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1 Deciphering the effects of 2D/3D urban morphology

2 on diurnal cooling efficiency of urban green space

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27 Keywords: Cooling efficiency, ECOSTRESS LST, Urban morphology metrics, Urban green space

29 Highlights

- We explore diurnal variations of cooling efficiency of urban green space in the megacity Paris.
- The relationship of cooling efficiency with urban forms varies across the day.
- The cooling efficiency is more influenced by the 2D urban form than the 3D counterpart.
- Taller trees had higher CEs, while more varied tree heights led to smaller CEs.
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1. Introduction

Urban heat island (UHI) is a phenomenon in which urban areas have higher air and surface 36 37 temperatures than surrounding non-urban areas (Oke 1988; Rizwan et al. 2008; Voogt and Oke 2003). 38 UHI has an adverse effect on society, economy, and ecology (Akbari et al. 1996; Battles and Kolbe 39 2019; Herbel et al. 2018; Youngsteadt et al. 2015). It also detrimentally affecting the physical and 40 mental well-being of its residents of citizens (Araujo et al. 2015; Robine et al. 2008; Rydin et al. 2012). Against the backdrop of global warming and rapid urbanization, the heightened occurrence of extreme 41 42 heat events has emerged as a primary concern in urban areas (Habeeb et al. 2015; Tuholske et al. 2021). 43 According to the World Health Organization, 166,000 people died from heat stress between 2000 and 2017 globally, and the number of people exposed to heat waves increased by 125 million between 2000 44 and 2016 (Howard and Krishna 2022). Consequently, addressing the mitigation of UHI has become an 45 important focus in the realm of urban resilient development and research on urban sustainability (Akbari 46 47 et al. 2016; Mohajerani et al. 2017; Wong et al. 2021).

48 Urban green space (UGS), recognized as nature-based solutions (Seddon et al. 2020), have been 49 demonstrated as effective means to lower urban temperatures through the processes of 50 evapotranspiration and shading (Akbari et al. 2016; Marando et al. 2022; Seddon et al. 2020). Cooling 51 efficiency (CE), defined as the degree of temperature decrease resulting from a one-unit increase in 52 vegetation abundance (such as a 1% increase in the extent of UGS) has been widely used to evaluate 53 the function of UGS cooling functions (Cheng et al. 2023; Zhou et al. 2017). CE variations can arise from disparities in social and ecological context (Myint et al. 2015; Wang et al. 2020; Zhou et al. 2017). 54 For instance, cities situated in hot and arid biomes exhibit higher CE values when contrasted with cities 55 in hot and humid environments (Myint et al. 2015; Wang et al. 2020; Zhou et al. 2017). Furthermore, 56 CE values can also vary across seasons (Zhou et al. 2017), different vegetation types (Pataki et al. 2011), 57 and varying elevations (Zhao et al. 2014). However, how CE varies during a 24-hour cycle remains 58 59 largely unexplored.

60 This gap can now be filled with the data from the Ecosystem Spaceborne Thermal Radiometer
61 Experiment on Space Station (ECOSTRESS), a mission that commenced on June 29, 2018. It provides

62 the highest spatial resolution $(38m \times 69m)$ thermal infrared data from space (Hulley et al. 2021). In contrast to polar satellites, ECOSTRESS can provide diurnal cycling of land surface temperature (LST) 63 (Wang et al. 2023), and it has been widely used for urban thermal environmental research (Chang et al. 64 2021; Kamaraj et al. 2021; Vo and Hu 2021). For example, ECOSTRESS LST data was employed to 65 66 investigate diurnal variations in LST and heat exposure in Xi'an, China across diverse local climate zones (Chang et al. 2021; Yuan et al. 2022). In a parallel study, ECOSTRESS LST data was harnessed 67 to evaluate and model diurnal temperature buffering in the forest restoration area (Hamberg et al. 68 69 2022).

70 Comprehending the impact of intricate urban morphology on the spatial heterogeneity of the LST is pivotal in the context of mitigating the UHI effect (Berger et al. 2017; Guo et al. 2023; Huang 71 and Wang 2019; Wu et al. 2022). Indeed, 2D urban morphology metrics, like UGS coverage and its 72 proportion to impervious surface area, have been established as primary contributors to the spatial 73 74 variability of LST (Henits et al. 2017; Kikon et al. 2016). Recent research has demonstrated that incorporating 3D urban morphology factors can enhance our explanatory power by up to 20% when 75 assessing variations in UHI intensity (Wu et al. 2022). Among these 3D factors, building height and 76 tree height stand out as significant determinants in delineating LST variations within urban 77 78 environments (Berger et al. 2017; Guo et al. 2023; Yu et al. 2020). The above studies advance our 79 comprehension of how urban morphology influences LST and the UHI effect. Nevertheless, the precise 80 influence of 2D and 3D urban morphology on the diurnal CE of UGS remains unclear.

81 Here, we examined the diurnal fluctuations in CE during summertime in Paris, utilizing highresolution LST data provided by ECOSTRESS. To unravel the connections between urban morphology 82 83 and CE, we derived various 2D and 3D urban morphology metrics from high-resolution land cover data and detailed information on building and tree heights. Moreover, a more in-depth analysis was 84 conducted to examine the varied impacts of distinct urban form characteristics on CE during different 85 86 time intervals. Our investigation aimed to address the following two questions: (1) What are the spatiotemporal patterns of diurnal CE variations in the mega-city Paris Region? (2) How does urban 87 88 morphology influence diurnal CE?

2. Study area and data collection

90 2.1 Study area

91 Paris, the capital of France, is a sprawling European megacity in Europe in the interior of the 92 country with an area of 12,012 km² and a population of 12 million residents. Paris has received a great deal of attention for its UHI problem, especially since the 2003 heatwave event in Paris resulted in many 93 94 losses (Fouillet et al. 2008). The urban boundary is defined as an essential property of cities, and the global urban boundary (GUB) product was generated based on 30m global artificial impervious area 95 96 data (Gong et al. 2020; Li et al. 2020). Here we use GUB in 2015 to extract the urban core area of Paris, which is used as our study area. GUB data are accessible and can be downloaded from the website 97 of https://data-starcloud.pcl.ac.cn/zh. 98

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Fig.1. (a) Study area and land-cover map, where IS means imperious surface. (b) 3D urban informationin a sample case.

103 2.2 Land cover data

The high-resolution land cover dataset utilized in this study originates from diverse existing products., including the MOS+ 2017-81 (IPR 2017), the green cadaster of the Department Hauts-de-Seine (DHS 2012), vegetation height (APUR 2017) and building footprints (IPR 2018), Copernicus Small and Woody features (Copernicus 2018a), and the Copernicus Street tree layer (Copernicus 2018b). To accomplish this, data types were merged according to a ruleset, where priority is given to the highest resolution or thematic content. To reduce the number of thematic classes in the dataset, land- use and land-cover classes were harmonized and aggregated. The output was then rasterized to land cover map with seven types and 5 m resolution (Figure 1a). The identified categories cover building, non-building imperious surface (IS), grass, shrub, tree, water, and agricultural land cover.

113 2.3 Building and tree height

Building height and tree heights for the urban core area of Paris in 2015 were determined through the analysis of aerial photographs (Louis-Lucas et al. 2021), which were created by contrasting the Digital Surface Model (DSM) with the Digital Terrain Model (DTM). Both building footprint and tree height data are available on the website (<u>https://opendata.apur.org/search</u>), stored in raster format with a spatial resolution of 1m.

119 2.4 ECOSTRESS LST data

120 EOSTRESS LST data were obtained from ECOSTRESS Level-2 product, which is generated by a physic-based Temperature Emissivity Separation (TES) algorithm (Hook et al. 2019; Mira et al. 121 2007). ECOSTRESS LST has a repeating cycle every three to five days and a spatial resolution of 70m, 122 123 which is resampled from the original pixel size (Xiao et al. 2021). Based on a global scale validation, ECOSTRESS LST has a high accuracy compared to ground-based observations with an overall root 124 mean square error of 1.07K, mean absolute error of 0.4K and high r^2 (>0.988) (Hulley et al. 2021). Here 125 we employed LST products from the warm summer months (June, July, August, September) from 2018 126 to 2023. Figure2 shows the diurnal LST patterns in the study area. All the cloud pixels were excluded 127 using ECOSTRESS L2 cloud mask (ECO2CLD) product (Anderson et al. 2021). Meanwhile, the time 128 of all data is converted to local time by adding 2 hours according to the time zone of Paris (UTC+2). 129 The ECOSTRESS LST and ECO2CLOUD products used in this study can be freely accessed from the 130 website of https://www.earthdata.nasa.gov/. 131



Fig.2. (a)-(j) Diurnal variations and spatial pattern of LST over the urban area of Paris Region, all times
are CEST time (UTC+2). (k) Regional average LST for different time slots (in 24h) in the GUB region
of Paris.

138 3. Method

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3.1 Cooling efficiency calculation

According to the definition of CE, it signifies the reduction in temperature associated with a 140 141 one-unit increase in vegetation abundance (e.g. 1% increase in UGS cover) (Wang et al. 2020; Zhou et al. 2017). The calculation of CE can be performed using the following equation. In this study, the urban 142 143 green space coverage (UGSC) is defined as the total area covered by various types of vegetation, encompassing trees, shrublands, and grasslands. To elucidate the dynamics of CE at the city scale and 144 145 discern the spatial nuances of CE, assessments were conducted at both city and grid scales using the Ordinary Least Squares (OLS) linear regression. At the city scale, UGSC and average LST were 146 computed for each 840 m × 840 m grid and CE are conducted based on OLS linear regression. For the 147 CE at grid-scale, the UGSC calculations were conducted for each 70m grid, with CE computed for each 148 149 840 m × 840 m grid.

$$CE = -\frac{\Delta LST}{\Delta UGSC}$$

151 3.2 Urban morphology metrics

152 To elucidate the influence of urban morphology metrics on CE, this study employs a diverse suite of urban morphology metrics. For 2D metrics, a nuanced consideration of patch density, edge 153 density, largest patch index, landscape shape index, and percentage of patch type are undertaken across 154 various land cover categories. The analysis encompasses four distinct land cover types: building, non-155 156 building impervious surface (IS), tree, and non-tree vegetation, leading to the inclusion of a comprehensive set of 20 unique 2D metrics. Concurrently, the investigation extends to the realm of 3D 157 metrics, where 17 distinct metrics are derived from building and tree height information. These 158 159 encompass fundamental measures such as the mean and standard deviation of height, along with an 160 array of metrics characterizing diverse aspects of building forms, including largest patch index based and landscape shape index based on building surface area, etc. Table 1 shows how all metrics are 161 calculated and defined. The computation of all 2D landscape metrics is executed through the utilization 162 163 of the "landscapemetrics" package (Hesselbarth et al. 2019) in R (version 4.0.2). This computational 164 process is conducted on an 840 m × 840 m grid size. Furthermore, the extraction of building landscape

165 metrics in 3D space is facilitated through the application of the LPA3D software (Guo et al. 2022).

Urban forms	Jrban Metrics Abb. forms		Description	
	Patch density	PD	the ratio between number of patches to the areas, it describes the fragmentation of the landscape.	
	Edge density	ED	all edges in the landscape in relation to the landscap area.	
2D	Largest patch index	LPI	the percentage of the landscape covered by corresponding largest patch of each class.	
	Landscape shape index	LSI	the ratio between the actual landscape edge length and the hypothetical minimum edge length.	
	Percentage of patch type	PLAND	percentage of the landscape belonging to class i.	
	Sub-high building coverage	Sub-high_BC	area percentage of subhigh-rise buildings (30m- 100m).	
	Middle building coverage	Middle_BC	area percentage of middle-rise buildings (20m-30m	
	Sub-low building coverage	Sub-low_BC	area percentage of sublow-rise buildings (10m-20n	
	Low building coverage	Low_BC	area percentage of low-rise buildings (below 10m).	
3D	Building Shade metrics	CNI	the ratio between building height and building spacing.	
	Landscape division index	LDI	aggregation degree of buildings. LDI =0 when the landscape consists of single patch.	
	Largest patch index	LPI	largest space occupation of single building.	
	Landscape shape index	LSI	deviation between patch shape and regular circle of square with same.	
	Proximity index	PROX	the ratio between building height and square of building spacing.	
	Surface area	SA	surface fluctuation compared with plane area.	

166 **Table1.** 2D and 3D urban morphology metrics

Surface developed ratio	SDR	deviation of building surface to projected plane.	
Building surface slope	SSL	integral slope of building surface, which is the sum	
Building surface slope		of surface fluctuation at adjacent building pixels.	
Mean Volume index	VOL	mean volume of buildings.	
Mean building height	BH	mean height of the buildings.	
Standard deviation of	DII Sta	underlaging af de sud en heithigt an offen	
building height	BH_Sta	undulation of the urban buildings surface.	
Mean tree height	TH	average tree height.	
Standard deviation of tree	TU Std	undulation of the trace appendix	
height	11_500	undulation of the tree canopy.	

167 3.3 Statistical analysis

The boosted regression tree (BRT) model has been widely employed for investigating intricate 168 and non-linear relationships between variables due to its capacity to mitigate the risk of overfitting 169 (Friedman 2002). Here, a series of BRT models was utilized to assess the relative importance of various 170 urban metrics on CE across different time intervals. Metrics exhibiting a substantial relative importance 171 172 value are indicative of their primary role in depicting the spatial variability of CE. The BRT models were fitted using the gbm.step function within the "dismo" package (Elith and Leathwick 2017) in R 173 (version 4.0.2). The four BRT parameters, learning rate, tree complexity, number of trees, and bag 174 fraction, were set at 0.001, 5, 1000, and 0.75, respectively. 175

To examine the relationship between 2D/3D urban morphology metrics and diurnal CE, a series 176 177 of generalized linear models (GLM) were constructed. Employing Akaike's Information Criterion corrected (AICc) through an information-theoretic approach, the best-fitting GLM model was 178 179 determined based on the smallest AICc value. The correlation coefficient was utilized to gain further insights into the relationship between various urban morphology metrics and CE. Specifically, negative 180 181 coefficients signify that a specific indicator is linked to a reduction in CE values. Conversely, positive coefficients indicate that an escalation in the indicator's value is correlated with an increase in CE values. 182 Notably, only relationships deemed statistically significant (p-value < 0.05) were considered in this 183 study. 184

185 4. Result

186 4.1 Diurnal CE in city scale

As expected, the CE shows obvious diurnal variation, Fig.3 depicts the intricate association 187 between UGSC and CE across an 840m × 840m grid on a city scale. The statistical analysis reveals a 188 significant correlation between UGSC and LST (p-value < 0.05). On this scale, continuous monitoring 189 exhibited a range of CE values from 0.008 °C (21:03) to 0.13 °C (12:43), showcasing a mean of 0.059 °C 190 and a standard deviation of 0.04 °C. Concurrently, the coefficient of determination (R²) exhibited a 191 192 range from 0.013 (21:03) to 0.46 (08:34), illustrating a mean of 0.25 with a standard deviation of 0.15. Remarkably, this implies that up to 46% of the variations in LST can be elucidated by UGSC in urban 193 areas. Examining the diurnal dynamic pattern, CE demonstrates an upward trajectory from the late-194 night hours (00:22) to the peak at midday (12:43), registering a maximum value of 0.13 °C, followed 195 196 by a declining trend in the afternoon and evening hours. Similarly, the diurnal pattern of R² mirrors this trend, with the highest R^2 occurring at 08:34 (0.46). 197



Fig.3. (a)-(j) Scatter plots of the relationship between UGSC and LST at different times and the fitted line based on the OLS linear regression model, each point represents the average of the LST in an 840m grid cell and the coverage of the UGS in that grid. (k) is the time series of CE values and model R² at the city scale.

4.2 Diurnal CE in grid scale

The diurnal variation of CE exhibits dynamic patterns at different times. During nighttime intervals such as 00:22, 02:50, and 21:03 (see Fig. 4a, b, j), substantial portions of the area, particularly in the core city center, manifest negative CE values. Interestingly, in the late afternoon, specifically at 18:12 (Fig. 4i), a sizable area in the city center also reveals negative CE values. Contrastingly, for daytime periods (Fig. 4c-h), most of the region registers positive CE values. Notably, parks and forested areas exhibit significantly higher CE values compared to densely built-up areas. The cumulative analysis (Fig. 4k) underscores that, on average, the mean CE predominantly assumes positive values 211 throughout the study area, with discernibly elevated CE levels in the sub-core urban regions. Table 2 presents the statistical outcomes of diurnal CE corresponding to distinct time intervals. The highest CE 212 value during day and night in 840m × 840m grid was 0.613 °C (13:38). At 12:43, the highest mean CE 213 value of 0.038 °C was observed, coupled with the maximum standard deviation value of 0.035 °C. 214 215 Conversely, the minimum CE value of 0.0005 °C occurred at 21:03. In terms of the proportion of the 216 area exhibiting positive CE values, the urban core area of the study demonstrated values ranging from 72.6% to 93.7% at daytime. Notably, at both 21:03 and 00:22, over 40% of the area exhibited negative 217 218 CE values. On average, the mean CE value across all time intervals was calculated to be 0.016 °C, accompanied by a standard deviation of 0.012 °C. Remarkably, 94.5% of the urban core area in Paris 219 220 Region exhibited positive average CE values.



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Fig.4. The diurnal variations and spatial pattern of CE over Paris with a resolution of 840m. (a)-(j)illustrate the diurnal variations of CE at different times of the day. (k) shows the average CE over time,

- while (1) presents the distribution of CE values for different times. The red dashed line indicates a CEvalue of 0.
- 226 Table2. Statistics of CE (°C) at different observation times as presented in this study, where SD
- 227 represent standard deviation of CE values.

times	Daytime/nighttime	Max	Min	Mean	SD	Fraction (CE>0)
00:22	Nighttime	0.044	-0.045	0.001	0.009	58.6%
02:50	Nighttime	0.052	-0.025	0.004	0.005	78.2%
06:18	Daytime	0.042	-0.011	0.01	0.007	93.7%
08:34	Daytime	0.231	-0.041	0.023	0.02	91.4%
12:43	Daytime	0.405	-0.053	0.038	0.035	91.5%
13:38	Daytime	0.613	-0.068	0.029	0.029	90.5%
14:02	Daytime	0.423	-0.055	0.024	0.024	88.8%
16:45	Daytime	0.104	-0.034	0.02	0.017	90.2%
18:12	Daytime	0.042	-0.095	0.004	0.007	72.6%
21:03	Nighttime	0.024	-0.022	0.0005	0.004	52.6%
Average	-	0.082	-0.014	0.016	0.012	94.5%

4.3 Relative importance or urban morphology metrics for CE

Fig. 5 presents an assessment of the relative importance of the different urban morphology 230 metrics in influencing diurnal CE. On average, the top five influential urban morphology metrics are 231 232 tree coverage (22.3%), tree largest patch index (19.3%), non-tree coverage (9.6%), tree height (7%), and tree edge density (4.4%). Notably, at 6:18, tree coverage exhibits the highest relative importance 233 value, reaching 49.6%. The pivotal role of 2D and 3D urban morphology metrics in influencing CE is 234 underscored in Fig. 6, where, on average, the relative importance of 2D and 3D urban morphology 235 metrics are 77.25% and 22.75%, respectively. To the relative importance of CE across the entire day. 236 The cumulative impact of 2D and 3D urban morphology metrics on CE varies temporally. Specifically, 237

238 2D metrics assume heightened importance at 13:38, reaching 86.6%, whereas 3D metrics peak at 00:22,



constituting 39.94% of the total relative importance.



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Fig.5. Relative importance of urban morphology metrics on diurnal CE.





Fig.6. Relative importance changes of 2D and 3D urban morphology metrics in the day.

4.4 Relationships between urban morphology metrics and CE

To enhance our comprehension of the intricate relationship between CE and diverse urban morphology metrics, we employed a series of GLM regression analyses encompassing 37 urban morphology metrics. Figure 7 illustrates the regression coefficients elucidating the associations between urban morphology metrics and diurnal CE.

In the realm of 2D urban morphology metrics, both land cover composition and configuration 249 250 exert substantial influence on diurnal CE. Robust associations were observed, particularly in relation to building coverage and vegetation coverage. Notably, during daytime hours, UGS exhibit heightened 251 cooling efficiency in areas characterized by elevated building coverage, contrasting with the nighttime 252 253 scenario. Regarding vegetation cover, a conspicuous positive correlation is evident between CE and 254 both non-tree vegetation and tree coverage during day and night. Land cover configuration metrics also 255 manifest notable effects on diurnal CE. For instance, an increase in the LSI of impervious surfaces during the daytime was found to be associated with a decrease in CE. The relationship between CE and 256 257 LSI exhibited variations between trees and non-tree vegetation. Specifically, heightened shape complexity of non-tree vegetation significantly increased CE, whereas the LSI of tree patches showed 258 259 a significant negative correlation with CE as does non-treen vegetation patch density. Simultaneously, tree patch density demonstrated a significant positive relationship with CE. 260

For 3D urban morphology metrics, CE values increase as low-rise building coverage increases during daytime. The SA shows a negative correlation with CE both during day and night, while the SDR exhibits a negative correlation with CE during the day but a positive correlation at night. SSL is positively correlated with CE during the day but negatively correlated at night. The imagery of CE is influenced similarly by building height and tree height, during the day, both tall buildings and tall trees contribute to increased CE values. However, higher heterogeneity in building and tree heights results in smaller CE values.



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Urban morphology metrics

269 Fig7. Coefficients of various urban morphology metrics based on a series of GLM regression models. Only model results that pass the significance test (p-value<0.05) are shown here. 270

To further understand the image of LST for CE, Figure 8 shows the scatterplot distribution of 271 272 LST vs. CE. The results show that there is a nonlinear relationship between LST and CE, which is 273 manifested by the fact that when the temperature is lower than a certain threshold, the higher the 274 temperature is, the larger the value of CE is, but when LST is higher than a certain value, the value of CE decreases as the temperature becomes higher. Based on the change-points estimation (with possibly 275 random effects) algorithm (Muggeo and Muggeo 2017), the threshold value of LST can be estimated 276 as 36.15°C. 277



Fig.8. Relationship between LST and CE. The green point represents the mean LST and CE in an 840m
grid. The purple dotted line indicates the turning point of LST change when CE starts a new trend. The
turning point was extracted using the 'segmented' package (Muggeo and Muggeo 2017) in R (version
4.02). The red line represents the linear fitted relationship in the two stages.

283 5. Discussion

5.1 Diurnal dynamics of CE

By means of using the ECOSTRESS LST, the study showed how CE changes throughout the 285 day and explored the influence of 2D and 3D urban structures. Unlike traditional analysis based on 286 287 polar orbiting satellites thermal data like from Landsat (Kong et al. 2014), ASTER (Myint et al. 2015), and MODIS (Yang et al. 2022), ECOSTRESS LST allows us to examine UGS cooling abilities 288 throughout the entire day because it's unfixed observation time (Hulley et al. 2019). By focusing on the 289 Paris region, our case study showcases ECOSTRESS LST's effectiveness in revealing the spatial 290 patterns of CE during a 24-hour cycle. Our findings enhance our understanding of intra-city urban heat 291 dynamics and highlight the role of UGS in cooling the city during over the course of day and night. CE 292

varies greatly over the course, ranging from 0.008 °C to 0.13 °C in the city scale, this is consistent with
previous studies (0.023 °C to 0.318 °C) (Wang et al. 2020; Yang et al. 2022). There was a large difference
between daytime and nighttime CE, with nighttime CE being smaller compared to daytime, possibly
due to the absence of transpiration and shading effects of vegetation at night.

In addition, unlike previous studies on the variation of CE at the regional scale (Kong et al. 297 2014; Wang et al. 2020). Based on refined temperature and land cover data, this study attempts to 298 compute CE in a smaller grid scale (840 m) to better understand the spatial distribution as well as the 299 diurnal variation characteristics of CE in a large city. The results demonstrated that CE is highly 300 301 heterogeneous in both time and space, for example, the average CE value in one day has a standard deviation of 0.012°C. The spatial distribution of CE at different times (Figure 4) also indicates that the 302 cooling benefit of UGS has a strong inequality in the spatial distribution, and the map of CE can provide 303 an effective reference to minimize this inequality. 304

305 5.2 Significant effects of urban morphology on CE

Based on explainable machine learning model, we find out that both 2D and 3D urban 306 morphology metrics have significant effects on the CE of UGS. Specifically, the 2D urban forms play 307 a more important role compared to 3D urban forms, because the 2D urban heterogeneity, such as land 308 cover composition, is the main cause of land surface temperature variations (Rahimi et al. 2021). 309 310 Meanwhile, the relative importance of 2D and 3D urban morphology metrics on CE shows a significant 311 diurnal variation, where especially the 3D urban morphology metrics increase their relative importance in the midday (13:38) and in the deep night (00:22). This could be attributed to the increased prominence 312 of tree shadows at 13:38, as the three-dimensional structure of the building significantly amplifies the 313 heterogeneity of thermal storage capacity during nighttime. 314

Buildings exert a pronounced influence on the cooling efficiency of urban green spaces. On one hand, areas with higher building coverage experience enhanced cooling efficiency of vegetation, possibly attributed to the elevated surface temperatures in densely built areas, enabling vegetation to yield more substantial cooling effects. However, it is noteworthy that the cooling impact of vegetation becomes subdued as temperatures reach a certain threshold (Figure 8), as illustrated in Figure 7 where
the correlation coefficient between building coverage and cooling efficiency peaks at 12:43.

Although CE can be significantly improved by increasing UGS cover, there are differences in 321 the effects of tree and non-tree vegetation patterns on cooling efficiency. For example, increasing patch 322 density or edge density of tree patches can significantly improve CE. This was mainly because that 323 increasing edge density has the potential to allow trees to provide greater shading on continuous 324 impervious surfaces (Wu et al. 2022). In addition, it expands energy flow and exchange between UGS 325 and their surroundings (Cadenasso et al. 2003). However, non-tree vegetation (e.g., grasses and shrubs) 326 327 lacks a shading effect, so increasing its degree of fragmentation (high ED and PD) rather reduces its cooling effect. 328

5.3 Implications for urban heat island effect mitigation

By engaging in this research, urban planners can gain important perspectives that will help them in designing urban areas and managing UGS to effectively counteract the impacts of urbanization on UHI. Based on the results of this study, we proposed the following recommendations for improving CE of UGS.

- (1) Increased shape complexity and irregularity of tree patches decreases the CE of UGS, but
 increased edge density and patch density improves the CE, so when constructing new tree,
 consider laying out trees in a decentralized manner, and, at the same time, pursuing regularity
 in the shapes of individual patches.
- (2) In contrast to trees, increasing the shape irregularity and complexity of non-tree vegetation can effectively optimize CE, while increasing ED or PD of non-tree vegetation can lead to lower
 CE. Therefore, when constructing new vegetation, such as grass and shrubs, continuous, complete, and more complex forms should be introduced rather than fragmented regular UGS.
 (3) Considering taller trees when selecting tree species and ensuring consistency of tree heights in patches can effectively optimize CE.

344 5.4 Limitations and future research directions

Limitations and uncertainties accompany this study. The ECOSTRESS LST data were acquired 345 from different days, even though we filtered the data from summertime, but there still have uncertainties 346 for diurnal variation of CE research. More investigations on the method for adjusting ECOSTRESS 347 LST data may need to be conducted in the future. CE of UGS can be affected by many factors such as 348 349 tree species, leaf size and color, climatic background of the region, and even anthropogenic heat emissions. More fieldwork and measurements of tree species and detailed climate records will support 350 a better understanding of factors affecting CE. Meanwhile, a machine learning model can elucidate the 351 intricate relationship between urban morphology and CE. However, it falls short in providing a 352 353 comprehensive understanding of the complete trajectory of these influences. Future research endeavors could explore the incorporation of indicators such as shadows and evapotranspiration (ET) as 354 interpermeates, aiming to deepen our comprehension of the intricate ways in which urban morphologies 355 impact CE. In this study, a grid-based method was employed, yet it focused solely on a single scale. It 356 357 is imperative to acknowledge that varying grid scales may yield divergent results (Guo et al. 2023). Consequently, the exploration of scale-effect phenomena should be prioritized in future investigations. 358 Lastly, owing to the unavailability of spatial-temporal continent air temperature data, our assessment 359 was confined to evaluating diurnal CE dynamics based on LST. The evaluation of spatial-temporal 360 361 changes in canopy temperature-based CE holds significant value and should be a focal point in forthcoming studies. 362

363 6. Conclusions

The main objective of this study is to investigate the diurnal dynamics of CE and, subsequently, to explore the impact of urban form (both 2D and 3D) on CE. To achieve this objective, we utilize ECOSTRESS LST data, overcoming the limitations associated with polar-orbiting satellites (e.g., Landsat and MODIS) observed in previous studies. The ECOSTRESS LST data allows us to acquire diurnal LST information, enabling a detailed analysis of the diurnal CE patterns in the study area, here in the Paris Region. To better understand the relationship between urban form and CE, we constructed 370 twenty 2D morphological indicators and seventeen 3D morphological indicators based on highresolution land cover data and urban 3D morphological information (including building and tree height 371 information). The relative importance and relevance of different urban morphological indicators for CE 372 are analyzed using BRT and GLM models. The results reveal a pronounced heterogeneity of CE both 373 temporally and spatially throughout the day. Additionally, the influence of urban structure on CE 374 exhibits significant variations at different times of the day. Drawing from the study's outcomes, practical 375 suggestions are presented to enhance the CE of UGS, with potential applications in real-world scenarios. 376 377 This study marks the first analysis of the diurnal variation of CE and its correlation with the 2D and 3D morphology of the city, introducing a novel perspective to the study of the urban thermal environment. 378 379 The insights gained from this research can be instrumental in optimizing the CE of UGS in urban areas, 380 thereby mitigating the adverse effects of extreme heat and enhancing urban resilience.

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Reference

387	Akbari, H., Cartalis, C., Kolokotsa, D., Muscio, A., Pisello, A.L., Rossi, F.,
388	Santamouris, M., Synnefa, A., Wong, N.H., & Zinzi, M. (2016). Local climate change and
389	urban heat island mitigation techniques-the state of the art. Journal of Civil Engineering and
390	Management, 22, 1-16
391	Akbari, H., Rosenfeld, A., Taha, H., & Gartland, L. (1996). Mitigation of summer
392	urban heat islands to save electricity and smog. In, 76th Annual Meteorological Society
393	Meeting, Atlanta, GA
394	Anderson, M.C., Yang, Y., Xue, J., Knipper, K.R., Yang, Y., Gao, F., Hain, C.R.,
395	Kustas, W.P., Cawse-Nicholson, K., & Hulley, G. (2021). Interoperability of ECOSTRESS
396	and Landsat for mapping evapotranspiration time series at sub-field scales. Remote Sensing of
397	Environment, 252, 112189
398	APUR (2017). Hauteur vegetation 2015. In: APUR
399	Araujo, R.V., Albertini, M.R., Costa-da-Silva, A.L., Suesdek, L., Franceschi, N.C.S.,
400	Bastos, N.M., Katz, G., Cardoso, V.A., Castro, B.C., & Capurro, M.L. (2015). São Paulo
401	urban heat islands have a higher incidence of dengue than other urban areas. Brazilian
402	Journal of Infectious Diseases, 19, 146-155
403	Battles, A.C., & Kolbe, J.J. (2019). Miami heat: urban heat islands influence the
404	thermal suitability of habitats for ectotherms. Global change biology, 25, 562-576
405	Berger, C., Rosentreter, J., Voltersen, M., Baumgart, C., Schmullius, C., & Hese, S.
406	(2017). Spatio-temporal analysis of the relationship between 2D/3D urban site characteristics
407	and land surface temperature. Remote Sensing of Environment, 193, 225-243
408	Cadenasso, M.L., Pickett, S.T., Weathers, K.C., & Jones, C.G. (2003). A framework
409	for a theory of ecological boundaries. BioScience, 53, 750-758
410	Chang, Y., Xiao, J., Li, X., Middel, A., Zhang, Y., Gu, Z., Wu, Y., & He, S. (2021).
411	Exploring diurnal thermal variations in urban local climate zones with ECOSTRESS land
412	surface temperature data. Remote Sensing of Environment, 263, 112544
413	Cheng, X., Liu, Y., Dong, J., Corcoran, J., & Peng, J. (2023). Opposite climate
414	impacts on urban green spaces' cooling efficiency around their coverage change thresholds in
415	major African cities. Sustainable Cities and Society, 88, 104254
416	Copernicus (2018a). Small Woody Features 2018. In: European Environment Agency
417	Copernicus (2018b). Urban Atlas Street Tree Layer 2018. In: European Environment
418	Agency

419	DHS (2012). Cadastre vert - Masses vertes. In: Department Hauts-de-Seine
420	Elith, J., & Leathwick, J. (2017). Boosted Regression Trees for ecological modeling.
421	R Documentation. Available online: <u>https://cran</u> . r-project.
422	org/web/packages/dismo/vignettes/brt. pdf (accessed on 12 June 2011)
423	Fouillet, A., Rey, G., Wagner, V., Laaidi, K., Empereur-Bissonnet, P., Le Tertre, A.,
424	Frayssinet, P., Bessemoulin, P., Laurent, F., & De Crouy-Chanel, P. (2008). Has the impact
425	of heat waves on mortality changed in France since the European heat wave of summer
426	2003? A study of the 2006 heat wave. International journal of epidemiology, 37, 309-317
427	Friedman, J.H. (2002). Stochastic gradient boosting. Computational statistics & data
428	analysis, 38, 367-378
429	Gong, P., Li, X., Wang, J., Bai, Y., Chen, B., Hu, T., Liu, X., Xu, B., Yang, J., &
430	Zhang, W. (2020). Annual maps of global artificial impervious area (GAIA) between 1985
431	and 2018. Remote Sensing of Environment, 236, 111510
432	Guo, F., Schlink, U., Wu, W., Hu, D., & Sun, J. (2023). Scale-dependent and season-
433	dependent impacts of 2D/3D building morphology on land surface temperature. Sustainable
434	Cities and Society, 97, 104788
435	Guo, F., Schlink, U., Wu, W., & Mohamdeen, A. (2022). Differences in Urban
436	Morphology between 77 Cities in China and Europe. Remote Sensing, 14, 5462
437	Habeeb, D., Vargo, J., & Stone, B. (2015). Rising heat wave trends in large US cities.
438	Natural Hazards, 76, 1651-1665
439	Hamberg, L.J., Fisher, J.B., Ruppert, J.L., Tureček, J., Rosen, D.H., & James, P.M.
440	(2022). Assessing and modeling diurnal temperature buffering and evapotranspiration
441	dynamics in forest restoration using ECOSTRESS thermal imaging. Remote Sensing of
442	Environment, 280, 113178
443	Henits, L., Mucsi, L., & Liska, C.M. (2017). Monitoring the changes in impervious
444	surface ratio and urban heat island intensity between 1987 and 2011 in Szeged, Hungary.
445	Environmental monitoring and assessment, 189, 1-13
446	Herbel, I., Croitoru, AE., Rus, A.V., Roșca, C.F., Harpa, G.V., Ciupertea, AF., &
447	Rus, I. (2018). The impact of heat waves on surface urban heat island and local economy in
448	Cluj-Napoca city, Romania. Theoretical and applied climatology, 133, 681-695
449	Hesselbarth, M.H., Sciaini, M., With, K.A., Wiegand, K., & Nowosad, J. (2019).
450	landscapemetrics: an open-source R tool to calculate landscape metrics. Ecography, 42,
451	1648-1657

452	Hook, S.J., Cawse-Nicholson, K., Barsi, J., Radocinski, R., Hulley, G.C., Johnson,
453	W.R., Rivera, G., & Markham, B. (2019). In-flight validation of the ECOSTRESS, Landsats
454	7 and 8 thermal infrared spectral channels using the Lake Tahoe CA/NV and Salton Sea CA
455	automated validation sites. IEEE Transactions on Geoscience and Remote Sensing, 58, 1294-
456	1302
457	Howard, S., & Krishna, G. (2022). How hot weather kills: the rising public health
458	dangers of extreme heat. <i>bmj</i> , 378
459	Huang, X., & Wang, Y. (2019). Investigating the effects of 3D urban morphology on
460	the surface urban heat island effect in urban functional zones by using high-resolution remote
461	sensing data: A case study of Wuhan, Central China. ISPRS Journal of Photogrammetry and
462	Remote Sensing, 152, 119-131
463	Hulley, G., Shivers, S., Wetherley, E., & Cudd, R. (2019). New ECOSTRESS and
464	MODIS land surface temperature data reveal fine-scale heat vulnerability in cities: A case
465	study for Los Angeles County, California. Remote Sensing, 11, 2136
466	Hulley, G.C., Göttsche, F.M., Rivera, G., Hook, S.J., Freepartner, R.J., Martin, M.A.,
467	Cawse-Nicholson, K., & Johnson, W.R. (2021). Validation and quality assessment of the
468	ECOSTRESS level-2 land surface temperature and emissivity product. IEEE Transactions on
469	Geoscience and Remote Sensing, 60, 1-23
470	IPR (2017). Mode d'occupation du sol (MOS) 2017 a 81 postes. In: L'Institut Paris
471	Region
472	IPR (2018). Densibati 2018. In: L'Institut Paris Region
473	Kamaraj, N.P., Shekhar, S., Sivashankari, V., Balasubramani, K., & Prasad, K.A.
474	(2021). Detecting heat-inducing urban built-up surface material with multi remote sensing
475	datasets using reflectance and emission spectroscopy. Remote Sensing of Environment, 264,
476	112591
477	Kikon, N., Singh, P., Singh, S.K., & Vyas, A. (2016). Assessment of urban heat
478	islands (UHI) of Noida City, India using multi-temporal satellite data. Sustainable Cities and
479	Society, 22, 19-28
480	Kong, F., Yin, H., James, P., Hutyra, L.R., & He, H.S. (2014). Effects of spatial
481	pattern of greenspace on urban cooling in a large metropolitan area of eastern China.
482	Landscape and Urban Planning, 128, 35-47
483	Li, X., Gong, P., Zhou, Y., Wang, J., Bai, Y., Chen, B., Hu, T., Xiao, Y., Xu, B., &
484	Yang, J. (2020). Mapping global urban boundaries from the global artificial impervious area
485	(GAIA) data. Environmental Research Letters, 15, 094044

486	Louis-Lucas, T., Mayrand, F., Clergeau, P., & Machon, N. (2021). Remote sensing
487	for assessing vegetated roofs with a new replicable method in Paris, France. Journal of
488	applied remote sensing, 15, 014501-014501
489	Marando, F., Heris, M.P., Zulian, G., Udías, A., Mentaschi, L., Chrysoulakis, N.,
490	Parastatidis, D., & Maes, J. (2022). Urban heat island mitigation by green infrastructure in
491	European Functional Urban Areas. Sustainable Cities and Society, 77, 103564
492	Mira, M., Valor, E., Boluda, R., Caselles, V., & Coll, C. (2007). Influence of soil
493	water content on the thermal infrared emissivity of bare soils: Implication for land surface
494	temperature determination. Journal of Geophysical Research: Earth Surface, 112
495	Mohajerani, A., Bakaric, J., & Jeffrey-Bailey, T. (2017). The urban heat island effect,
496	its causes, and mitigation, with reference to the thermal properties of asphalt concrete.
497	Journal of environmental management, 197, 522-538
498	Muggeo, V.M., & Muggeo, M.V.M. (2017). Package 'segmented'. Biometrika, 58,
499	516
500	Myint, S.W., Zheng, B., Talen, E., Fan, C., Kaplan, S., Middel, A., Smith, M., Huang,
501	HP., & Brazel, A. (2015). Does the spatial arrangement of urban landscape matter?
502	Examples of urban warming and cooling in Phoenix and Las Vegas. Ecosystem Health and
503	Sustainability, 1, 1-15
504	Oke, T.R. (1988). The urban energy balance. Progress in Physical geography, 12,
505	471-508
506	Pataki, D.E., McCarthy, H.R., Litvak, E., & Pincetl, S. (2011). Transpiration of urban
507	forests in the Los Angeles metropolitan area. Ecological Applications, 21, 661-677
508	Rahimi, E., Barghjelveh, S., & Dong, P. (2021). Quantifying how urban landscape
509	heterogeneity affects land surface temperature at multiple scales. Journal of Ecology and
510	Environment, 45, 1-13
511	Rizwan, A.M., Dennis, L.Y., & Chunho, L. (2008). A review on the generation,
512	determination and mitigation of Urban Heat Island. Journal of environmental sciences, 20,
513	120-128
514	Robine, JM., Cheung, S.L.K., Le Roy, S., Van Oyen, H., Griffiths, C., Michel, JP.,
515	& Herrmann, F.R. (2008). Death toll exceeded 70,000 in Europe during the summer of 2003.
516	Comptes rendus biologies, 331, 171-178
517	Rydin, Y., Bleahu, A., Davies, M., Dávila, J.D., Friel, S., De Grandis, G., Groce, N.,
518	Hallal, P.C., Hamilton, I., & Howden-Chapman, P. (2012). Shaping cities for health:

519	complexity and the planning of urban environments in the 21st century. <i>The lancet, 379</i> ,
520	2079-2108
521	Seddon, N., Chausson, A., Berry, P., Girardin, C.A., Smith, A., & Turner, B. (2020).
522	Understanding the value and limits of nature-based solutions to climate change and other
523	global challenges. Philosophical Transactions of the Royal Society B, 375, 20190120
524	Tuholske, C., Caylor, K., Funk, C., Verdin, A., Sweeney, S., Grace, K., Peterson, P.,
525	& Evans, T. (2021). Global urban population exposure to extreme heat. Proceedings of the
526	National Academy of Sciences, 118, e2024792118
527	Vo, T.T., & Hu, L. (2021). Diurnal evolution of urban tree temperature at a city scale.
528	Scientific Reports, 11, 10491
529	Voogt, J.A., & Oke, T.R. (2003). Thermal remote sensing of urban climates. Remote
530	Sensing of Environment, 86, 370-384
531	Wang, J., Zhou, W., Jiao, M., Zheng, Z., Ren, T., & Zhang, Q. (2020). Significant
532	effects of ecological context on urban trees' cooling efficiency. ISPRS Journal of
533	Photogrammetry and Remote Sensing, 159, 78-89
534	Wang, Q., Wang, X., Meng, Y., Zhou, Y., & Wang, H. (2023). Exploring the impact
535	of urban features on the spatial variation of land surface temperature within the diurnal cycle.
536	Sustainable Cities and Society, 91, 104432
537	Wong, N.H., Tan, C.L., Kolokotsa, D.D., & Takebayashi, H. (2021). Greenery as a
538	mitigation and adaptation strategy to urban heat. Nature Reviews Earth & Environment, 2,
539	166-181
540	Wu, WB., Yu, ZW., Ma, J., & Zhao, B. (2022). Quantifying the influence of 2D
541	and 3D urban morphology on the thermal environment across climatic zones. Landscape and
542	Urban Planning, 226, 104499
543	Xiao, J., Fisher, J.B., Hashimoto, H., Ichii, K., & Parazoo, N.C. (2021). Emerging
544	satellite observations for diurnal cycling of ecosystem processes. Nature Plants, 7, 877-887
545	Yang, Q., Huang, X., Tong, X., Xiao, C., Yang, J., Liu, Y., & Cao, Y. (2022). Global
546	assessment of urban trees' cooling efficiency based on satellite observations. Environmental
547	Research Letters, 17, 034029
548	Youngsteadt, E., Dale, A.G., Terando, A.J., Dunn, R.R., & Frank, S.D. (2015). Do
549	cities simulate climate change? A comparison of herbivore response to urban and global
550	warming. Global change biology, 21, 97-105
551	Yu, S., Chen, Z., Yu, B., Wang, L., Wu, B., Wu, J., & Zhao, F. (2020). Exploring the
552	relationship between 2D/3D landscape pattern and land surface temperature based on

- 553 explainable eXtreme Gradient Boosting tree: A case study of Shanghai, China. *Science of the*
- 554 *Total Environment*, 725, 138229
- 555 Yuan, B., Zhou, L., Hu, F., & Zhang, Q. (2022). Diurnal dynamics of heat exposure in
- 556 Xi'an: A perspective from local climate zone. *Building and Environment, 222*, 109400
- 557 Zhao, L., Lee, X., Smith, R.B., & Oleson, K. (2014). Strong contributions of local
- background climate to urban heat islands. *Nature*, *511*, 216-219
- 559 Zhou, W., Wang, J., & Cadenasso, M.L. (2017). Effects of the spatial configuration of
- trees on urban heat mitigation: A comparative study. Remote Sensing of Environment, 195, 1-
- 561 12