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Abstract

CONTEXT

Differences in land tenure regimes are one challenge to implement soil conservation practices in agricultural systems. It is frequently assumed that tenants are less prone to adopt soil conservation strategies than owners, given a shorter-term engagement with the field. Also, the field's environmental potential (i.e., potential for agricultural production) may influence farmers' investment decisions, since high-potential fields increase the chances of achieving a return of the investment.

OBJETIVE

Understand the effect of land tenure regimes and environmental potential on fertilization rates and balance of nitrogen, phosphorus, and sulfur in soybean and maize crops of Argentina.

METHODS

We applied mixed-effects models on a database of 52,588 fields of soybean and maize farms, covering a total area of 3.8 M ha in Argentina during the period of 2017-2022.

RESULTS AND CONCLUSIONS

In general, the balance of nitrogen, phosphorus, and sulfur were (mean \pm SE) -29.11 \pm 0.15, -2.58 \pm 0.38, and 8.26 \pm 0.044 kg ha⁻¹yr⁻¹, respectively. Despite 8.04 and 0.63 kg ha⁻¹yr⁻¹ more nitrogen and phosphorus were applied in high-potential than in low-potential maize fields, nutrient outputs were still higher, therefore, net nutrient exports of the most productive fields increased by 9.99 and 2.06 kg ha⁻¹yr⁻¹ for nitrogen and phosphorus, respectively. In soybean fields, environmental potential had no effect on nutrient application, but nitrogen and phosphorus net nutrient exports were 9.85 and 2.14 kg ha⁻¹yr⁻¹ higher in high-potential fields compared to low-potential fields. Tenure regime had a weak effect, mainly on phosphorus. On average, owners applied 0.37 kg ha⁻¹yr⁻¹ more and exported 0.28 kg ha⁻¹yr⁻¹ less phosphorus than tenants in both crops. Sulfur application and balance was weakly affected by the studied variables, and the positive balance suggests overfertilization under the assumptions of this paper. We conclude that the Argentine farming system depletes some of the main nutrients, regardless of the field's environmental potential or the land tenure system. The effect of the tenure regime is overwhelmed by the impact of environmental potential on

farmers' fertilization management, with high-potential fields degrading due to soil mining at a faster pace than low-potential fields, putting future yields at risk.

SIGNIFICANCE

By exploring a farming system based on nutrient depletion, our results contribute to the general understanding of tenure regime consequences on soil degradation. Argentinean farmers should consider increasing N and P application and contemplate environmental heterogeneity to avoid nutrient mining and degradation of one of the most productive areas of the world.

1 Title page

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3	application and soil nutrient mining in soybean and maize crops
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- 46 covering a total area of 3.8 M ha in Argentina during the period of 2017-2022.

47 RESULTS AND CONCLUSIONS

In general, the balance of nitrogen, phosphorus, and sulfur were (mean \pm SE) -29.11 \pm 0.15, -48 2.58 \pm 0.38, and 8.26 \pm 0.044 kg ha⁻¹yr⁻¹, respectively. Despite 8.04 and 0.63 kg ha⁻¹yr⁻¹ more 49 50 nitrogen and phosphorus were applied in high-potential than in low-potential maize fields, 51 nutrient outputs were still higher, therefore, net nutrient exports of the most productive fields increased by 9.99 and 2.06 kg ha⁻¹yr⁻¹ for nitrogen and phosphorus, respectively. In soybean 52 53 fields, environmental potential had no effect on nutrient application, but nitrogen and phosphorus net nutrient exports were 9.85 and 2.14 kg ha⁻¹yr⁻¹ higher in high-potential fields 54 compared to low-potential fields. Tenure regime had a weak effect, mainly on phosphorus. On 55 56 average, owners applied 0.37 kg ha⁻¹yr⁻¹ more and exported 0.28 kg ha⁻¹yr⁻¹ less phosphorus than tenants in both crops. Sulfur application and balance was weakly affected by the studied 57 58 variables, and the positive balance suggests overfertilization under the assumptions of this 59 paper. We conclude that the Argentine farming system depletes some of the main nutrients, 60 regardless of the field's environmental potential or the land tenure system. The effect of the 61 tenure regime is overwhelmed by the impact of environmental potential on farmers' fertilization management, with high-potential fields degrading due to soil mining at a faster pace than low-62 63 potential fields, putting future yields at risk.

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By exploring a farming system based on nutrient depletion, our results contribute to the general
understanding of tenure regime consequences on soil degradation. Argentinean farmers should
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69 **Graphical abstract**



71 Keywords: field crops, crop nutrition, decision making, soil conservation, nutrient mining,
 72 tenure regime

73 Highlights

- The influence of land tenure and environmental potential for crop production on
 fertilization and nutrient balance was studied.
- Net nutrient exportation prevailed regardless of tenure regime and environmental
 potential, except for S.
- Land tenure had a low effect on fertilization and nutrient balance.
- In fields of high environmental potential, more nutrients were depleted than in fields of
 low environmental potential.
- Environmental heterogeneity should be considered when fertilizing.

82 1. Introduction

83 Field crops need regular nitrogen (N), phosphorus (P), and sulfur (S), among other nutrients, to 84 achieve yields. When the quantity of fertilizer additions used for crop production does not 85 compensate for the nutrient exported when harvesting, soil nutrient content declines, leading 86 to soil degradation and yields decrease over time (Nawara et al., 2018; Rawal et al., 2022). 87 Nutrient-rich soils, such as mollisols, can withstand negative nutrient balances for a longer time 88 without a significant impact on yield, while for other soils type, fertilization is essential for long-89 term farming (Jobbágy et al., 2021). Whereas overfertilization is a major concern in several 90 countries, in Latin-American the current trend is soil nutrient mining (i.e., soil nutrient depletion) 91 (FAO, 2019; Vitousek et al., 2009), favored by the capability of soils to produce with low fertilizer addition. Therefore, the implication of this nutrient deficit can have negative consequences for
the environment. For example, insufficient fertilizer addition reduces soil organic matter, which
is essential for CO² atmosphere sequestration, soil quality, and sustainability (Lal, 2009; Liu et
al., 2006).

96 One of the main obstacles to the implementation of soil conservation practices, such as 97 fertilization, is assumed to be land leasing. If tenants cultivate fields in the short term, they may 98 not be interested or financially capable in investing in soil conservation, as its benefits may only 99 materialize in the medium and long-term (Arora et al., 2015; Eder et al., 2021). Furthermore, 100 tenants have to pay rent, which limits their investment capacity when sowing, besides they are 101 at greater risk of losing the investment, profitability, and business viability (Bert et al., 2011). 102 Several studies have documented the negative consequences of land tenancy on soil 103 conservation, such as erosion events (Sklenicka et al., 2022), overexploitation of soils (Eder et 104 al., 2021), more soil compaction, and less soil organic matter content (Walmsley et al., 2020). 105 However, most studies were carried on high fertilization context, it is important to understand 106 the dynamics of farming systems based on nutrient depletion to achieve a more general 107 knowledge of tenure regime implications on land degradation.

Decisions on the fertilization regime applied at a field depend on multiple factors, such 108 109 as the regions climate, expected climatic conditions during the cropping season (for example 110 ENSO events), crop species and cultivar, farm-gate fertilizer and agricultural products prices, soil 111 properties and fertility, all of which can be resumed in the expected yield of a given field (Zhou 112 et al., 2010). The economic aspect, such as fertilization costs, is a key component of the decision-113 making process (Brunelle et al., 2015). These two decision dimensions are expected to interact, 114 for example, when evaluating the amount of fertilizer applied by owners versus tenants. For 115 example, differences could be expected when analyzing the field's environmental potential (i.e., 116 potential of the field for agricultural production), where owners and tenants would take 117 different decisions based on their priorities (economic return, extreme climate events risks, 118 expected product price, etc.). Although the impact of land tenure regime on soil conservation 119 investment has been previously studied (Higgins et al., 2018), few works have evaluated the 120 interaction between a field's environmental potential and tenure regimes over the amount of 121 fertilizer applied.

Argentina is mainly an agricultural country in which maize (*Zea mays L.*) and soybean (*Glycine max L.*) are the two dominant field crops. During the 2020/2021 campaign 58.4 Mt of maize and 48.8 Mt of soybean grain were produced. This represents the 5% and 13% of global 125 maize and soybean production (FAOSTAT, 2023). Furthermore, approximately 28% of the land 126 under extensive agriculture is managed by tenants, in which contracts are mostly short-term 127 (one to three years) (INDEC, 2019). Although tenants renew contracts in 85% of the cases, they 128 manage fields as if they were not, focusing on maximizing short-term economic gains at the 129 expense of field conservation (Arora et al., 2015). As a result, the tenure regime makes soil 130 conservation difficult, which is a major problem in Argentina. Agricultural production net 131 extracts an annual average of 26 kg ha⁻¹ of N and 5 kg ha⁻¹ of P per year (Díaz de Astarloa and Pengue, 2018), and particularly, summer crops have a negative nutrient balance of -31.7 kg N 132 133 ha⁻¹ yr⁻¹, -7.6 kg P ha⁻¹ yr⁻¹, and -4.69 kg S ha⁻¹ yr⁻¹ (Koritschoner et al., 2023). This soil fertility 134 loss is expected to have negative effects on crop yield (Eltun et al., 2002). Therefore, 135 understanding which variables limit adequate fertilizer application in Argentina is essential in 136 terms of conservation and production.

137 The main objective of this work was to evaluate the effect of land tenure regimes and 138 environmental potential on the amount of agricultural input used and soil degradation. 139 Specifically, we focused on fertilization and balance of three elements: N, P, and S in soybean 140 and maize fields. We hypothesize that tenants' investment in soil conservation practices differs 141 from owners since tenants have to pay rent and possess the lands for a short period of time, 142 while any benefits occur in the long term. We expect higher fertilizer additions and less nutrient 143 depletion on owner-managed fields. Secondly, we hypothesize that high-potential fields (i.e., 144 capacity to obtain high yield) have a better response in yields to fertilizer addition than low-145 potential fields guaranteeing the recouping of the investment. Therefore, we expect an increase 146 in fertilizer addition and a decrease in nutrient mining as environmental potential increases.

147 **2.** Material and methods

148 2.1 Description of the study area

149 Argentina's extensive agriculture is widely distributed going from the northern to the central part of the country (lat 22° S - lat 40° W), this includes the Pampas and Chaco plains (Fig. 1). This 150 151 area has a warm temperate climate, mean annual precipitation declines from 1200 mm in the 152 central-east to 400 mm in central-west and to the north, and annual mean temperature declines 153 uniformly from 23 °C in the north to 13 °C in the central part (Cravero et al., 2017). Soils are 154 mainly Mollisols, being Argiudols and Haplustols the most represented groups (Caviglia and 155 Andrade, 2010). The most widespread cropping systems focus on no-tillage summer crops, mainly in glyphosate-resistant soybeans, which actually represent 40% of the total cropped area, 156 157 and secondly glyphosate-resistant maize (MAGyP, 2022). Some crop rotation schemes, like maize/soybean, or wheat-soybean/maize are widely spread, and soybean monoculture is occasionally used (de Abelleyra and Verón, 2020). Soybean cropping is mainly inoculated with the symbiont and has a very low fertilization rate (Austin et al., 2006) (Fig. 2). Biological nitrogen fixation is reduced by N fertilization; however, sometimes a starter N is applied (Gan et al., 2002).

162 2.2. Data collection

163 We analyzed a total of 52,588 fields from seasons 2017 to 2022 from almost all extensive 164 agriculture areas of Argentina, covering a total area of 3,791,516 ha (Fig. 1). Of the total fields, 165 27,143 were cropping with soybean as the summer crop and 25,445 fields were cropping with 166 maize as the summer crop. Fields cropping with soybean preceded by a winter crop were 167 excluded, because this soybean has a lower yield and different management than soybean 168 preceded by fallow. The data were gathered and systematized by CREA 169 (https://www.crea.org.ar/), a non-profit civil association integrated by more than 1,800 farming 170 companies that share farming experiences and knowledge. On average, a CREA farming 171 company manages 737 ha, while Argentine farming companies manage an average of 686 ha 172 (INDEC, 2019). For each field, we had data on the tenure regime (owned or rented), field 173 location, region (15 regions as defined by CREA), season, farm company identity, crop variety, environmental potential (low, medium, and high), previous crop, N, P, and S fertilization (kg ha-174 175 ¹), and crop yield (kg ha⁻¹). Environmental potential of the field is defined by CREA experts based 176 on soil toposequence and the historic yield. High-potential fields are more prone to yield more 177 than medium and low-potential fields (Goldenberg et al., 2022). The environmental potential 178 assessment is intrinsic to each region, or so the same climatic regime.

179 *2.3. Nutrients balances*

180 We estimated the nutrient balance as the difference between nutrient inputs and outputs,181 detailed below:

2.3.1. Nutrient Input: These included fertilizer applied per unit of area (kg ha⁻¹), biological
N fixation, and atmospheric deposition. Biological N fixation was considered just for soybean
crops, and we estimated it as 60% of the total N harvests in soybean grains (Collino et al., 2015).
We used the atmospheric depositions estimated in a long-term study carried out on a central
location in the Rolling Pampa that measured an average annual atmospheric deposition of 7.2
kg S ha⁻¹ and 9 kg N ha⁻¹ (Carnelos et al., 2019).

189 2.3.2. Nutrients output: These included nutrients withdrawn through crop harvests and 190 direct and indirect N emissions. N losses through leaching are insignificant because of the low 191 fertilization rate and the flatness of the landscape in the study area (Portela et al., 2006). 192 Nutrients withdrawn through crop harvests were estimated with the yield per unit of area (kg 193 ha⁻¹) and the standard values of nutrients that the crop requires to produce grain (nutrient 194 content in the grain). The nutritional requirements are: 1 ton of maize requires 13.1 kg N, 2.64 195 kg of P, and 1.22 kg of S, 1 ton of soybean requires 48.5 kg of N, 5.4 kg of P, and 2.8 kg of S 196 (Cruzate and Casas, 2009). We used the N₂O emissions factors calculated by Koritschoner et al., 197 2023 based on IPCC guidelines (2019) for the Chaco-Pampean plain. The direct emission factor 198 of N₂O considered was 0.01 kg N₂O-N per kg N applied, and the indirect emission factor of N₂O 199 from volatilization considered was 0.1 kg NH₃-N and NO_x-N per kg of N applied or deposited via 200 crop residues. Since most of the fields are based on a no-tillage system, we assumed that crop 201 residues remain in the soil. We considered that for 1 ton of maize and soybean produced, 7 kg 202 and 20 kg of N, respectively, remain in crop residues (Ciampitti and García, 2007).

203 2.4. Data analysis

204 We established separate models for N, P and S addition and balance for maize and soybean. We 205 applied linear mixed-effects models, assuming a Gaussian error distribution (R version 4.1.3, 206 Ime4 package, Imer function) (Bates et al., 2015, R Core Team, 2022). All models considered 207 region, farm company identity, year, previous crop, and crop variety as non-nested random 208 effects, and tenancy, environmental potential, and their interactions as fixed-effects. We used 209 log transformation in N applied to soybean and S applied to maize and soybean models, and 210 root-square transformation in P applied to maize and soybean models to reduce the 211 heteroscedasticity of residuals. We performed a type III two-way ANOVA for each model 212 (package car, Anova function) (Fox & Weisberg, 2019), and a no orthogonal a priori comparison 213 between environmental potential levels with Dunn-Šidák test (multcomp package, cld function) 214 (Šidák, 1967) only for the models that evidence differences between means in the ANOVA.

215 3. Results

Of the total fields, 41% (21,718) were managed by tenants and 59% (30,870) were managed by owners (Fig. 1). High-potential fields were managed mainly by owners, while medium-potential fields were more frequently managed by tenants, and low-potential fields were equally distributed. Soybean was more frequent than maize and had a similar proportion for both tenure regimes. Weak differences in field size were detected with an average size of 72.1 ha and 73.3 ha for owners and tenants, respectively (p-value = 0.0753). Nutrient balance was negative for N

- and P, and positive for S in both crops. In maize fields, the mean ± standard error (SE) balances
- 223 of N, P, and S were -14.1 ± 0.19 kg ha⁻¹ yr⁻¹, -1.1± 0.052 kg ha⁻¹ yr⁻¹, and 9.1 ± 0.051 kg ha⁻¹ yr⁻¹,
- respectively, while in soybean fields the mean ± SE balances of N, P, and S were -46.3 ± 0.01, -
- 225 3.64 ± 0.041 kg ha⁻¹ yr⁻¹, and 7.14 ± 0.054 kg ha⁻¹ yr⁻¹, respectively.

226 3.1. Maize fields

227 We found strong evidence that N fertilization increased with the environmental potential but 228 there was no evidence that the tenure regime affected N fertilization or balance (Fig. 3, table 1). 229 In contrast, tenants applied less P than owners in high-potential fields, but there was no 230 difference in low and medium-potential fields (Fig. 3). Also, P fertilization in owners-managed 231 fields increased with the environmental potential, but in tenant-managed fields, there was a 232 difference in P application only between low and medium-potential fields (Fig. 3). Regarding S 233 application, tenants fertilized less than owners just in low-potential fields (Fig. 3). Differences in 234 application of S between environmental potential levels were found just for tenants-managed 235 fields, where high-potential fields were fertilizer more than low and medium-potential fields 236 (Fig. 3).

237 Since productivity increased more than fertilization with the environmental potential, 238 nutrient exports increased as well, except for S. N net export increased strongly with the 239 environmental potential, however, no differences were found between tenure regimes (Fig. 3, 240 table 1, table 2). P net export increased moderately with the environmental potential, except 241 for low-potential fields managed for owners where the balance was positive (Fig. 3). Also, 242 tenants net exported more P than owners just in high-potential fields (Fig. 3). S balance was 243 positive and no differences were found between tenure regimes (Fig. 3, table 1). High-potential 244 fields managed by owners had less S net import than low-potential fields (Fig. 3).

245 *3.2. Soybean fields*

No evidence was found that the environmental potential had an effect on fertilizer application, whereas the tenure regime had a moderate influence (Fig. 4, table 1). Tenants applied less N in low and high-potential fields, and more N in medium-potential fields than owners, but the effect size was weak (Fig. 4, table 2). Also, tenants applied less P than owners in all environmental potential levels (Fig. 4). For S application, no evidence of differences between tenure regimes was found (Fig. 3, table 1). 252 Similar to maize nutrient balance tendency, in soybean crops nutrient outputs exceed 253 inputs, except for S (Fig. 4). Particularly, N net export increased strongly with the environmental 254 potential (Fig. 4, table 2). Further, tenants net exported more N than owners in low-potential 255 fields, whereas no differences were found in medium and high-potential fields (Fig. 4, table 2). 256 P net export also increased with the environmental potential, but the magnitude of the effect 257 was smaller than N (Table 2). Tenants net exported more P than owners for the three 258 environmental potential levels (Fig. 4, table 2). In contrast, S balance was positive (Fig. 4). High-259 potential fields had less S net import than low-potential fields (Fig. 4, table 2). No effect of tenure 260 regime in S balance was found (Fig. 4, table 1, table 2).

261 **4. Discussion**

262 Besides the strong nutrient depletion currently taking place in soybean and maize crop 263 in Argentina, our study shows that environmental potential is the most important variable 264 explaining the dynamics of fertilizer use and the resulting nutrients balances at the field scale. 265 We found a strong general effect size of between 8% to 12% more fertilizer addition in high-266 potential fields than in low-potential fields (just in maize crops), and, opposite to our prediction, 267 between 25% to 146% more nutrient net exported in high-potential fields compared to low-268 potential fields (Fig. 3, Fig. 4). Regarding tenure regime, the only nutrient affected was P in maize 269 and soybean crops, tenants applied less and net exported more P than owners in both crops, 270 however, the effect size was weak (between 2% to 3% more fertilization and 6.5% to 37% more 271 nutrient net exported) compared with environmental potential effect (Fig. 3, Fig. 4).

272 *4.1. Tenure regime and nutrients dynamic*

273 We found that the tenure regime has a low effect on fertilization application and soil 274 conservation. The lack of effect of tenure regime on maize N fertilization and balance can be 275 explained by the fact that N has low residuality in soil (Glendining et al., 2001), so the majority 276 N applied is used by the current crop or is lost (Cassman et al., 2002), thus N fertilization could 277 be considered as a short-term practice. Further, maize has to be fertilized with N to obtain 278 economically profitable yields (Gregoret et al., 2011), therefore farmers probably cannot skimp 279 on N fertilizer application. In the opposite, the tenure regime had a weak effect on soybean N 280 fertilization and balance. Should be noted that both tenures had low N fertilization rates in 281 soybean and the environmental potential had no effect on the amount of nutrients applied 282 (Fig. 2, Fig. 4, Table 1). These results suggest that soybean crops received a standard amount 283 of N. This may be explained by the fact that soybean is inoculated with the N-fixing symbiont, 284 and this process is inhibited by N fertilization (Gan et al., 2002). Furthermore, soybeans show

high yields despite receiving very little fertilization (Austin et al. 2006). Finally, soybean fields
should have other soil conservation practices, such as more diversified crop rotation or
overfertilization in winter crops, that compensate N exportation.

288 The balance and application of P was affected moderately by the tenure regime. This 289 could be explained by the fact that a fraction of P applied in fertilization accumulates as residues 290 in soil, thus tenants could not be willing to apply a high amount of P if they will not take benefit 291 of it (Eghball et al., 1990; Syers et al., 2008). Hence, adequate P fertilization could be considered 292 a long-term practice. Furthermore, the largest P fertilization gap between owners and tenants 293 was in high-potential fields in both crops (Fig. 3, Fig. 4). This suggests that owners have better 294 knowledge of the heterogeneity of their lands, or they are concerned to conserve the high-295 potential of their fields.

296 An unexpected result was the positive balance of S, or rather S overfertilization. Our 297 results contrast with a previous study that found a negative nutrient balance of S in the same 298 studied area (Koritschoner et al., 2023). However, the S atmospheric deposition that we used 299 comes from a long-term measurement from a single location in the Pampa Argentina (Carnelos 300 et al., 2019), while Koritschoner et al. (2023) used a worldwide study that established a low 301 amount of S atmospheric deposition (Vet et al., 2014). This difference between S atmospheric 302 deposition estimation may explain the contrasting results. It should be considered that S 303 atmospheric depositions show a great variability with time and the distance to sea or a big city 304 (Carnelos et al., 2019). Local studies and models to predict S atmospheric deposition are needed 305 to improve S fertilization efficiency. On the other hand, the generally low adoption of S fertilizer 306 application could explain the absence of the effect of the tenure regime and the environmental 307 potential. Also, crop response to S fertilization is still unclear for both farmers and the scientific 308 community (Torres Duggan et al., 2012).

309 Several studies found that owners are more willing to invest in soil conservation and 310 long-term benefit practices, such as crop rotation with perennials and forage legumes, bunds, 311 and compost investment (Eder et al., 2021; Fraser, 2004; Sklenicka et al., 2015; Teshome et al., 312 2016). On the contrary, the studies that examined whether inorganic fertilization depends on 313 the tenure regime, show that there was no difference between the amount of inorganic fertilizer 314 applied by owners and fixed tenants (Akram et al., 2019; Teshome et al., 2016), partially in 315 concordance with our results. We found that the tenure regime affects P fertilization, but this 316 effect is low. Myyrä et al. (2007) show that P fertilization increases with lease contract time. 317 Further, a previous study in the Pampa region established that when tenants have legal lease terms of a field for a longer time period than one year, they manage that land similarly to an owned land (Arora et al., 2015). In our data is not established the lease term or the period of time that the field is managed by a unique farmer, if several tenants of our data have a longterm lease, this could explain the low effect size on P fertilization.

322 4.2. Nutrients export and environmental heterogeneity

323 The most significant finding of our study are the high rates of N and P exportation in Argentinean 324 soybean and maize crops (Fig. 3, Fig. 4). This general pattern is consistent with previous findings 325 that identify Argentina as one of the major nutrient exporter worldwide (Díaz de Astarloa and 326 Pengue, 2018; Guareschi et al., 2019; Koritschoner et al., 2023), and, particularly, the country 327 with the major P depletion process of the world (Schipanski and Bennett, 2012). Soybean crops 328 exported large quantities of N, and as shown in previous studies, is one of the major N soil mining 329 in Argentina (Austin et al., 2006; Koritschoner et al., 2023). This may be explained by the fact 330 that biological N fixation supplies 60% of that required by soybean crops, and the rest is 331 absorbed from the soil (Collino et al., 2015). We found that nutrient depletion is highly 332 dependent on environmental potential, this may be explained by the fact that high-potential 333 fields yield more than low-potential fields. Also, high-potential fields respond to fertilization with 334 a higher yield than low-potential fields, which is not compensated by higher fertilizer application 335 in high-potential fields, in consequence, the nutrient net export is higher (Fig. 3, Fig. 4). 336 Furthermore, this pattern of increased fertilizer application with the increase of the 337 environmental potential may be explained by the fact that high-potential fields secure high yield 338 profitability, hence farmers will recoup their investment in fertilization. An alternative 339 explanation to the trend of fertilization increased with the environmental potential is that high-340 potential fields conserve their good features due to receiving more nutrients, while medium or 341 low-potential fields degrade because they were not supplied with enough fertilizer (Rawal et al., 342 2022). However, the field's environmental potential assessment is based on the soil type and 343 surface topography, which strongly determine fields' yield (and partially, the nutrient balance), 344 so this last hypothesis could be partially correct.

In contrast to Argentina, other grain exporting countries, such as Brazil, have positive N and P balances (Guareschi et al., 2019). N-fixing bacteria in Brazil's fields fix 80% of the N needed by soybean crops, while in Argentina the N-fixation is 60% (Collino et al., 2015). However, Brazil's soils are poor in P, so fertilization is indispensable for farming, whereas Argentine farmers may save fertilize and deplete soils for the longest period without a significant effect on yield (Jobbágy et al., 2021). It should be noted that the synthetic fertilizer's price has risen in the last

few years, and the tendency is to continue increasing (Brunelle et al., 2015), so positive nutrient balance based on synthetic fertilizers could be non-viable in the future. To avoid nutrient loss and land degradation sustainable strategies, such as agroecological principles, need to be implemented in Argentine and world agricultural systems (Brunelle et al., 2015). This paper highlights that environmental potential is a key variable in farmer management and considers soil heterogeneity to ensure appropriate fertilization practices and reduce soil degradation is essential.

358 While the CREA farm data provides an exceptionally rich data set covering 3.8 M ha, the 359 CREA farm data can be prone to some biases. CREA provides management standards to all 360 farmers, which could level out differences in management between owners and tenants, which 361 might be more pronounced in the more diverse Argentinean agriculture situation. Further, this 362 study did not differentiate between short-term and long-term field tenancy. It could be plausibly 363 assumed that long-term tenant management does not differ from owner management. In future 364 studies, it would be interesting to consider the tenancy time period to better understand this 365 complex variable. Further, other sources of nutrient gains and losses are expected to modify the 366 calculated nutrient balances over time. For example, winter legumes or other crops preceding 367 summer crops can incorporate nitrogen into the soil or be over-fertilized to improve soil fertility 368 (Koritschoner et al., 2023).

369 **5.** Conclusions

370 Agriculture can have a massive impact on the environment and on the long-term 371 capacity of the natural capital to produce goods and services. In this study, using maize and 372 soybean in Argentina as an example, we show how environmental potential can have important 373 effects on farmers' fertilization decisions and the nutrient mining process. High-potential maize 374 fields were more fertilized but net exported more N and P than low-potential fields, implying 375 greater soil degradation in the most productive fields. The tenure regime had a weak effect on 376 nutrient application and balance. The highest differences were for P in high-potential fields, 377 where tenants applied and net exported more P than owners, suggesting that owners have a 378 better knowledge of fields heterogeneity and are more concerned with soil conservation. 379 However, soybean fertilization rates responded weakly to the environmental potential and the 380 tenure regime, and presented the highest nutrient mining, principally N, with high-potential 381 fields being the most affected in this study. S fertilization may not be needed in extensive 382 agriculture in Argentina; however, this is a global result and may vary between regions. This 383 work reveals the necessity of a better knowledge of soil and environmental heterogeneity of fields to achieve an adequate fertilization plan that avoids soil degradation. Soybean fields should implement extra soil conservation practices, such recover soil nutrient fertilizing in another crop of the rotation sequence. Also, more studies addressing the S cycle in agroecosystems are needed to optimize S fertilization. Both tenants and owners should increase nutrient input on average of N and P by 19%, 1% in maize crops, and 574% and 32% of N and P in soybean crop to achieve more sustainable agricultural systems in Argentina.

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394 References

- Akram, N., Akram, M.W., Wang, H., Mehmood, A., 2019. Does land tenure systems affect
 sustainable agricultural development? Sustain. 11. https://doi.org/10.3390/su11143925
- Arora, P., Bert, F., Podesta, G., Krantz, D.H., 2015. Ownership effect in the wild: Influence of
 land ownership on agribusiness goals and decisions in the Argentine Pampas. J. Behav.
 Exp. Econ. 58, 162–170. https://doi.org/10.1016/j.socec.2015.02.007
- Austin, A.T., Piñeiro, G., Gonzalez-Polo, M., 2006. More is less: Agricultural impacts on the N
 cycle in Argentina. Biogeochemistry 79, 45–60. https://doi.org/10.1007/s10533-0069002-1
- Bert, F.E., Podestá, G.P., Rovere, S.L., Menéndez, Á.N., North, M., Tatara, E., Laciana, C.E.,
 Weber, E., Toranzo, F.R., 2011. An agent based model to simulate structural and land use
 changes in agricultural systems of the argentine pampas. Ecol. Modell. 222, 3486–3499.
 https://doi.org/10.1016/j.ecolmodel.2011.08.007
- Brunelle, T., Dumas, P., Souty, F., Dorin, B., Nadaud, F., 2015. Evaluating the impact of rising
 fertilizer prices on crop yields. Agric. Econ. (United Kingdom) 46, 653–666.
- 409 https://doi.org/10.1111/agec.12161
- 410 Carnelos, D.A., Portela, S.I., Jobbágy, E.G., Jackson, R.B., Di Bella, C.M., Panario, D., Fagúndez,
- 411 C., Piñeiro-Guerra, J.M., Grion, L., Piñeiro, G., 2019. A first record of bulk atmospheric
- 412 deposition patterns of major ions in southern South America. Biogeochemistry 144, 261–

413 271. https://doi.org/10.1007/s10533-019-00584-3

- 414 Cassman, K.G., Dobermann, A., Walters, D.T., 2002. Agroecosystems, Nitrogen-use Efficiency,
- and Nitrogen Management. AMBIO A J. Hum. Environ. 31, 132.
- 416 https://doi.org/10.1639/0044-7447(2002)031[0132:anuean]2.0.co;2

417 Caviglia, O.P., Andrade, F.H., 2010. Sustainable Intensification of agriculture in the Argentinean

418 Pampas: Capture and use efficiency of environmental resources. Am. J. Plant Sci.

419 Biotechnol. 87, 117–129.

- 420 Ciampitti, I.A., García, F.O., 2007. Requerimientos Nutricionales. Absorción y Extracción de
 421 macronutrientes y nutrientes secundarios. I. Cereales, Oleaginosas e Industriales. Arch.
 422 Agron. 11, 13–16.
- 423 Collino, D.J., Salvagiotti, F., Perticari, A., Piccinetti, C., Ovando, G., Urquiaga, S., Racca, R.W.,
 424 2015. Biological nitrogen fixation in soybean in Argentina: relationships with crop, soil,
 425 and meteorological factors. Plant Soil 392, 239–252. https://doi.org/10.1007/s11104426 015-2459-8
- 427 Cravero, S. A. C., Bianchi, C. L., Elena, H. J., & Bianchi, A. R., 2017. Clima de Argentina: Mapas
 428 digitales mensuales de precipitación y precipitación menos evapotranspiración potencial.
 429 Adenda del Atlas climático digital de la República Argentina.
- 430 Cruzate, G., Casas, R.R., 2009. Extracción de Nutrientes en la Agricultura Argentina. Inf.
 431 Agronómicas del Cono Sur 44, 21–26.
- 432 de Abelleyra, D., Verón, S., 2020. Crop rotations in the Rolling Pampas: Characterization,
- 433 spatial pattern and its potential controls. Remote Sens. Appl. Soc. Environ. 18.
- 434 https://doi.org/10.1016/j.rsase.2020.100320
- 435 Díaz de Astarloa, D.A., Pengue, W.A., 2018. Nutrients Metabolism of Agricultural Production in
- 436 Argentina: NPK Input and Output Flows from 1961 to 2015. Ecol. Econ. 147, 74–83.
- 437 https://doi.org/10.1016/j.ecolecon.2018.01.001
- 438 Eder, A., Salhofer, K., Scheichel, E., 2021. Land tenure, soil conservation, and farm
- 439 performance: An eco-efficiency analysis of Austrian crop farms. Ecol. Econ. 180.
- 440 https://doi.org/10.1016/j.ecolecon.2020.106861
- 441 Eghball, B., Sander, D.H., Skopp, J., 1990. Diffusion, Adsorption, and Predicted Longevity of
- 442 Banded Phosphorus Fertilizer in Three Soils. Soil Sci. Soc. Am. J. 54, 1161–1165.
- 443 https://doi.org/10.2136/sssaj1990.03615995005400040041x
- Eltun, R., Korsæth, A., Nordheim, O., 2002. A comparison of environmental, soil fertility, yield,
- and economical effects in six cropping systems based on an 8-year experiment in Norway.
- 446 Agric. Ecosyst. Environ. 90, 155–168. https://doi.org/10.1016/S0167-8809(01)00198-0
- 447 FAO, 2019. World fertilizer trends and outlook to 2022.
- FAO, 2023. FAOSTAT: Production: Crops and livestock products. In: FAO. Rome. Cited Febrery
 2023. Available in: https://www.fao.org/faostat/en/#data/QCL

- 450 Fox, J., Weisberg, S., 2019. An R Companion to Applied Regression, Third Edition, Sage.
- 451 Fraser, E.D.G., 2004. Land tenure and agricultural management: Soil conservation on rented
 452 and owned fields in southwest British Columbia. Agric. Human Values 21, 73–79.
 453 https://doi.org/10.1023/B
- Gan, Y., Stulen, I., Posthumus, F., Van Keulen, H., Kuiper, P., 2002. Effects of N management on
 growth, N2 fixation and yield of soybean. Nutr. Cycl. Agroecosystems 62, 163–174.
 https://doi.org/10.1023/A:1015528132642
- 457 Glendining, M.J., Poulton, P.R., Powlson, D.S., Macdonald, A.J., Jenkinson, D.S., 2001.
- 458 Availability of the residual nitrogen from a single application of 15N-labelled fertilizer to 459 subsequent crops in a long-term continuous barley experiment. Plant Soil 233, 231–239.
- 460 https://doi.org/10.1023/A:1010508914895
- 461 Goldenberg, M.G., Burian, A., Seppelt, R., Santibañez Ossa, F.A., Bagnato, C.E., Satorre, E.H.,
- 462 Martini, G.D., Garibaldi, L.A., 2022. Effects of natural habitat composition and
- 463 configuration, environment and agricultural input on soybean and maize yields in
 464 Argentina. Agric. Ecosyst. Environ. 339. https://doi.org/10.1016/j.agee.2022.108133
- 465 Gregoret, M.C., Díaz Zorita, M., Dardanelli, J., Bongiovanni, R.G., 2011. Regional model for
- 466 nitrogen fertilization of site-specific rainfed corn in haplustolls of the central Pampas,
- 467 Argentina. Precis. Agric. 12, 831–849. https://doi.org/10.1007/s11119-011-9224-7
- 468 Guareschi, R.F., Boddey, R.M., Rodrigues Alves, B.J., Sarkis, L.F., dos Reis Martins, M., Jantalia,
- 469 C.P., Peña Cabriales, J.J., Vera Núñez, J.A., Urquiaga, S., 2019. Nitrogen, phosphorus and
- 470 potassium balance in agriculture of Latin America and the Caribbean. Terra Latinoam. 37,
- 471 105–119. https://doi.org/10.28940/terra.v37i2.423
- Higgins, D., Balint, T., Liversage, H., Winters, P., 2018. Investigating the impacts of increased
 rural land tenure security: A systematic review of the evidence. J. Rural Stud. 61, 34–62.
 https://doi.org/10.1016/j.jrurstud.2018.05.001
- 475 Instituto Nacional de Estadística y Censos INDEC, 2019. Censo Nacional Agropecuario 2018:
 476 resultados preliminares, 1ra edició. ed. Ciudad Autónoma de Buenos Aires.
- Jobbágy, E.G., Aguiar, S., Piñeiro, G., Garibaldi, L.A., 2021. Impronta ambiental de la agricultura
 de granos en Argentina: revisando desafíos propios y ajenos. Cienc. Hoy 29, 55–64.
- Koritschoner, J.J., Whitworth Hulse, J.I., Cuchietti, A., Arrieta, E.M., 2023. Spatial patterns of
 nutrients balance of major crops in Argentina. Sci. Total Environ. 858.

481

https://doi.org/10.1016/j.scitotenv.2022.159863

- 482 Lal, R., 2009. Challenges and opportunities in soil organic matter research. Eur. J. Soil Sci. 1–12.
 483 https://doi.org/10.1111/j.1365-2389.2008.01114.x
- Liu, X., Herbert, S.J., Hashemi, A.M., Zhang, X., Ding, G., 2006. Effects of agricultural
- 485 management on soil organic matter and carbon transformation a review. PLANT SOIL
 486 Environ. 52, 531–543.
- 487 Ministerio de agricultura, ganadería y pesca (2022). Available in:
- 488 https://datosestimaciones.magyp.gob.ar/reportes.php?reporte=Estimaciones. Access:
 489 20/12/2022.
- Myyrä, S., Pietola, K., Yli-Halla, M., 2007. Exploring long-term land improvements under land
 tenure insecurity. Agric. Syst. 92, 63–75. https://doi.org/10.1016/j.agsy.2006.02.009
- 492 Nawara, S., van Dael, T., De Cooman, E., Elsen, A., Merckx, R., Smolders, E., Amery, F., 2018.
- 493 Testing soil phosphorus in a depleting P scenario: an accelerated soil mining experiment.
- 494 Eur. J. Soil Sci. 69, 804–815. https://doi.org/10.1111/ejss.12684
- R Core Team, 2022. R: A language and environment for statistical computing. R Foundation for
 Statistical Computing, Vienna, Austria. URL https://www.R-project.org/.
- 497 Portela, S.I., Andriulo, A.E., Sasal, M.C., Mary, B., Jobbágy, E.G., 2006. Fertilizer vs. organic
 498 matter contributions to nitrogen leaching in cropping systems of the Pampas: 15N
- application in field lysimeters. Plant Soil 289, 265–277. https://doi.org/10.1007/s11104006-9134-z
- Rawal, N., Pande, K.R., Shrestha, R., Vista, S.P., 2022. Soil Nutrient Balance and Soil Fertility
 Status under the Influence of Fertilization in Maize-Wheat Cropping System in Nepal.
 Appl. Environ. Soil Sci. 2022, 1–11. https://doi.org/10.1155/2022/2607468
- Schipanski, M.E., Bennett, E.M., 2012. The Influence of Agricultural Trade and Livestock
 Production on the Global Phosphorus Cycle. Ecosystems 15, 256–268.
- 506 https://doi.org/10.1007/s10021-011-9507-x
- Šidák, Z., 1967. Rectangular confidence regions for the means of multivariate normal
 distributions. Journal of the American Statistical Association 62(318):626-633
- 509 Sklenicka, P., Efthimiou, N., Zouhar, J., van den Brink, A., Kottova, B., Vopravil, J., Zastera, V.,
 510 Gebhart, M., Bohnet, I.C., Molnarova, K.J., Azadi, H., 2022. Impact of sustainable land

- 511 management practices on controlling water erosion events: The case of hillslopes in the
- 512 Czech Republic. J. Clean. Prod. 337. https://doi.org/10.1016/j.jclepro.2022.130416
- 513 Sklenicka, P., Molnarova, K.J., Salek, M., Simova, P., Vlasak, J., Sekac, P., Janovska, V., 2015.
- 514 Owner or tenant: Who adopts better soil conservation practices? Land use policy 47,
- 515 253–261. https://doi.org/10.1016/j.landusepol.2015.04.017
- Syers, J.K., Johnston, A.., Curtin, D., 2008. Efficiency of soil and fertilizer phosphorus use, Food
 and Agriculture Organization of the United Nations.
- 518 Teshome, A., de Graaff, J., Ritsema, C., Kassie, M., 2016. Farmers' Perceptions about the
- 519Influence of Land Quality, Land Fragmentation and Tenure Systems on Sustainable Land520Management in the North Western Ethiopian Highlands. L. Degrad. Dev. 27, 884–898.
- 521 https://doi.org/10.1002/ldr.2298
- 522 Torres Duggan, M., Melgar, R., Rodríguez, M.B., Lavado, R.S., Ciampitti, I.A., 2012. Sulfur
 523 fertilization technology in the Argentine Pampa region: A review. Rev. Agron. Ambient.
 524 61 72
- 52432, 61–73.
- 525 Vet, R., Artz, R.S., Carou, S., Shaw, M., Ro, C.U., Aas, W., Baker, A., Bowersox, V.C., Dentener,
- 526 F., Galy-Lacaux, C., Hou, A., Pienaar, J.J., Gillett, R., Forti, M.C., Gromov, S., Hara, H.,
- 527 Khodzher, T., Mahowald, N.M., Nickovic, S., Rao, P.S.P., Reid, N.W., 2014. A global
- 528 assessment of precipitation chemistry and deposition of sulfur, nitrogen, sea salt, base
- 529 cations, organic acids, acidity and pH, and phosphorus. Atmos. Environ. 93, 3–100.
- 530 https://doi.org/10.1016/j.atmosenv.2013.10.060
- 531 Vitousek, P.M., Naylor, R., Crews, T., David, M.B., Drinkwater, L.E., Holland, E., Johnes, P.J.,
- 532 Katzenberger, J., Martinelli, L.A., Matson, P.A., Nziguheba, G., Ojima, D., Palm, C.A.,
- 533 Robertson, G.P., Sanchez, P.A., Townsend, A.R., Zhang, F.S., 2009. Nutrient imbalances in
- agricultural development. Science (80-.). 324, 1519–1520.
- 535 https://doi.org/10.1126/science.1170261
- Walmsley, A., Azadi, H., Tomeckova, K., Sklenicka, P., 2020. Contrasting effects of land tenure
 on degradation of Cambisols and Luvisols: The case of Central Bohemia Region in the
- 538 Czech Republic. Land use policy 99. https://doi.org/10.1016/j.landusepol.2020.104956
- 539 Zhou, Y., Yang, H., Mosler, H.J., Abbaspour, K.C., 2010. Factors affecting farmers ' decisions on
- 540 fertilizer use : A case study for the Chaobai watershed in Northern China. Cons. J.
- 541 Sustain. Dev. 4, 80–102.

Table 1: ANOVA table of each model, N, P, and S applied and balance in maize and soybean crops. Table shows chis-q values and p-value in brackets for each fixed variable, number of observations (n=), and R² conditional and marginal for each model. Degrees' freedom were 1 for tenancy, and 2 for environmental potential (EP) and tenancy:EP.

Crop	Response variable		Predictors	R ²		
		Tenancy	EP	Tenancy: EP	Conditional	Marginal
Maize	N applied	0.12 (0.73)	166.9 (<.0001)	1.76 (0.41)	0.67	0.015
	(n= 16,056)					
	N balance	2.85 (0.091)	133.9 (<.0001)	1.91 (0.38)	0.62	0.013
	(n= 16 <i>,</i> 056)					
	P applied	8.49 (0.0036)	38.9 (<.0001)	31.9 (<.0001)	0.63	0.008
	(n= 13,937)					
	P balance	3.24 (0.071)	140.8 (<.0001)	13.7 (0.001)	0.55	0.016
	(n= 13,937)					
	S applied	12.8 (0.0003)	35.9 (<.0001)	11.0 (0.0039)	0.71	0.005
	(n= 7,523)					
	S balance	0.24 (0.62)	8.68 (0.013)	11.0 (0.004)	0.61	0.002
	(n= 7,523)	/				
Soybean	N applied	12.6 (0.0004)	3.51 (0.17)	0.90 (0.63)	0.81	0.004
	(n= 7,436)	/		/		
	N balance	9.93 (0.001)	194.4 (<.0001)	16.8 (0.0002)	0.72	0.005
	(n= 7,436)					
	P applied	15.5 (<.0001)	0.87 (0.64)	4.32 (0.11)	0.71	0.013
	(n=11,610)					
	P balance	14.1 (0.0001)	150.0 (<.0001)	6.07 (0.048)	0.68	0.021
	(n=11,610)					
	S applied	6.11 (0.013)	5.48 (0.064)	1.91 (0.38)	0.84	0.003
	(n=4,222)					
	S balance (n=4,222)	3.25 (0.071)	12.9 (0.001)	0.90 (0.64)	0.72	0.005
	(11-7,222)					

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548

549 **Table 2:** Summary table showing fixed effects estimates, standard error, and P-value in brackets

of each parameter of the models explaining nutrients applied and balance as responses to land

- tenure regime and environmental potential (EP). Standard deviations of random effects are also
- 552 included.

	Maize						Soybean					
Fixed effects	N applied	N balance	P applied	P balance	S applied	S balance	N applied	N balance	P applied	P balance	S applied	S balance
Intercept	65.1±5.8 (<.0001)	-2.55±8.6 (0.77)	3.68±0.11 (<.0001)	0.28±1.4 (0.85)	1.97±0.088 (<.0001)	9.65±0.9 (<.0001)	1.85±0.075 (<.0001)	-39.8±5.8 (<.0001)	3.18±0.11 (<.0001)	-2.22±1.5 (0.18)	1.73±0.14 (<.0001)	7.86±1.2 (<.0001)
Owner	0.44±1.3 (0.73)	-2.72±1.6 (0.091)	0.12±0.042 (0.0035)	0.81±0.45 (0.071)	0.13±0.038 (0.0003)	0.24±0.48 (0.62)	0.093±0.026 (0.0004)	4.21±1.3 (0.001)	0.14±0.035 (<.0001)	1.36±0.36 (0.0001)	0.11±0.047 (0.013)	1.05±0.58 (0.071)
Medium Ep	7.08±1.1 (<.0001)	-10.5±1.3 (<.0001)	0.22±0.035 (<.0001)	-1.14±0.38 (0.002)	0.08±0.03 (0.008)	-0.9±0.38 (0.018)	0.040±0.021 (0.066)	-5.12±1.1 (<.0001)	0.015±0.029 (0.59)	-1.77±0.3 (<.0001)	0.05±0.036 (0.16)	-0.90±0.45 (0.045)
High Ep	13.6±1.1 (<.0001)	-15.9±1.4 (<.0001)	0.16±0.037 (<.0001)	-3.81±0.40 (<.0001)	0.17±0.032 (<.0001)	-0.28±0.40 (0.49)	0.039±0.024 (0.10)	-14.8±1.2 (<.0001)	0.029±0.031 (0.36)	-3.77±0.32 (<.0001)	0.08±0.037 (0.022)	-1.61±0.46 (0.0005)
Owner: medium	-1.51±1.3 (0.25)	1.84±1.6 (0.26)	-0.08±0.04 (0.045)	-0.47±0.47 (0.31)	-0.08±0.039 (0.037)	0.08±0.5 (0.86)	-0.024±0.025 (0.34)	-5.22±1.3 (0.0001)	0.03±0.035 (0.40)	-0.50±0.37 (0.18)	-0.05±0.05 (0.29)	-0.54±0.60 (0.36)
Owner: high	-0.64±1.4 (0.64)	0.63±1.7 (0.71)	0.07±0.04 (0.080)	0.70±0.49 (0.15)	-0.13±0.039 (0.001)	-1.0±0.50 (0.047)	-0.020±0.027 (0.46)	-2.92±1.4 (0.044)	0.072±0.038 (0.059)	0.13±0.4 (0.72)	-0.07±0.05 (0.16)	-0.32±0.61 (0.60)
Random effects	n Standard deviation											
Farm	18.2	18.5	0.64	5.70	0.49	5.25	0.40	11.8	0.56	4.64	0.49	5.17
Crop variety	4.92	7.41	0.15	1.04	0.13	1.60	0.10	5.55	0.15	1.47	0.15	1.28
Previous crop	3.16	5.86	0.09	0.64	0.07	0.60	0.065	3.86	0.20	0.50	0.20	0.26
Region	20.2	18.7	0.37	3.42	0.25	2.66	0.22	13.0	0.30	3.87	0.43	3.19
Year	4.64	15.3	0.08	2.31	0.08	0.60	0.06	9.5	0.057	2.42	0.08	1.34
							1					



Figure 1: Study sites with the proportion of tenancy regime and number of observations for eachregion.



Figure 2: Density plots show the data distribution of nitrogen, phosphorus and sulfur applied
and balance, field sizes, and yield of soybean (green) and maize crops (yellow).



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Figure 3: Mean observed (± SE) of nitrogen applied (A) and balance (B), phosphorus applied (C) and balance (D), and sulfur applied (E) and balance (F) in maize crops managed by tenants (orange) and owners (cyan) in low, medium, and high-potential fields. Means with different lowercase letters differ statistically at α = 0.05 according to a pairwise post hoc Dunn-Šidák test.



Figure 4: Mean observed (± SE) of nitrogen applied (A) and balance (B), phosphorus applied (C) and balance (D), and sulfur applied (E) and balance (F) in soybean crops managed by tenants (orange) and owners (cyan) in low, medium, and high-potential fields. Means with different lowercase letters differ statistically at $\alpha = 0.05$ according to a pairwise post hoc Dunn-Šidák test.