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1	Non-targeted analysis of lipidic extracts by high-resolution mass spectrometry to characterise the
2	chemical exposome: comparison of four clean-up strategies applied to egg
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17 Abstract

Biota samples are used to monitor chemical stressors and their impact on the ecosystem and to 18 19 describe dietary chemical exposure. These complex matrices require an extraction step followed by 20 clean-up to avoid damaging sensitive analytical instruments based on chromatography coupled to 21 mass spectrometry. While interest for non-targeted analysis (NTA) is increasing, there is no versatile 22 or generic sample preparation for a wide range of contaminants suitable for a diversity of biotic 23 matrices. Among the contaminants' variety, persistent contaminants are mostly hydrophobic (mid- to 24 non-polar) and bio-magnify through the lipidic fraction. During their extraction, lipids are generally co-25 extracted, which may cause matrix effect during the analysis such as hindering the acquired signal. The 26 aim of this study was to evaluate the efficacy of four clean-up methods to selectively remove lipids 27 from extracts prior to NTA. We evaluated (i) gel permeation chromatography (GPC), (ii) Captiva EMR-28 lipid cartridge (EMR), (iii) sulphuric acid degradation (H₂SO₄) and (iv) polydimethyl siloxane (PDMS) for 29 their efficiency to remove lipids from hen egg extracts. Gas and liquid chromatography coupled with 30 high-resolution mass spectrometry fitted with either electron ionisation or electrospray ionisation 31 sources operating in positive and negative modes were used to determine the performances of the 32 clean-up methods. A set of 102 chemicals with a wide range of physico-chemical properties that covers 33 the chemical space of mid- to non-polar contaminants, was used to assess and compare recoveries and 34 matrix effects. Matrix effects, that could hinder the mass spectrometer signal, were lower for extracts 35 cleaned-up with H₂SO₄ than for the ones cleaned-up with PDMS, EMR and GPC. The recoveries were 36 satisfactory for both GPC and EMR while those determined for PDMS and H₂SO₄ were low due to poor 37 partitioning and degradation/dissociation of the compounds, respectively. The choice of the clean-up methods, among those assessed, should be a compromise that takes into account the matrix under 38 39 consideration, the levels and the physico-chemical properties of the contaminants.

40 <u>Keywords</u>

- 41 Non-targeted analysis; Clean-up; Lipid removal; High-resolution mass spectrometry; Recoveries;
- 42 Matrix effects.

43 1. Introduction

Over the past 30 years, more than 150,000 chemicals have been registered for commercial use in 44 45 Europe, the United States and Canada [1]. The Chemicals Abstracts Service has assigned over 204 46 million registry numbers as of June 2023, i.e. 64 million more than in September 2018 and 164 million 47 more than in 2002 [2]. These chemicals' hazard and fate are less well characterised than the legacy 48 contaminants. A portion of them may be released into the environment in various ways and may enter 49 food chains with potential health issues for wildlife and humans [3,4]. Also, the transformation 50 products, which may in certain cases be more persistent or toxic than the parent compound [5], 51 broaden the spectrum of physico-chemical properties of the compounds to be characterised.

52 Hazardous and persistent chemicals need to be monitored to manage the risks they pose and ensure healthy ecosystems. Their analysis in biotic samples generally involves sample preparation, combining 53 54 extraction and purification followed by data acquisition [6]. Chromatography coupled with mass 55 spectrometry makes it possible to separate, identify and quantify contaminants. Non-targeted 56 strategies, including suspect screening, are designed to describe samples' comprehensive fingerprints. 57 As opposed to targeted analysis, non-targeted analysis (NTA), based on full scan mode high-resolution 58 mass spectrometry (HRMS), is characterised by its non-selective data acquisition for thousands of 59 chemicals over a wide mass range. A major drawback of full scan mode is its lower sensitivity compared 60 to acquisition modes used for targeted analysis such as multiple reaction monitoring [7]. Furthermore, 61 when performing mass spectrometry analysis on biotic samples, the significant presence of matrix 62 interferents can cause ion suppression or enhancement, hindering the signal acquisition [8,9].

The development and implementation of sample preparation for contaminants analysis in biota by NTA is a challenge due to their wide physico-chemical properties (octanol-water partition coefficient *K*_{ow}, molecular mass, stability) and the complexity of these matrices (essentially lipids, proteins, water). Nevertheless, due to the lack of generic sample preparation methods for NTA, most studies available in the literature used sample preparation inspired from targeted methods to extract (solvent

68 extraction, QuEChERS) and clean-up (GPC, SPE, multilayer silica) the samples [10]. In NTA workflows, 69 mid- to non-polar persistent contaminants are generally extracted from biotic environmental and food 70 samples with solvent using pressurised liquid extraction [11–14], ultrasonication [15,16] or maceration 71 [17,18]. These extractions involve the use of non-polar solvent or a mixture of non-polar and slightly 72 polar solvents resulting in the co-extraction of matrix components such as lipids (triglycerides, 73 phospholipids etc.) which could cause analytical problems such as poor chromatographic separation 74 or alteration of the signal acquired. To preserve most of the contaminants in the extract while ensuring 75 compatibility with analytical instruments based on chromatography coupled with HRMS, selective lipid 76 clean-up methods should be used after biota extractions. Recently Dubocq et al. [19] compared several 77 extractions and clean-up methods for fish tissue, a fatty matrix, and concluded that ultrasonication 78 extraction followed by deactivated silica clean-up presented the best recoveries and repeatability for 79 mid- to non-polar contaminants (pesticides, flame retardants). Since deactivated silica clean-up involves hydrophobic interactions, it is selective for both lipids but also hydrophobic contaminants that 80 81 can be removed from the extract. In this study, four less selective clean-ups methods for which the 82 efficiency of isolating lipids and other matrix components from contaminants in lipid extracts were 83 assessed. These methods, as detailed below, were specifically selected for their potential to remove 84 lipids, via their ability to take advantage of the physico-chemical properties of lipids such as their 85 molecular size and reactive chemical group.

Most classes of lipids consist of linked fatty acids with 18 carbon atoms in their hydrocarbon chains, although this number could theoretically range between 4 and 26 carbon atoms [20]. These lipids, which may be larger than the chemical contaminants, can thus be separated using gel permeation chromatography (GPC), a separation technique based on size exclusion. GPC was used as early as 1972 to isolate pesticides and PCBs from fish lipids [21]. Since then, GPC has been used for both targeted analysis [22] and NTA [13] on complex fatty extracts from biotic samples by collecting cleaned-up extracts after the elution of the largest lipids amount and other large matrix components. However,

93 when using this method, chemicals similar in molecular size to lipids are lost in the process, such as
94 contaminants with long carbon chain (e.g., phthalates, chlorinated paraffins, PFAS).

Faced with this challenge associated to the absence of a satisfactory protocol, and to respond to this 95 96 demand, the analytical industry took into account the problems posed by lipid extracts and proposed 97 an easy-to-implement solution. In 2017, Agilent commercialised Enhanced Matrix Removal (EMR)-98 Lipid, a phase, initially used as a dispersive phase, embedded into a cartridge that combines size 99 exclusion and hydrophobic interactions to selectively remove lipids from extracts. To improve the 100 efficiency of lipid removal, it has been recommended to load an extract containing at least 10% water 101 [23,24]. The lipid selectivity of this cartridge appears promising for cleaning-up lipidic extracts with 102 recoveries between 60% and 120% and minor matrix effects (8% of mean signal enhancement) [25,26].

As fatty acids are mostly bound to glycerol or phosphoric acid in the form of esters they can be hydrolysed with sulphuric acid and water [27,28]. The literature reports examples of comparing this strategy with GPC after the cleaning step, concluding that acidification had a negative effect on several compounds such as bromoindole and halogenated methyl bipyrroles [29]. It is assumed that only nondissociated compounds under acidic conditions, acid-resistant compounds (POPs and other persistent chemicals) and degradation products generated during the process could be detected after such destructive clean-up [30].

110 Other strategies such as the use of polydimethylsiloxane (PDMS) as a sampler, used to extract 111 contaminants from adipose tissues in order to avoid cleaning steps, have also been reported. Indeed, 112 contaminants are extracted from tissues by passive diffusion towards the PDMS. PDMS has been used 113 as passive sampler in several geometries: solid phase microextraction coating [31], microtube [32] and 114 thin-film [33,34]. The kinetics equilibrium in diverse mammalian tissues was extensively reviewed [35] 115 and the more lipid-rich a tissue is, the faster is time to equilibrium with lipid-rich tissues and pure lipid reaching steady state within 24 h [36]. As the partition coefficient between PDMS and lipid varies by 116 117 less than one to two orders of magnitude for chemicals covering 10 orders of magnitude in

hydrophobicity [37,38], partition with PDMS is in principle amenable to NTA because all chemicals are
extracted with similar efficacy.

120 In order to comprehensively address the problem linked to the presence of lipids in extracts dedicated 121 to the analysis of non-polar contaminants, this study aimed to compare the performances of these 122 four techniques by adapting the published methods (GPC, EMR, H₂SO₄ and PDMS). Hen eggs were used 123 as a model matrix due to their high lipid content but also the diversity of lipid classes present [39]. A 124 set of 102 chemicals exhibiting a broad range of physico-chemical properties were selected to evaluate 125 recoveries and matrix effects. Cleaned-up extracts were analysed by GC-EI(+)-HRMS and LC-ESI(±)-126 HRMS. Linearity, limit of detection (LOD) and combination of the analytical methods were evaluated 127 prior the methods performances assessment.

128

129 2. Materials and methods

130 **2.1. Chemicals and materials**

131 Toluene, acetone, cyclohexane, ethyl acetate (LV-GC SuperTrace grade) and dichloromethane (Dioxins, 132 Pesti-S, Furans, PCB's Analysis grade) were purchased from Biosolve (Valkenswaard, The Netherlands). 133 Water and acetonitrile LC-MS grade were obtained from VWR (Radnor, PA, USA) and Riedel-de-Haën 134 (Seelze, Germany), respectively. Concentrated sulphuric acid (98%) was provided by Panreac 135 (Barcelona, Spain). Anhydrous sodium sulphate was purchased from Merck (Emsure® grade, 136 Darmstadt, Germany). Captiva EMR-lipid cartridges (6 mL, 600 mg) were obtained from Agilent 137 Technologies (Santa Clara, CA, USA). PDMS sheets (SSP-M823, Special Silicone Products, Ballston, USA) with thickness of 1 mm and a density of 1.17 g cm⁻³ were provided by Shieldings Solutions (Great 138 139 Notley, Great Britain).

The performance of the methods was assessed using a spiking solution constituted of 102 compounds at concentration range from 0.05 to 1 ng.μL⁻¹ depending on the concentration of the individual reference standards (Table B.1). In order to simplify the discussion, we will refer to compound

concentrations as the dilution of the spiking solution hereafter. Extensive details such as acronyms,
 molecular formula, InChiKeys, compound class and supplier are also provided in Table B.1. ¹³C₁₀-anti Dechlorane Plus (anti-DP), used as external standard for the fraction analysed by GC-EI(+)-HRMS, was
 purchased from Cambridge Isotope Laboratories (Andover, MA, USA). ¹³C₁₂–Tetrabromobisphenol A
 (TBBPA) and ²H₁₈-β-hexabromocyclododecane (HBCDD), used as external standard for the fraction
 analysed by LC-ESI(±)-HRMS, were obtained from Wellington Laboratories (Guelph, Ontario, Canada).

To minimize procedural contamination, several precautions were taken. All glassware was heated at
400 °C for 4 h before use. Teflon caps and magnetic bar were rinsed with dichloromethane before use.
Handling of samples and extracts was carried out in an overpressure room as much as possible.

152

2.2. Model lipid extract solution

153 Eggs from caged hens were purchased at a local store (Nantes, France) in November 2021. Whole eggs 154 were pooled and freeze-dried resulting in a weight loss of 75.4%. The fat content of the pool was 155 determined gravimetrically to be 10.4% (*l.w./w.w.*) following a method detailed by Bichon et al. [40]. 156 Microwave assisted extraction (Anton Paar, Graz, Austria) was used to extract the lipids from the freeze 157 dried samples for subsequent assessment of the clean-up methods. For each glass tube, approximately 158 1 ± 0.2 g freeze dried sample was extracted with 20 mL toluene/acetone mixture (70:30, v/v). Within 159 2 min, the temperature reached 130 °C and was maintained for 20 min. Agitation was done by a 160 magnetic bar, at 600 rpm during the entire extraction. Centrifugation was applied to separate the solid 161 and the liquid phases (1000 q, 10 min, 20 °C). Liquid extracts, containing lipids, were collected, pooled 162 and evaporated to obtain 9 g of dried lipid extract. The nine grams of lipids were suspended in a cyclohexane/ethyl acetate mixture (50:50, v/v) at 200 mg_{lipid}.g_{extract}⁻¹. 163

164

2.3. Clean-up methods

165 The clean-up methods were selected for their efficiency to remove lipids. The workflow depicted in 166 Figure 1 was developed to evaluate their performances to preserve contaminants in the extract while 167 remove the matrix interferents. Considering that the GPC device capacity was limited to 200 mg of

168 lipids, we selected this sample amount for all clean-up methods in order to avoid biases in the 169 comparison interpretations.

170 2.3.1. Size exclusion using gel permeation chromatography (GPC)

This clean-up procedure was adapted and modified from Abdel Malak et al. [41]. Aliquots of 1 g of the lipidic extract at 200 mg_{lipid}.g_{extract}⁻¹ were dried under a gentle stream of nitrogen and suspended in 500 μ L of cyclohexane/ethyl acetate mixture (50:50, *v/v*) and fractionated through a GPC column (58 cm × 24.4 mm) packed with Bio-Beads SX-3 (Bio-Rad, Philadelphia, PA, USA) using cyclohexane/ethyl acetate mixture (50:50, *v/v*) as mobile phase at a flow rate of 5 mL min⁻¹ during 70 min.

A preliminary experiment aimed at optimising the collection of fractions to remove the lipids was carried out. To do this, the extract was fractioned into 28 eluates collected every minute between 5 and 33 min, then dried under a gentle stream of nitrogen and weighted. Gravimetric determination of dried residues showed that lipids were eluted within the first 21 min of analysis. Therefore, the fraction between 22 min and 70 min (end of elution) was collected in a round-bottom flask. The collected extracts were concentrated and transferred to new glass tubes.

182 2.3.2. Size exclusion and hydrophobic interaction using Captiva EMR-lipid (EMR) cartridges

183 This clean-up procedure was adapted and modified from Zhao et al. [26]. Aliquots of 1 g of the extract 184 containing 200 mglipid.gextract⁻¹ were dried under a gentle stream of nitrogen, suspended in 2 mL 185 acetonitrile/ethyl acetate mixture (75:25, v/v) and vortexed. The extracts were warmed up to 40 °C for 186 1 h to accelerate contaminants partitioning between the lipids and the solvent phases. After 187 centrifugation under cooled temperature to make easier the phase separation and to minimize the lipid amount solubilize into the solvent layer (700 g, 10 min, 0 °C), the solvent phase was separated. 188 189 This procedure was repeated once and solvent phases were combined. Water (1 mL) was added to the 190 solvent phase. Captiva EMR-lipid cartridges were rinsed and conditioned with 4 mL acetonitrile/ethyl 191 acetate/water (20:60:20, v/v/v). The solvent phases were loaded onto the cartridges and eluates collected by gravity without adding any further eluent. Eluates were concentrated until 192

approximatively 10 μ L under at gentle stream of nitrogen and suspended with 2 mL toluene. Anhydrous sodium sulphate was added to eliminate water traces. The eluates were centrifuged (700 g, 2 min, 20 °C) and the organic layer was transferred to new glass tube.

196 2.3.3. Ester hydrolysis using sulphuric acid (H₂SO₄)

197 This clean-up procedure was adapted and modified from Cariou et al. [11]. Aliquots of 1 g of the extract solution at 200 mglipid.gextract⁻¹ were dried under a gentle stream of nitrogen and suspended in 12 mL 198 199 cyclohexane. The extracts were partitioned with 2 mL concentrated sulphuric acid. After centrifugation 200 (700 g, 10 min, 20 °C), the organic layer was separated. This procedure was repeated four times at 201 more drastic conditions as the ester hydrolysis is a slow chemical reaction (60 °C, 6 h, frequent vortex-202 mixing). The organic layer was neutralised three times with 2 mL water to obtain pH = 7 in the aqueous 203 layer. The organic layer was then dried with anhydrous sodium sulphate, centrifuged (700 g, 2 min, 20 204 °C) and transferred to new glass tube.

205

2.3.4. Differential partitioning using polydimethylsiloxane (PDMS)

206 This clean-up procedure was adapted and modified from Baumer et al. [35]. While the authors had 207 directly extracted lipid-rich tissue with PDMS, here we used the method to partition from the already 208 enriched lipids. The advantage is that a pure lipid phase reaches equilibrium faster than tissue with 209 lower lipid content. For egg yolk, it would be possible to extract directly but, for applicability to lean 210 tissue and comparison with the other three methods, we applied PDMS to the lipidic extracts. PDMS 211 disks were cleaned by Soxhlet extraction using ethyl acetate for 24 h and stored at room temperature 212 in amber bottles containing ethyl acetate. The PDMS disks were air dried during 2 h and the initial 213 masses were determined prior to the experiment.

Aliquots of 1 g of the extract at 200 mg_{lipid}.g_{extract}⁻¹ were dried under a gentle stream of nitrogen. Solvent cleaned PDMS disks (125 mg, 12 mm \times 1 mm) were embedded in the viscous lipid phase and dynamically exposed to the dried extracts on an orbital shaker during 24 h at 40 °C. The PDMS disks were cleaned with lint-free paper wipes and water, then weighted to determine the coextracted matrix

- component. Then PDMS disks were extracted twice with 2.5 mL ethyl acetate during 2 h and the two
 fractions were combined into new glass tube.
- 220 2.4. Analysis and processing
- 221

2.4.1. Cleaned-up extract conditioning

Each cleaned-up extract was reconstituted in 40 μ L toluene with 5 ng ¹³C₁₂-TBBPA and 10 ng ¹³C₁₀-anti-DP for GC-HRMS analysis (GC fraction). A 20 μ L aliquot of the GC fraction was further dried under a gentle stream of nitrogen and reconstituted in 20 μ L acetonitrile with 4 ng ²H₁₈- β -HBCDD for LC-HRMS analysis (LC fraction).

226 2.4.2. GC-EI-HRMS analysis

227 The GC fractions were analysed in a single sequence according to the method developed by Simonnet-228 Laprade et al. [42] with minor modifications. Briefly, a Trace 1310 gas chromatograph coupled to an 229 Orbitrap Q Exactive GC mass spectrometer (Thermo Scientific, San José, CA, USA) was used. Injection 230 (1 μ L) was performed at 300 °C in the splitless mode onto a DB-5MS column (30 m × 0.25 mm, 0.25 µm, Agilent, Palo Alto, CA, USA). The carrier gas was helium at a flow rate of 1 mL min⁻¹. The 231 232 temperature gradient started at 100 °C (held for 2 min) and raised at 10 °C min⁻¹ to a final temperature 233 of 325 °C (held for 10 min). Electron ionization source was operated at an electron energy set at 70 eV 234 and data were acquired in full-scan mode over the range of 120-800 m/z at a resolving power of 235 120,000 at *m/z* 200 (acquisition window = 5-32 min). The automatic gain control target was set at 5.10⁵ 236 and the maximum injection time at 200 ms.

237 2.4.3. LC-ESI(±

2.4.3. LC-ESI(±)-HRMS analysis

The LC fractions were analysed in two sequences according to the method developed by Cariou et al. [11] with minor modifications. Briefly, an Ultimate 3000 UHPLC pumping system coupled to an Orbitrap Q Exactive mass spectrometer was used. Five μ L were loaded on a reverse phase C₁₈-like column (Hypersil Gold, 100 × 2.1 mm, 1.9 μ m fitted with a Hypersil Gold guard column, 10 × 2.1 mm, 1.9 μ m,

Thermo Fisher Scientific) that was kept at 45 °C. The mobile phase consisted of water (A) and 242 acetonitrile/water (99:1, v/v, B) each containing 10 mM ammonium acetate with a flow rate set at 0.4 243 244 mL.min⁻¹. The gradient started with 20% B (held for 2 min), ramped linearly at 2.5%.min⁻¹ to 40% B then 245 at 2%.min⁻¹ to 100% B (held for 6 min) before returning to 20% B at 40%.min⁻¹ for equilibration (held 246 for 4 min). The total run time was 52 min. Data were recorded in both negative and positive modes in 247 the same run with heated ESI source parameters as follows: sheath gas flow = 50 arbitrary units (AU), 248 auxiliary gas flow = 5 AU, auxiliary gas temperature = 150 °C, capillary temperature = 350 °C, spray 249 voltage = 2.5 kV, s-lens radio frequency = 50 AU. HRMS system was operated in full scan mode over 250 the range 120-1000 m/z at a resolving power of 70,000 at m/z 200 (acquisition window = 0-48 min). The automatic gain control target was set at 5.10^5 and the maximum injection time at 250 ms. 251

252 2.4.4. QA/QC

Procedural blanks were prepared in triplicates at the same time and using the same methods in order
to assess only the clean-up procedural contamination. Water was used as blank for the PDMS
procedure.

To assess the mass spectrometer detector response during the sequence, a pooled QC sample constituted with aliquots of each final extract was prepared. For GC-HRMS, the pooled QC sample was injected every 10 randomised samples. For LC-HRMS, the pooled QC sample was injected five times before the samples (column conditioning) and then every 10 randomised samples.

260 2.4.5. Data processing

Peak integration was performed using Skyline software [43,44] for both GC- and LC-HRMS data. For
each compound, 3 ions were searched in GC-HRMS data (1 quantifier and 2 qualifiers) and 2 ions were
searched in LC-HRMS data (1 quantifier and 1 qualifier) (Table B.1).

In order to correct the detector response variation during the sequences and considering that mimetic
internal standards were not available for all the compounds, peak areas were corrected by the pooled

QC sample using the "batch correction" module implemented into Workflow4Metabolomics [45,46].
Loess pool regression method was used for both analytical methods [47].

268

2.5. Methods performance assessment

269 The performance of the methods was assessed using the spiking solution containing diverse set of 102 270 mid to non-polar compounds including pesticides and their transformation products, polycyclic 271 aromatic hydrocarbons (PAHs), phthalates, POPs and flame retardants (Table B.1), at concentrations in the range of 6.25-125 ng.g⁻¹ lipids (0.625 to 12.5 ng.g⁻¹ w.w.). These spike levels are higher than 272 273 those reported in food analysis, but are in the same concentration ranges of those reported in the 274 literature for environmental eggs like sea turtle eggs [48], terrestrial bird eggs [28,49] and seabird eggs 275 [50,51]. The selected compounds covered a wide range of physico-chemical properties: molecular 276 weight (122.6-706.1 Da), water solubility expressed as log K_{ow} (-0.9 to 10.6), molecule size expressed as collisional cross section (CCS) value (119.4 to 267.4 Å), halogenation degree and detectability in GC-277 278 El and/or LC-ESI(+) and/or LC-ESI(-). We assumed that the previously mentioned extraction method is 279 ideal for extracting hydrophobic compounds from the sample. However, we opted for mid-to non-280 polar compounds to encompass a slightly wider range of the chemical space.

281 Before assessing the performance of the clean-up methods, instrumental sensitivity and linearity, 282 expressed by the limit of detection (LOD) and the coefficient of determination (R^2) respectively, were 283 checked with an external calibration curve at dilution of 1/40, 1/20, 1/10, 1/5, 1/2.86, 1/2, 1/1.6, 1/1.33 and 1/1 of the spiking solution concentration (Table B.1). The LOD was determined as the lowest 284 285 concentration which allows to detect chromatographic peak defined with at least 5 consecutives scans, 286 considering the slow acquisition rate of HRMS in full-scan mode, and S/N higher than 3. Spiked 287 compounds were included to the performance assessment if their LOD was greater than 1/5 dilution. 288 The linearity was accepted if R² was greater than 0.9 with at least 5 calibration concentrations. This R² 289 value is less stringent in NTA than the acceptation criterion commonly used in targeted analysis (R²>0.98 or 0.99) because the signal variation in NTA is not normalised with fully mimetic internal
 standards.

Absolute recovery and matrix effect were selected criteria for the method performances assessment. Calculation was achieved for the compounds of the spiking solution detected in the pooled QC sample for which LOD and linearity were accepted. If a compound was detected by several techniques (GC-EI and/or by LC-ESI(+) and/or LC-ESI(-)), all values were considered and checked for consistency. Recoveries of compounds detected by several techniques appeared globally consistent. Recovery and matrix effect were calculated with equations (1) and (2), respectively. Equation (2) corrects for the endogenous levels.

299 Recovery (%) =
$$\left(\frac{\text{Area}_{\text{comp. spiked be. ext.}}}{\text{Area}_{\text{comp. spiked af. ext.}}}\right) \times 100$$
 (1)

300 Matrix effect (%) =
$$\left(\frac{\text{Area}_{\text{comp. spiked af. ext.}} - \text{Area}_{\text{comp. non-nspiked sample}}}{\text{Area}_{\text{standard}}} - 1\right) \times 100$$
 (2)

With "comp" the considered compound, "be. ext." and "af. ext." meaning before and after extraction,respectively.

For the recovery assessment, an aliquot (15 g) of the lipidic extract solution at 200 mg_{lipid}.g_{extract}⁻¹ was spiked with 375 μ L of the spiking solution (Table B.1), resulting in 25 μ L spiking solution per g of lipidic extract solution. This solution was vortex-mixed during 5 min and stored at -20 °C to equilibration for a night. For the matrix effects, extracts were spiked with 25 μ L of the spiking solution just prior to the reconstitution in toluene. Non-spiked extracts were prepared with the same clean-up methods to check endogenous contamination.

309

310 3. Results and discussion

311 **3.1. Detection and instrumental performances**

The selected compounds of the spiking solution (n = 102) cover a wide range of physico-chemical properties as indicated above. This was confirmed by comparing the chemical space covered by these 102 compounds with those covered by compounds listed in the CECscreen database with molecular weight below 1000 g.mol⁻¹ [52] (Figure 2A).

The complementarity between selected separation and ionisation techniques is illustrated Figure 2B. In total, 97 out of the 102 compounds were detected with at least one technique in the spiking solution. Naphthalene, acenaphthylene and acenaphthene were not detected, likely due to losses during evaporation process (volatile compounds) or inappropriate GC parameters. Malathion dicarboxylic acid and 6-chloronicotinic acid, two pesticides transformation products, were not detected, likely due to inappropriate ionisation parameters.

322 GC-EI-HRMS led to the detection of 64 compounds including 39 that were also detected by LC-ESI(±)-323 HRMS. LC-ESI(±)-HRMS led to the detection of 73 compounds, 35 for the positive polarity only, 31 for 324 the negative polarity only and 7 for both polarities. Most of the phthalate compounds could be 325 detected both by GC-EI and LC-ESI(+) and their LODs were quite similar (Table B.1). All pesticides and pesticide transformation products could be detected by LC-ESI(±) with low LODs except deltamethrin, 326 327 4,4'-DDE, β-HCH and HCB that could be detected only by GC-EI. Flame retardants, PAHs and POPs could 328 only be detected by GC-EI, except α -HBCDD, TBBPA and 2,4-dibromophenol, while hydroxy-PBDE have 329 only been detected by LC-ESI(-). Globally, phenolic compounds were more sensitive to ESI(-) due to the 330 hydrogen lability. Although expected, these observations confirmed that the combination of GC-EI and 331 LC-ESI(±) allowed covering a broad range of compounds for the suspect screening or NTA. A majority 332 of compounds was detected at the highest dilution factor with at least one of the analysis techniques. 333 Only a few compounds, including PAHs (fluorene, fluoranthene, pyrene and dibenzo (a,e) pyrene), 334 phthalates (DMT, DAP, DiNcH, TBR-DEHP), hydroxyl-BDEs (OH-BDE-28, OH-BDE-85 and OH-BDE-137), 335 BPF, deltamethrin, mecoprop, chlormequat, DETP, triclosan sulphate, triclosan glucuronide and FPrPA, 336 presented LODs greater than the highest dilution factor used for all analytical conditions.

337 In the present study, suspect screening was applied to assess the performances of the four tested 338 clean-up methods for a wide range of contaminants by direct comparison of the peak areas between 339 samples and only one calibration dilution. For this purpose, linearity was accepted if R² for the external 340 calibration curve was greater than 0.9. In that respect, linearity appeared acceptable for a majority of 341 compounds with at least one of the analysis techniques (Table B.1). DMT, chlormequat, FPrPA, 2,4-D 342 and DMDTP were the only compounds with R² below 0.9 whatever the analysis techniques.

343 Since signal area was batch corrected using the pool QC samples, only the compounds detected in all 344 the pool QC samples were used for assessing the clean-up methods performances. Consequently, DiNA 345 was not considered in GC-EI-HRMS, BPF and p-toluensulfonamide were not considered in LC-ESI(-)-346 HRMS and benzophenone-3, DAP, DPhP and fenhexamid were not considered in LC-ESI(+)-HRMS.

347 3.2. Clean-up methods performances assessment

348 Eggs from caged hens were a relevant model matrix due to their high lipid content associated with a 349 wide diversity of lipid classes [39]. In addition, low contamination levels were described for this type 350 of egg, which helps prevent any potential bias involving spiked compounds that may arise from 351 endogenous contamination [53,54]. Before applying equations (1) and (2) to calculate recoveries and 352 matrix effects, respectively, the endogenous contamination level in non-spiked extracts was checked 353 to minimise bias in the calculations of recovery.

354

3.2.1. Co-extracted matrix after clean-up

355 Extract appearance could present an overview of the clean-up performance based on the matrix 356 elimination. While extracts cleaned-up by H₂SO₄ were colourless and transparent, extracts cleaned-up by PDMS, GPC and EMR were coloured with a gradient from yellowish to orange, respectively. 357 358 However, when these extracts were reconstituted in acetonitrile before LC-ESI(±)-HRMS analysis, the 359 extracts were totally solubilised in one liquid phase, suggesting that triglycerides and phospholipids, 360 the main lipophilic lipid classes in poultry eggs, were removed [39,55].

361 The PDMS disks gained about 0.6 mg after the partitioning. This suggests that 0.3% of the initial lipid 362 mass was taken up into the PDMS disks. This weight gain represents $0.48 \pm 0.1\%$ of the PDMS disk 363 mass, which is in the range of that recorded by for PDMS exposed to pork adipose tissue [35] and 364 dugong blubber [34]. The LLE step prior the EMR-lipid cartridge extracted 32 mg of lipids and 96.4% of 365 these lipids were removed by the EMR-lipid cartridge resulting of a 99.6% overall lipid removal. These 366 values were in accordance and even better than those reported in the literature. Pedersen et al. [24] 367 reported that 97.2% of the matrix was removed from QuECHERS extracts of marine mammal bubbler 368 cleaned-up with one Captiva EMR lipid cartridge. Zhao et al. [26] also determined that 85% of the 369 matrix was removed from olive oil extract obtained with a similar LLE prior the EMR-lipid cartridge 370 loading. Muz et al. [25] reported an average of 98.2% of lipids were removed from a salmon lipid 371 extract using an LLE process involving acetonitrile. We estimated that 13.5 mg (27%) of salmon lipids 372 were extracted in this LLE condition. We hypothesized that if more lipids are extracted during an LLE 373 involving a more non-polar solvent mixture, more hydrophobic contaminants are also extracted.

Weights of residual matrix components were not determined for H₂SO₄, and GPC. The preliminary experiments carried out to determine the GPC fraction collection indicated that no measurable mass remained after 22 min (Figure A.1). However, a slight deposit was visible after drying.

377

3.2.2. Matrix effect assessment

Matrix effect results from unremoved matrix components that may cause signal suppression or enhancement in mass spectrometry [56–58]. If signal alteration can be characterised and corrected with the addition of mimetic labelled standards with targeted analysis, such matrix effect may impede compound detection with NTA. NTA interpretation is particularly vulnerable to matrix effect as a few standards, not necessarily mimetic of the matrix effect, are added to the sample. Matrix effects disturbing compounds detected by several techniques appeared independent and variable from one technique to another. Indeed, the presence of interferences and their competition for charges at a

target compound retention time greatly depend on the chromatographic separation and the ionisationmode.

387 For GC-EI-HRMS, the ratio of the spiked compounds presenting signal suppression or enhancement 388 between \pm 20% was 60% with H₂SO₄ clean-up, while this ratio decreased to 20%, 14% and 6% with 389 PDMS, EMR and GPC, respectively (Figure 3A, Table B.2). This observation was in line with the matrix 390 components detected among the chromatograms for the 4 tested clean-ups methods since signal 391 intensity was lower for H₂SO₄ modality (Figure 4A). For instance, unknown matrix components were 392 detected in the extracts from the four clean-up procedures at 7.85 min (m/z = 145.0648), 12.47 min 393 (m/z = 186.1039), 13.08 min (m/z = 236.1772), 21.24 min (m/z = 316.2394). These signals could be 394 related to hydrocarbon compounds with unsaturation between 4 and 6 and might be representative 395 of remaining unsaturated fatty acids. However, H₂SO₄ modality exhibited prominent chromatographic 396 hump between 19 and 25 min, mostly resulting from signals corresponding to $[C_xH_y]^{+\bullet}$ ions. Although 397 the origin of these ions could not be determined, they did not affect the detection of co-eluted spiked 398 compounds more than other spiked compounds.

For LC-ESI(±)-HRMS, the ratio of the spiked compounds presented a signal suppression or enhancement between ± 20% was 50% with H₂SO₄ clean-up (Figure 3B, Table B.3). For GPC and PDMS this ratio was around 33% of spiked compounds (75% of common compounds among them) whereas for EMR this ratio dropped to 25%. EMR, GPC and PDMS total ion chromatograms (TICs) showed intense interferences, mostly from hydrophilic matrix components, while H₂SO₄ TIC profile was similar to the *1/1.6* spiking solution dilution (Figure 4B and 4C). These observations confirmed that H₂SO₄ procedure remove more matrix components than the EMR, GPC and PDMS.

406 Matrix effects resulting from EMR were compared to those reported by Pourchet et al. [59] for breast 407 milk samples as the same instruments were used for both studies. The authors observed no significant 408 matrix effect for LC-ESI(±)-HRMS whereas significant signal alterations were observed for GC-EI(+)-409 HRMS. Results from both studies were in the same range for only a few common spiked contaminants

410 (in both studies about 20% and -25% for metolachlor and β -HCH, respectively, by GC-EI-HRMS analysis, 411 and about 0% for fenhexamid and fipronil by LC-ESI(±)-HRMS). Matrix effects calculated were not 412 within the same range for all the other common contaminants in both studies. This statement might 413 indicate that the matrix removal with EMR procedure depends on the matrix composition such as 414 lipids, free fatty acids or cholesterol. In addition, chromatographic conditions might be involved as 415 spiked contaminants and matrix components' retentions were different. Since GPC matrix removal 416 depends on the fraction collected, the lipid composition and the mobile phase composition, direct 417 comparison is at risk. Last, to the best of our knowledge, there are no studies focusing on matrix effects 418 using H₂SO₄ and PDMS to compare with.

419 *3.2.3. Recoveries*

420 Matrix component removal during the sample preparation is important for NTA to minimise matrix 421 effects. Furthermore, the maximisation of the compounds' preservation in the extract during this 422 process is also important to acquire a comprehensive fingerprint of the samples. Recoveries calculated 423 with equation (1) were independent of the matrix effect since the latter applies equally to extracts 424 spiked before and after clean-up.

425 With both GC- and LC- based techniques, recoveries could be calculated for only 36 detected 426 compounds after H₂SO₄ clean-up procedure (Figure 5, Tables B.2 and B.3 for detailed results). Acidic 427 conditions likely degraded many compounds sensitive to low pH, such as most pesticides and 428 phthalates. In addition, hydrophilic compounds with log $K_{ow} \leq 3$ or dissociated species under acidic or 429 neutral conditions (e.g. simazine, 3-phenoxybenzoic acid, mecoprop, fipronil, triclosan, prochloraz) 430 were probably back-extracted into the aqueous phase and thus removed. Because phthalates are acid-431 sensitive, their presence was mostly due to a procedural contamination occurring after the 432 neutralisation step. However, flame retardants, which are hydrophobic (log $K_{ow} > 3$), were resistant to 433 acidic conditions as their recoveries were 50% in average. The use of the H₂SO₄ procedure is limited to

the analysis of contaminants that are acid-resistant, hydrophobic and not prone to dissociation, whichprevents a comprehensive screening.

436 PDMS clean-up led to the detection of 69 spiked compounds. Most of them (n = 63) exhibited 437 recoveries lower than 20%, in accordance with the estimated partition coefficient between lipid and 438 PDMS *K*_{lipid/PDMS} at the equilibrium (Text A.1). The uptake did not seem to be related to molecular weight 439 or log K_{ow} (Figure A.3), which is also consistent with the partitioning theory. PDMS was expected to 440 partition a maximum fraction in PDMS f_{PDMS} of 6% for $K_{lipid/PDMS}$ of 10 and 2% for for $K_{lipid/PDMS}$ of 30, 441 which are the typical range of literature data [60] (calculation is detailed in Text A.1). The applicability 442 to partitioning contaminants with PDMS in the tested condition appeared to be confined to the highly 443 concentrated chemicals when applying NTA. To overcome this issue, experimental conditions could be 444 modified to increase the PDMS uptake quantity. According to Smedes et al. [61], increasing 445 extract/PDMS mass ratio increases the equilibrium concentration in the PDMS but also increases the 446 time for the equilibrium to be attained *i.e.* enhances the PDMS exposition time.

447 Size exclusion influenced both EMR and GPC to remove lipids characterised by long alkyl chains. 448 Additionally, EMR involved hydrophobic interactions to retain compounds [26]. EMR and GPC led to 449 the detection of 88 and 86 spiked compounds, respectively. Compounds characterised by the highest 450 CCS values presented lower recoveries than the other ones with GPC, in line with the theory [62] 451 (Figure A.4). Due to the size of these compounds, their elution might overlay with lipids. Due to their 452 sizes, the elution of these compounds likely overlapped with the discarded fraction containing lipids. 453 Conversely, regarding the EMR protocol, the recoveries do not seem to be correlated with the CCS 454 values or log K_{ow} (Figure A.5), although the retaining phenomena involves these physico-chemical 455 properties. Although these two methods presented similar average recoveries, the repeatability, 456 expressed in relative standard deviation (RSD), appeared better with EMR than with GPC (Table B.2 457 and B.3).

458 The recoveries after EMR reported in the literature [25,26] are higher than those presented here and 459 could be explained by the combination of two experimental considerations. The first consideration 460 might be that the solvent strength of the acetonitrile/ethyl acetate (75:25, v/v) mixture was not strong 461 enough for partitioning the spiked compounds between the lipids and the solvent. Indeed, in the two 462 studies mentioned above, the spike was performed directly in a unique solvent phase to be loaded on 463 the SPE cartridge and containing the lipid, so that no losses could occur prior to that loading step. The 464 second consideration might be that no vacuum was applied to collect the solvent from the cartridge. 465 In order to avoid cross-contamination that could occur at the port of the vacuum manifold, in-house 466 made cartridge rack was used with which it was not possible to apply vacuum. The solvent volume 467 remaining into the cartridge was evaluated to 1 mL. Further optimisation of these two parameters 468 could increase the recoveries of the EMR procedure.

469

470 4. <u>Conclusion</u>

Effective sample preparation is necessary to take advantage of the recent advances in chromatography and high-resolution mass spectrometry which can detect thousands of chemicals in a single analysis. NTA requires non-selective sample preparation to isolate compounds from matrix components such as lipids. Since it is difficult to isolate compounds with a wide range of physico-chemical properties, as present in samples yet, we suggested a clean-up that selectively remove matrix components from lipid extracts.

The procedure involving H₂SO₄ demonstrated matrix removal efficiency and reduced the procedural contamination. On the other hand, only acid-resistant compounds, such as POPs, could be detected following such rather destructive clean-up procedure impeding the acquisition of a comprehensive fingerprint. In this clean-up method, the lipid mass can be increased to a few grams, allowing a concentration that may facilitate the detection of less concentrated acid-resistant compounds.

482 Nevertheless, concentrated H₂SO₄ is hazardous for the operator and the applied procedure remains
483 labour-intensive (5 days).

484 PDMS absorbed the compounds of interest while a small amount of lipids was taken up into it, 485 demonstrating the effectiveness of matrix removal. Under the conditions tested, the partition ratio of 486 compounds enriched in PDMS was low, so many compounds were not identified. However, increasing 487 the *extract/PDMS* mass ratio could increase the amount of analyte partitioned to reach instrumental 488 LOD. This method required relatively low volumes of solvent per samples (5 mL without the PDMS 489 cleaning solvent). The matrix removal mainly depends on the operator's care to remove components 490 adsorbed to the surface of the PDMS material. Advantages of PDMS are its ease of operation and 491 affordability, requiring minimal technical equipment and allowing high sample throughput.

492 Since both the Captiva EMR-lipid cartridge and GPC involved size exclusion, their matrix removal and 493 recovery were similar and satisfactory for most of the spiked compounds. EMR recoveries can be 494 higher by eluting solvent that remains adsorbed on the phase, although there is a potential risk of 495 eluting matrix components as well. The LLE step preceding loading onto EMR lipid could also be 496 optimised to partition a larger quantity of compounds but, again, attention must be paid to the co-497 extracted matrix components. Conversely, the fraction collection could start later to enhance the 498 matrix components removal, but the largest compounds of interest might be excluded. The first major 499 difference between EMR and GPC was repeatability that was better using EMR. The second major 500 difference between both procedures was the solvent consumption. It was reduced using EMR (about 501 10 mL) whereas over 350 mL per sample were used with GPC clean-up.

If EMR and GPC seem suitable for NTA, attention needs to be paid to the lipid composition (lipid class, fatty acid chain length, unsaturation), to extend these procedures to fatty matrices other than hen eggs. Using EMR, Zhao et al. [26] showed that the matrix removal may vary depending to the type of natural oil (olive, corn, soybean and canola). GPC collection start may vary depending to the matrix components such as lipid composition as well.

507 With the knowledge gained in this study on the performance and limitations of several sample 508 preparation strategies, appropriate approaches could be selected to perform non-targeted analysis of 509 contaminants in fatty matrices, in order to better describe the chemical universe to which wildlife and 510 human are exposed.

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517	
518	Associated content

Supporting material. The following files are available free of charge.

520 5. <u>References</u>

- 521 [1] D. Muir, X. Zhang, C.A. de Wit, K. Vorkamp, S. Wilson, Identifying further chemicals of emerging
 522 arctic concern based on 'in silico' screening of chemical inventories, Emerging Contaminants. 5
 523 (2019) 201–210. https://doi.org/10.1016/j.emcon.2019.05.005.
- 524 [2] B.I. Escher, H.M. Stapleton, E.L. Schymanski, Tracking complex mixtures of chemicals in our 525 changing environment, Science. 367 (2020) 388–392. https://doi.org/10.1126/science.aay6636.
- [3] H. Chen, C. Wang, H. Li, R. Ma, Z. Yu, L. Li, M. Xiang, X. Chen, X. Hua, Y. Yu, A review of toxicity
 induced by persistent organic pollutants (POPs) and endocrine-disrupting chemicals (EDCs) in
 the nematode caenorhabditis elegans, Journal of Environmental Management. 237 (2019) 519–
 525. https://doi.org/10.1016/j.jenvman.2019.02.102.
- F. Vasseur, C. Cossu-Leguille, Linking molecular interactions to consequent effects of persistent
 organic pollutants (POPs) upon populations, Chemosphere. 62 (2006) 1033–1042.
 https://doi.org/10.1016/j.chemosphere.2005.05.043.
- 533 [5] A.B.A. Boxall, C.J. Sinclair, K. Fenner, D. Kolpin, S.J. Maund, When synthetic chemicals degrade
 534 in the environment, Environ. Sci. Technol. 38 (2004) 368A-375A.
 535 https://doi.org/10.1021/es040624v.
- A.C. Dirtu, A. Covaci, A.C. Dirtu, M. Abdallah, Advances in the sample preparation of
 brominated flame retardants and other brominated compounds, TrAC Trends in Analytical
 Chemistry. 43 (2013) 189–203. https://doi.org/10.1016/j.trac.2012.10.004.
- 539 [7] A. Kaufmann, High-resolution mass spectrometry for bioanalytical applications: Is this the new
 540 gold standard?, Journal of Mass Spectrometry. 55 (2020) e4533.
 541 https://doi.org/10.1002/jms.4533.
- J.-P. Antignac, K. de Wasch, F. Monteau, H. De Brabander, F. Andre, B. Le Bizec, The ion
 suppression phenomenon in liquid chromatography–mass spectrometry and its consequences
 in the field of residue analysis, Analytica Chimica Acta. 529 (2005) 129–136.
 https://doi.org/10.1016/j.aca.2004.08.055.
- [9] N.S. Chatterjee, S. Utture, K. Banerjee, T.P. Ahammed Shabeer, N. Kamble, S. Mathew, K. Ashok
 Kumar, Multiresidue analysis of multiclass pesticides and polyaromatic hydrocarbons in fatty
 fish by gas chromatography tandem mass spectrometry and evaluation of matrix effect, Food
 Chemistry. 196 (2016) 1–8. https://doi.org/10.1016/j.foodchem.2015.09.014.
- [10] P. Hajeb, L. Zhu, R. Bossi, K. Vorkamp, Sample preparation techniques for suspect and non-target screening of emerging contaminants, Chemosphere. (2021) 132306.
 https://doi.org/10.1016/j.chemosphere.2021.132306.
- [11] R. Cariou, P. Méndez-Fernandez, S. Hutinet, Y. Guitton, F. Caurant, B. Le Bizec, J. Spitz, W.
 Vetter, G. Dervilly, Nontargeted LC/ESI-HRMS detection of polyhalogenated compounds in marine mammals stranded on french atlantic coasts, ACS EST Water. 1 (2021) 309–318.
 https://doi.org/10.1021/acsestwater.0c00091.
- [12] B. Du, J.M. Lofton, K.T. Peter, A.D. Gipe, C.A. James, J.K. McIntyre, N.L. Scholz, J.E. Baker, E.P.
 Kolodziej, Development of suspect and non-target screening methods for detection of organic
 contaminants in highway runoff and fish tissue with high-resolution time-of-flight mass
 spectrometry, Environ. Sci.: Processes Impacts. 19 (2017) 1185–1196.
 https://doi.org/10.1039/C7EM00243B.
- [13] N.J. Shaul, N.G. Dodder, L.I. Aluwihare, S.A. Mackintosh, K.A. Maruya, S.J. Chivers, K. Danil, D.W.
 Weller, E. Hoh, Nontargeted biomonitoring of halogenated organic compounds in two ecotypes
 of bottlenose dolphins (Tursiops truncatus) from the southern California bight, Environ. Sci.
 Technol. 49 (2015) 1328–1338. https://doi.org/10.1021/es505156q.
- [14] Q. Wu, C. Munschy, Y. Aminot, N. Bodin, W. Vetter, High levels of halogenated natural products
 in large pelagic fish from the Western Indian Ocean, Environ Sci Pollut Res. (2021).
 https://doi.org/10.1007/s11356-021-14738-0.

- 569 [15] W. Dürig, A. Kintzi, O. Golovko, K. Wiberg, L. Ahrens, New extraction method prior to screening
 570 of organic micropollutants in various biota matrices using liquid chromatography coupled to
 571 high-resolution time-of-flight mass spectrometry, Talanta. 219 (2020) 121294.
 572 https://doi.org/10.1016/j.talanta.2020.121294.
- [16] Y. Li, L. Wang, M. Zheng, Y. Lin, H. Xu, A. Liu, Y. Hua, Y. Jiang, K. Ning, S. Hu, Thin-layer
 chromatography coupled with HPLC-DAD/UHPLC-HRMS for target and non-target
 determination of emerging halogenated organic contaminants in animal-derived foods, Food
- 576 Chem. 404 (2023) 134678. https://doi.org/10.1016/j.foodchem.2022.134678.
- 577 [17] A. Rebryk, P. Haglund, Non-targeted screening workflows for gas chromatography–high578 resolution mass spectrometry analysis and identification of biomagnifying contaminants in
 579 biota samples, Anal Bioanal Chem. 413 (2021) 479–501. https://doi.org/10.1007/s00216-020580 03018-4.
- [18] Q. Wu, N. Eisenhardt, S.S. Holbert, J.R. Pawlik, J.R. Kucklick, W. Vetter, Naturally occurring
 organobromine compounds (OBCs) including polybrominated dibenzo-p-dioxins in the marine
 sponge Hyrtios proteus from The Bahamas, Marine Pollution Bulletin. 172 (2021) 112872.
 https://doi.org/10.1016/j.marpolbul.2021.112872.
- 585 [19] F. Dubocq, B.B. Bæringsdóttir, T. Wang, A. Kärrman, Comparison of extraction and clean-up
 586 methods for comprehensive screening of organic micropollutants in fish using gas
 587 chromatography coupled to high-resolution mass spectrometry, Chemosphere. 286 (2022)
 588 131743. https://doi.org/10.1016/j.chemosphere.2021.131743.
- [20] J. Pokorný, J. Dostálová, Food lipids, in: Food Quality and Standards Volume III, EOLSS
 Publications, 2009: pp. 178–203.
- 591 [21] D.L. Stalling, R.C. Tindle, J.L. Johnson, Cleanup of pesticide and polychlorinated biphenyl
 592 residues in fish extracts by gel permeation chromatography, Journal of AOAC INTERNATIONAL.
 593 55 (1972) 32–38. https://doi.org/10.1093/jaoac/55.1.32.
- 594 [22] C. Munschy, N. Guiot, K. Héas-Moisan, C. Tixier, J. Tronczyński, Polychlorinated dibenzo-p595 dioxins and dibenzofurans (PCDD/Fs) in marine mussels from French coasts: levels, patterns
 596 and temporal trends from 1981 to 2005, Chemosphere. 73 (2008) 945–953.
 597 https://doi.org/10.1016/j.chemosphere.2008.06.062.
- J.-H. Zhao, L.-X. Hu, L.-X. He, Y.-Q. Wang, J. Liu, J.-L. Zhao, Y.-S. Liu, G.-G. Ying, Rapid target and
 non-target screening method for determination of emerging organic chemicals in fish, Journal
 of Chromatography A. 1676 (2022) 463185. https://doi.org/10.1016/j.chroma.2022.463185.
- A.F. Pedersen, R. Dietz, C. Sonne, L. Liu, A. Rosing-Asvid, M.A. McKinney, Development and
 validation of a modified QuEChERS method for extracting polychlorinated biphenyls and
 organochlorine pesticides from marine mammal blubber, Chemosphere. 312 (2023) 137245.
 https://doi.org/10.1016/j.chemosphere.2022.137245.
- M. Muz, E. Rojo-Nieto, A. Jahnke, Removing disturbing matrix constituents from biota extracts
 from total extraction and silicone-based passive sampling, Environ Toxicol Chem. 40 (2021)
 2693–2704. https://doi.org/10.1002/etc.5153.
- L. Zhao, T. Szakas, M. Churley, D. Lucas, Multi-class multi-residue analysis of pesticides in edible
 oils by gas chromatography-tandem mass spectrometry using liquid-liquid extraction and
 enhanced matrix removal lipid cartridge cleanup, Journal of Chromatography A. 1584 (2019) 1–
 12. https://doi.org/10.1016/j.chroma.2018.11.022.
- [27] R. Cariou, E. Omer, A. Léon, G. Dervilly-Pinel, B. Le Bizec, Screening halogenated environmental
 contaminants in biota based on isotopic pattern and mass defect provided by high resolution
 mass spectrometry profiling, Analytica Chimica Acta. 936 (2016) 130–138.
 https://doi.org/10.1016/j.aca.2016.06.053.
- 616 [28] W. Vetter, B. Luckas, M. Oehme, Isolation and purification of the two main toxaphene 617 congeners 1n marine organisms, Chemosphere. 25 (1992) 1643–1652.
- 618 [29] E. Hoh, S.J. Lehotay, K. Mastovska, H.L. Ngo, W. Vetter, K.C. Pangallo, C.M. Reddy, Capabilities
- of Direct Sample Introduction–Comprehensive Two-Dimensional Gas Chromatography–Time-

- 620of-Flight Mass Spectrometry to Analyze Organic Chemicals of Interest in Fish Oils, Environ. Sci.621Technol. 43 (2009) 3240–3247. https://doi.org/10.1021/es803486x.
- 622 [30] E.F. Houtz, D.L. Sedlak, Oxidative Conversion as a Means of Detecting Precursors to
 623 Perfluoroalkyl Acids in Urban Runoff, Environ. Sci. Technol. 46 (2012) 9342–9349.
 624 https://doi.org/10.1021/es302274g.
- [31] L. Ossiander, F. Reichenberg, M.S. McLachlan, P. Mayer, Immersed solid phase microextraction
 to measure chemical activity of lipophilic organic contaminants in fatty tissue samples,
- 627 Chemosphere. 71 (2008) 1502–1510. https://doi.org/10.1016/j.chemosphere.2007.11.060.
- [32] P. Mayer, L. Toräng, N. Glæsner, J.Å. Jönsson, Silicone membrane equilibrator: measuring
 chemical activity of nonpolar chemicals with poly(dimethylsiloxane) microtubes immersed
 directly in tissue and lipids, Anal. Chem. 81 (2009) 1536–1542.
 https://doi.org/10.1021/ac802261z.
- 632 [33] A. Jahnke, M.S. McLachlan, P. Mayer, Equilibrium sampling: partitioning of organochlorine
 633 compounds from lipids into polydimethylsiloxane, Chemosphere. 73 (2008) 1575–1581.
 634 https://doi.org/10.1016/j.chemosphere.2008.09.017.
- https://doi.org/10.1016/j.chemosphere.2008.09.017.
 [34] L. Jin, C. Gaus, L. van Mourik, B.I. Escher, Applicability of passive sampling to bioanalytical
 screening of bioaccumulative chemicals in marine wildlife, Environ Sci Technol. 47 (2013) 7982–
 7988. https://doi.org/10.1021/es401014b.
- A. Baumer, S. Jäsch, N. Ulrich, I. Bechmann, J. Landmann, B.I. Escher, Kinetics of equilibrium
 passive sampling of organic chemicals with polymers in diverse mammalian tissues, Environ.
 Sci. Technol. 55 (2021) 9097–9108. https://doi.org/10.1021/acs.est.1c01836.
- [36] A. Jahnke, P. Mayer, D. Broman, M.S. McLachlan, Possibilities and limitations of equilibrium
 sampling using polydimethylsiloxane in fish tissue, Chemosphere. 77 (2009) 764–770.
 https://doi.org/10.1016/j.chemosphere.2009.08.025.
- [37] F. Smedes, SSP silicone–, lipid– and SPMD–water partition coefficients of seventy hydrophobic
 organic contaminants and evaluation of the water concentration calculator for SPMD,
 Chemosphere. 223 (2019) 748–757. https://doi.org/10.1016/j.chemosphere.2019.01.164.
- 647 [38] A. Jahnke, P. Mayer, S. Schäfer, G. Witt, N. Haase, B.I. Escher, Strategies for transferring
 648 mixtures of organic contaminants from aquatic environments into bioassays, Environ. Sci.
 649 Technol. 50 (2016) 5424–5431. https://doi.org/10.1021/acs.est.5b04687.
- [39] J. Haedrich, C. Stumpf, M.S. Denison, Rapid extraction of total lipids and lipophilic POPs from all
 EU-regulated foods of animal origin: Smedes' method revisited and enhanced, Environmental
 Sciences Europe. 32 (2020) 118. https://doi.org/10.1186/s12302-020-00396-5.
- [40] E. Bichon, I. Guiffard, A. Vénisseau, E. Lesquin, V. Vaccher, P. Marchand, B. Le Bizec,
 Simultaneous analysis of historical, emerging and novel brominated flame retardants in food
 and feed using a common extraction and purification method, Chemosphere. 205 (2018) 31–
 40. https://doi.org/10.1016/j.chemosphere.2018.04.070.
- [41] I. Abdel Malak, R. Cariou, A. Venisseau, G. Dervilly-Pinel, F. Jaber, M. Babut, B. Le Bizec,
 Occurrence of Dechlorane Plus and related compounds in catfish (Silurus spp.) from rivers in
 France, Chemosphere. 207 (2018) 413–420.
- 660 https://doi.org/10.1016/j.chemosphere.2018.05.101.
- 661 [42] C. Simonnet-Laprade, S. Bayen, B. Le Bizec, G. Dervilly, Data analysis strategies for the
 662 characterization of chemical contaminant mixtures. Fish as a case study, Environment
 663 International. 155 (2021) 106610. https://doi.org/10.1016/j.envint.2021.106610.
- [43] B. MacLean, D.M. Tomazela, N. Shulman, M. Chambers, G.L. Finney, B. Frewen, R. Kern, D.L.
 Tabb, D.C. Liebler, M.J. MacCoss, Skyline: an open source document editor for creating and analyzing targeted proteomics experiments, Bioinformatics. 26 (2010) 966–968.
 https://doi.org/10.1093/bioinformatics/btq054.
- [44] K.J. Adams, B. Pratt, N. Bose, L.G. Dubois, L. St. John-Williams, K.M. Perrott, K. Ky, P. Kapahi, V.
 Sharma, M.J. MacCoss, M.A. Moseley, C.A. Colton, B.X. MacLean, B. Schilling, J.W. Thompson,
 Skyline for small molecules: a unifying software package for quantitative metabolomics, J.
- 671 Proteome Res. 19 (2020) 1447–1458. https://doi.org/10.1021/acs.jproteome.9b00640.

- 672 [45] F. Giacomoni, G. Le Corguillé, M. Monsoor, M. Landi, P. Pericard, M. Pétéra, C. Duperier, M.
- 673 Tremblay-Franco, J.-F. Martin, D. Jacob, S. Goulitquer, E.A. Thévenot, C. Caron,
- 674 Workflow4Metabolomics: a collaborative research infrastructure for computational
- 675 metabolomics, Bioinformatics. 31 (2015) 1493-1495.
- 676 https://doi.org/10.1093/bioinformatics/btu813.
- 677 [46] Y. Guitton, M. Tremblay-Franco, G. Le Corguillé, J.-F. Martin, M. Pétéra, P. Roger-Mele, A. Delabrière, S. Goulitquer, M. Monsoor, C. Duperier, C. Canlet, R. Servien, P. Tardivel, C. Caron, 678 679 F. Giacomoni, E.A. Thévenot, Create, run, share, publish, and reference your LC–MS, FIA–MS, 680 GC–MS, and NMR data analysis workflows with the Workflow4Metabolomics 3.0 Galaxy online 681 infrastructure for metabolomics, The International Journal of Biochemistry & Cell Biology. 93 682 (2017) 89–101. https://doi.org/10.1016/j.biocel.2017.07.002.
- 683 [47] F.M. van der Kloet, I. Bobeldijk, E.R. Verheij, R.H. Jellema, Analytical error reduction using single 684 point calibration for accurate and precise metabolomic phenotyping, J Proteome Res. 8 (2009) 685 5132–5141. https://doi.org/10.1021/pr900499r.
- 686 [48] D. Savoca, M. Arculeo, L. Vecchioni, I. Cambera, G. Visconti, R. Melfi, V. Arizza, A. Palumbo 687 Piccionello, S. Buscemi, A. Pace, Can phthalates move into the eggs of the loggerhead sea turtle 688 Caretta caretta? The case of the nests on the Linosa Island in the Mediterranean Sea, Marine 689 Pollution Bulletin. 168 (2021) 112395. https://doi.org/10.1016/j.marpolbul.2021.112395.
- 690 [49] D. Venugopal, M. Subramanian, J. Rajamani, J. Palaniyappan, J. Samidurai, A. Arumugam, Levels 691 and distribution pattern of organochlorine pesticide residues in eggs of 22 terrestrial birds from 692 Tamil Nadu, India, Environ Sci Pollut Res. 27 (2020) 39253-39264. 693 https://doi.org/10.1007/s11356-020-09978-5.
- 694 [50] S. Huber, N.A. Warner, T. Nygård, M. Remberger, M. Harju, H.T. Uggerud, L. Kaj, L. Hanssen, A 695 broad cocktail of environmental pollutants found in eggs of three seabird species from remote 696 colonies in Norway, Environmental Toxicology and Chemistry. 34 (2015) 1296–1308. 697 https://doi.org/10.1002/etc.2956.
- 698 [51] S.R. de Solla, D.V.C. Weseloh, K.D. Hughes, D.J. Moore, Forty-Year Decline of Organic 699 Contaminants in Eggs of Herring Gulls (Larus argentatus) from the Great Lakes, 1974 to 2013, 700 Cowa. 39 (2016) 166–179. https://doi.org/10.1675/063.039.sp117.
- [52] J. Meijer, M. Lamoree, T. Hamers, J.-P. Antignac, S. Hutinet, L. Debrauwer, A. Covaci, C. Huber, 701 702 M. Krauss, D.I. Walker, E.L. Schymanski, R. Vermeulen, J. Vlaanderen, An annotation database 703 for chemicals of emerging concern in exposome research, Environment International. 152 704 (2021) 106511. https://doi.org/10.1016/j.envint.2021.106511.
- 705 [53] J.-F. Hsu, C. Chen, P.-C. Liao, Elevated PCDD/F levels and distinctive PCDD/F congener profiles in 706 free range eggs, J. Agric. Food Chem. 58 (2010) 7708–7714. https://doi.org/10.1021/jf100456b.
- 707 [54] G. Schoeters, R. Hoogenboom, Contamination of free-range chicken eggs with dioxins and 708 dioxin-like polychlorinated biphenyls, Molecular Nutrition & Food Research. 50 (2006) 908–914. 709 https://doi.org/10.1002/mnfr.200500201.
- 710 [55] O.S. Privett, M.L. Blank, J.A. Schmit, Studies on the composition of egg lipid, Journal of Food 711 Science. 27 (1962) 463-468. https://doi.org/10.1111/j.1365-2621.1962.tb00128.x.
- 712 [56] J. Hajšlová, J. Zrostlíková, Matrix effects in (ultra)trace analysis of pesticide residues in food and 713 biotic matrices, Journal of Chromatography A. 1000 (2003) 181–197. 714
 - https://doi.org/10.1016/S0021-9673(03)00539-9.
- 715 [57] Liang. Tang, Paul. Kebarle, Dependence of ion intensity in electrospray mass spectrometry on 716 the concentration of the analytes in the electrosprayed solution, Anal. Chem. 65 (1993) 3654-717 3668. https://doi.org/10.1021/ac00072a020.
- 718 [58] P.J. Taylor, Matrix effects: the Achilles heel of quantitative high-performance liquid 719 chromatography–electrospray–tandem mass spectrometry, Clinical Biochemistry. 38 (2005) 720 328–334. https://doi.org/10.1016/j.clinbiochem.2004.11.007.
- 721 [59] M. Pourchet, L. Narduzzi, A. Jean, I. Guiffard, E. Bichon, R. Cariou, Y. Guitton, S. Hutinet, J.
- 722 Vlaanderen, J. Meijer, B. Le Bizec, J.-P. Antignac, Non-targeted screening methodology to

- characterise human internal chemical exposure: application to halogenated compounds in
 human milk, Talanta. 225 (2021) 121979. https://doi.org/10.1016/j.talanta.2020.121979.
- F. Smedes, T.P. Rusina, H. Beeltje, P. Mayer, Partitioning of hydrophobic organic contaminants
 between polymer and lipids for two silicones and low density polyethylene, Chemosphere. 186
 (2017) 948–957. https://doi.org/10.1016/j.chemosphere.2017.08.044.
- F. Smedes, L.A. van Vliet, K. Booij, Multi-ratio equilibrium passive sampling method to estimate
 accessible and pore water concentrations of polycyclic aromatic hydrocarbons and
 polychlorinated biphenyls in sediment, Environ. Sci. Technol. 47 (2013) 510–517.
- 731 https://doi.org/10.1021/es3040945.
- 732 [62] D.E. Lahmanov, Y.I. Varakina, A short review of sample preparation methods for the pesticide
 733 residue analysis in fatty samples, IOP Conf. Ser.: Earth Environ. Sci. 263 (2019) 012061.
- 734 https://doi.org/10.1088/1755-1315/263/1/012061.
- 735

736 Figures



- 738 Figure 1: Detailed analytical workflow used to assess the clean-up performances. GPC: size exclusion
- 739 using gel permeation chromatography method; EMR: size exclusion and hydrophobic interaction
- 740 using Captiva EMR-Lipid cartridges; H₂SO₄: ester hydrolysis using sulphuric acid; PDMS: differential
- 741 *partitioning using polydimethylsiloxane.*



Figure 2: Chemical space covered by the 102 compounds of the spiking solution (black diamonds) within
the compounds with a molecular mass under 1000 g mol⁻¹ from the CECscreen database (grey crosses,
n = 69,704 compounds) (A) and Venn diagram of the 98 detected compounds by GC-EI and LC-ESI(±)
(B).



Figure 3: Matrix effects observed for the analysis of the spiked compounds by GC-EI-HRMS (A) and LCESI(±)-HRMS (B) with the four clean-up methods (n = 3). Compounds are grouped according to their
classes and ESI polarity. Whiskers: min and max values; boxes: interquartile range; crosses: mean
values; lines: median values; dots: outliers (data that are 1.5 times larger than the 3rd quartile or 1.5
times smaller than the 1st quartile).





756 *Figure 4: Representative TICs acquired by GC-EI-HRMS (A), LC-ESI(+)-HRMS (B) and LC-ESI(-)-HRMS (C)*

- 757 for the spiking solution at 1/1.6 dilution (black) and egg samples spiked after the H₂SO₄ (red), Captiva-
- 758 EMR-lipid (green), GPC (blue) and PDMS (yellow) clean-up procedures. NL: normalised level.



Figure 5: Absolute recoveries of spiked compounds analysed by GC-EI-HRMS (A) and LC-ESI(±)-HRMS
(B) with the four clean-up methods (n = 3). Compounds are grouped according to their class and ESI
polarity. Whiskers: min and max values; boxes: interquartile range; crosses: mean values; lines: median
values; dots: outliers (data that are 1.5 times larger than the 3rd quartile or 1.5 times smaller than the
1st quartile).