidelberg, pp. 87-112.
- Ruckelshaus M, McKenzie E, Tallis H, Guerry A, Daily G, Kareiva P, et al., 2015. Notes from the field:
   Lessons learned from using ecosystem service approaches to inform real-world decisions.
- 999 Ecological Economics, 115, 11-21. doi: <u>http://dx.doi.org/10.1016/j.ecolecon.2013.07.009</u>
- 1000 Saarikoski H., Jax K., Harrison P.A., Mononen L., Primmer E., Barton D.N., Vihervaara P., Furman E.,
- 1001 2015. Exploring operational ecosystem service definitions: the case of boreal forests. *Ecosystem*
- 1002 Services, 14, 144-157. doi: <u>http://dx.doi.org/10.1016/j.ecoser.2015.03.006</u>

- 1003 Saarikoski, H., Mustajoki, J., Barton, D., Geneletti, D., Langemeyer, J., Gomez-Baggethun, E.,
- Marttunen, M., Antunes, P., Keune, H. & Santos, R. 2016. Multi-Criteria Decision Analysis and
   Cost-Benefit Analysis: Comparing alternative frameworks for integrated valuation of ecosystem
- 1006 services. Ecosystem Services 22: 238–249.
- 1007 Saarikoski, H. et al., 2017...Ecosystem Services [this issue]
- Salomaa A., & Paloniemi R., Hujala T., Rantala S., Arponen A, Niemelä J., 2016. The use of knowledge
   in evidence-informed voluntary conservation of Finnish forests. Forest Policy and Economics, 73,
   90-98
- 1011 Schleyer, C., Görg, C., Hauck, J.; Winkler, K.J., 2015. Opportunities and challenges for mainstreaming
- 1012 the ecosystem services concept in the multi-level policy-making within the EU. Ecosystem
- 1013 Services, Ecosystem Services, 16, 174-81, <u>http://dx.doi.org/10.1016/j.ecoser.2015.10.014i</u>
- 1014 Schleyer, C., Lux, A., Mehring, M., Görg C., 2017. Ecosystem Services as a boundary concept: Social-
- 1015 ecological perspectives on a contested concept and the need for inter- and transdisciplinary
- 1016 collaboration, accepted for Sustainability, Special Issue on Social Ecology
- 1017 Schnegg, M., R. Rieprich, and M. Pröpper. 2014. Culture, Nature, and the Valuation of Ecosystem
- 1018 Services in Northern Namibia. *Ecology and Society* **19**(4): 26.
- 1019 http://dx.doi.org/10.5751/ES-06896-190426
- Sikor, T., Fisher, J., Few, R., Martin, A., Zeitou, M., 2013. The justices and injustices of ecosystem
   services, in: Sikor, T. (Ed.), The justices and injustices of ecosystem services. Routledge, London,
   pp. 187-200.
- 1023 Smith, R. et al. (2017) .... Ecosystem Services [this issue]
- 1024 Spangenberg, J.H., Christoph. G., Truong, D.T., Tekken, V., Bustamante, J.V., Settele, J., 2014.
- 1025 Provision of ecosystem services is determined by human agency, not ecosystem functions. Four
- 1026 case studies. International Journal of Biodiversity Science, Ecosystem Services & Management,
- 1027 10(1), 40-53. doi: 10.1080/21513732.2014.884166
- 1028 Spangenberg, J.H., Görg, C., Settele, J., 2015. Stakeholder involvement in ESS research and
- 1029 governance: between conceptual ambition and practical experiences risks, challenges and
- 1030 tested tools. Ecosystem Services, 16, 201-211. doi:
- 1031 http://dx.doi.org/10.1016/j.ecoser.2015.10.006
- 1032 Stone, R., 2006. After the Tsunami: A Scientist's Dilemma. Science, 313 (5783), 32-35. doi:
- 1033 10.1126/science.313.5783.32
- TEEB, 2010. The economics of ecosystems and biodiversity: ecological and economic foundations.
   Earthscan, London and Washington D.C., p. 410.
- 1036 Turkelboom, F. et al. (2017) Ecosystem Services [this issue]

- Turnhout, E., Waterton, C., Neves, K., Buizer, M., 2013. Rethinking biodiversity: From goods and
   services to "living with". Conservation Letters 6, 154-161.
- 1039 Ulenaers P., Turkelboom F., Simoens I., Keune H., Deneef H., Stevens K., 2014. Participatieve
- 1040 gebiedsvisieontwikkeling voor De Wijers via het ecosysteemdienstenconcept Procesarchitectuur
- 1041 en procesevaluatie. Rapporten van het Instituut voor Natuur- en Bosonderzoek 2014
- 1042 (INBO.R.2014.2853501). Instituut voor Natuur- en Bosonderzoek, Brussels.
- 1043 UNEP, 2014. Conceptual framework for the Intergovernmental Science-Policy Platform on
- 1044 Biodiversity and Ecosystem Services. Decison IPBES-2/4., Report of the second session of the
- 1045 Plenary of the Intergovernmental Science-Policy Platform on Biodiversity and Ecosystem Services.
- 1046 Van Der Steen, W. J., 1990. Concepts in biology: a survey of practical methodological principles.1047 Journal of theoretical Biology, 143, 383-403.
- 1048 Vlachopoulou, M., Coughlin, D., Forrow, D., Kirk, S., Logan, P., Voulvoulis, N., 2014. The potential of
   1049 using the Ecosystem Approach in the implementation of the EU Water Framework Directive.
   1050 Science of the Total Environment 470-471, 684-694.
- 1051 von Haaren, C., Albert, C., 2011. Integrating ecosystem services and environmental planning:
- 1052 Limitations and synergies. International Journal of Biodiversity Science, Ecosystems Services and1053 Management 7, 150-167.
- 1054 Wijnja, H., van Uden, G., Delbaere, B. (eds.), 2016. Ecosystem services in operation: case studies.
- 1055 European Commission FP7. Available at:
- 1056 https://issuu.com/ecnc.org/docs/openness\_casestudies\_brochure
- 1057 Wittmer, H. Gundimedia, H., 2012. The economics of ecosystems and biodiversity in local and1058 regional policy and management. Earthscan, Oxon. 351 p.
- 1059 WWF (World Wide Fund for Nature) 2014. The Growth of Soy: Impacts and Solutions. WWF
- 1060 International, Gland, Switzerland. 94 p.
- 1061 Zulian, G. et al. (2017) Ecosystem Services [this issue]

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1063 Table 1: Ten Principles to consider when putting ecosystem services concepts into practice.

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- A) As real-world problems involve and affect people, it will in most cases be necessary to involve stakeholders from the very beginning, when the problem is defined and the entire problem space is laid out.
- *B)* The role of scientists in approaching the problem should be clarified.
- C) The complexity and often "wickedness" of the problem should be acknowledged and the nature of the problem, including its social, ecological, administrative and economic spheres should be charted. Simplistic understandings of problems should be avoided to ensure that the problem and how an ES approach might contribute to its solution is clearly expressed. Hidden links between the different spheres need to be exposed.
- D) Political space and influence spheres should be defined. The manoeuvring space (what is possible within the boundaries of e.g. a legal or societal situation) should be defined.
- E) Concepts and language should be adapted to the specific situations and stakeholders.
- *F)* When applying the ES concept in a specific situation, hidden and neglected issues, hidden values, and hidden links between issues should be revealed and made transparent. Hidden and suppressed issues and values may be the most important in terms of conflicts and conflict resolution between stakeholders.
- *G)* To avoid socially unacceptable results or decisions arising from application of the ES concept the social and political compatibility of outcomes should be assessed and potential winners and losers identified during the implementation process.
- H) ES tools and methods may not always be the only or even the best tool. To determine what kind of assessment is needed for decision-making, deliberative tools for scoping the problem space and the most appropriate tools for problem solving can be useful.
- It is important to connect various data pools with each other. This requires including all relevant sectors (such as research disciplines, policy fields). However, not all ES need to be assessed, which leaves space to focus on the most relevant, assuming all relevant stakeholders are consulted.
- J) Different methods, or combinations of methods, can be useful for answering different questions, and therefore it is important to identify as precisely as possible what is wanted (in iterations with knowledge users) before starting to search for appropriate methods.

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#### 1067 Fig. 1 The ESAST scheme. See text for explanation.

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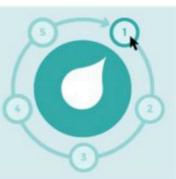
#### Ecosystem service assessment is a process that is carried out in a close interaction with key stakeholders to ensure that they find the results reliable and relevant for decision-making.

Putting in practice

Policy instruments and best practice examples e.g. on nature based solutions in urban areas

#### Valuation

What is the importance of ES for people in monetary or non-monetary terms, and what kind of meanings do people assign to ES, including intrinsic and relational values?



#### **Biophysical assessment**

What are the links between the ES and underlying ecosystem structures and processes and how do drivers of change influence the flow of ES?

Setting the scene

What is the purpose of the assessment, the decision-making context and who are the key actors to be involved in the assessment?

Identification of ES

What are the key ES and related benefits, and beneficiaries, in the issue at hand?

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