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1 From spectra to plant functional traits: Transferable multi-trait models 2 from heterogeneous and sparse data

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20

21 Abstract

22 Large-scale information on several vegetation properties ('plant traits') is critical to assess ecosystem
23 functioning, functional diversity and their role in the Earth system. Hyperspectral remote sensing of plant
24 canopies offers a key tool to map multiple plant traits. However, we are still lacking generalized methods
25 to translate hyperspectral reflectance into a suite of relevant plant traits across biomes, land cover and
26 sensor types. The absence of globally representative data sets and the gap between the available
27 reflectance data with corresponding in-situ measurements have hampered such approaches. In recent
28 years, the scientific community acquired multiple data sets encompassing canopy hyperspectral reflectance
29 and plant traits from different plant types and sensors. To combine these heterogeneous data sets, we
30 propose three multi-trait modeling approaches based on Convolutional Neural Networks (CNNs) to
31 simultaneously infer a broad set of 20 structural and chemical traits (e.g. leaf mass per area, leaf area
32 index, pigments, nitrogen). The performance of these multi-trait CNN models predicting these traits was

33 compared against single-trait CNN as well as single-trait partial least squares regression (PLSR). We found
34 that the multi-trait CNNs performances significantly increased from single-trait CNNs (nRMSE = 0.027-
35 19.61%) and the state-of-the-art PLSR models (nRMSE = 1.94 - 40.07%) across a broad range of
36 vegetation types (crops, forest, tundra, grassland, shrubland) and sensor types. Thus, providing a single
37 model for multiple traits not only proved to be computationally more efficient, but also more accurate, since
38 it enabled the model to incorporate traits' co-variation. Despite the data heterogeneity of the merged data
39 set, our models performances' were comparable or exceeded those of previous studies. Overall, this study
40 highlights the potential of weakly supervised approaches to overcome the scarcity of in-situ measurements
41 and take a step forward in creating efficient predictive models of multiple biochemical and biophysical
42 vegetation properties.

43

44 **Keywords:** Hyperspectral remote sensing, Plant trait retrieval, Deep learning, Biophysical variables,
45 Imaging spectroscopy, Canopy properties, Weakly supervised learning, Multi-task regression

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60 **1. Introduction**

61 Plant functional traits are key for assessing and monitoring terrestrial ecosystem properties. They provide
62 insights on functional diversity and can enhance our understanding of ecosystem functioning (Lavorel and
63 Garnier, 2002; Migliavacca et al., 2021). Traits determine plant productivity and stress resistance and thus
64 also how plants compete for growth and survival in different environments (Funk et al., 2017). For example,
65 leaf mass per area (LMA) is positively related to photosynthetic productivity and negatively to structural
66 robustness and depends on resource availability and environmental conditions (Díaz et al., 2016; Grime,
67 1988; Poorter et al., 2009). Leaf pigments (e.g., chlorophyll, carotenoids) determine photosynthetic
68 capacities and their variations can indicate changes in plant health due to stress (Feret et al., 2008; Zarco-
69 Tejada et al., 2019, 2018, Berger et al., 2022). Other leaf constituents such as nitrogen and carbon are
70 directly linked to biosphere-atmosphere cycles (de Bello et al., 2010) and are important to parameterize
71 vegetation in Earth system models (Yang et al., 2015). A comprehensive set of quantitative trait
72 measurements is thus desirable to understand the functioning of ecosystems.

73

74 Still, despite the efforts towards compiling field observations from a myriad of studies into global databases
75 (e.g. TRY, Kattge et al., 2020), the available data are sparse in terms of geographical coverage, species
76 and range of traits (Asner et al., 2015; Kattge et al., 2020). In this context, hyperspectral remote sensing
77 data offer an efficient proxy to map plant traits (Cavender-Bares et al., 2020; Jetz et al., 2016). Such data
78 enable repeatable and non-destructive optical observations using numerous platforms and sensors
79 providing information on spectral reflectance across a wide range of the electromagnetic spectrum via
80 continuous narrow bands. Given the mechanistic interactions of light with leaf and canopy traits (Billings
81 and Morris, 1951; Gates et al., 1965; Kattenborn and Schmidtlein, 2019; Ustin and Gamon, 2010),
82 hyperspectral observations have a high potential to reveal plant traits over remote and large areas (Hank
83 et al., 2019, Asner and Martin, 2016; Homolová et al., 2013; Singh et al., 2015; van Cleemput et al., 2018,
84 Danner et al., 2021; Woche et al., 2022). Recently launched and forthcoming hyperspectral space missions
85 such as PRecursores IperSpettrale della Missione Applicativa (PRISMA, Cogliati et al., 2021),
86 Environmental Mapping and Analysis Program (EnMAP, Guanter et al., 2015) and Surface Biology and
87 Geology (SBG, Cawse-Nicholson et al., 2021) along with the higher-resolution proximal and airborne
88 instruments, support this potential and will provide an unprecedented source of data. However, in view of

89 the varieties of these hyperspectral data sources and potential applications, we are missing transferable
90 retrieval methods across sensors, acquisition settings, ecosystems and plant functional types.

91
92 From a methodological perspective, available retrieval methods range from data-driven statistical methods
93 to the inversion of radiative transfer models (RTM) to hybrid methods (see Verrelst et al., 2019 for a review).
94 RTMs simulate the interaction of light with vegetation properties and thus their inversion can represent a
95 promising approach for plant trait retrieval (Berger et al., 2018; Dorigo et al., 2007; Feilhauer et al., 2018,
96 2017; Jacquemoud et al., 2009). Yet, plant trait retrieval by RTM inversion is only possible for traits that are
97 considered in the RTM itself. Moreover, RTM inversions are very sensitive to the RTM's configuration and
98 thus have to be specifically parameterized for different vegetation types, canopy structures, phenological
99 stages or use cases (Dorigo et al., 2007; Atzberger and Richter, 2012; Verrelst et al., 2013). Conversely,
100 data-driven approaches automatically learn the statistical relation between the spectral data and plant traits.
101 Partial Least Squares regression (PLSR) (Geladi and Kowalski, 1986; Wold et al., 2001) can be considered
102 as the benchmark approach given its long history in imaging spectroscopy (Asner and Martin, 2008;
103 Feilhauer et al., 2010; Singh et al., 2015; Wang et al., 2020). In recent years, new machine learning
104 algorithms emerged as powerful approaches to solve retrieval tasks from hyperspectral data (Wang et al.,
105 2020, Prilianti et al., 2021, Shi et al., 2022).

106
107 Despite the potentials of data-driven methods, there are multiple constraints:
108 1) Commonly, data-driven models are trained with data sets representing limited variation in ecosystem
109 properties, plant functional types, sensor systems and acquisition settings, thus limiting their transferability.
110 For instance, previous studies (Asner et al., 2015; Berger et al., 2020; Wang et al., 2019) have concentrated
111 on individual ecosystems such as croplands, forests, or grasslands using specific data sets. However,
112 models developed from these data sets may produce significant uncertainties when employed on a new
113 data set, making them less transferable to other ecosystems or alternative data sets (Wang et al. 2020).
114 2) Data-driven models are often built independently for different traits. This prevents exploiting
115 interrelationships between certain traits. For example, different traits may be driven by the same processes
116 or may manifest in overlapping absorption features such as pigments or resource-investment related traits.
117 Consequently, taking the trait interrelations into account might improve the overall retrieval quality.

118 Moreover, the simultaneous prediction of multiple traits may also enlarge computational efficiency. It is thus
119 compelling to aim for a data-driven approach that is capable of predicting a set of traits simultaneously. We
120 further refer to such an approach as 'multi-trait' model.

121 3) Furthermore, most data-driven approaches for plant trait retrieval cannot easily be extended with new
122 training data, which hinders continuous model improvements and knowledge transferability.

123 Deep learning and particularly Convolutional Neural Networks (CNNs) may pave new avenues to alleviate
124 such issues (Sosnin et al., 2019; Yosinski et al., 2014; Zhang and Yang, 2021). CNNs are a powerful
125 method for automatic feature engineering and are increasingly being applied to remote sensing data
126 (Kattenborn et al., 2021; Zhu et al., 2017). Due to their depth and large number of neurons such models
127 are capable of learning complex relationships. Accordingly, given sufficient representativeness of the input
128 data, such models may learn transferable relationships across application domains, sensor types and
129 acquisition settings. Moreover, CNNs are commonly trained iteratively, enabling to exploit very large data
130 sets and allowing for continuous updating and fine-tuning of models with new, unseen data (Shin et al.,
131 2016).

132 The availability of canopy spectra and their corresponding trait observations from different studies
133 encompassing different plant types and sensors constantly increases (e.g. Rogers et al., 2021; van
134 Cleemput et al., 2019). This opens a way to harness the scalability of deep learning and test the robustness
135 of the models when integrating multi-source hyperspectral and plant trait data (e.g. EcoSIS, Wagner et al.,
136 2018). However, due to the different context of these studies, a combination of such data sets is naturally
137 sparse, meaning not all potential traits are covered across different data sets. Therefore, the objective of
138 this study is to explore the potential of weakly supervised approaches to train models on sparse data for
139 simultaneously predicting multiple traits ($n = 20$) from canopy hyperspectral data. This analysis is based on
140 a combination of 42 data sets from heterogeneous data of different vegetation and sensor types. We
141 implement three weakly-supervised multi-trait CNN approaches to investigate the hypothesis of whether
142 the incorporation of trait-trait correlation in models' calibration would improve the trait estimations. The
143 performance of these strategies is compared to common single-trait PLSR and single-trait CNN models.

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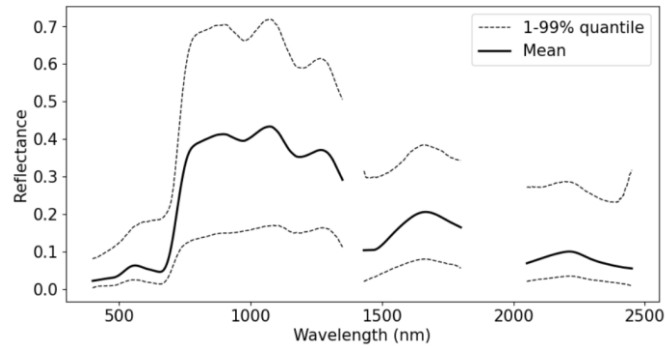
147 **2. Material and Methods**

148 *2.1 Data merging and cleaning*

149 We employed 42 data sets of full range canopy spectra (400-2500 nm) with corresponding trait
150 measurements (e.g. LMA, pigments) that were available upon request or from archives (e.g. EcoSIS,
151 PANGEA) (Burnett et al., 2021; Cerasoli et al., 2018; Chlus et al., 2020; Ewald et al., 2018b, 2020; Hank
152 et al., 2016, 2015; Herrmann et al., 2011; Kattenborn et al., 2019a; Pottier et al., 2014; Rogers et al., 2021;
153 Singh et al., 2015; van Cleemput et al., 2019; Wang et al., 2020; Wocheer et al., 2018, Dao et al., 2021).
154 The sites of the collected samples are distributed across different continents (America, Asia and Europe
155 see Fig. SA.1) and vary in climate and vegetation type (see details Table SA.1). The data comprise
156 observations from different natural and semi-natural ecosystems (forest, grassland, tundra and shrubland),
157 agricultural sites (crops and pastures), as well as plant-pot-experiments. Hence, the resulting data
158 represent an aggregation of large and heterogeneous multi-site and multi-ecosystem sources.

159 *2.1.1 Hyperspectral data*

160 Canopy reflectance spectra were acquired with proximal and airborne hyperspectral sensors (e.g. ASD
161 FieldSpec, Spectra Vista Corporation, SVC; AVIRIS; NEON Airborne Observation Platform AOP) and have
162 different spectral properties. Still, they cover a comparable wavelength range of the solar electromagnetic
163 spectrum (see Table SA.1). A forward and backward linear interpolation was performed to unify the diverse
164 measurements in the full range of 400-2500 nm in 1 nm steps. To deal with the known issues of atmospheric
165 water absorption in open-sky canopy reflectance spectra, we removed water absorption regions (1351-
166 1430, 1801-2050 and 2451-2501 nm) and independently smoothed the three remaining parts of the spectra
167 with a Savitzky-Golay filter (Savitzky and Golay, 1964) using a window size of 65 nm. Finally, 1721
168 interpolated spectral bands were retained for the analysis. Given the heterogeneity of the different data
169 sets, the 5573 processed spectra cover a wide range of reflectance values (Fig. 1). The data were checked
170 for overall spectral artifacts or inconsistencies (Supplement B). Despite the heterogeneity in land cover and
171 vegetation types, we observed smooth transitions between the spectral features of the different data sets
172 originating from sensor or pre-processing procedures (Fig. SB.1).



173

174

Fig. 1. Distribution of canopy reflectance of the available samples (N = 5573).

175 2.1.2. Leaf and canopy traits

176 From the available reference data, we selected a variety of traits (Table 1) related to light harvesting and
 177 growth, e.g. leaf pigments, nitrogen, structure and resource investments, leaf area index (LAI), equivalent
 178 water thickness (EWT) and leaf mass per area (LMA). Where necessary, leaf traits were converted to area
 179 basis, in contrast to mass-based measurements, to avoid the high correlation with LMA (see also Fig. SB.2)
 180 and to facilitate the model to learn the actual absorption features of the respective constituent (Hill et al.,
 181 2019; Kattenborn et al., 2019b; Ustin and Gamon, 2010, Zhao et al., 2021) (Fig. SB.2). Main conversions
 182 were based on the common relationships from literature (Kattenborn and Schmidlein, 2019; Lichtenthaler,
 183 1987) (Table SA.2). Table 1 summarizes the quantity of observations for each trait, their value ranges and
 184 the number of associated data sets. For the further analyses and for the sake of the training stability and
 185 computational efficiency, the trait values were rescaled. For this, we used the power transformation, which
 186 is a monotonic transformation to transform and normalize the data to a more-Gaussian-like distribution.

187

188 **Table 1.** Statistics of 20 selected functional traits available across 42 data sets. More details on the data
 189 sets can be found in Table SA.1. N = Number of samples, N Data sets = Number of data sets including
 190 the trait, Std = standard deviation, Min = minimum, Max = maximum.

191

Trait name	Trait description	Unit	N	N Data sets	Mean	Std	Min	Max
Anth	Anthocyanin content	($\mu\text{g}/\text{cm}^2$)	644	2	1.27	0.41	0.56	2.98

Boron	Boron content	($\mu\text{g}/\text{cm}^2$)	1086	14	0.39	0.26	0.01	2.34
C	Carbon content	(mg/cm^2)	1876	23	5.84	4.44	0.10	37.29
Ca	Calcium content	($\mu\text{g}/\text{cm}^2$)	1045	16	107.25	101.97	0.69	988.73
Car	Carotenoid content	($\mu\text{g}/\text{cm}^2$)	1859	21	8.75	2.77	1.18	40.44
Cellulose	Cellulose	(mg/cm^2)	1402	15	2.35	1.87	0.35	15.22
Chl	Chlorophyll content	($\mu\text{g}/\text{cm}^2$)	2141	24	38.57	14.53	4.45	229.50
Copper	Copper content	($\mu\text{g}/\text{cm}^2$)	1101	14	0.07	0.03	0.01	0.28
EWT	Equivalent Water Thickness	(mg/cm^2)	1918	19	15.65	9.27	0.23	80.62
Fiber	Fiber	(mg/cm^2)	1385	15	5.23	4.57	0.14	29.81
LAI	Leaf Area Index	(m^2/m^2)	1643	15	3.35	1.64	0.06	7.67
Lignin	Lignin	(mg/cm^2)	1415	16	2.69	2.41	0.05	14.58
LMA	Leaf Mass per Area	(g/m^2)	3328	32	92.05	68.08	5.72	663.81
Magnesium	Magnesium content	($\mu\text{g}/\text{cm}^2$)	1099	15	24.09	16.16	0.25	141.54
Manganese	Manganese content	($\mu\text{g}/\text{cm}^2$)	894	14	3.09	2.31	0.01	15.19
N	Nitrogen content	(mg/cm^2)	2193	26	0.19	0.10	0.01	0.95
NSC	Non-Structural Carbohydrates	(mg/cm^2)	1093	14	3.21	2.85	0.28	21.83
Phosphorus	Phosphorus content	($\mu\text{g}/\text{cm}^2$)	1289	16	14.42	9.45	0.29	73.43
Potassium	Potassium content	($\mu\text{g}/\text{cm}^2$)	1008	15	102.64	62.73	0.40	470.07
Sulfur	Sulfur content	($\mu\text{g}/\text{cm}^2$)	1039	14	13.31	9.13	0.62	57.23

192 2.2. Multi-trait model development

193 2.2.1. CNN implementation and training

194 Given the one-dimensional nature of the spectral data, we used one-dimensional Convolutional Neural
 195 Networks (1D-CNN). The context of neighboring wavelengths of the spectra makes CNN-based models

196 preferable here to the naive multilayer perceptron architecture (MLP). CNNs can have a deep structure and
 197 conventionally include blocks (convolutional block) of successive layers including convolution, pooling, and
 198 activation layers. The convolution operation is a sliding dot product of a filter (kernel) applied to the spectral
 199 signal. Several filters are used in the convolutional layer where they serve as a feature extractor and are
 200 iteratively learned during the CNN training process. The kernel's sliding fashion enables feature detection
 201 to be applied across the full signal range. Subsequently, the pooling layers enable to condense information
 202 from the output of the convolutional layers and facilitate a hierarchical feature extraction at multiple
 203 wavelength scales. For more details about CNNs, the reader is referred to Goodfellow et al. (2016).

204 As backbone architecture, we used an adapted version of EfficientNet-B0 (Tan and Le, 2019), which is
 205 modified for one dimensional input data. EfficientNet architectures are composed of a sequence of the
 206 previously described CNN blocks with skip connections. They are designed to improve accuracy and
 207 efficiency by using a scalable structure that allows the network to learn effectively from larger resolutions
 208 while reducing computation costs. This is done through a combination of depthwise separable convolutions,
 209 1x1 convolutions and network scaling methods (Tan and Le, 2019). The output layer of the implemented
 210 architecture comprised 20 units corresponding to the number of traits to be predicted.

211 The learning process of the model was based on the stochastic gradient descent algorithm, where the
 212 Adaptive Momentum estimation (Adam) optimizer was used to update the weights (Kingma and Ba, 2014).
 213 The number of epochs was set to 300 with a batch size of 32. We employed the Hubert loss function to
 214 reduce the effect of outliers. Given sparsity and resulting imbalance of trait observations in the merged data
 215 set, we used a weighted loss version. The weights of the samples w (%) were calculated for each sample
 216 as the complement of the number of non-null trait samples n_{nn} to the total number of samples n_{tot} in the
 217 corresponding original data set (Eq 1). Additionally, a random up-sampling with replacement was
 218 performed to have an equal number of samples from each data set on the training set.

$$219 \quad w = 100 - n_{nn} / n_{tot} * 100 \quad (1)$$

220 To avoid over-fitting, two conventional regularization techniques were used: data augmentation and drop
 221 out. Data augmentation introduces artificial variation in the data to help regulate the learning process. We

222 applied two random modifications for every epoch (a training cycle using all observations) with a 15%
223 chance. This included 1) an addition of random noise with +/-30% of the spectral standard deviation per
224 wavelength derived from all training samples and 2) an amplitude multiplication of the entire reflectance
225 spectra with a random value between 0.98 and 1.02. As additional model regularization, we applied dropout
226 (Hinton et al., 2012) after each block, which randomly drops learning units with a defined probability.

227 Within the 300 epochs, we selected the final model according to the lowest root mean squared error of a
228 20% hold-out from the training data. All CNN models were implemented in Python (3.9.5) with the
229 TensorFlow (2.7.0) and Keras (2.7.0) frameworks.

230 *2.2.2. CNN multi-trait and weakly supervised learning*

231 *In view of* the sparsity of the merged data set (Table. 1), we tested three different strategies to train multi-
232 trait models using the above-mentioned CNN architecture: The first strategy, $\text{CNN}_{\text{multiIncomplete}}$, was trained
233 on the original sparse data set. To overcome data sparsity, we modified the loss function to only update
234 the weights according to traits where a corresponding reference observation was present (i.e. not a missing
235 value). This approach falls within the incomplete supervision category in the context of weakly supervised
236 learning (Zhou, 2018). This strategy is considered as the baseline approach in this study.

237 The second strategy, $\text{CNN}_{\text{multiInexact}}$ aims to maximize the identification of trait-trait relations during the
238 learning process from all data samples and, hence, includes a gap-filling of missing trait values. The gap-
239 filling process is based on the predictions of the $\text{CNN}_{\text{multiIncomplete}}$. To avoid unrealistic values, trait predictions
240 lower than the 1% quantile and exceeding the 99% quantile of the original data set (Table 1) were not
241 considered for gap filling. This automated gap-filling approach does not require data on species or
242 ecosystem characteristics, which might be missing or hard to define (Schrodt et al., 2015; Shan et al.,
243 2012). Instead, it directly learns trait-trait relationships from available hyperspectral data. $\text{CNN}_{\text{multiInexact}}$ falls
244 within the two weak supervision categories: incomplete and inexact supervision. The incomplete
245 supervision is related to the gap-filling procedure, and the inexact supervision is performed when training
246 on the completed but noisy labels (i.e. reference data with gap-filled values).

247 The third strategy, $\text{CNN}_{\text{multiincompleteTRY}}$ aims to fill data gaps with trait observations obtained from the TRY
248 plant trait database (Kattge et al., 2020). The TRY database (version 5), includes more than 11.8 million
249 trait observations across more than 270.000 taxa. For each dominant species found in the reference data,
250 trait observations were queried from TRY using the species name. We applied fuzzy matching to deal with
251 minor inconsistencies in the spelling of the species names with a Damerau-Levenshtein-Edit distance >89
252 (Damerau, 1964; Konstantinidis, 2005). The dominant species mapping resulted therefore in 144
253 correspondences with TRY species. For these species, the mean median trait values were then used to fill
254 the missing values. This gap-filled data set was then used to train the multi-trait CNN models
255 ($\text{CNN}_{\text{multiincompleteTRY}}$). This strategy falls also within the inexact and incomplete weak supervision categories
256 as the model is trained on sparse and noisy labels (i.e. median trait values within species).

257 *2.3. Comparison to single-trait models*

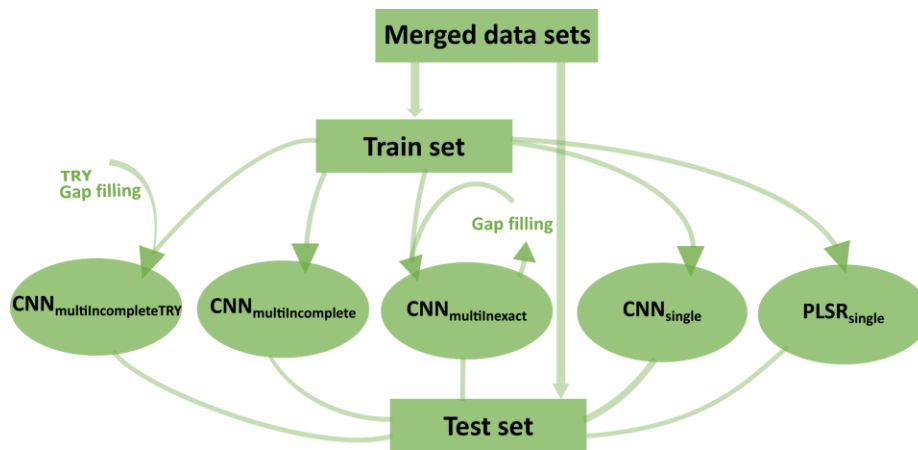
258 To evaluate the benefit of the multi-trait models and the uncertainty introduced from the weakly supervised
259 approaches (i.e. inexact and incomplete), we additionally trained single-trait CNN ($\text{CNN}_{\text{single}}$) models, where
260 a separate model was trained for each individual trait. Apart from the final layer (number of output units),
261 the architecture for these models was the same as for the multi-trait models (Section 2.2.1). Moreover, we
262 compared the CNN-based single and multi-trait models to partial least squares regression (Wold et al.,
263 1984). PLSR is currently one of the most frequently applied algorithms for imaging spectroscopy (Feilhauer
264 et al., 2010; Homolová et al., 2013). For training PLSR models for each trait ($\text{PLSR}_{\text{single}}$), we used scikit-
265 learn (version 0.24.2) Python libraries. To avoid over-fitting, the optimal PLSR number of latent components
266 was selected by minimizing the predicted residual sum of squares (PRESS) in cross-validation (Chen et al.,
267 2004).

268 *2.4. Model evaluation*

269 Using trait measurements and the canopy reflectance data from 42 data sets described in (section 2.1), we
270 compared the predictive performance of the 1) multi-trait CNN models to 2) single-trait CNN and 3) single-
271 trait PLSR models (Fig. 2). For a fair comparison, the same input data settings were adopted for the training
272 and evaluation of all modeling approaches including data splitting, transformation and up-sampling. The up-
273 sampling procedure is a random sampling with replacement and was applied to all samples in the training

274 set to make sure that a comparable number of samples is included from each data set and to reduce the
 275 effect of bias towards data sets with more samples.

276 After training, the models were evaluated for their performance 1) within the domain of the training data
 277 (internal validation) using randomly sampled hold-outs, and 2) with regard to their transferability to new
 278 domains (external validation), where each individual data set was once retained from model training. For
 279 the internal evaluation, we adopted a 5-fold cross-validation (CV) for all models. Given the unbalanced
 280 sampling frequency of the individual data sets, we performed a stratified cross-validation based on the data
 281 set provenance (data sets). This procedure ensures equal distribution of trait samples across the folds. For
 282 the hold-out test sets, only the original (and not the gap-filled) samples were used. The external validation
 283 consisted of training the models repetitively on 41 out of 42 data sets while keeping one data set as hold-
 284 out for testing. To reduce computational load, the data set-CV was only applied for $CNN_{\text{multiIncomplete}}$ and
 285 $PLSR_{\text{single}}$. We evaluated the model performances using the coefficient of determination R^2 and the
 286 normalized root mean squared error (nRMSE, %). The nRMSE was derived by normalizing the root mean
 287 square error over the range of the observations (1-99% quantile). The final model performance was obtained
 288 by averaging the R^2 and nRMSE values over the 5 folds of the CV.



289
 290 **Fig. 2.** Model performance assessment (internal validation) of multi-trait and single-trait models. The
 291 evaluation is based on a stratified 5-fold cross-validation.

292 2.5. Feature attribution

293 To visualize the spectral features learned by the $\text{CNN}_{\text{multiIncomplete}}$ model, we estimated the feature
294 importance of each input wavelength, which were interpreted and compared with known spectral plant
295 features. As feature importance metrics, we derived medians of SHapley Additive exPlanations (SHAP,
296 Lundberg and Lee, 2017) absolute values. The SHAP values present a unified approach to explain model
297 predictions based on the optimal game theory Shapley values. The Shapley values represent the local
298 marginal contribution (i.e. for individual samples in the data) of each feature in the input for a specific
299 prediction. They attribute the change in the expected model prediction when conditioning on one feature
300 by calculating the difference from the prediction in which the feature in question is not included (Lundberg
301 and Lee, 2017). These values can be approximated with different algorithms. We used the gradient
302 explainer class, which combines the integrated gradients (Sundararajan et al., 2017), SHAP and
303 SmoothGrad (Smilkov et al., 2017) methods. All SHAP coefficients were rescaled between 0-1 and
304 normalized by the mean SHAP value of all traits to eliminate the effect of the learned trait covariance. For
305 comparison, we also displayed the PLSR regression coefficients.

306 *2.6. Uncertainty estimation*

307 As indicated by earlier studies, transferability of machine learning-based models to new, unseen data
308 depends on the distance in feature space (Kattenborn et al., 2022; Ludwig et al., 2023; Mila et al., 2022).
309 Therefore, we implemented an uncertainty estimation procedure to reveal the effect of spectral dissimilarity
310 between new data and data used in model training. Such a procedure is particularly valuable in view of
311 large-scale mapping across ecosystems and sensors.

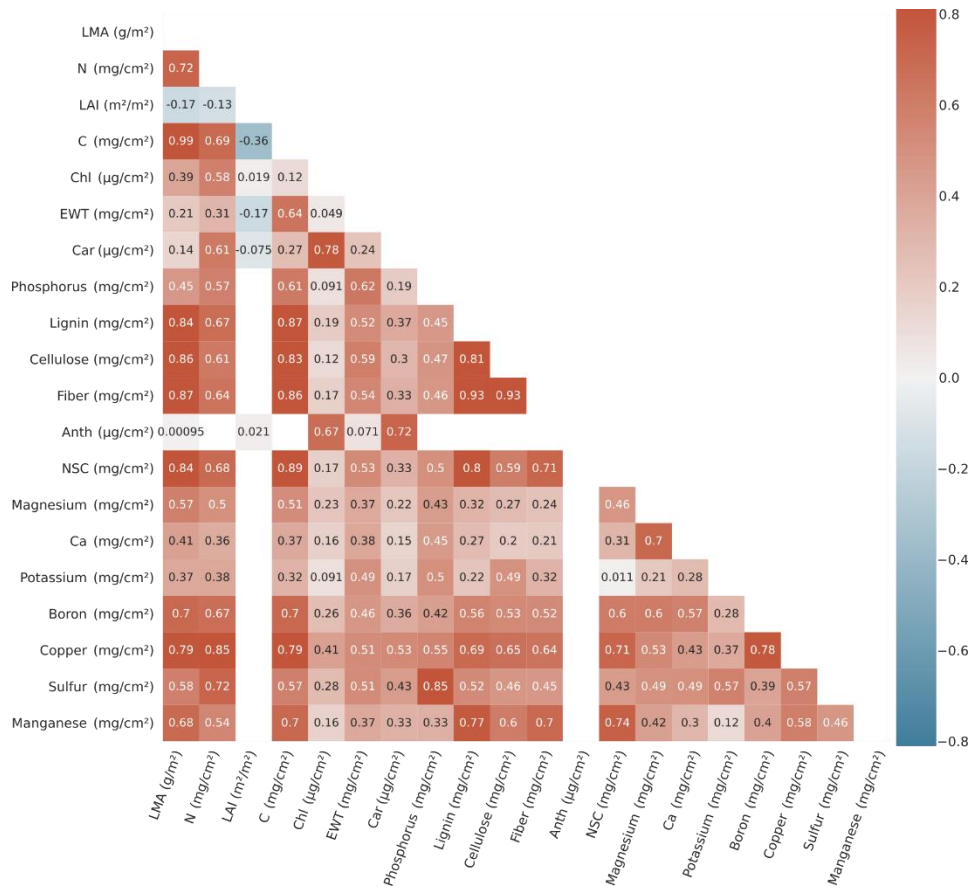
312 Inspired by Janet et al. (2019) and Meyer and Pebesma (2021), the implemented uncertainty estimation
313 was based on the relationship between 1) CNN model residuals obtained from the internal evaluation and
314 2) the distance in feature space (dissimilarity of training vs. test sets). To reveal spectral dissimilarity from
315 the eye of the CNN, the feature space was obtained from the CNN model embedding space of the global
316 pooling of the last convolutional layer. Based on this feature space, the dissimilarity for each test sample
317 was calculated as the average distance to the five nearest neighbors of the training data. The model
318 uncertainty was then estimated using the calculated dissimilarity as predictor in a 95% quantile regression.
319 The predicted values can be seen as the worst-case error prediction of the model. This procedure was
320 tested for the $\text{CNN}_{\text{multiIncomplete}}$ model.

321 **3. Results**

322 *3.1. Summary of the merged data set*

323 *3.1.1. Trait variations*

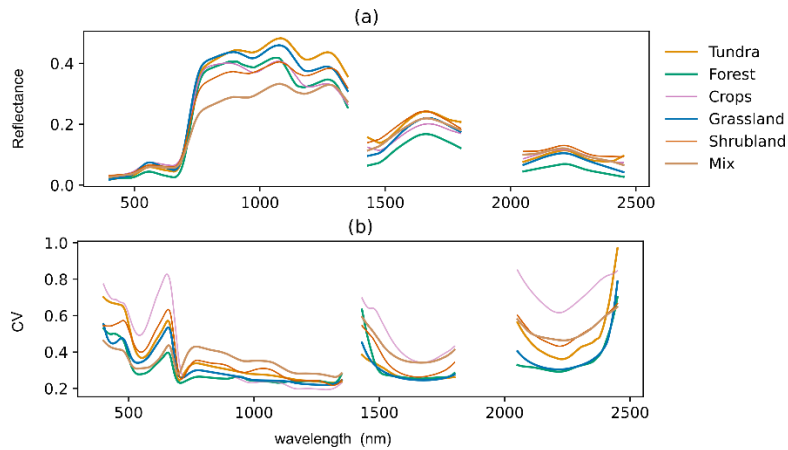
324 The trait values across the merged data sets varied highly due to the heterogeneity in vegetation types and
325 species (Table 1). This yields a large range in the trait values. LMA, Chl and EWT showed the highest
326 variability (Coefficient of Variation CV = 47.97, 42.91, 38.44%, see Fig SB. 3, 4) in the original data while
327 all other traits had similar variations (on average 35%). The correlation analysis based on Spearman
328 coefficient of the merged data set revealed high correlation between several traits (Fig. 3). As expected,
329 leaf constituents related to plant resource investments showed a large correlation (e.g. LMA, Carbon,
330 Lignin, Fiber, Cellulose). These resource-investment related traits were rather independent from leaf
331 pigments, which in turn were highly correlated among each other (Chl, Car, Anth). Both resource investment
332 related traits and pigments showed a considerable correlation with leaf N. Overall, rather weak correlations
333 were found for LAI and leaf constituents, whereas for N and C a positive relationship was observed. Water
334 content overall also showed a positive correlation with other leaf constituents.



335
 336 **Fig. 3.** Correlation plot of traits based on Spearman's rank correlation coefficient. Refer to Table 1 for an
 337 explanation of the traits. A correlation of leaf traits on a mass-basis is given in Fig. SB.2.

338
 339 *3.1.2. Canopy spectra*

340 The canopy reflectance spectra were relatively similar when averaged across land cover types (Fig. 4) and
 341 we found smooth transitions across data sets and biomes (Fig. SB.1). Higher reflectance values were
 342 observed for the Tundra data in the NIR region (Fig. 4a). Largest coefficients of variations were found in the
 343 SWIR 2 region (2000-2500 nm) followed by the VIS region (400-750 nm). Most of the spectral variation was
 344 found in the crop related samples whilst forest samples had the lowest spectral variation (Fig. 4b).



345
 346 **Fig. 4.** Canopy average reflectance and the corresponding spectral variation (Coefficient of Variation, CV)
 347 across the different land cover types

348 3.2. Trait predictions

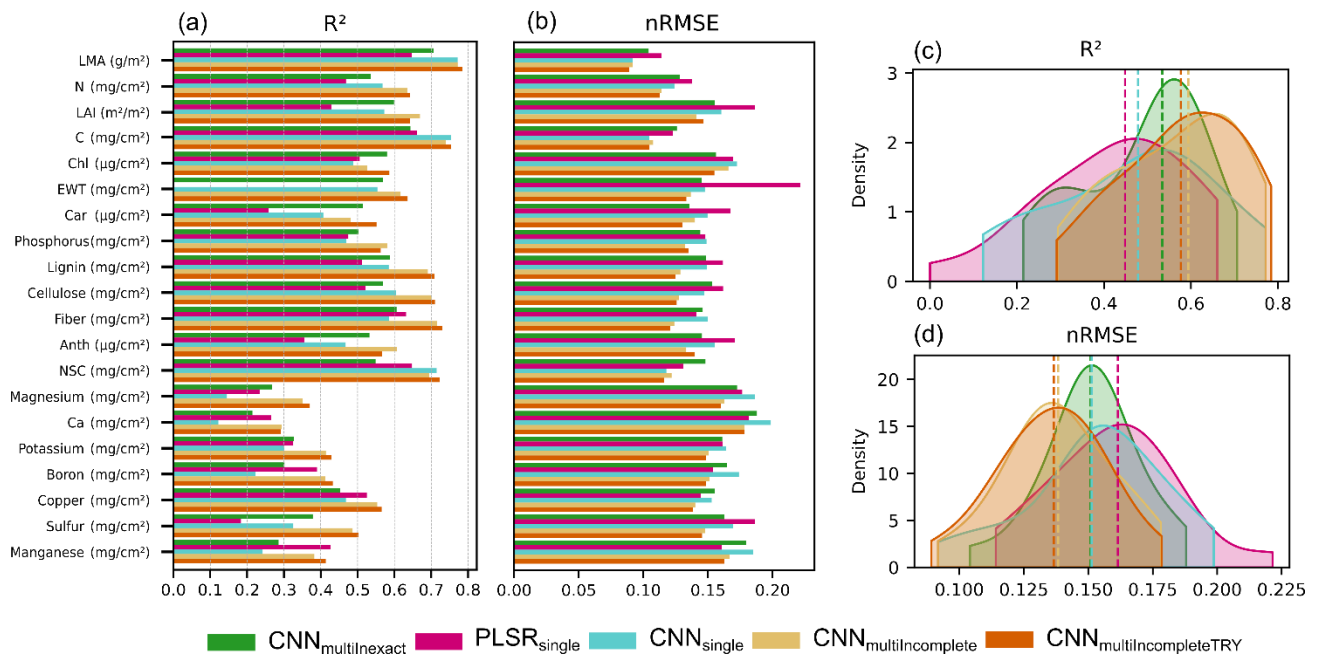
349 3.2.1. Prediction performances

351 The model performances derived from the 5-fold cross-validation showed the overall predictive performance
 352 varied greatly for the different traits (Fig. 5). With all CNN-based models, the goodness-of-fit of the
 353 predictions was higher for LMA, C, NSC (Non-structural carbon) ($R^2 > 0.69$). Lower predictive performances
 354 of these models were obtained for EWT, N, Pigments, LAI, Cellulose, Lignin, Fiber, Copper and Phosphorus
 355 (R^2 : 0.46 – 0.69 and nRMSE: 12 – 17%). Overall, the trait estimation performances of the CNN-based
 356 models exceeded those of the PLSR models (R^2 : 0.18 to 0.66 and nRMSE: 11 – 22%). The PLSR models
 357 showed bias with high values for some traits, including LMA, Pigments and Carbon related traits (See Fig.
 358 SD.1). Only for a few traits, i.e. Boron, Ca and Manganese, the PLSR models showed higher performances
 359 than CNN models. Ca, Boron, Magnesium, Sulfur, Potassium and Manganese obtained the lowest validation
 360 performance for all models, especially with single-trait models ($R^2 < 0.44$ and nRMSE >15 %).

361
 362 According to a Wilcoxon signed-rank test, the multi-trait models performed significantly better than single-
 363 trait models across all traits (e.g. $\text{CNN}_{\text{multiIncomplete}}$ $p < 0.001$, $w = 205$, details see Supplement C). In
 364 comparison to $\text{CNN}_{\text{single}}$, CNN-based, multi-trait models clearly improved the prediction performance for
 365 most of the traits. The prediction performance was particularly improved for traits where fewer samples were
 366 available or where a comparably lower correlation with spectral bands was observed (Fig SB.4), including

367 Anth, Sulfur, Ca and Potassium (Fig. 5b). Overall, the R^2 across all traits was higher for CNN multi-trait
 368 models than for CNN_{single} except for LMA, C and NSC (Fig. 5c, d).

369 Similar performance was obtained among the different CNN-based multi-trait models, i.e. CNN_{multiIncomplete},
 370 CNN_{multiInexact} and CNN_{multiIncompleteTRY}. The predictive performance for the CNN_{multiInexact} ranged from R^2 of
 371 0.21 – 0.70 and nRMSE of 10.41 – 18.79% , for CNN_{multiIncomplete} R^2 of 0.29 – 0.77 and nRMSE of 9.17 –
 372 17.81% and CNN_{multiIncompleteTRY} R^2 of 0.29 – 0.78 and nRMSE of 8.92 – 17.85%. Overall, the
 373 CNN_{multiIncompleteTRY} performed slightly better than the other two multi-trait strategies for most of the traits (Fig.
 374 5). The CNN_{multiIncompleteTRY} procedure is further discussed in section 3.2.2.



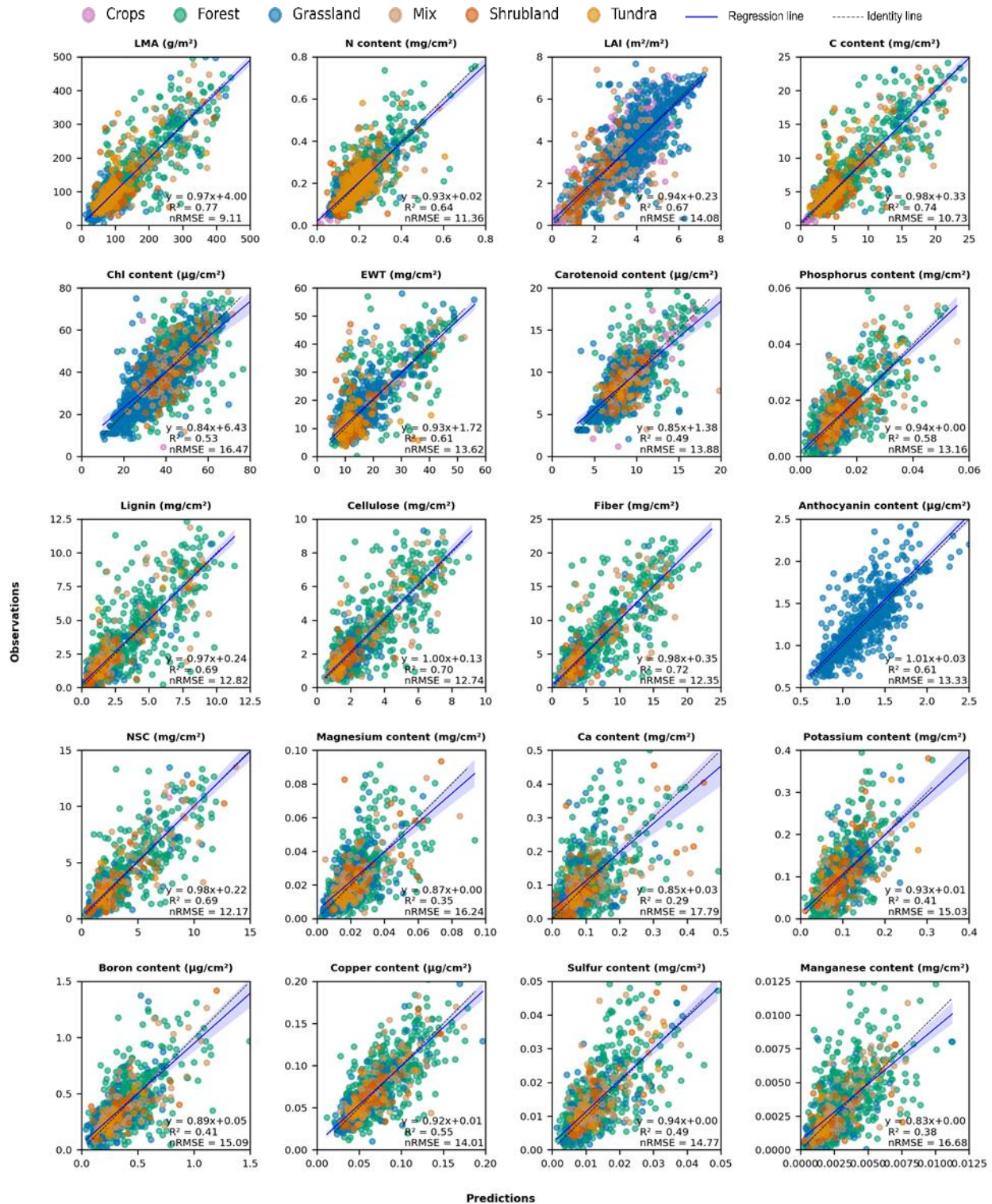
375
 376 **Fig. 5.** (a) and (b): Comparative predictive accuracies (R^2 (a) and nRMSE (b)) from the stratified 5-fold
 377 cross validation of the CNN_{multiIncomplete}, CNN_{multiInexact} and CNN_{multiIncompleteTRY} models as well as PLSR_{single}
 378 and CNN_{single} models for 20 traits. (c) and (d): The kernel density estimate (KDE) of the trait-based metric
 379 distributions (R^2 (c) and nRMSE (d)) with the associated median values (dashed lines). Refer to Table 1 for
 380 an explanation of the traits. Detailed performances can be found in Table SC.1 and 2 Supplement C.

381
 382 All multi-trait approaches resulted in relatively robust and similar prediction performances across the
 383 different vegetation types (Fig. 6, SD.3, 4). For some traits (e.g. LMA, N, EWT), the values are slightly
 384 clustered according to vegetation types, but we did not observe a prominent or systematic bias in predictive

385 performance across these classes. For most traits, the model predictions are evenly scattered around the
386 1:1 line, which is also underlined by slopes of the linear fit close to 1 between the predicted and observed
387 trait values (Fig. 6, SD.3, 4).

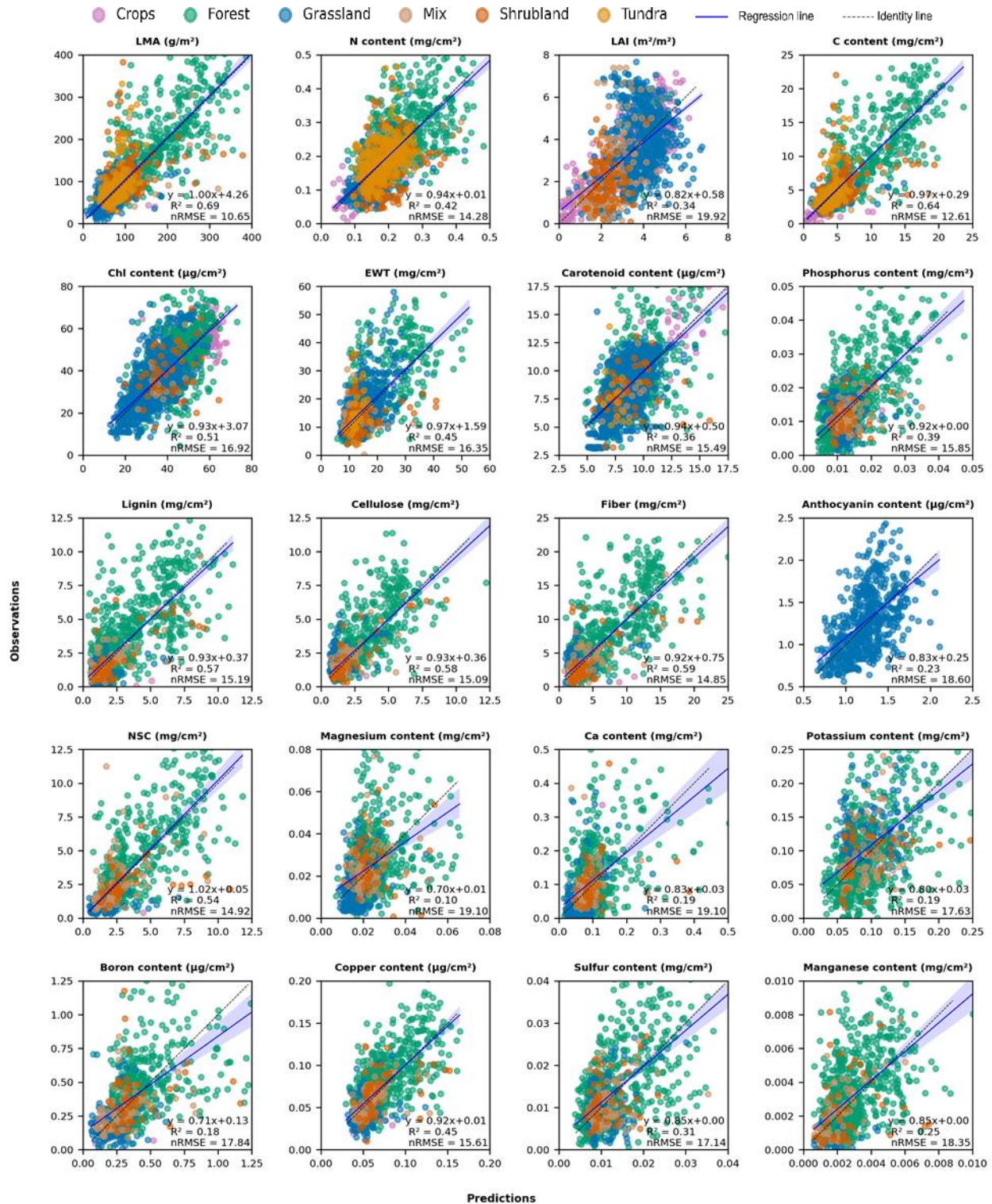
388

389



390
 391 **Fig. 6.** Internal validation: Correlation between observed and predicted values of 20 traits from the multi-
 392 trait model $CNN_{\text{multilIncomplete}}$. The shown vegetation types only refer to the available types in the original
 393 associated data sets (not all land cover types are covered for each trait). Refer to Table 1 for an explanation
 394 of the traits. Scatter plots for the other models are given in supplement D.

395 Similar performance results were followed with the external evaluation, where CNN multi-trait model
396 surpassed the performance of PLSR_{single} models (Fig. 7, SE.1). With both modeling approaches, the
397 performance across all traits with the external validation was lower than the internal validation, especially
398 with PLSR_{single} (Fig. 7, SE.1, Table SE.1). For CNN_{multiIncomplete} LMA and C were the most transferable traits
399 with R^2 higher than 0.6 which is consistent with the internal validation, while for PLSR Copper and Chl had
400 the highest goodness-of-fit with $R^2 > 0.39$. However, the baseline multi-trait model (CNN_{multiIncomplete}) showed
401 a bias in high trait values with N and LAI for example.



402
 403 **Fig. 7.** External validation: Correlation between observed and predicted values of 20 traits from the multi-
 404 trait model CNN_{multiIncomplete}. The shown vegetation types only refer to the available types in the original
 405 associated data sets. Scatter plots for PLSR_{single} are given in supplement E.

406 3.2.2. Details on trait database integration

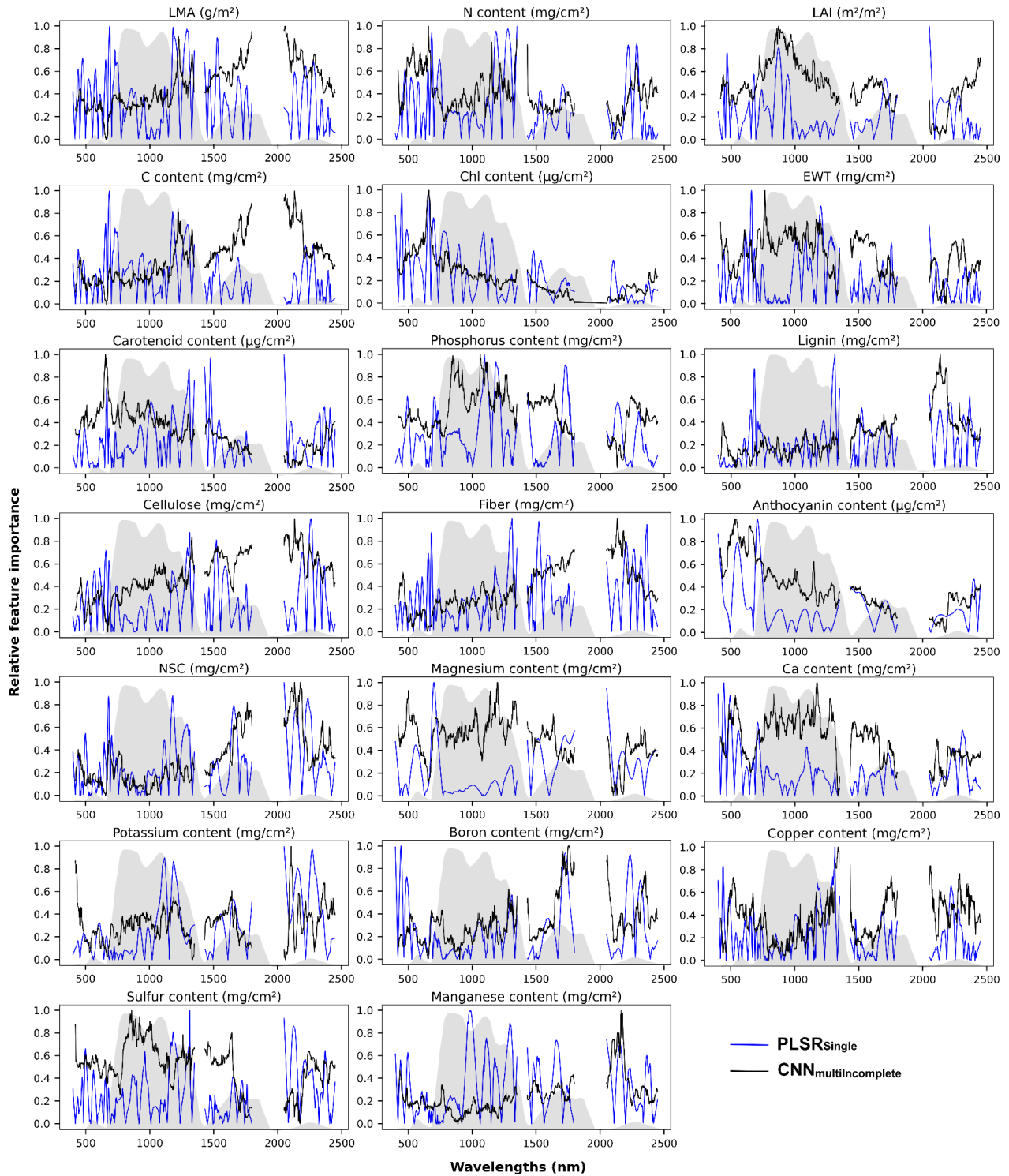
407 Due to data availability, the gap filling of the $\text{CNN}_{\text{multiIncompleteTRY}}$ procedure was limited to 13 out of 20 traits
 408 (Tab. 3). The model performance significantly improved for all the gap-filled traits ($p = 0.004$, $w = 82$,
 409 Wilcoxon signed-rank test). Surprisingly, the $\text{CNN}_{\text{multiIncompleteTRY}}$ approach resulted even in significantly
 410 improved performance for traits where no gap-filling could be performed, i.e. EWT, Car, Fiber, NSC and S
 411 ($p = 0.0313$, $w = 15$). While the filling rate was not an important factor for model improvement, the
 412 introduced variation from the species-based trait values had the largest effect on traits that already had
 413 less sparse trait observations samples in the data set. For instance, Chl had the highest improvement in
 414 performance and even surpasses the results of the baseline model $\text{CNN}_{\text{multiIncomplete}}$ (Table 3).

415
 416 **Table 3.** Comparative nRMSE values (%) of the $\text{CNN}_{\text{multiIncompleteTRY}}$ with $\text{CNN}_{\text{multiIncomplete}}$ and $\text{CNN}_{\text{single}}$
 417 models. $\text{CNN}_{\text{multiIncompleteTRY}}$. Filling rates = $(n \text{ obs. after} - n \text{ obs. before}) * 100 / \text{old metric}$. Refer to Table. 1
 418 for an explanation of the traits and to Table. SC.3 for more detailed metrics.

Traits	Filling rate (%)	nRMSE (%) ($\text{CNN}_{\text{single}}$)	nRMSE (%) ($\text{CNN}_{\text{multiIncomplete}}$)	nRMSE (%) ($\text{CNN}_{\text{multiIncompleteTRY}}$)
Potassium	118.14	16.42	15.04	14.84
Phosphorus	99.07	14.89	13.23	13.51
Ca	97.42	19.87	17.82	17.85
Magnesium	93.64	18.65	16.26	16.00
C	92.33	10.45	10.76	10.48
Manganese	64.54	18.49	16.69	16.26
N	50.69	12.40	11.39	11.29
Copper	50.27	15.29	14.02	13.83
Chl	34.13	17.25	16.58	15.50
LMA	23.20	9.18	9.18	8.92
Lignin	12.43	14.91	12.86	12.48
Cellulose	7.77	14.71	12.78	12.58
Boron	0.55	17.39	15.11	14.86

419
 420 3.2.3. Feature importance

421 The feature importance for $CNN_{\text{multiIncomplete}}$ and $PLSR_{\text{Single}}$ showed a clear correspondence in the overall
422 patterns (Fig. 8). For LMA, the relevant wavelengths in the CNN multi-trait model were spread across the
423 entire spectrum, with higher values in the longer wavelengths of the NIR SWIR regions (1200 - 2450 nm).
424 As expected, very similar patterns were found for traits that directly contribute to LMA, namely C, Cellulose,
425 Fiber and Lignin. The CNN multi-trait estimation of Chl and Car mostly relied on spectral bands in the VIS
426 and red-edge region (approx. 500 - 800 nm). For LAI, high SHAP values were found in the NIR.



427
 428 **Fig. 8.** Relative importance of spectral bands for the prediction of 20 traits using the CNN_{multiIncomplete} and
 429 PLSR_{single} models. The importance metric of CNN_{multiInComplete} (Black) is based on the SHAP scores with the
 430 gradient explainer, as for PLSR_{single} the regression coefficients are shown (Blue). The gray shaded polygon
 431 represents a sample vegetation spectrum for orientation.

432 **4. Discussion**

433 *4.1. Considerations on the merged data set*

434 The transferability of statistical models to predict plant traits from new reflectance spectra is a major
435 challenge (Ainsworth et al., 2014; Heckmann et al., 2017; Silva-Perez et al., 2018). Few previous studies
436 have demonstrated that the transferability of models can be enhanced when the model training includes
437 plots from different species and sites (Asner et al. 2015, Serbin et al., 2019; Wang et al., 2020, Kothari et
438 al., 2022). Here, we merged 42 canopy reflectance data sets (from 28 studies) to assess the robustness of
439 retrieval models when calibrated on heterogeneous data not only from different ecosystem types but also
440 experimental settings (e.g. hyperspectral data acquisition and in-situ measurement protocols). This
441 procedure provides an opportunity to address common shortages of reference data while also increasing
442 the representativeness in terms of geographical coverage and diversity in vegetation type in the training
443 data. Yet, it should be noted that the temporal coverage of the data is biased towards the peak of the
444 vegetation period, while the senescence is underrepresented. This may affect for example the inter-
445 correlations between traits as displayed in Fig. 3.

446
447 Merging the data sets required expert knowledge and a considerable effort for checking, cleaning, and
448 converting trait observations. Although a large share of the data used here was acquired from the EcoSIS
449 database, the available data often include errors and inconsistencies, e.g. assignment of wrong dimensions
450 or units. In consideration of future initiatives for data integration, these experiences emphasize the need
451 for a harmonization of plant trait observations, including units and dimensions, e.g. area or mass based, as
452 well as quality assessments, terminology and sampling protocols.

453
454 As this merged data set incorporates various ecosystems and land cover types, its trait variability exceeds
455 those of previous studies (Table 1, Fig. 3, Asner et al., 2015; Schiefer et al., 2021; Serbin et al., 2019;
456 Wang et al., 2020, 2019). We assume that merging the different data sets is a compelling requirement for
457 developing models that are transferable and robust across different traits, ecosystems, and vegetation
458 types in the context of global mapping. Here, the baseline multi-trait model ($CNN_{\text{multiIncomplete}}$) appeared to
459 generalize well over the individual data sets (Fig. 7). It should be noted, however, that the data only
460 represent a small portion of the Earth's flora and its spatio-temporal variation. Hence, despite the

461 unprecedented trait variability realized here, the presented study should be regarded as a pioneering study
462 in terms of model transferability and performance.

463
464 In the merged data set, not only the trait values but also the reflectance data showed considerable
465 variability, which could be attributed not only to the spectral properties of the vegetation itself but also to
466 differences in pre-processing modes with related uncertainties (e.g. during atmospheric correction
467 procedures), remote sensing data acquisition settings (e.g. sun-observer-relationship) and instruments
468 (e.g. airborne vs. field spectrometer data). We could not investigate in depth to what extent such factors
469 limited the transferability of the models as information on such factors was not available for all individual
470 data sets. Yet, we did not observe a significant difference in performance of our baseline multi-trait model
471 ($\text{CNN}_{\text{multitraitincomplete}}$) across the different remote sensing platforms ($p = 0.17$, $u = 72$, Mann–Whitney-U test)
472 (Fig. SE.3, 4).

473
474 Merging data from multiple sources may improve model performances and transferability, but the sparsity
475 and imbalance of trait observations challenged the model evaluations. For example, the number of data
476 sets per trait ranges from 2 to 32 (Table 1), as most studies are application-specific and, hence, trait-
477 specific. Likewise, the number of observations per data set ranged from 22 to 549. Thus, the relative
478 performance of the model for the different traits is not necessarily directly comparable. Similarly, for some
479 ecosystems or vegetation types only a few samples were available, which limited a conclusive performance
480 evaluation in this regard. These challenges are expected to be resolved as more data may become openly
481 available in the future.

482
483 *4.2. Comparison of modeling approaches*

484 Overall, the model performances of CNN-based models outperformed the widely used PLSR based
485 models. This is consistent with previous studies that used hyperspectral data to retrieve vegetation and soil
486 properties (Cui and Fearn, 2018; Ng et al., 2019; Pullanagari et al., 2021). The increased performance of
487 CNN over PLSR may be explained by its ability to represent nonlinear relationships with an overall
488 increased number of parameters, enabling the algorithm to learn more complex relationships. For example,
489 the large trait-ranges of the merged data set presented in this study may inherit several non-linearities

490 between spectra-trait-relationships. Such nonlinearities may result from saturation effects, where a change
491 in high trait values results in little change of spectral reflectance, as observed in the present study for
492 chlorophyll, LMA or LAI. The linearity of PLSR models appeared to be less suitable to resolve such effects,
493 as indicated by a clear saturation of PLSR-based predictions for high values for these traits (Fig. SD.1,
494 SE.1). In such cases PLSR models tended to include more predictors (latent vectors) but this did not
495 necessarily improve the model performance. Similar issues with PLSR-based models and saturation effects
496 were also reported with leaf-scale reflectance data in Kothari et al. (2022). In contrast to the PLSR-based
497 predictions, the predictions of the CNN models did not show saturation effects and no obvious systematic
498 biases could be observed across the trait range (Fig. SD.2 – 4).

499
500 In addition to the model performance, CNNs are known to be less reliant on feature engineering and are
501 effective to identify automatically relevant features from the input data (Goodfellow et al. 2016). Previous
502 studies in the context of variable retrieval from hyperspectral data showed that shallower machine learning
503 methods were more dependent on pre-processing of input data (Cui and Fearn, 2018; Ng et al., 2019).
504 Another advantage of CNNs and other batch-compatible deep learning methods over previous machine
505 learning methods (e.g. PLSR, Random Forest) is that the data are exposed iteratively to the model, which
506 potentially enables training models with an infinite amount of data without exceeding the memory. The latter
507 aspect may become very relevant in the near future that promises an increase in data availability, e.g. via
508 more data acquisitions from spaceborne spectrometers and a growing culture of open data through
509 initiatives such as ecosis.org.

510
511 The multi-trait CNN models clearly outperformed the single-trait models. This is consistent with other studies
512 in different areas which employed multi-task CNN models (Ng et al., 2019; Padarian et al., 2019; Ramsundar
513 et al., 2015; Tsakiridis et al., 2020). In comparison of the CNN_{single} model the retrieval of pigments, N, LAI,
514 EWT, Phosphorus, Lignin, Cellulose, Fiber, Magnesium, Ca, Potassium, Boron, Copper, Sulfur was
515 improved with our baseline multi-trait model (CNN_{multiIncomplete}). Even for traits that were only represented in
516 a few data sets, the multi-trait models performed better than the single-trait models (e.g. Anth, Sulfur,
517 Copper, Boron, Magnesium). We assume that multi-trait models not only allow for simultaneous and thus
518 efficient trait retrieval, but also allow the model to indirectly learn trait-trait relationships.

519 Such trait-trait relationships may also explain the observed feature importances (Fig. 8). For instance, the
520 spectral features for N were consistent with known protein features in the SWIR region (Féret et al., 2021)
521 and others near the red-edge region related to pigments (Ustin et al., 2009). As expected, we also observed
522 very similar spectral features across all wavelengths among traits related to leaf resource investments (LMA,
523 Lignin, Fiber, Cellulose, and C; compare Kokaly et al., 2009), which may also explain higher model
524 performance for several of these traits when predicted in a multi-trait setting. For Anth, we observed
525 relatively accurate predictions and rather broad absorption features, although previous studies revealed that
526 Anth pigments have rather subtle and narrow spectral absorption properties (Féret et al., 2017). We assume
527 that the broad features obtained here result from the high correlation with Chl and Car (Fig. 3), which in turn
528 have more broad spectral absorption features and may indirectly facilitate Anth estimation (Jacquemoud
529 and Ustin, 2019, Ollinger, 2011; Ustin et al., 2009). Similarly, nutrients such as Copper, Sulfur, Potassium
530 and Boron do not have distinct spectral absorption features in canopy spectra, but their surprisingly high
531 retrieval performance may be explained by their correlation with other leaf traits that are related to leaf
532 resource investments (Fig. 3, 6) and that have a more explicit spectral response, such as LMA or C
533 (Domínguez et al., 2012, Kothari and Schweiger, 2022).

534
535 Largest improvements from single- to multi-trait estimates were found for Lignin, Cellulose and Fiber (Fig.
536 5), which can be attributed to the high correlation with LMA and C (Fig. 3). Conversely, for LMA, C and NSC
537 the multi-trait approaches did not result in notable improvements. This may be explained by the fact that
538 these three traits are already very tightly related (chemically and spectrally) and a covariance among these
539 traits does not add further benefit. Moreover, compared to other traits, LMA, C and NSC can be predicted
540 most accurately (Fig. 6), so the covariance with other traits that cannot be predicted as accurately is less
541 likely to facilitate the predictive performance. Similar findings for LMA estimation were found by Furbank et
542 al. (2021) when including the inter-correlation with photosynthetic traits.

543
544 We tested three weakly supervised strategies for training the multi-trait models in the context of the data
545 sparsity, i.e. $\text{CNN}_{\text{multiIncompleteTRY}}$, $\text{CNN}_{\text{multiIncomplete}}$ and $\text{CNN}_{\text{multiInexact}}$. The three strategies resulted in similar
546 model performance across the traits (e.g. for each strategy, LMA, C and NSC were most accurate and
547 macronutrients least accurate). Yet, $\text{CNN}_{\text{multiInexact}}$ resulted in the lowest model performance. This is

548 explained by the uncertainty introduced during the spectrally-based gap-filling procedure. However, even
549 with the propagated uncertainty from the gap-filling process, $\text{CNN}_{\text{multiInexact}}$ outperformed the single-trait
550 models. This demonstrates that such gap-filling strategies are promising to enrich existing sparse data sets,
551 especially as no external knowledge on species or ecosystem type is required. Future attempts may apply
552 a more conservative gap-filling, where data gaps are only filled if the estimated traits are assumed to have
553 a low uncertainty. The uncertainty assessment presented in this study (see Fig SG.1 for details) may be a
554 promising avenue.

555
556 The gap filling strategy based on trait databases ($\text{CNN}_{\text{multiIncompleteTRY}}$) significantly improved the
557 performance (compared to $\text{CNN}_{\text{multiIncomplete}}$) for those traits that were gap-filled ($p = 0.004$, $w = 82$ Wilcoxon
558 signed-rank test, Table 3, SC.3), even when using median trait values by species which do not account for
559 the within-species trait variations. Nonetheless, for the scope of this analysis this does not affect the
560 interpretation of the results as most of the collected samples were taken in the growing season and the
561 results were only evaluated with the original trait observations (i.e. no gap filling). Interestingly,
562 $\text{CNN}_{\text{multiIncompleteTRY}}$ even improved the model performance for those traits where no gap-filling was
563 performed (due to missing observations in the TRY database, $p = 0.0313$, $w = 15$, Wilcoxon signed-rank
564 test, Fig. 5, Table SC.3). This not only underlines the potential of incorporating ancillary trait information,
565 but also highlights the overall value of the multi-trait and corresponding trait-trait relationship. For instance,
566 this has surprisingly influenced the retrieval of Chl and Car, with an improvement of 12 – 16 % in R^2 and 7
567 – 7.88 % in nRMSE; as well as EWT, Fiber, NSC and Sulfur by 2.00 – 4.10% in R^2 and 1.68 – 4.95% in
568 nRMSE. We assume that the growth of trait databases as TRY will even increase the potential of this gap-
569 filling approach.

570

571 *4.3. Model performance across plant traits*

572 Across all traits, highest model performance was observed for LMA (Fig. 5). This is in line with a series of
573 previous studies highlighting the transferability of models for estimating LMA across data sets at leaf and
574 canopy scale (Serbin et al., 2019; Silva-Perez et al., 2018; Wang et al., 2019, 2020, Helsen et al., 2021,
575 Schiefer et al. 2021, Kothari et al., 2022). In contrast to these previous studies, the CNN models used here
576 resulted in comparable or even higher model performances although we tested our models using a more

577 diverse data set and exclusively on canopy spectra. The high performance of the LMA estimation is partly
578 a surprise given its broad and overlapping absorption features with water content and scattering
579 components at the canopy scale (Homolová et al., 2013). The high performance of LMA may be partially
580 supported by the ample samples across most of the used data sets (32 data sets out of 42 had LMA
581 observations). Moreover, the robustness of the LMA estimation may also be explained by the overall high
582 correlation of LMA with individual bands across the entire spectrum (Fig. SB. 4).

583
584 Particularly for LMA but also for most of the other traits, our results suggest that the performances of the
585 multi-trait models are often on par to those of previous studies. For instance, for LAI, Chl, Car and EWT,
586 our models obtained higher performances than Schiefer et al. (2021), who used PLSR models on a data
587 set of canopy spectra across grassland species, which was also integrated in our study. Overall, model
588 performances were comparable to Wang et al. (2020), who used airborne canopy spectra across biomes
589 and to Wang et al. (2019), who used canopy spectra in grasslands. EWT performances were lower than in
590 Wang et al. 2020, where water content was one of the most accurately retrieved traits. The fact that the
591 estimation of EWT was comparably low in the present study may result from the different protocols used
592 across the merged data sets.

593
594 In this study we focused on area-based leaf traits due to multiple reasons: Firstly, as highlighted across
595 different studies in the context of the radiative transfer theory (Dawson et al., 1998; Ganapol et al., 1998;
596 Jacquemoud and Baret, 1990; Vilfan et al., 2016), the retrieval of leaf constituents from spectral signals
597 depends on how much of a leaf constituent (mass) in a given leaf area interacts with light (area-based). In
598 contrast, relative ratios of leaf constituents to LMA (mass-based traits) are not directly related to spectral
599 absorption features (also discussed in Kattenborn et al. 2019b, Zhao et al. 2021). Secondly, normalizing
600 traits on a mass-basis may overshadow the original variation of leaf traits and introduce unrealistic trait-
601 trait-relationships. For instance, photosynthetic traits (e.g. pigments) are generally assumed to be largely
602 independent of leaf resource investments (LMA) (Lloyd et al., 2013; Osnas et al., 2013). This was confirmed
603 for the present data set (Spearman $\rho < 0.4$) - but only if the data was scaled on an area-basis (Fig. SB.2).
604 As soon as pigments were scaled on a mass-basis, ill-founded correlations were introduced (Spearman
605 $\rho < -0.73$, see Lloyd et al. 2013 for a statistical rationale). Likewise, traits that directly contribute to the

606 total leaf mass were obviously highly correlated to LMA when compared on an area-basis (Spearman's
607 $\rho > 0.84$ for Carbon, NSC, Lignin, Fiber, Cellulose), while a comparably weak relationship was found on
608 a mass-basis (Spearman's $\rho < 0.51$). Moreover, we found unrealistically high variation of these LMA-
609 related traits (Carbon, NSC, Lignin, Fiber, Cellulose) when assessed on a mass-basis, which may have
610 mis-lead model calibration (Fig. SB.3). Thus, to comply with the physical principles of radiative transfer
611 theory but also reasonable trait-trait relationships, the modeling in the present study was performed
612 exclusively on an area basis.

613
614 Note, however, that our proposed models can also be used to derive mass-based traits through
615 normalizing the respective trait prediction by LMA predictions ($\text{trait}_{\text{mass}} = \text{trait}_{\text{area}} / \text{LMA}$). We applied this
616 procedure to compare our model performances to previous studies that performed trait retrieval on a mass-
617 basis (Supplement F). The performances of our baseline multi-trait model ($\text{CNN}_{\text{multiIncomplete}}$) with mass-
618 based N and Phosphorus were comparable to studies reviewed in Homolová et al. (2013), while exceeding
619 those of Wang et al. (2020), Asner et al. (2015), Chadwick and Asner (2016), Ewald et al. (2018a) and
620 Wang et al. (2019). The predictive performance for the converted pigments, Fiber, Lignin and Cellulose
621 was lower or comparable to Wang et al. (2020) and Singh et al. (2015) and exceeded those of Asner et al.
622 (2015) and Martin et al. (2018) for tropical forest.

623
624 Nevertheless, it should be highlighted that it is often not possible to directly and quantitatively compare
625 model performances across studies, since they frequently differ in vegetation type, modeling approach,
626 model performance metrics and validation strategy, remote sensing platform and sensor, temporal and
627 spatial resolution and extent, simulated and real data, plant traits or a combination of these. Also, the aim
628 and thus the setting of the individual modeling attempts largely differs: some studies aimed to predict traits
629 in a very specific domain and from a very specific platform, while here we aimed to predict traits across
630 different platforms, sensors and vegetation types.

631
632 *4.4 Model performance across data sets (transferability)*
633 While the 5-fold CV evaluated the model performance with observations that are similar to those
634 observations used in training (internal validation), the model transferability specifically estimated the model

635 performance towards entirely unseen data sets (external validation). The model performances for the
636 transferability evaluation were lower than the internal 5-fold CV (decline of 32% R^2 and 18% nRMSE (mean
637 across traits), Fig. 7, Table SE.1). This decline in performance is expected given the large heterogeneity
638 among the data sets (Tab. 1, Fig.4) which might largely differ from the training data, e.g. in terms of sensor,
639 platform, illumination conditions, calibration procedure, trait sampling protocol or vegetation type. Overall,
640 in terms of transferability the CNN_{multiIncomplete} model clearly outperformed the PLSR_{single} model (Fig. 7, SE.1).
641 This may be explained by the larger number of parameters in the CNN-based models, which may facilitate
642 learning more abstract spectral features and to resolve spectral features across different sensor or
643 calibration settings. Both CNN- and PLSR-based traits, whose predictions had higher performances with
644 the random internal CV, corresponded to those that had on average the most accurate prediction with the
645 site transferability evaluation. Similar findings have been obtained in Kothari et al. (2022) at the leaf-scale.
646

647 Overall, the CNN-based transferability across data sets in this study can be considered as relatively high
648 when compared with previous studies. Even at the leaf-level where spectrally-based trait retrieval is
649 generally less challenging than at the canopy-scale, several studies reported similar or larger drops in
650 performance across traits (Serbin et al. 2019, Helsen et al. 2021, Kothari et al. 2022). For instance, the
651 LMA PLSR multi-biome model of Serbin et al. (2019) resulted in R^2 of 0.89 for the internal calibration and
652 dropped to 0.66 when validated externally with LOPEX (Hosgood et al., 1995) and ANGERS data sets
653 (Feret et al., 2008) and to 0.68 with the CABO data set (<https://data.caboscience.org/leaf>, Kothari et al.,
654 2022, Kothari et al. 2022a). Wang et al. (2020) showed a very high model transferability with PLSR models
655 across different vegetation types particularly for LMA and EWT. Likewise, the CNN-based model in
656 Pullanagari et al. (2021) resulted in a robust transferability performance for N retrieval from grasslands
657 where the authors claimed that this can be attributed to the richness of samples from multi-year and multi-
658 site in the training set. However, these studies were based on a consistent sensor and data calibration and
659 processing procedure. This underscores the challenge to train models that are transferable across remote
660 sensing data acquisition settings. However, despite these challenges stemming from the diversity of
661 integrated data sets, the transferability in this study is surprisingly high and we anticipate that with ever
662 increasing data availability more generalized models can be trained in the future.

663

664 Eventually, the transferability of models will depend on the feature space distance between the new, unseen
665 data to the training data (Ludwig et al. 2023). This is confirmed by the model uncertainty estimation
666 procedure developed in this study (Fig. SG1), which is based on this principle and estimates the model
667 uncertainty from the internal CNN embedding, i.e. the feature space viewed from the perspective of the
668 model itself. Such an approach is assumed to be very promising to reveal the area of applicability of a
669 model to new observations and domains (Meyer and Pebesma, 2021).

670

671 *4.5. Outlook*

672 As demonstrated in the present study, multi-trait models may not only facilitate high model performances
673 due to the incorporated trait interrelationships, but also provide a tool to simultaneously and, hence,
674 efficiently track multiple traits from remotely sensed hyperspectral data. The multi-trait approach presented
675 here is expandable to more traits and can continuously be improved as new data become available.
676 Instead of retraining the model from scratch, the model weights can be easily updated by retraining the
677 model on new data. In the near future, a large increase in the availability of hyperspectral and trait data can
678 be expected through the availability of operationally scheduled large-scale hyperspectral observations from
679 spaceborne platforms. This goes along with a generally increased incentive for data sharing by the
680 community and institutions and will include future in-situ and airborne campaigns that contribute to the
681 success of global missions such as PRISMA, EnMAP, CHIME and SBG (Guanter et al., 2015; Labate et
682 al., 2009). Upcoming approaches may also test the integration of simulated data from soil-leaf-canopy
683 RTMs, in the context of hybrid retrieval models (e.g. Wocheer et al., 2022; Verrelst et al., 2021). Such an
684 approach might be particularly promising for traits, vegetation types or states for which only few data are
685 available. In addition, such a physically based approach also takes information about the soil background
686 as well as viewing and observation geometries into account, which may be neglected by empirical
687 approaches.

688

689 **5. Conclusion**

690 From terrestrial platforms up to satellites, hyperspectral remote sensing is advancing as an important tool
691 for future global monitoring applications. Currently, a significant bottleneck to unleash this potential is the
692 lack of scalable and transferable models. Here, we compiled a large and sparse data set with a wide

693 variability in vegetation types and traits. Our results showed that multi-trait CNN models trained on these
694 data can be more performant than CNN models trained for single traits individually. All tested CNN model
695 approaches outperformed widely-used PLSR models. For multiple traits, the model performances obtained
696 using the CNN multi-trait models were on par to those obtained in previous studies – although the model
697 performances here were estimated from a more diverse data set. This highlights that building robust models
698 requires substantial data variability and only a collaborative effort by the remote sensing community can
699 significantly advance our ability to create models that are transferable across sensors, scales, domains,
700 and ecosystems.

701

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703 Eya Cherif: Conceptualization, Methodology, Formal analysis, Data Curation, Writing - Original Draft,
704 Visualization

705 Hannes Feilhauer: Data collection, Conceptualization, Writing - Original Draft, Supervision

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729

730 **Declaration of Competing Interest**

731 The authors declare that they have no known competing financial interests or personal relationships.

732

733 **Code and data availability**

734 The code, the data and trained model objects are available on Gitlab: <https://gitlab.com/eya95/multi->

735 [traitretrieval](#)

736

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