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1 **Title: Spatial patterns of extracellular enzymes: combining X-ray computed micro-**  
2 **tomography and 2D zymography**

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34 **Abstract**

35           Linking the distribution of enzyme activity to the size and properties of soil pores is a  
36 necessary prerequisite for mechanistic understanding of soil biochemical processes. In this study  
37 we used soil 2D zymography and X-ray computed tomography ( $\mu$ CT) to assess the relationship  
38 between enzymes and pores. The objectives of the study were (i) to assess spatial distribution  
39 patterns in the activity of six enzymes contributing to C, N and P cycles, namely,  
40 cellobiohydrolase,  $\beta$ -glucosidase, xylanase acid phosphatase, leucine aminopeptidase, and N-  
41 acetylglucosaminidase, in soils from five long-term land use and management practices, (ii) to  
42 study the correlation between enzyme activities and  $\mu$ CT information, i.e., pore characteristics  
43 and image grayscale values, and (iii) to explore the potential use of soil 2D zymography in  
44 predicting enzyme activities within 3D soil cores. 3D pore-size distributions were obtained from  
45  $\mu$ CT images of 13 intact soil cores and then 8-15 2D zymography maps were taken from each  
46 core. Spatial distributions in the activities of all studied enzymes were auto-correlated; the spatial  
47 correlation ranges were equal to  $\sim$ 7-8 mm. The relative activity of all enzymes was positively  
48 associated within 60-180  $\mu$ m  $\varnothing$  pores. Combining 3D  $\mu$ CT information with 2D zymography  
49 maps visualized the overall patterns of enzyme activity distributions with respect to soil pores  
50 and particulate organic matter locations. Based on the findings we propose a conceptual scheme  
51 relating localization of microorganisms, enzymes and substrates to pores of different size ranges.  
52 Specifically, we suggest that pores in the tens of microns size range represent optimal microbial  
53 habitats, and as such are associated with greater microbial abundance, leading to high enzyme  
54 production and activity.

55

56

57 **Keywords:** Enzyme activities, Spatial statistics, Soil zymography, Pore distribution, Microbial  
58 habitats, Microorganisms' localization

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## 63 **1. Introduction**

64 Extracellular enzymes (EEs) produced by roots and microorganisms in order to meet their  
65 nutrient and energy demands, play a major role in biochemical processes, including soil organic  
66 matter transformations (Burns et al., 2013). Yet, despite substantial efforts in studying EEs, the  
67 drivers of their production and subsequent fate in soil are not fully understood (Nannipieri et al.,  
68 2012; Burns, 2013). One of the reasons is the extremely high spatial heterogeneity of EE activity  
69 (Baldrian, 2014) which can differ by as much as an order of magnitude within a distance of only  
70 a few millimeters (Razavi et al., 2016; Ma et al., 2017; Kuzyakov and Razavi, 2019). At a field  
71 scale (10 – 100 m), EE spatial patterns are related to patterns in soil fertility (Banerjee et al.,  
72 2016), landscape topography (Wickings et al., 2015; Mganga et al., 2016), and land use and  
73 management practices (Stursova and Baldrian, 2011; Baldrian, 2014). At a scale of 10  $\mu\text{m}$  – 10  
74 mm, spatial patterns in EE are related to fungal or bacterial colony sizes (Baldrian and  
75 Vetrovsky, 2012) and the activity of enzyme producers (Banerjee et al., 2016; Hoang et al.,  
76 2016b; Stursova et al., 2016; Navratilova et al., 2017). Hot-spots of EE presence are not only  
77 associated with microbial colonies, but also with soil macro- and micro-fauna (Hoang et al.,  
78 2016a; Hoang et al., 2016b), plant roots (Razavi et al., 2016; Ge et al., 2017; Razavi et al., 2017)  
79 and plant residues (Hoang et al., 2016b; Liu et al., 2017), as well as influxes of fresh organic  
80 inputs (Heitkotter and Marschner, 2018). To complicate matters, EEs can react with organic  
81 sources and become anchored in the soil matrix while still preserving a certain degree of activity  
82 (Nannipieri et al., 2012; Burns, 2013). Moreover, activity of EEs can last surprisingly long after  
83 the disappearance of their original microbial producers (Schimel et al., 2017). Quantifying the  
84 spatial variability patterns in EE and linking them with those of soil and root characteristics and  
85 microorganisms is crucial for understanding what drives EE activities and functions.

86 The majority of previous studies related EE to soil biological and/or chemical properties  
87 as well as to substrate inputs. Chemical and biological processes occur within the physical frame  
88 defined by the soil pore network (Young and Crawford, 2004; Or et al., 2007; Tecon and Or,  
89 2017), however, the influences of soil pores on EE have so far been largely overlooked. We  
90 hypothesize that variations in soil physical properties, especially the pore presence and  
91 characteristics, can also contribute to the distribution of spatial patterns in EEs.

92 Pores can impact EE distribution within the soil matrix via several mechanisms. They can  
93 affect spatial patterns in inputs of the substrates for microbial decomposition directly, i.e., by

94 driving localization of roots, rhizodeposits, and earthworms (Baldrian et al., 2010a; Baldrian et  
95 al., 2010b; Athmann et al., 2017; Banfield et al., 2017a; Banfield et al., 2017b; Navratilova et al.,  
96 2017); as well as indirectly, i.e., by influencing diffusion and convective transport of soluble  
97 organic compounds (Allison, 2005). Pores can also define micro-environmental conditions, e.g.,  
98 water regime and O<sub>2</sub> supply (Keiluweit et al., 2016; Keiluweit et al., 2018), which in turn  
99 influence the ability of microorganisms to function and produce EEs. Moreover, pores provide  
100 the physical space necessary to host microbial colonies, which can range in size from a few  
101 dozen to some hundreds of μm (Nunan et al., 2003). The combination of these mechanisms can  
102 result in the creation of optimal microbial habitats. Such habitats largely determine the regions of  
103 microbial EE production; and higher EE activity can be expected to correspond to these prime  
104 habitats.

105 Visualization of the connections between EE and soil pores could provide valuable input  
106 in understanding the *in situ* biochemical processes taking place within an intact soil matrix. One  
107 of the possible techniques is coupling soil 2D zymography, which enables EE activity mapping  
108 (Razavi et al., 2016), with X-ray computed micro-tomography (μCT), which allows for the 3D  
109 characterization of soil pores (Peth et al., 2008; Helliwell et al., 2013). Combining of 3D X-ray  
110 μCT information with 2D soil data was pioneered by Hapca et al. (Hapca et al., 2015), who  
111 created 3D maps of soil elements' contents using 2D SEM-EDX.

112 We considered two types of μCT data of potential use in EE spatial variability  
113 predictions: (i) the presence and abundance of pores of different sizes and (ii) the grayscale  
114 values of μCT soil images. Grayscale values reflect the attenuation of X-rays as they pass  
115 through scanned material. The attenuation within solid soil matrix reflect variations in soil  
116 mineralogy, presence of pores with sizes below the image resolution, and presence of organic  
117 materials, e.g., plant residues and particulate organic matter (POM). They also highlight areas  
118 with high soil organic matter (SOM) levels (Kravchenko et al., 2014; Quigley et al., 2018a).

119 In this study we explored mm-scale, i.e. one to tens of mm, spatial patterns in the  
120 distribution of six EEs involved in soil C, N, and P cycling: cellobiohydrolase, β-glucosidase,  
121 xylanase, acid phosphatase, leucine aminopeptidase, and N-acetylglucosaminidase (chitinase  
122 NAG). Cellobiohydrolase and β-glucosidase are involved in consecutive stages of cellulose  
123 degradation (German et al., 2011). Xylanase is responsible for breaking down hemicelluloses  
124 (German et al., 2011). Acid phosphatase mineralizes organic P into phosphate by hydrolyzing

125 phosphoric (mono) ester bonds under acidic conditions (Eivazi and Tabatabai, 1977; Malcolm,  
126 1983; German et al., 2011). Leucine aminopeptidase facilitates hydrolysis of leucine residues  
127 from the amino-termini of protein or peptide substrates (Rawlings et al., 2004). N-  
128 acetylglucosaminidase (NAG) decomposes chitin to low molecular weight chitooligomer  
129 (Baldrian and Stursova, 2011), and decomposes bacterial peptidoglycan. We explored EE  
130 activity in soil from several long-term land use and management practices, which over time  
131 developed substantial differences in their SOM levels and pore characteristics (Kravchenko et  
132 al., 2018).

133 Our objectives were (i) to assess the spatial variability of the activity of six EEs in intact  
134 soil cores, which represent a diverse range of long-term land use and management practices, (ii)  
135 to study correlations between EE activity and physical soil properties, i.e., pore characteristics  
136 and grayscale values from X-ray  $\mu$ CT scanning, and (iii) to explore the potential of using soil 2D  
137 zymography to predict enzyme activity within 3D soil cores with/without  $\mu$ CT information.

138

## 139 **2. Materials and methods**

### 140 *2.1. Land use and management systems*

141 The five studied land use and management systems are a part of the Great Lake  
142 Bioenergy Center experiment, Kellogg Biological Station, Michigan, USA. The experiment was  
143 established in 2008. The soils of the site are well-drained Alfisols of Oshtemo and Kalamazoo  
144 series (mesic Typic Hapludalf) (Robertson and Hamilton, 2015). The experimental design is a  
145 randomized complete block with five replicated 0.12 ha experimental plots randomly assigned to  
146 each land use system. The five studied systems are: continuous corn (*Zea mays* L.) (G1) and  
147 continuous corn with winter cover crop of cereal rye (*Secale cereale* L.) (G2), a monoculture  
148 switchgrass (*Panicum virgatum* L.) (G5), a hybrid poplar (*Populus nigra*  $\times$  *P. maximowiczii*  
149 ‘NM6’) with herbaceous understory (Sprunger and Robertson, 2018) (G8), and an early  
150 successional community (G9). Detailed description of the experimental site and management  
151 practices can be found at [https://lter.kbs.msu.edu/research/long-term-experiments/glbrc-](https://lter.kbs.msu.edu/research/long-term-experiments/glbrc-intensive-experiment/)  
152 [intensive-experiment/](https://lter.kbs.msu.edu/research/long-term-experiments/glbrc-intensive-experiment/) (verified on April 20, 2018). The intact soil cores were collected from the  
153 replicated plots of each system from 5-10 cm depth. A total of 13 cores (2-3 cores per system)  
154 were used for zymography analyses.

155

## 156 2.2 X-ray $\mu$ CT scanning and image analysis

157 Soil pore characteristics and grayscale values were obtained via X-ray  $\mu$ CT image  
158 analyses. For that, the soil cores were subjected to X-ray scanning using a GE Phoenix v|tome|x  
159 at the Institute of Soil and Environment at the Swedish University of Agricultural Sciences in  
160 Uppsala. 3D  $\mu$ CT X-ray images were reconstructed using the GE software datos|x. Detailed  
161 description of the scanning specifications is reported in Kravchenko et al. (2018). Each image  
162 had a resolution 29  $\mu$ m in all directions.

163 The image processing was conducted in ImageJ/Fiji software (Schindelin et al., 2012).  
164 Preprocessing consisted of a 3D median filtering with a radius of two in all directions to reduce  
165 random noise. We removed 0.5 cm border part around each core to avoid artifacts associated  
166 with sample wall effects. Based on the scanning resolution of  $\mu$ CT analysis, we identified pores  
167 with diameters  $>60$   $\mu$ m, referred to as visible pores. The thresholds were computed using  
168 minimum error thresholding approach (Kittler and Illingworth, 1986). Following (Nakagawa and  
169 Rosenfeld, 1979) the two-Gaussian fits were applied to sequences of grayscale histograms of 2D  
170 images separately for each soil core. For these computations we used the Regression Wizard tool  
171 of the SigmaPlot software (Systat Software, Inc). Then, pore size distributions were obtained  
172 using the Pore size distribution tool of Xlib plugin for ImageJ, based on the maximum inscribed  
173 spheres approach (Munch and Holzer, 2008). On the studied images we also identified fragments  
174 of particulate organic matter using the approach outlined in Kravchenko et al (2014).

175 Another employed  $\mu$ CT image characteristic was  $\mu$ CT grayscale values of the soil solid  
176 matrix voxels. The grayscale values reflect the attenuation of X-rays as they pass through the soil  
177 sample; they are driven by the density of the material and by the atomic numbers of the  
178 constituting elements (Ketcham, 2005; Peth, 2010). On 8-bit images, the voxels that contain  
179 primarily pore space (air) appear dark and have grayscale values close to zero, while the voxels  
180 that contain primarily solid material dominated by elements with high atomic number, e.g., iron,  
181 appear bright and have grayscale values close to 255. Here we only used the grayscale values of  
182 the image voxels that were classified as solids; the gray scale values from the image voxels that  
183 were classified as pores were not used in this analysis. Thus, the darker grayscale values of the  
184 studied solid voxels correspond to the greater abundance of elements with low atomic numbers,  
185 notably, carbon (Quigley et al., 2018b). Please note that darker values could also be related to  
186 greater presence of pores smaller than the scanning resolution ( $<60$   $\mu$ m  $\varnothing$  pores).

187

### 188 2.3. 2D zymography of soil core slices

189 Mapping of soil enzyme activities was conducted via 2D soil zymography (Spohn and  
190 Kuzyakov, 2014; Razavi et al., 2016), as described in detail in Razavi et al (Razavi et al., 2016).  
191 In a course of 2D zymography a membrane saturated with an enzyme-specific substrate is placed  
192 on a soil surface. Contact between substrate and enzyme releases a fluorescent product (e.g.  
193 MUF: methylumbelliferon, AMC: 7-amido-4-methylcoumarin) and the resulting fluorescing  
194 patterns reflect spatial distribution of active EE (Guber et al., 2018).

195 Hydrophilic polyamide filters (0.45  $\mu\text{m}$  pore size; 100  $\mu\text{m}$  thick, Tao Yuan, China) were  
196 used as membranes (Razavi et al., 2016; Sanauallah et al., 2016). Photos of the membrane on the  
197 soil surface were taken using Nikon D90 camera (Nikon Inc.) with a Sigma 18-250 mm f/3.5-6.3  
198 DC Macro OS HSM lens (Sigma Corp. of America) installed on a Rocwing Pro Copy Stand  
199 (Rocwing Co., UK). The source of UV light was a 22W Blue Fluorescent Circline Lamp -  
200 FC8T9/BLB/RS (Damar Worldwide 4 LLC.).

201 Six enzymes were studied:  $\beta$ -glucosidase, cellobiohydrolase, xylanase, N-acetyl-beta-  
202 glucosaminidase (chitinase, NAG), leucine aminopeptidase, and acid phosphomonoesterase Acid  
203 phosphatase). The respective enzyme-specific substrates used were: 4-Methylumbelliferyl- $\beta$ -D-  
204 Glucoside, 4-Methylumbelliferyl- $\beta$ -D-Cellobioside, 4-Methylumbelliferyl- $\beta$ -D-Xylopyranoside,  
205 4-Methylumbelliferyl-N-Acetyl- $\beta$ -D-Glucosaminide, L-leucine-7-amido-4-methylcoumarin  
206 hydrochloride, and 4-methylumbelliferyl-phosphate (Razavi et al., 2017). Each substrate was  
207 dissolved in a concentration of 6 mM in either TRIZMA buffer (was used for AMC-based  
208 substrate - leucine aminopeptidase, pH: 7.2) or MES (2-(N-morpholino)ethanesulfonic acid)  
209 buffer (was used for MUF-based substrates - all other enzymes, pH: 6.5): [MES(pH: 6.5)  
210 ( $\text{C}_6\text{H}_{13}\text{NO}_4\text{SNa}_{0.5}$ ) TRIZMA (pH: 7.2) ( $\text{C}_4\text{H}_{11}\text{NO}_3\cdot\text{HCl}$ ,  $\text{C}_4\text{H}_{11}\text{NO}_3$ )] (Razavi et al., 2017).

211 We obtained 8-13 enzyme maps per each intact soil core (1-3 maps of each individual  
212 enzyme from each core) for a total of 180 enzyme maps. One enzyme map was obtained per each  
213 soil slice. The order in which specific enzymes were measured within the core was randomized.  
214 For the measurements, each core was placed within a cutting table (Supplement Fig. 1). A  
215 calibrated handle at the bottom of the table allowed pushing the core out of the sample cylinder  
216 in 0.5 mm increments. At each 0.5 mm increment the soil layer pushed above the table was  
217 removed manually using a microtome knife. Care was taken to minimize disturbance to the soil



218 surface while cutting, by removing stones or large sand grains with tweezers from the surface  
219 prior/during cutting. Enzyme maps were obtained on soil surfaces in 2 mm increments  
220 (Supplement Fig. 1). For that, a polyamide membrane ( $\varnothing$  4.5 cm) was saturated in 240  $\mu$ l of the  
221 MUF/AMC-based substrate solution and placed on top of the prepared soil surface. Additional  
222 120  $\mu$ l of the substrate solution was added on top of the membrane with a pipette and evenly  
223 spread with a fine brush. The membrane was covered by a layer of aluminum foil followed by a  
224 100 g sandbag weight. The membrane was incubated on the soil surface for 30 minutes at room  
225 temperature, then it was placed within a light-proof zymography chamber and a photo was taken  
226 in UV light as described above. The membrane was then removed from the soil surface.

227 Because of unevenness of the soil surface, only a portion of it was in full contact with  
228 the membrane; thus reliable enzyme activity data could be obtained only from the portions of the  
229 membrane (Guber et al., 2018). In order to estimate presence and strength of the contact we used  
230 MUF-staining approach (Guber et al., 2018b). For that, immediately after removing the  
231 membrane with the substrate, we applied to the soil surface a membrane fully saturated with 6  
232 mM MUF solution. The membrane was covered with a 100 g sandbag and kept for 30 s. Then,  
233 the soil surface was photographed in UV light. The bright areas on the image indicated the  
234 localities on the soil surface that received MUF from the membrane, and thus could be regarded  
235 as such that were in contact with the membrane. The image processing was conducted in ImageJ  
236 and the images were converted into an 8-bit format. The image from the substrate membrane was  
237 matched with the image of the MUF-stained surface, and the areas with minimal contact, that is,  
238 with MUF-stained grayscale values of  $<30$  (0-255 grayscale scale), were excluded from further  
239 enzyme map analyses.

240

#### 241 *2.4. Matching enzyme maps with $\mu$ CT information*

242 The enzyme maps were obtained from the surfaces of the individual soil slices  
243 (Supplement Fig.1b). The enzyme map was covered by a 1 mm<sup>2</sup> grid and EE readings from  
244 zymograms were used to produce a single value per each 1 mm<sup>2</sup> grid cell. For that, for each 1  
245 mm<sup>2</sup> pixel of the enzyme map we, first, calculated the average grayscale value corresponding to  
246 it, then, the pixel averages were further standardized based on the mean and standard deviations  
247 of the entire map. The latter step was necessary to enable comparisons among different enzymes  
248 and systems.

249 Then, for each soil slice we identified the corresponding layer from the  $\mu$ CT image, such  
250 that the center of the layer corresponded to the soil slice (Supplement Fig. 1c). The  $\mu$ CT layers  
251 were 1 mm in height, 0.5 mm above and below the soil surface layer. The  $\mu$ CT information was  
252 aggregated to 1 mm<sup>3</sup> grid cells. For each 1 mm<sup>3</sup> grid cell we calculated the total volumes of the  
253 pores of the studied sizes. For example, for each given 1 mm<sup>3</sup> grid cell we had the number of 29-  
254  $\mu$ m voxels of the original  $\mu$ CT image that belonged to that cell and that were occupied by 60  $\mu$ m  
255 diameter pores, and that number was used to calculate the volume of the 60  $\mu$ m pores in that grid  
256 cell. The gray scale values from all 29- $\mu$ m voxels of the original  $\mu$ CT image that belonged to that  
257 1 mm<sup>3</sup> cell were used to calculate the average gray scale value of the cell.

258 To match enzyme maps from soil slices with 3D information from  $\mu$ CT scans  
259 horizontally, we used the mark placed on an acrylic tube of each soil core prior to  $\mu$ CT scanning.  
260 The cores were located within the cutting table so as to ensure a match between the mark and the  
261 position of the zymography membrane on the soil surface. To match them vertically, we used  
262 visual observations from the 3D images and pictures of soil surface taken at each soil cut, and the  
263 height of the soil remaining after all the desired enzyme slices were cut from the soil core.

264 The aggregation of the data to the 1 mm scale conducted here was a conservative  
265 measure to address the uncertainties associated with some movement of soil during surface  
266 cutting as well as with matching enzyme maps with  $\mu$ CT images. However, it did introduce  
267 smoothing into the resulting pore data.

268

## 269 *2.5. Data analysis*

### 270 Removal of artificial spatial trends

271 Despite the best efforts, during placing the substrate membranes on the soil surface there  
272 was some unevenness in the redistribution of the liquid substrate through the membrane. These  
273 spatial trends could distort the assessments of spatial variability patterns via variography. To  
274 ensure that such redistribution is not affecting the variogram estimates we removed the trends  
275 using multiple regression models. Application of polynomial models with various degrees of  
276 complexity was explored and the model that appeared to adequately describe the spatial trend in  
277 most of the studied 180 samples was the model with linear effects of x and y spatial coordinates.  
278 Specifically, for each enzyme map we fitted the regression model to the EE data, obtained

279 residuals from the regression analysis, and then used the residuals in all further analyses as a  
280 measure of relative EE activity corrected for presence of linear spatial trends.

281

## 282 Variography

283 Sample variograms of the residuals, obtained from the trend removal procedures  
284 described above, were calculated individually for each 2D enzyme map. We used lag distance of  
285 1 mm and considered 30 lag distances, thus, covered 30 mm of the sample, avoiding border  
286 effects. The number of point pairs in the considered lags always exceeded 500, hence was  
287 sufficient for a reliable estimation of the sample variogram value (Goovaerts, 1998). Variogram  
288 calculations were conducted using PROC VARIOGRAM in SAS (SAS 9.4).

289 The spherical model was used in variogram fitting for kriging, as the model that was  
290 adequately fitting most of the obtained sample variograms. Automated fitting of all 180 sample  
291 variograms was not possible because of convergence problems, thus manual fitting was  
292 performed. As a nugget, we selected the sample variogram value at the first lag; and the sill was  
293 "eye-ball" selected as the value corresponding to the plateau. The spatial auto-correlation range  
294 was set as the lag distance corresponding to the sill. As a measure of proportion of the variability  
295 occurring at distances  $< 1,000 \mu\text{m}$  we used the ratio of nugget and sill (N/S) expressed as  
296 percent.

297

## 298 Ordinary and regression kriging

299 Prediction accuracy of 2D zymography data was tested using test-model data set  
300 approach (Goovaerts, 1998). That is, the data were divided into two sub-sets, a model data set  
301 and a test data set. The model data set was used for generating kriging predictions for the test  
302 data set. The test data set consisted of a total of 120 randomly selected data points (Supplement  
303 Fig 2). The remaining data constituted the model data set. We used data from all individual 2D  
304 enzyme maps sequentially

305 For ordinary kriging (OK), the model data set was used to compute a sample variogram  
306 and determine the variogram model parameters, as described above. Then, the model parameters  
307 were used in 2D OK to generate predictions for the test data points. SAS procedure PROC  
308 KRIGE2D was used to perform OK. The search radius was set to 14 mm and the minimum and  
309 maximum number of data points used in kriging estimation was set to 4 and 10, respectively. For

310 regression kriging (RK), the model data set was first used to relate EE values with auxiliary  
311 variables from  $\mu$ CT images, i.e., abundances of pores of different sizes and grayscale values.  
312 Predicted values for the test data points were recorded. Then the residuals from the fitted model  
313 were used in computing sample variogram, fitting variogram model, and obtaining kriging  
314 predictions at the test data points. The final predictions for each test data point were obtained by  
315 adding predictions from the linear regression model and the kriging predicted residuals. For both  
316 OK and RK the correspondence between true and predicted values for the test data sets were  
317 assessed using  $R^2$  values, MSE, along with parameters (slope and intercept) of the regression  
318 equation relating true values with predictions.

319

### 320 Statistical analysis

321 Statistical analyses, ANOVA and ANCOVA, were conducted in SAS using PROC  
322 MIXED and PROC GLIMMIX tools (Milliken and Johnson, 2001, 2009). Statistical models for  
323 exploring spatial parameters for the studied enzymes included the fixed effects of the enzymes  
324 and the land use systems and the interaction between them, and the random effects of blocks,  
325 plots, and cores (nested within systems and plots). Normality of the residuals and homogeneity  
326 of variances were checked for each variable. In case of marked deviations from normality the  
327 data were log-transformed, while in case of variance heterogeneity, unequal variance analysis  
328 was performed (Milliken and Johnson, 2001, 2009). Significant interactions were examined  
329 using analysis of simple effects, and, when significant, were followed by multiple comparisons  
330 via t-tests. The results are reported as statistically significant at 0.05 level.

331 To assess the associations between EE and auxiliary variables from  $\mu$ CT images,  
332 correlation analysis and ANCOVA were applied. For ANCOVA we used the statistical model  
333 described above and added to it the linear effects of the auxiliary variables of interest. The  
334 resulting linear coefficients were then used to explore the patterns in relationships among six EE  
335 in the studied land use systems. To facilitate comparisons we report standardized coefficients  
336 from these analyses, i.e.,  $t$ -values.

337

## 338 **3. Results**

### 339 *3.1. Spatial variability patterns*

340 Spatial autocorrelation was present in the activity distributions in all soil slices from all  
341 studied EE and land use systems. Examples of  $\beta$ -glucosidase and acid phosphatase maps of EE  
342 activity are shown on Fig. 1 along with the corresponding sample variograms fitted with  
343 spherical models.

344 ANOVA indicated no interactions between EE and land use systems and no land use  
345 effects ( $p>0.05$ ) (Supplement Table 1), thus we focused the analyses on the main effect of EE.  
346 Acid phosphatase and leucine aminopeptidase had much higher nuggets and sills than the other  
347 EE (Fig. 2a and b), indicating an overall greater variability in their activity. However, their N/S  
348 ratios were lower than for the other enzymes. The N/S ratio was 47% for acid phosphatase and  
349 54% for leucine aminopeptidase, for other EE it exceeded 60%. Since N/S ratios here represent  
350 the proportion of the random variability occurring at distances  $<1000 \mu\text{m}$ , the lower values in  
351 phosphatase and leucine aminopeptidase suggest that the distribution patterns of these enzymes  
352 had greater spatial continuity than that of the other EE, while small random patches were more  
353 abundant in the other EE's distributions. Overall, N/S ratios were quite substantial and ranged  
354 from 17% to 80% in individual soil slices. The average spatial correlation range across all  
355 enzymes and land use systems was equal to 7.5 mm; no significant differences among the  
356 enzymes and among systems were observed.

357

### 358 *3.2. Correlations of EE activities with X-ray $\mu\text{CT}$ information*

359 In most soil slices the studied EE were negatively correlated to grayscale values from X-  
360 ray  $\mu\text{CT}$  images (Fig. 3). NAG and  $\beta$ -glucosidase were the two enzymes with the strongest  
361 associations with the grayscale values (Fig. 3). For the other two enzymes involved in C cycle,  
362 i.e., cellobiohydrolase and xylonase, the associations with the grayscale values were relatively  
363 weak. With the exception of a few slices, acid phosphatase was not correlated with grayscale  
364 values.

365 Associations with pores of various sizes varied among the six enzymes, but positive  
366 relationships with 60-180  $\mu\text{m}$   $\emptyset$  pores and subsequent decrease with further increasing pore sizes  
367 was present in all EEs (Fig. 4). For 60  $\mu\text{m}$  pores, the associations were the highest for  $\beta$ -  
368 glucosidase, closely followed by acid phosphatase, and then by NAG. Associations with 120  $\mu\text{m}$   
369 pores were the strongest for phosphatase, followed by  $\beta$ -glucosidase, and NAG. Associations

370 with pores of 180-300  $\mu\text{m}$  size range were substantially stronger for acid phosphatase as  
371 compared to the other EEs. All enzymes were negatively associated with pores  $>360 \mu\text{m}$  (Fig. 4).

372

### 373 *3.3. Kriging predictions*

374 As expected, the  $R^2$  values for test data set predictions were higher (Fig. 5) for the  
375 enzymes with greater spatial autocorrelation, i.e., acid phosphatase and leucine aminopeptidase.  
376 The  $R^2$  values for these two enzymes were around 0.45, while for the other enzymes they were in  
377 0.30-0.35 range. Adding auxiliary X-ray information, i.e., grayscale values did not lead to a  
378 substantial improvement in mapping accuracy during regression kriging (results not shown).

379

## 380 **4. Discussion**

381 Our findings can be interpreted at two spatial scales: tens-of-mm spatial scale for the EE  
382 spatial variability data and tens-of-micron scale for correlations between EE and  $\mu\text{CT}$  data. Note  
383 that the soil cores here were investigated in the absence of live plant roots, thus the observed  
384 relationships can be regarded as typical for a soil matrix outside of the actively functioning  
385 rhizosphere. Despite our expectations, the spatial distributions of EE activities and their  
386 associations with pores did not differ among the studied land use systems.

387

### 388 *4.1. Spatial variability patterns of enzyme activities*

389 Spatial distributions of EE activity at tens-of-mm spatial scale were auto-correlated (Fig.  
390 2), but between 47% and 60% of EE variability took place at distances  $<1000 \mu\text{m}$ . The strengths  
391 of spatial auto-correlation in distribution of EE at the studied spatial scale are likely related to: (i)  
392 spatial patterns of EE's microbial producers, and (ii) spatial patterns and diffusion rates of the  
393 substrates subject to the enzymes. In the absence of live roots, microbial activities are the main  
394 driver of EE production (Nannipieri et al., 2012; Burns, 2013). However, past history, such as  
395 former presence and activity of live plant roots, affected the spatial distribution patterns of both  
396 microorganisms and EE substrates.

397 Even though the spatial resolution of our study was too coarse to conduct in-depth  
398 assessment of the spatial patterns in microbial producers, the highly patchy distributions of  
399 enzymes observed here are consistent with the typically reported, very sparse distributions of  
400 microorganisms (Nunan et al., 2003; Franklin and Mills, 2009; Baldrian and Vetrovsky, 2012).

401 Presence of bacterial colonies and individual cells are highly sporadic even on plant roots, which  
402 are the sites of the greatest microbial activity in soils (Schmidt and Eickhorst, 2013). The typical  
403 reported sizes of microbial colonies and spatial correlation ranges in their distributions are much  
404 smaller than the studied resolution (1000  $\mu\text{m}$ ). For example, Nunan et al. (Nunan et al., 2002;  
405 Nunan et al., 2003; Nunan et al., 2006) reported 100-600  $\mu\text{m}$  spatial correlation ranges for  
406 bacteria presence. Probandt et al. (Probandt et al., 2018) observed that the average distances  
407 between individual bacteria cells on sand grains varied from 0 to 29  $\mu\text{m}$ . It can be surmised that  
408 sporadic patterns in microbial distributions played a major role in EE variability and were  
409 responsible for the high spatially unexplained component in the EE variograms.

410 Spatial patterns and diffusion rates of EE substrates were the other possible source of the  
411 observed auto-correlations. Specifically, comparisons among the studied EE enabled insights  
412 into how differences in their substrates can be the potential source of the observed differences in  
413 EE spatial patterns. For example, lower N/S ratios for leucine aminopeptidase and acid  
414 phosphatase as compared to the other EE could be related to the fact that they have relatively low  
415 substrate specificity and can act on different substrates including a variety of compounds within  
416 non-particulate SOM (Alef et al., 1995). These enzymes can be expressed by a wide range of  
417 producers (Dick and Tabatabai, 1984; Blagodatskaya and Kuzyakov, 2008; Nannipieri et al.,  
418 2012). Also, it is assumed that the mobilization of organic P by phosphatases is necessary over  
419 larger soil volumes compared to the enzymes responsible for other nutrients (Kuzyakov and  
420 Razavi, 2019), because P delivery to roots is strongly controlled by diffusion, which is very slow  
421 for P (Nye and Tinker, 1977). Mancarella et al (Mancarella et al., 1981) suggested that the  
422 soluble form of aminopeptidase in soil is a result of proteolytic and non-proteolytic processing of  
423 epithelial cell membranes and that it is not a true secretory product. This is also valid for  
424 intercellular phosphatase which can be released from the cells of plants, fungi and bacteria after  
425 lyses, and can react as EE in soil matrix. However, the substrates for the other studied enzymes  
426 are primarily plant and fungal residues and spatial continuity in their distribution patterns is  
427 typically quite low. This makes their distribution random and sporadic, which would explain the  
428 observed differences in enzyme auto-correlations.

429 Differences between  $\beta$ -glucosidase and cellobiohydrolase in the strength of association  
430 with pores, namely, strong positive correlation for  $\beta$ -glucosidase and relatively weak correlations  
431 for cellobiohydrolase, might be an indication of the importance of substrate diffusion rates. Even

432 though  $\beta$ -glucosidase and cellobiohydrolase are both involved in cellulose degradation  
433 (Nannipieri et al., 2012), the substrate of cellobiohydrolase activity is insoluble cellulose while  
434 the substrates of  $\beta$ -glucosidase are soluble compounds, e.g. disaccharides (Alef and Nannipieri,  
435 1995). The latter can easily diffuse within the soil matrix, with pores being the avenues for such  
436 diffusion. The spatial patterns in  $\beta$ -glucosidase likely follow the spatial patterns of its soluble  
437 substrates which are then reflected in positive correlations of  $\beta$ -glucosidase activities with pores.  
438 Work by Bailey et al. (2017) demonstrated that fine pores with  $\sim 6 \mu\text{m}$  neck diameters contained  
439 more complex organic compounds than large pores with  $\sim 200 \mu\text{m}$  necks, while large pores had  
440 greater presence of simple soluble organic. Note that the enzymes diffusion is negligible (Guber  
441 et al., 2018), thus, the reactions between enzymes and substrates are driven solely by substrate  
442 diffusion and substrate mass flow with water.

443

#### 444 *4.2. Associations with pores*

445 Even though the pore size distribution data were aggregated to 1 mm spatial scale, the EE  
446 associations with pores of different sizes are indicative of the processes taking place at tens-of-  
447 micron scale, i.e., the scale corresponding to  $\mu\text{CT}$  resolution. Presence of most enzymes was  
448 positively associated with 60-180  $\mu\text{m}$  pores, while negatively associated with presence of 360  
449  $\mu\text{m}$  pores (Fig. 4). Positive associations of EE with 60-180  $\mu\text{m}$   $\emptyset$  pores (Fig. 4) are consistent  
450 with a substantial body of experimental evidence suggesting that such pores are of particular  
451 significance for soil microbial functioning.

452 We suggest that pores of this size range are optimal microbial habitats, and as such are  
453 associated with greater microbial abundance, leading to high enzyme production and activity  
454 (Fig. 6). For once, the micro-environmental conditions within these pores are better for  
455 microbial functioning than those in pores of other sizes. Pores of this range are associated with  
456 greater presence of fine roots (Pagliai and Denobili, 1993), and thus with greater new C inputs  
457 (Quigley et al., 2018a). These pores are also the main transport avenues for soluble organic  
458 substances released by litter decomposition (e.g. in O horizon) and/or from detritosphere. Pores  
459 of this size range likely provide optimal water availability and do not often experience a lack of  
460  $\text{O}_2$  (Keiluweit et al., 2017). These pores probably also supply an optimally-sized physical space  
461 for the formation of active microbial colonies, since pores of  $<10 \mu\text{m}$   $\emptyset$  are too small to furnish  
462 sufficient space for colony formation. Better micro-environmental conditions likely lead to more



463 active microbial communities populating these pores. Wright et al., (1995) reported higher  
464 activity of bacteria introduced into 6-30  $\mu\text{m}$  as opposed to small ( $<6 \mu\text{m}$ ) pores. Carbon newly  
465 added to soil via plant roots is most actively consumed in pores of this size range (Quigley et al.,  
466 2018a). They are the locations of faster C turnover and greater decomposition of newly added  
467 organics (Strong et al., 2004; Ruamps et al., 2011; Ruamps et al., 2013). Dissolved organic  
468 carbon extracted from pores of this size group was found to be more recalcitrant than that from  $<$   
469  $10 \mu\text{m}$  pores, suggesting quick consumption of labile compounds by resident microbes (Bailey et  
470 al., 2017). Indeed, these pores were found to differ in terms of their microbial community  
471 composition (Ruamps et al., 2011; Ruamps et al., 2013).

472 Based on their C inputs, micro-environmental conditions, and size, the tens-of- $\mu\text{m}$  pores  
473 constitute prime microbial habitats (Fig. 6) and are, potentially, the areas with the greatest  
474 microbial presence and activity within the non-rhizosphere soil matrix. Higher EE activity  
475 associated with greater abundance of such pores (Fig. 4) in all tested enzymes supports the prime  
476 microbial habitat theory.

477

#### 478 *Enzyme mapping*

479 While the observed associations are providing new insights into factors driving spatial  
480 patterns in EEs distributions within intact soil matrix at few mm spatial scale, the strength of the  
481 observed relationships was not sufficient to achieve high accuracy for EE micro-scale mapping.  
482 The biggest roadblock was an extremely high variability of EE' activity at  $<1000 \mu\text{m}$  distances.  
483 N/S ratios, which are the quantitative representations of that variability, exceeded 50% in most  
484 soil slices and were as high as 80% in some of them. Nevertheless, the  $R^2$  values obtained from  
485 ordinary kriging were in a 35-45% range (Fig. 5a). The stronger spatial autocorrelation, as in  
486 acid phosphatase and leucine aminopeptidase distributions, the higher accuracy in kriging maps.

487 Correlations of EE activities with X-ray  $\mu\text{CT}$  data, while statistically significant, and  
488 meaningful from biogeochemical standpoint, numerically were relatively weak. In cases when  
489 the auxiliary variable is only weakly correlated to the main variable of interest the incorporation  
490 of the auxiliary information via either regression kriging or co-kriging typically does not lead to  
491 sizeable improvement in accuracy. For example, a reliable improvement in prediction accuracy  
492 via regression kriging was only possible when the  $R^2$  for the linear regression between the main  
493 and auxiliary variables exceeded 0.6 (Zhu and Lin, 2010), the condition that was barely present

494 among the soil zymography slices. Even when the  $R^2$  values are relatively high, the improvement  
495 in mapping accuracy due to regression kriging may still be only minor if spatial autocorrelations  
496 of the main and auxiliary variables are interrelated (Kravchenko and Robertson, 2007).

497 The mentioned above very high patchiness in microbial community locations and  
498 activities (Nunan et al., 2002; Franklin and Mills, 2009; Baldrian and Vetrovsky, 2012) suggests  
499 that only predictions and mapping at scales less than a hundred micron would be able to explain  
500 significant portion of EE variability. In this study the resolution of the  $\mu$ CT data was 30  $\mu$ m. The  
501 resolution of zymograms is not possible to establish precisely due to the diffusion of products of  
502 enzyme catalysis (Guber et al., 2018b), but an estimate of 100  $\mu$ m seems plausible. Yet, because  
503 of the inaccuracies potentially involved in matching the two sources, the 1000  $\mu$ m resolution for  
504 the final joint data sets combining zymography and  $\mu$ CT was the only reasonable option for this  
505 study.

506 Even though the predictions of specific locations of EE activity were not possible, the  
507 approach did offer the possibility of exploring general associations between EEs and  
508 characteristics of soil micro-environments with potential relevance to EE distributions as well as  
509 their visualizations (Fig. 5b, Supplemental video 1).

510 In further work, greater reduction of uncertainties in spatial coupling of the two data  
511 sources can be achieved by improving zymography and X-ray  $\mu$ CT image matches (Guber et al.,  
512 2018a). The high random component in spatial patterns of soil microorganisms, i.e. sources of  
513 EE in the soil, suggests that additional factors/drivers, e.g., fresh C inputs, variations in pH and  
514 soil moisture, need to be considered and better matching approaches are necessary for  
515 improvements at the finer spatial resolutions.

516

## 517 **5. Conclusions**

518 At the studied scales of a few mm, within non-rhizosphere soil matrix, we found spatial  
519 autocorrelations in distributions of all studied EE. Observed spatial patterns are a function of  
520 spatial patterns in the distributions of microbial producers, but also of nature, availability, and  
521 diffusion properties of EE substrates.

522 Positive associations between pores with 60-180  $\mu$ m diameters and relative EE activities  
523 were found for all studied enzymes, across all five studied land use and management practices.  
524 Apparently, micro-scale areas with prevalence of such pores experience elevated levels of

525 microbial activities leading to EE production and are the potential hotspots of C, N and P  
526 cycling. The results suggest that in the studied soils pores of this size range serve as a prime  
527 habitat for soil microbial communities.

528

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542

543 **Figure 1.** Examples of zymograms for  $\beta$ -glucosidase (top) and acid phosphatase (bottom) and  
544 their corresponding variograms of detrended values for one of the studied soil cores. Color  
545 gradient represent the range of enzyme activity from low (blue) to high (yellow). White scale bar  
546 represents 1 cm. Vertical lines mark spatial correlation ranges.

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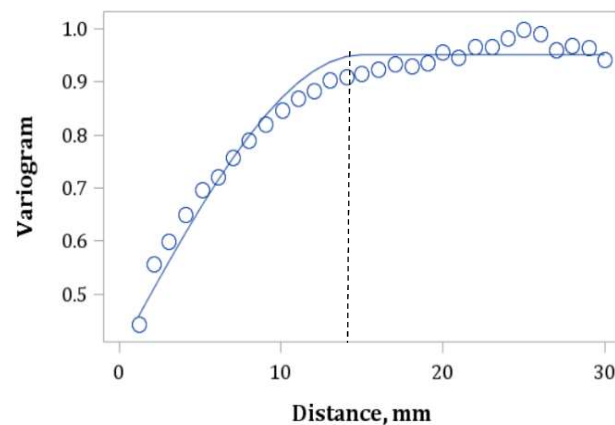
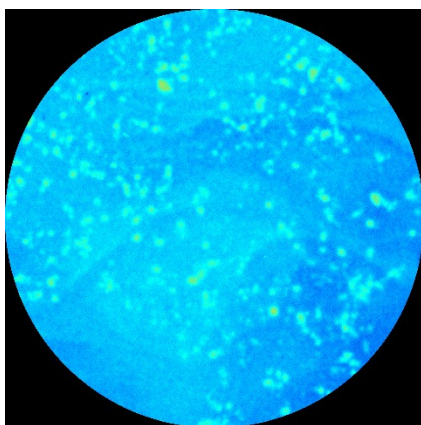
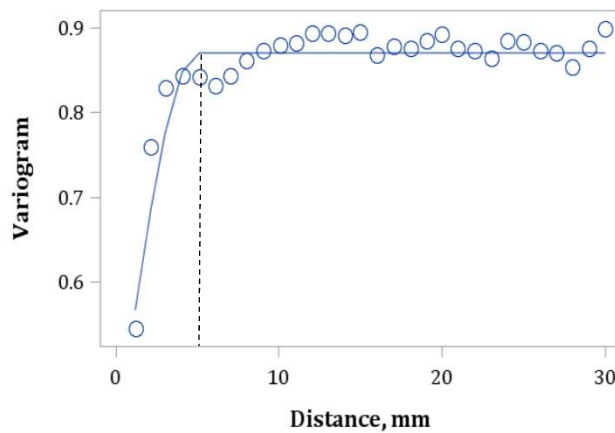
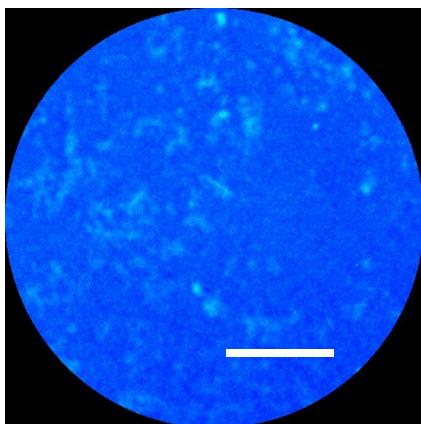
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563 **Figure 2.** Nuggets (a), sills (b), and N/S ratios (c) for the studied enzymes across all land use  
564 systems. Shown are means and standard errors. Letters mark significant differences (at  $p < 0.05$ ).

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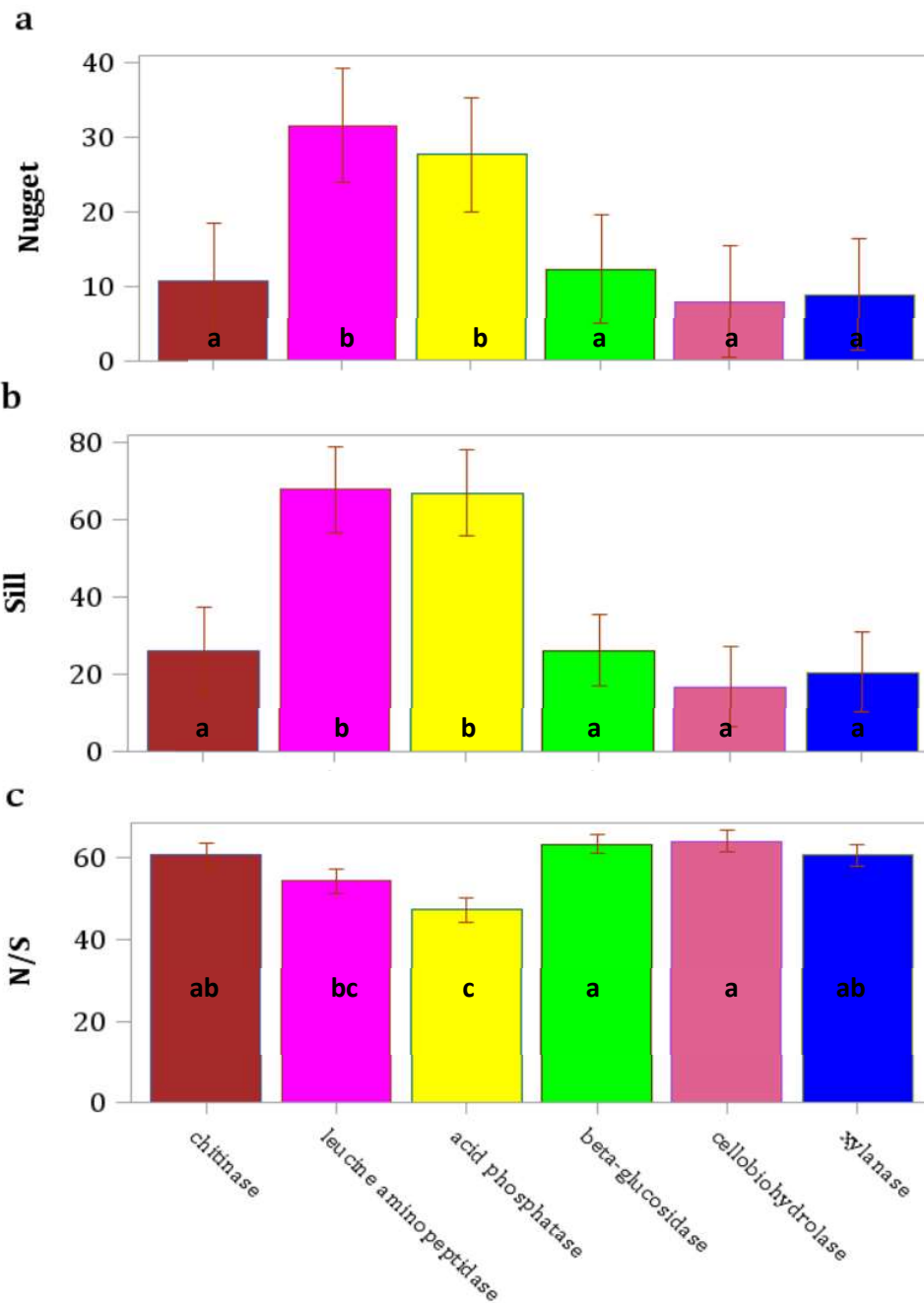
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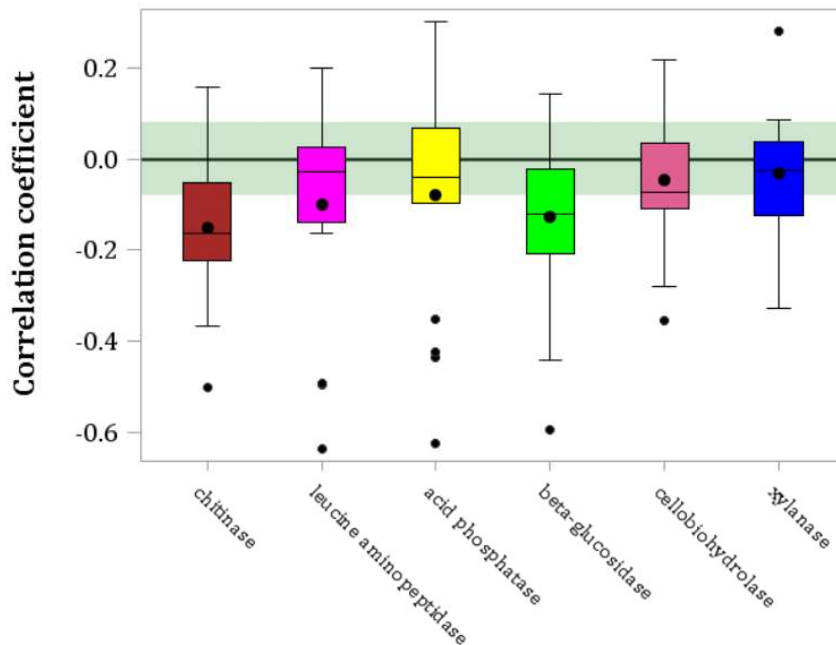
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596 **Figure 3.** Correlation coefficients between  $\mu$ CT grayscale values from solid voxels and  
597 standardized values of the studied six enzymes. Horizontal lines and dots within the boxes mark  
598 medians and means, respectively, while outside dots mark outliers. Shaded area marks  
599 correlation coefficients that are not significantly different from zero ( $p < 0.05$ ). Note that negative  
600 correlations signify that higher EE levels were present in darker (lower grayscale value) areas of  
601  $\mu$ CT images, which are in part associated with greater presence of organic materials.

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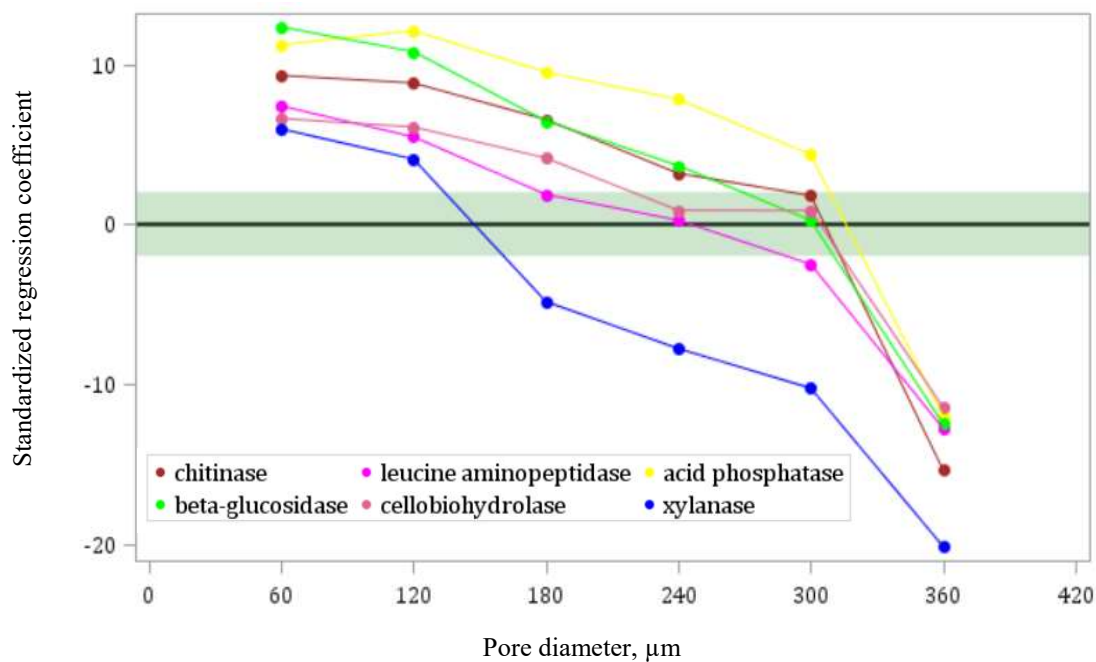
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622 **Figure 4.** Standardized linear regression slopes from ANCOVA relating enzyme activity with  
623 volumes of pores of different sizes across all studied sites. Shaded area marks correlation  
624 coefficients that are not significantly different from zero ( $p < 0.05$ ).



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644 **Figure 5.** (A)  $R^2$  values from predicting test data set values using ordinary kriging. A total of  
645 120 randomly selected observations from the independent test data set are predicted in every soil  
646 slice. Letters mark significant differences among the enzymes (at  $p < 0.05$ ). (B) A section of a 3D  
647 map of  $\beta$ -glucosidase (pink) distribution within an intact soil core obtained from regression  
648 kriging along with soil pores (blue) and particulate organic matter (green). White scale bar  
649 represents 5 mm.

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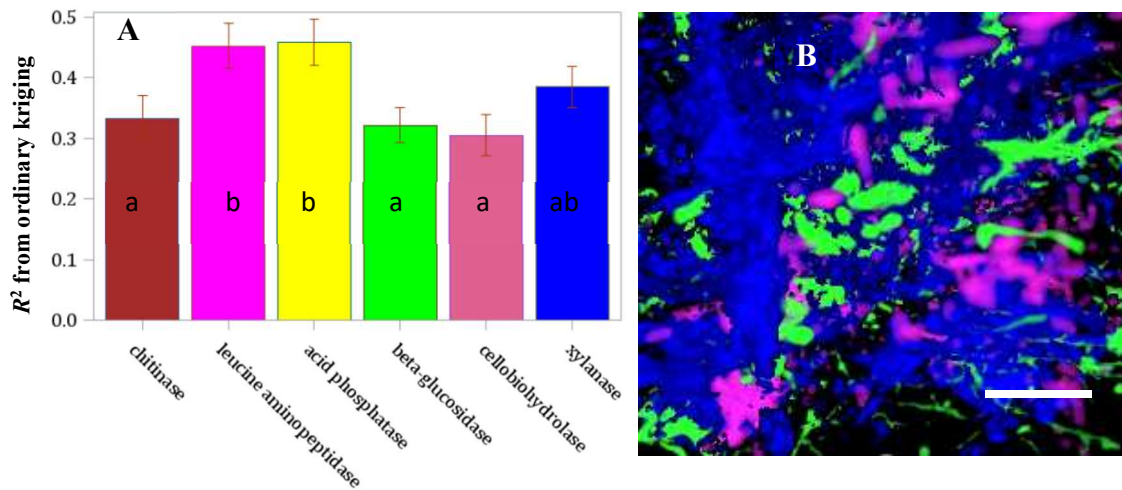
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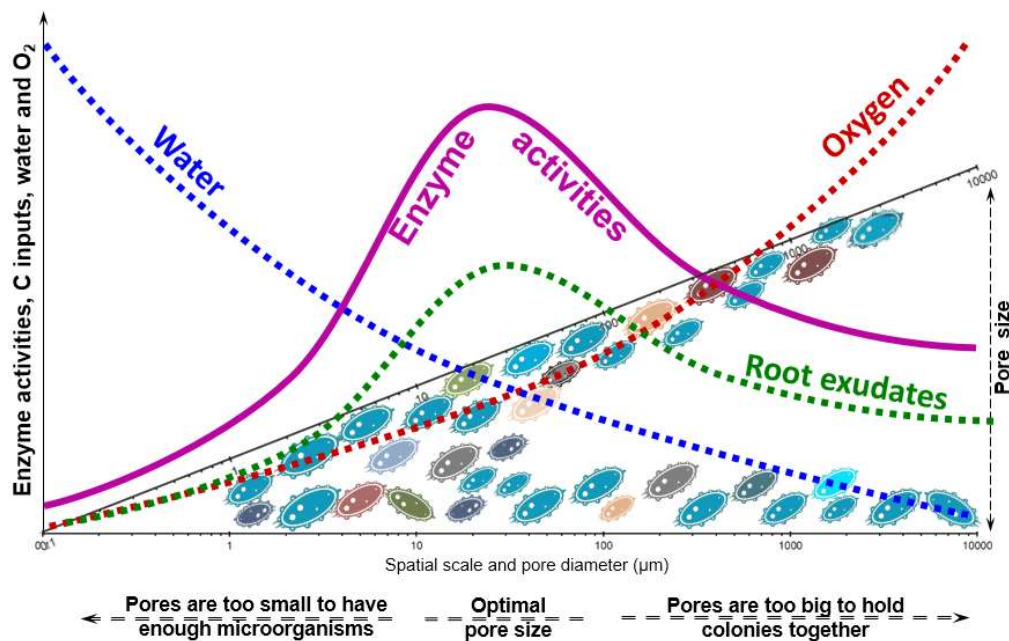
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664 **Figure 6.** Hypothesized relationships between extracellular enzyme (EE) activities, abundance of  
665 microorganisms, new carbon inputs from root exudates, and availability of water and O<sub>2</sub> in pores  
666 of different sizes. The highest EE activity corresponds to the optimum between water and O<sub>2</sub>  
667 availability and high level of root exudation. Low EE activity in pores <10 μm is also related to  
668 these pores being too small to maintain sizeable microbial colonies, while pores > 300 μm are  
669 too large to provide 3D connectivity between colonies located on the pore surfaces and too  
670 accessible for grazing by the predators.



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684 **Appendix**

685

686 **Supplement Figure 1.** Intact soil core within a cutting table: a view from the top (a) and a view  
687 from a side with a 2 mm soil layer pushed out using the calibrated handle below (b); and  
688 schematic representation of the soil cutting and subsequent matching of soil surface slices where  
689 zymography was conducted with corresponding 1mm deep layer of  $\mu$ CT image data (c).

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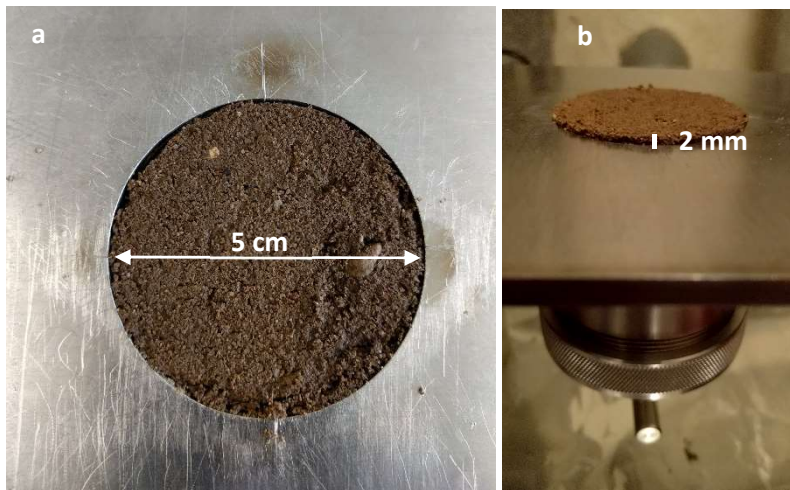
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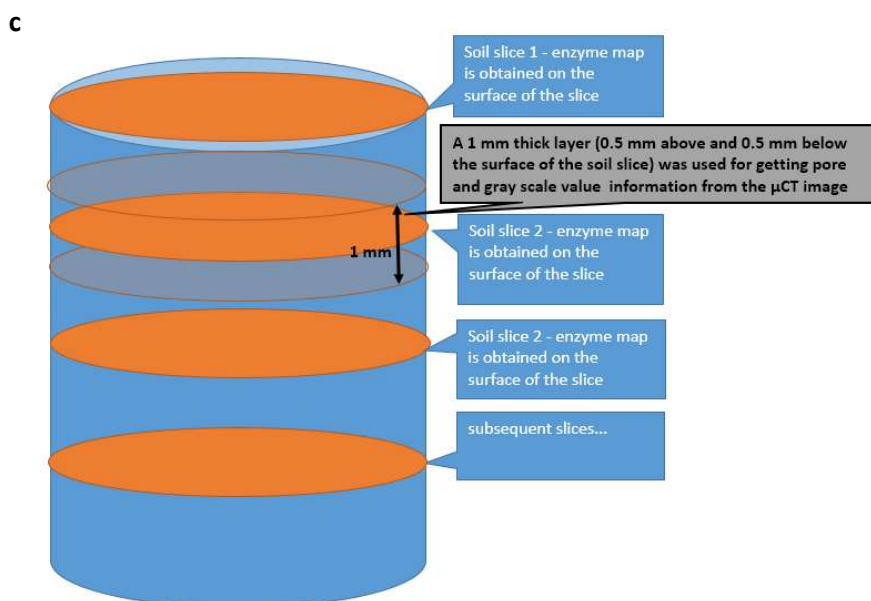
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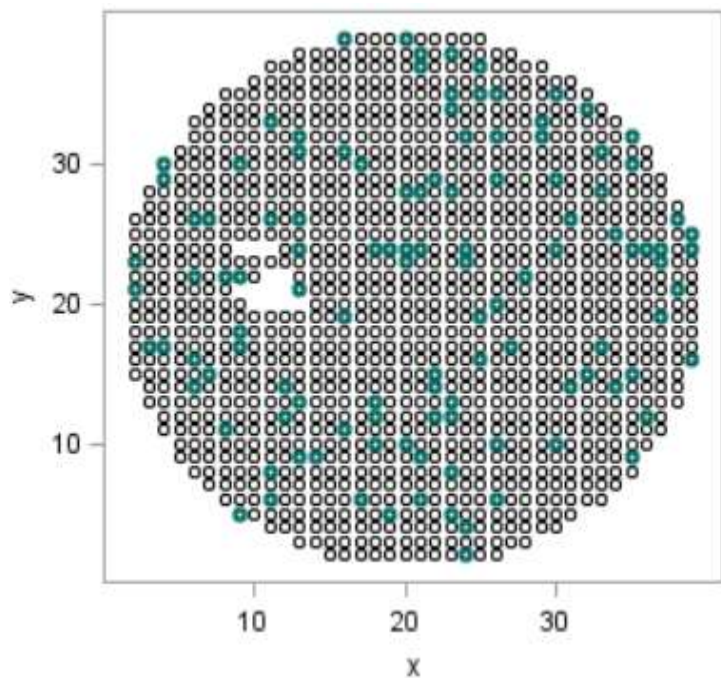
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706 **Supplement Figure 2.** An illustration of a data from one soil slice used in assessing accuracy of  
707 kriging mapping. The black circles are the locations of the data points that were used for  
708 mapping (model data set) and the green circles are the locations to be predicted (test data set).



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710 **Supplement Table 1.** Results of ANOVA for the effects of the land use systems, enzymes, and  
 711 system by enzyme interactions on geostatistical parameters, namely, nugget, sill, range, and  
 712 nugget-to-sill ratio (N/S). Shown are F values for the effects and the estimates of the error  
 713 variances for cores and the residuals. F-values significant at 0.05, and 0.01 levels are marked  
 714 with \*\*, and \*\*\*, respectively.

715

Effect	Geostatistical parameter			
	Nugget	Sill	Range, mm	N/S
Land use system	0.7	1.0	3.1	2.8
Enzyme	11.8***	6.1***	1.5	5.6***
System*Enzyme	0.4	0.6	1.0	0.9
Core(Land use system) variance	93	409	0.1	9.4
Residual variance	162	1561	7.5	141

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719 **Supplement Table 2.** Summary of pore-size distribution data obtained from  $\mu$ CT images with  
720 scanning resolution of 29  $\mu\text{m}$ .

Pore radius, $\mu\text{m}$	Pore volume, % of total			
	Mean	Standard deviation	Minimum	Maximum
30	0.17	0.09	0.05	0.42
60	0.69	0.38	0.22	2.01
90	0.78	0.48	0.17	2.60
120	0.49	0.33	0.08	1.67
150	0.45	0.32	0.06	1.53
180	0.23	0.16	0.03	0.75
210	0.21	0.15	0.02	0.68
240	0.13	0.09	0.01	0.40
270	0.12	0.08	0.01	0.36
300	0.08	0.06	0.01	0.28
330	1.82	1.38	0.06	6.45

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