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2 **MODELING THE CO-EVOLUTION OF NATURAL, ECONOMIC AND**
3 **GOVERNANCE SUBSYSTEMS IN INTEGRATED AGRI-ECOLOGICAL**
4 **SYSTEMS: PERSPECTIVES AND CHALLENGES**

5 *Viewpoint article*

6 Gerling, C.^{a*}, Wätzold, F.^a, Theesfeld, I.^b, Drechsler, M.^c, Nixdorf, B.^d, Isselstein, J.^e,
7 Pirscher, F.^b, Rücker, J.^d, Sturm, A.^a.

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9
10 ^a Brandenburg University of Technology Cottbus – Senftenberg, Chair of Environmental
11 Economics, Erich-Weinert-Str. 1, Lecture building 10, Room 532, 03046 Cottbus, Germany

12 ^b Martin-Luther-University Halle-Wittenberg, Von-Seckendorff-Platz 4, 06120 Halle (Saale),
13 Germany

14 ^c UFZ – Helmholtz Centre for Environmental Research, Department of Ecological Modelling,
15 Permoserstr. 15, 04318 Leipzig, Germany

16 ^d Brandenburg University of Technology Cottbus – Senftenberg, Department of Freshwater
17 Conservation, Research Station Bad Saarow, Seestraße 45, 15526 Bad Saarow, Germany

18 ^e Institute of Grassland Science, Georg-August-Universität Göttingen, Wilhelmsplatz 1,
19 37073 Göttingen, Germany

20 * Corresponding author, charlotte.gerling@b-tu.de

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30 **Modeling the Co-Evolution of Natural, Economic and Governance Subsystems in**
31 **Integrated Agri-Ecological Systems: Perspectives and Challenges**

32 *Viewpoint article*

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34 F., Rucker, J., Sturm, A.

35 **ABSTRACT**

36 Current agri-ecological systems face the twin challenge of providing sufficient food for a
37 growing global population and of mitigating severe negative environmental impacts from
38 agricultural land-use in terms of biodiversity loss, greenhouse gas emissions, nutrient run-off,
39 soil degradation, and water pollution. Disciplinary research provides detailed answers to
40 specific questions related to the challenges faced by agri-ecological systems. However, it fails
41 to consider the complex interrelationships and dynamics of the different economic, natural
42 and governance subsystems of which agri-ecological systems consist and that need to be
43 considered to address these challenges. In principle, it is possible to develop models that
44 integrate knowledge from the fields of ecology, economics and governance, and consider
45 dynamic system features such as feedbacks between subsystems as well as tipping points. In
46 this viewpoint article, we scrutinize selected integrated agri-ecological system models and
47 find that only very few models address the challenges mentioned above. We suggest further
48 research in three areas: (I) in-depth integration of the governance subsystem in integrated
49 models, (II) more comprehensive inclusion of tipping points in integrated models, and (III)
50 integration of cascading effects where one system change stimulates another system change.
51 Finally, we briefly discuss the challenges of complex integrated modeling in relation to
52 computational power and the necessity to gather expertise from different disciplines.

53 Keywords: agri-ecological systems; integrated modeling; socio-ecological systems

54 **HIGHLIGHTS**

- 55 - Agri-ecological systems are highly complex
56 - Their analysis requires integration of diverse disciplines
57 - Models are crucial for this purpose
58 - Future research may include tipping points, cascading effects, governance subsystem

59 1. INTRODUCTION

60 Current agri-ecological systems face the challenge of providing food security for an estimated
61 9 billion people worldwide by the middle of the century (Godfray et al. 2010). At the same
62 time, they generate severe negative environmental impacts such as biodiversity loss
63 (Tscharrntke et al. 2012), greenhouse gas emissions (Godfray et al. 2010; Lal 2004; West &
64 Marland 2002), nutrient run-off (Korsaeth & Eltun 2000), soil degradation (Godfray et al.
65 2010; Lal 2004), and water shortages and pollution (Godfray et al. 2010) which need to be
66 mitigated. Disciplinary research provides detailed answers to specific questions related to the
67 challenges faced by agri-ecological systems. However, it fails to consider the complex
68 interrelationships and dynamics of the different economic, natural and governance subsystems
69 that interact. These interactions need to be considered jointly to provide comprehensive
70 solutions to the enormous food production and environmental challenges that confront agri-
71 ecological systems (Doyen et al. 2013; Pulver et al. 2018; van Riper et al. 2018).

72 Whereas natural scientists are becoming increasingly familiar with economic knowledge
73 (Cooke et al. 2009), there is often little understanding about the realm of governance. We
74 understand the governance subsystem here to include the full range of regulatory processes,
75 social mechanisms and organizations through which political actors influence agri-ecological
76 actions and outcomes. Although the action of state actors plays a considerable role, according
77 to Lemos and Agrawal (2006), non-state actors' power and mechanisms to regulate, manage
78 and steer collective action have to be considered too. Governance therefore includes all
79 coordination mechanisms that guide the behavior of individuals, groups and organizations in
80 formal (e.g. through ordinances and regulations) and informal ways (e.g. through social and
81 cultural norms). Coordination may be organized by hierarchical-driven modes of governance
82 constraining, supporting or supervising actors or sanctioning actors' non-compliance but also
83 by information and knowledge sharing mechanisms, monitoring mechanisms, conflict
84 resolution mechanisms and self-governance networks. Often, market relations are also
85 considered to be part of governance (Lemos & Agrawal 2006). However, given the explicit
86 distinction made here between the economic and governance subsystem we consider that they,
87 together with the agri-ecological production system, make up the economic subsystem.

88 To address the complexity of agri-ecological systems, knowledge from different disciplines
89 needs to be integrated (Mauser et al. 2013), and models may be useful for this (Wätzold et al.
90 2006; Filatova et al. 2016; Polhill et al. 2016a). Statistical, system dynamics, equilibrium and
91 agent-based models are already commonly used to model complex socio-environmental
92 systems (Filatova et al. 2016), and they are also useful for modeling agri-ecological systems.

93 Yet, there are two important aspects that add further complexity to the study of agri-
94 ecological systems and impede their proper understanding if not considered. First, the various
95 natural, economic and governance subsystems are intricately linked via feedback loops.
96 Feedback loops describe situations in which a change in a subsystem causes a change in a
97 different subsystem. However, this change triggers a reaction back in the original subsystem,
98 and thus a feedback-loop emerges (Wätzold et al. 2006). For example, unsustainable cropping
99 practices (economic subsystem) can lead to soil compaction (natural subsystem), which in
100 turn reduces the options for productive use (economic subsystem) in the future.

101 Second, marginal alterations in the subsystems frequently drive the system towards a tipping
102 point, after which the system changes rapidly (Crépin et al. 2012; Vandermeer & Perfecto

103 2012). A tipping point happens when a system exhibits threshold behavior where at some
104 point incremental changes cause a change in the state of the system (Crépin et al. 2012) with
105 limited reversibility (Brook et al. 2013). Typically, tipping points refer to changes in natural
106 systems, but social systems, in particular their governance subsystems, can also reach tipping
107 points (Theesfeld & MacKinnon 2014) – passing a threshold that leads to a new governance
108 structure to manage a natural resource. For example, public concern over the environmental
109 impact of a pesticide can reach such a high level of political relevance that the pesticide is
110 banned. This policy can hardly be reversed, even if it turns out that the pesticide is less
111 harmful than previously thought. Reasons for this de facto irreversibility may be path
112 dependence and bureaucratic inertia (Dobusch & Kapeller 2013). Sometimes tipping points
113 also interact (Filatova et al. 2016) as cascading effects, where one threshold triggers the
114 crossing of other thresholds (Kinzig et al. 2006).

115 An example of where interacting feedback loops and tipping points may occur in agri-
116 ecological systems is the high release of nitrogen compounds into the environment due to the
117 intensive application of nitrogen fertilizer, manure and – more recently – the fermentation
118 residues from biomass production (Weiland 2010). Tipping points occur: (I) from a
119 governance perspective, if water can no longer be used as drinking water due to the violation
120 of maximum permissible values of nitrate (Kastens & Newig 2007), and (II) from an
121 ecological perspective if agri-ecological systems lose their suitability as a habitat for specific
122 species and provider of ecosystem services due to eutrophication in areas that are particularly
123 vulnerable to nitrogen inputs (Glibert et al. 2014). A policy reaction to increasing nitrogen
124 fertilization might be to implement a regulation to curb nitrogen emissions accompanied by
125 comprehensive documentation duties with resulting cost increases for farms. This in turn may
126 feed back to the farming system with partly intended, partly unintended consequences. An
127 intended consequence would be that farms which manage their land intensively have to curb
128 emissions. An unintended consequence would be cost increases for all farms due to the new
129 documentation duties, including farms with low profitability on extensively managed
130 marginal land. This added difficulty may be enough to cause some of these farms to close
131 down resulting in an abandonment of marginal farm land. This negative feedback loop has an
132 unwanted effect, as extensively managed farmland is often highly valuable for biodiversity
133 (Isselstein et al. 2005).

134 In order to comprehensively understand the dynamics of agri-ecological systems and develop
135 appropriate policy recommendations, it is therefore necessary to develop models that combine
136 natural, economic and governance analyses (henceforth referred to as integrated models) and
137 to consider tipping points and feedback loops between the governance, economic and natural
138 subsystems. Reviews on modeling socio-ecological systems (Schlüter et al. 2011; An 2012,
139 Polhill et al. 2017; Schulze et al. 2017) and tipping points (Filatova et al. 2016) exist.
140 However, this work is not specific to agri-ecological systems and does not consider integrated
141 modeling jointly with tipping points and feedback loops.

142 The purpose of this paper is to scrutinize selected research in agri-ecological system modeling
143 focusing on the three fields of (1) integration of different disciplines in agri-ecological
144 models, (2) consideration of feedback loops, and (3) consideration of tipping points. Given
145 the framework of a viewpoint article and the broadness of the three research fields, we
146 refrained from carrying out a systematic literature review. We have rather drawn on the
147 expertise of the authors who cover relevant disciplines for these three fields, including
148 environmental economics, political science, ecology, agronomy, and integrated modelling.

149 Each author brought in his or her individual experience but has also done a careful literature
150 screening in their respective field. This literature screening was conducted using different
151 methods, including electronic searches with keywords relevant to each field on Google
152 Scholar and the Web of Knowledge, snowballing, and e-mail alerts on new research published
153 in relevant journals. A broad range of model types has been considered because of the
154 different definitions of what constitutes a “model” in the different disciplines (cf. Drechsler et
155 al. 2007).

156 We identified areas for which we suggest that more research is fruitful and provide
157 recommendations for further work to successfully address the co-evolution of natural
158 economic and governance subsystems in complex agri-ecological systems. Finally, we
159 address two key challenges for such a critical endeavor: model complexity and the necessity
160 to integrate expertise from different disciplines.

161 2. INTEGRATED MODELING IN AGRICULTURAL SYSTEMS

162 Regarding models that integrate knowledge from economics and natural sciences, there is a
163 growing body of literature that has been developed over the past twenty years or so (cf. Vatn
164 et al. 1999 for an early example). Many of these ecological-economic or hydrological-
165 economic models gather economic and natural scientific information or data on small spatial
166 units. This information is then combined in a spatially explicit manner to simulate policies
167 and other drivers or to optimize agri-environment measures and policy instruments. Following
168 this general approach, Wätzold and Drechsler (2014) developed modeling procedures to
169 spatially optimize agri-environment schemes, and Wätzold et al. (2016) to spatio-temporally
170 optimize them. Spatial optimization of agri-environment measures is also an important
171 research topic in the field of water pollution by nutrients (Konrad et al. 2014). An example of
172 research applying models integrating economics and natural sciences in other areas is Esteve
173 et al. (2015), who integrated spatially explicit economic, hydrological and climate information
174 to evaluate climate adaptation measures in agriculture.

175 In contrast to models that integrate economic and natural sciences subsystems, much fewer
176 models exist that explicitly integrate the governance subsystem. However, a few models try to
177 analyze the impacts of the governance subsystem on the natural subsystem or the agricultural
178 production unit (Archer et al. 2008; Hendrickson et al. 2008; Daloğlu 2014; Hauck et al.
179 2019). Others make isolated governance aspects measurable and operational. Examples
180 include an aggregated farmers’ mental model (Jabbour et al. 2014), and the measurement of
181 power aspects in agricultural water distribution (Theesfeld 2011). While making these factors
182 measurable is an important step towards including them in models, the above-mentioned
183 studies do not develop models.

184 The few agri-ecological models that explicitly include the governance subsystem typically
185 consider economic and, less frequently, natural subsystems. Happe et al. (2006) used a meta-
186 modeling approach, based on agent-based modeling, to analyze impacts of reform of the EU’s
187 Common Agricultural Policy on farmers’ production decisions and agricultural structural
188 change. Schouten et al. (2013) applied an agent-based model to analyze farmers’ decision-
189 making regarding conventional milk production and the implementation of agri-environment
190 measures that benefit biodiversity. The governance aspect examines a flexible compensation
191 payment for an agri-environment measure based on the location of a land parcel. Depending
192 on location, the ecological benefit may be higher or lower. Finally, the economic subsystem is

193 included when examining the impact of price volatility. Dace et al. (2015) developed and
194 applied a system dynamics model to assess the impact that policy measures have on
195 greenhouse gas emissions from agri-ecological systems, thereby integrating the governance,
196 ecological and economic subsystems.

197 Overall, we found that the large majority of agri-ecological models that integrate knowledge
198 from different disciplines combine knowledge from the natural sciences with economic
199 knowledge. Some of these models also address aspects relevant to the governance subsystem
200 such as the simulation and optimization of agri-environment policies. However, these models
201 implicitly adopt the perspective of a social planner assumed to be willing and able to optimize
202 policies (Sugden 2013). While such research is relevant to understand the direct impact of
203 policies it is obviously a simplistic view that does not consider the richness of the political
204 actors, processes and mechanisms that make up the governance subsystem (Lemos and
205 Agrawal 2006), and the resulting interactions with the economic and natural subsystems. Only
206 few models exist that combine natural sciences and governance knowledge or knowledge
207 from all three subsystems.

208 3. FEEDBACK LOOPS IN INTEGRATED MODELING

209 Some studies model feedback loops to connect different subsystems that are traditionally
210 investigated by different disciplines. Elshafei et al. (2015, 2016) modeled feedback loops to
211 connect natural and human subsystems in their models on agriculturally used catchment areas.
212 They developed a model considering catchment hydrology, population, economic and
213 environmental indicators, quality of life and collective responses. Dace et al. (2015) included
214 different feedback loops to connect natural, economic and governance subsystems. Zavalloni
215 et al. (2015) discussed the interaction between farmers and tourists, and their response to rural
216 policies. One such policy rewards “green farming”, which in turn leads to increased tourism.
217 This increase leads to additional income for farmers in areas of “green farming”. Depending
218 on this benefit, a loop feeds back to the farmer’s decision-making for the next period, as
219 depending on the economic benefit of increased tourism farmers decide whether or not to
220 expand “green farming” in the next period.

221 Often feedback loops are used to integrate human decision-making in models integrating at
222 least two disciplines. In a seminal paper, Berger (2001) developed an agent-based spatial
223 model to analyze household-level decisions related to technology diffusion, resource use
224 changes and policy in the field of water irrigation in agriculture. In the model, individuals try
225 to maximize their expected income taking into account previous experiences and the feedback
226 they receive. Le et al. (2012) also used an agent-based model to study land-use change. This
227 model includes feedback loops between the human and environmental subsystems in the short
228 and long term. In short-term feedback loops, individuals react according to their environment,
229 while the long-term feedback loops describe cumulative, longer-term changes on a larger
230 scale and include learning effects. Similarly, Tsai et al. (2015) used an agent-based model to
231 examine land-use changes. Feedback loops are included as land users adapt their behavior
232 depending on the outcome of the previous period.

233 Huber et al. (2013) also considered human decision-making by modeling the implementation
234 of payments for environmental services based on feedback loops between vegetation
235 dynamics, farm-based decision-making, and policy decisions on the national level. Henderson
236 et al. (2016) developed a model to study the mosaic landscapes of forest, grassland and

237 agricultural land in Brazil. Based on penalties, profits and conservation values, humans
238 choose certain actions which influence the environment. This in turn influences the penalties,
239 profits and conservation values that underlie human decision-making. Finally, Chapman and
240 Darby (2016) developed a system dynamics model to assess rice cultivation strategies. This
241 model includes physical, economic, and decision-making characteristics linked by loops
242 connecting aspects such as fertilization, yield, nutrients, crop selection and technical capacity.

243 In summary, we found that feedback loops are mainly used for two purposes in integrated
244 models of agri-ecological systems: to link subsystems that are traditionally studied by
245 different scientific disciplines and to integrate human decision-making in the agri-ecological
246 system models.

247 4. TIPPING POINTS IN INTEGRATED MODELING

248 One main area where tipping points are considered in agri-ecological system models relates to
249 climate change and agricultural practices. Kenny et al. (2000) addressed the issue of climate
250 thresholds in their paper on modeling the impacts of climate change on certain agricultural
251 produce in New Zealand. However, the authors conclude that determining a critical threshold
252 is not straightforward and requires more research. Nelson et al. (2014) combined climate and
253 biophysical models with different economic models in order to investigate the role of
254 economic models in interdisciplinary analyses of the uncertainty of the impact of ‘climate
255 shocks’ on agriculture. They find that the inclusion of (different) economic models influences
256 the scale of the considered climate change impact.

257 The other main area where tipping points are taken into account relates to pollution thresholds
258 as tipping points in agri-ecological systems. Stoorvogel et al. (2004) developed a model of
259 agri-ecological systems that can be used to calculate the probabilities of crossing pesticide
260 thresholds as defined by the United States Environmental Protection Agency. The model
261 represents land-use and input use decisions based on changing biophysical and economic
262 conditions. Bennett et al. (2008) combined multiple models predicting the risk of exceeding
263 pollutant (phosphorus) thresholds and find that this combination leads to much better
264 prediction than the use of the single best model. Similarly, Wang and Burke (2017) developed
265 a model that facilitates the understanding of local nitrate-legacy issues for policy makers. It
266 specifically addresses long-term impacts to support water quality compliance, where passing
267 the regional compliance level can be considered a tipping point.

268 Some research also focuses on other topics. Tzanopoulos et al. (2007) developed a grazing
269 model, where crossing a threshold of grazing intensity leads to different vegetation structures.
270 They analyze at which grazing intensity the threshold is crossed and predict the likely impact
271 of policy reforms. Concerning prevailing governance structures, Theesfeld and McKinnon
272 (2014) discussed governance tipping points in agricultural water management in Wyoming.
273 The tipping point occurs where pending management reforms are held back until they allow
274 the pent-up pressure of unmet climate adaptation needs to push the governance subsystem
275 over a threshold. Nyborg et al. (2016) found that there are social behavior tipping points that
276 turn a vicious cycle of a damaging behavior into a virtuous one. They developed a model to
277 investigate whether policies, based on observability of the behavior, or informal institutions
278 such as social norms are better suited to push society towards crossing a desirable tipping
279 point.

280 Finally, only very few models integrate both tipping points and feedback loops. Lafuite and
281 Loreau (2017) developed a model to determine thresholds for land conversion from
282 biodiverse ecosystems to biodiversity-poor agricultural landscapes depending on human
283 population size. These thresholds are linked via feedback loops in the economic,
284 demographic, technological and ecological subsystems. Passing these thresholds leads to
285 ecological crises. Fernald et al. (2012) modeled tipping points of sustainable and
286 unsustainable water usage in traditional irrigation communities. Their analysis covers
287 hydrological, ecosystem, land-use, economic, and sociocultural subsystems which are linked
288 via feedback loops. Hossain et al. (2017) developed a system dynamics model to analyze the
289 safe operating space on a regional scale beyond which human wellbeing in the case study area
290 decreases dramatically, considering for example income and predicted deaths from natural
291 disasters. The model includes complex factors such as non-linearity, feedbacks and
292 interactions. In this model, they include the impacts of natural subsystems (i.e. predicted
293 changes in the climate and water subsystems) and economic subsystems (including changes in
294 subsidies, as well as indicators such as income and gross domestic product).

295 Although it has been recognized that tipping points are important when modeling agri-
296 ecological systems (Vandermeer & Perfecto 2012), we found that only few models consider
297 them and those that do mainly focus on two topics: interactions between climate change and
298 agricultural practices and pollution thresholds in agri-ecological systems. Finally, there are
299 only very few models that jointly consider tipping points and feedback loops. An analysis
300 where tipping points interact as cascading effects has not been carried out in any of the
301 integrated models investigated by the authors.

302 5. DISCUSSION: KEY CHALLENGES AND WAYS FORWARD

303 We have provided insights into the last two decades of research considering the co-evolution
304 of the natural, economic and governance subsystems that make up complex agri-ecological
305 systems. We focused on integrated modeling, the modeling of feedback loops and of tipping
306 points. Based on our scrutiny of the literature, we think that three areas would particularly
307 benefit from further research.

308 (I) Given our finding that integrated models that include the governance subsystem are rare,
309 together with the fact that governance has a key impact on agri-ecological systems (von Braun
310 & Birner 2016), we see particular benefits in developing models that consider the interactions
311 of the governance subsystem with the economic and natural subsystems. For this purpose,
312 knowledge from a broad range of disciplines including sociology, psychology and political
313 science with relevance for the governance subsystem may be integrated into the models. This
314 knowledge may allow analyzing aspects that have hitherto been neglected in integrated
315 modeling such as the impact of cultural traditions or nudging on farmers' behavior and hence
316 the natural subsystem. Generally, the integration of the governance subsystem in integrated
317 models would enable new dynamic analyses, which could lead to better tailored policy
318 recommendations. Such integrated models could, for example, provide quantitative
319 assessments of the ecological impacts of changes in the governance structure that are
320 conveyed through production changes in the economic subsystem and manifest themselves in
321 the ecological subsystem. Quantitative assessments cannot be undertaken from a purely
322 governance perspective as they require a detailed understanding of the economic and natural
323 subsystem. However, quantitative analyses are highly relevant for formulating policies with

324 adequate incentives that lead to sufficient food production and consider societal demands for
325 environmentally sound production.

326 (II) We further found that the analysis of tipping points is limited to few research fields and,
327 additionally, rarely considers the governance subsystem. However, tipping points are
328 important phenomena in agri-ecological systems and their emergence can only be
329 comprehensively understood and avoided if the governance subsystem is considered.
330 Similarly, Filatova et al. (2016) conclude that tipping point models rarely integrate all aspects
331 commonly considered important in empirical studies. Therefore, we think that developing
332 integrated models to better understand the emergence of tipping points is a promising avenue
333 of further research. In our opinion, and in line with Polhill et al. (2016b), tipping-point models
334 should include the possible influence of gradual, internal dynamics and exogenous
335 disturbances that may occur in the natural as well as the economic and governance
336 subsystems. Related to that we see a further benefit if governance research also employs the
337 concept of tipping points at a conceptual level, although we also acknowledge that models of
338 policy change include ideas of tipping dynamics, for example, the punctuated equilibrium
339 framework of policy change (Kuhlmann & van der Heijden 2018). Using the same
340 terminology may facilitate interdisciplinary research in the field of agriculture, as other
341 relevant disciplines (such as ecology) already apply the concept of tipping points. In turn,
342 their research may stimulate further research in other disciplines as governance scientists may
343 come up with novel ideas about the conceptualization and emergence of tipping points.
344 However, one needs to be careful when different disciplines use the same terminology in
345 order to avoid misunderstandings (Drechsler et al. 2007).

346 (III) We find that cascading effects, where the crossing of one system's threshold triggers the
347 crossing of thresholds in other systems triggering a "domino effect", are rarely implemented
348 in integrated models in agri-ecological systems research. However, only by considering
349 cascading effects is it possible to understand the whole impact of a policy or other drivers.
350 The consideration of only the direct tipping point resulting from a driver may address some
351 problems, but vastly underestimate the driver's overall impact if cascading effects are not
352 considered. Therefore, in our opinion, a further promising approach is to include cascading
353 effects in integrated modeling of agri-ecological systems. Such research may help to detect at
354 an early stage important environmental consequences of drivers that would not be discovered
355 if research were to focus only on one tipping point and its direct consequences.

356 Complex models integrating natural, economic and governance subsystems and considering
357 factors of complexity such as feedback loops, tipping points and cascading effects may
358 greatly improve the relevance of models to understand the dynamics of agri-ecological
359 systems and to develop policy-relevant recommendations to overcome the challenges they
360 face. However, one has to consider trade-offs between an improved understanding of system
361 complexity and the associated cost increases in terms of data collection (cf. Jones et al. 2017)
362 and modelling efforts that come with higher model complexity. We wish to highlight two key
363 challenges regarding modelling efforts.

364 1. Model complexity. Integrating natural, economic and governance subsystems in agri-
365 ecological system models is likely to increase model complexity considerably, since not only
366 each of the three system components must be modeled adequately but also the interactions
367 between those components. In general, high model complexity hampers the systematic
368 analysis of a model, as it becomes increasingly difficult to understand the behavior of the

369 model and the modeled system. However, the growth in computing power over recent decades
370 and the development of better tools and approaches for model analysis now allow the
371 development and analysis of models that some time ago would have been deemed too
372 complex.

373 We highlight two examples where the growth in computational modeling power has made
374 more complex modeling approaches possible: inverse modeling and sensitivity analysis.
375 Inverse modeling allows us to estimate uncertain model parameters (whose number may
376 appear prohibitively large in complex models) by comparing model outputs with observed
377 system behavior. An intuitive variant of inverse modeling is pattern-oriented modeling (see
378 Grimm et al. (2005) for general considerations and Magliocca & Ellis (2013) for a recent
379 application), in which as many as possible macroscopic patterns like the spatial distribution of
380 a species, a land-use pattern in response to some policy instrument, or the decision of a policy
381 maker are generated by the model. Regarding socio-economic patterns, examples include the
382 observable behavior of land users (Nelson et al. 2008). Since these patterns are likely to
383 depend on the model parameters, different combinations of (uncertain) model parameter
384 values will lead to different patterns. Model parameter combinations that generate patterns
385 similar to the observed ones are likely to be the true ones valid in the real system. In a
386 sensitivity analysis (Saltelli et al. 2009), the behavior of a model is explored in response to
387 variations in the model parameters, which allows the development of understanding and
388 system predictions even if model parameters are uncertain. Statistical methods like
389 classification (e.g. Surun & Drechsler 2017) or regression trees (e.g. Guisan & Zimmermann
390 2000) further enhance the effectiveness of sensitivity analysis in the analysis of complex
391 models.

392 2. A high degree of expertise in several disciplines and appropriate cooperation. If the
393 different subsystems of governance, economics and ecology are to be integrated, a wide range
394 of experts need to be involved in developing the models. Including what is perceived
395 knowledge from other disciplines without relying on experts from that field should be avoided
396 (Armsworth 2014). Otherwise, the opportunity to make full use of the richness of knowledge
397 that exists in a specific discipline is likely to be missed (Wätzold et al. 2006). However,
398 researchers from different disciplines have distinct educational backgrounds and are trained to
399 analyse problems from different perspectives. For interdisciplinary work, researchers will
400 have to become aware of and overcome problems caused by different, unconscious patterns of
401 thought (Bauer 1990). For example, different questions regarding a common problem may be
402 considered interesting by the distinct disciplines, the underlying assumptions of this problem
403 may vary between the disciplines, and differences in epistemology may result in contrasting
404 methods being acceptable (Lélé & Norgaard 2005). Against this background, it is problematic
405 if researchers from one discipline develop interdisciplinary research projects, and other
406 disciplines are added as “add-ons” because, for example, their participation is required by a
407 call from a funding agency (Wätzold et al. 2006). This “add-on approach” may not capture all
408 possible benefits that could be realized by developing projects jointly from the beginning.
409 Moreover, different disciplines may use distinct concepts of terminology, scale and models
410 (Drechsler et al. 2007). If these differences are not addressed from the beginning, frustration
411 may lead to researchers avoiding interdisciplinary research (Wätzold et al. 2006).

412 Considering the subsystems of ecology, economics and governance, most integrated models
413 are developed jointly by natural scientists, in particular ecologists, and by economists.
414 Cooperation is facilitated because both disciplines have a strong tradition in mathematical

415 modeling, such as the modeling of ecosystem processes or population viability analysis in
416 ecology, and the modeling of the behavior and interaction of economic agents in economics.
417 Furthermore, researchers in both disciplines have made attempts to integrate knowledge from
418 the respective other discipline (Cooke et al. 2009). Agricultural economists, for example,
419 integrate knowledge from the natural sciences when applying the production function
420 approach, which requires natural science knowledge to quantify the impact of changes in the
421 natural system on yield (cf. Brady et al. 2015 for an example). In the field of conservation
422 biology, a prominent question is how to optimally allocate conservation reserves to maximize
423 the number of protected species under a budget constraint, which is heavily based on the
424 economic concept of cost-effectiveness (Kukkala & Moilanen 2013). While ecologists and
425 economists do share a number of central concepts, researchers working in the field of
426 governance typically rest on different concepts and address problems that cannot be easily
427 captured by mathematical modeling or mathematical optimization. Hence, they apply
428 different approaches such as qualitative research or statistical methods.

429 There are, however, concepts that may be shared by all three disciplines. One example is the
430 concept of “tipping points”. This term was originally used in studies on racial segregation
431 (Grodzins 1958). After publication of the book “The tipping point” by Malcolm Gladwell
432 (Gladwell 2000), the use of the term in the scientific literature has multiplied. It is now mainly
433 used by natural scientists, but the diverse examples given in Gladwell’s book, including
434 fashion trends, crime rates and the financial crisis show that the concept is highly relevant for
435 social scientists too (van Nes et al. 2016).

436 Researchers in the field of governance also work towards developing measurable proxies to
437 evaluate certain governance structures that may be integrated in models. For instance, the
438 governance aspect “bargaining power of farmer associations” may be expressed by the
439 indicator “membership in farmer associations” which is easy to measure as a ratio of number
440 of farmers that are members in a farmer association / number of farms * 100. This might be
441 accompanied by measures on the importance of the agricultural sector in an economy (ratio of
442 agricultural area / total area and agricultural employment / total employment, cf. Theesfeld et
443 al. 2010). Similarly, according to Rabinowicz and Swinnen (1997), the use of the indicator
444 “share of agricultural votes” can be used as an indicator to assess the bargaining power of
445 farmers’ associations. Further, running an associations office in Brussels (“yes” or “no”) can
446 be included as a binary variable to assess the same aspect.

447 In addition to the development of such quantifiable indicators, several other developments
448 may facilitate the future integration of the governance subsystem in integrated models and the
449 development of such models in general. First, there is an increasing use of agent-based
450 models. These models are able to integrate knowledge from different disciplines and to
451 capture ideas based on behavioral theories that go beyond profit-maximizing behavior of
452 agents traditionally assumed in economics (Schlüter et al. 2017). This may lower the barriers
453 for governance research towards participation in integrated modelling. Second, today many
454 calls from funding agencies require an interdisciplinary consortium. In our opinion, this has
455 helped, over time, to improve the understanding between disciplines, as only well-composed
456 and balanced project proposals have a chance of being successful. Finally, it is our impression
457 that younger researchers are more open and curious towards integrated research than previous
458 researcher generations.

459 Despite the described challenges, we strongly argue for the development of complex models
460 that combine the natural, economic and governance subsystems and include feedback loops,
461 tipping points and cascading effects. Developing such models has the potential to greatly
462 enhance our understanding of agri-ecological systems and our capability to address the huge
463 challenges they face.

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