# This is the accepted manuscript version of the contribution published as:

Goldenberg, M.G., **Burian, A., Seppelt, R.**, Santibañez Ossa, F.A., Bagnato, C.E., Satorre, E.H., Martini, G.D., Garibaldi, L.A. (2022): Effects of natural habitat composition and configuration, environment and agricultural input on soybean and maize yields in Argentina *Agric. Ecosyst. Environ.* **339**, art. 108133

## The publisher's version is available at:

http://dx.doi.org/10.1016/j.agee.2022.108133

## Highlights

- Agricultural input correlated negatively with the percentage of natural habitat.
- Environmental and crop management variables explained spatial variation in yield.
- Fields yields increased when they were recently transformed from natural habitat.
- Effects of natural habitat on yield were not detected, probably due to high input.

## **Tables**

**Table 1.** Environmental, crop management and landscape variables considered in this

study. Quantitative variables data is summarized in Table A.1.

Variable class	Variable	Туре	Units/categories
Environment	Latitude	Quantitative	degrees
	Longitude	Quantitative	degrees
	Region	Qualitative	11 categories
	Environmental potential	Quantitative	1-3
Management	Nitrogen fertilization	Quantitative	kg ha <sup>-1</sup>
	Phosphorus fertilization	Quantitative	kg ha <sup>-1</sup>
	Sowing date	Quantitative	days
	Seed density	Quantitative	seeds ha <sup>-1</sup>
	Seed treatment	Qualitative	no treatment, field treatment,
			professional seed treatment
	Previous crop	Qualitative	double crop, service crop, same
			crop, different crop, and natural
			area (i.e., recently converted
			into agricultural land)
	Fungicide application	Qualitative	yes/no
	Irrigation	Qualitative	yes/no
	Farm	Qualitative	farm ID
	Crop cultivar	Qualitative	cultivar ID
Landscape	Natural habitat	Quantitative	%
	Edge density	Quantitative	m ha <sup>-1</sup>

\* Environmental potential is a variable that summarizes the environmental (climatic and soil) conditions and yield potential of each field. This is established by experienced agronomists directly involved in crop management decisions related to the fields. For simplification, this variable was converted to a quantitative value ranging from 1 to 3.

 Table 2. Spearman correlation coefficients between the main environmental and input

 quantitative variables and the landscape variables for soybean and maize. Numbers in

 bold indicate statistically significant coefficients (*p*-value \*< 0.05, \*\*<0.01,</td>

 \*\*\*<<0.001).</td>

	Intensification Landscape variables		pe variables
Crop	variables	Natural habitat (%)	Edge density (m ha <sup>-1</sup> )
	Environmental potential	-0.163***	0.020
ean	Phosphorus fertilization	-0.212***	-0.053
Soybe	Seed density	0.035	-0.019
	Sowing date	-0.031	0.083*
	Environmental potential	-0.065	-0.046
	Nitrogen fertilization	-0.147***	-0.075
laize	Phosphorus fertilization	-0.101**	-0.073
2	Seed density	-0.222***	-0.097*
	Sowing date	-0.182***	-0.198***

## **Figures**



**Figure 1.** Soybean and maize field distribution across Argentina, covering more than 324,000 hectares of farmland. The images below show contrasting landscapes from different regions. The density plots on the right show the data distribution of some continuous variables of 2,858 soybean fields and 1,548 maize fields for 2018-2019.



Figure 2. The left panel shows normalized parameter estimates and confidence intervals for the fixed part of the soybean yield model (Intercept = 0.480). Random effects had a standard deviation of 0.080 for region, 0.027 for cultivar and 0.074 for farm. The right panel shows the observed versus fitted values from the model ( $r^2 = 0.82$ ).



**Figure 3.** The left panel shows normalized parameter estimates and confidence intervals for the fixed part of the maize yield model (Intercept = 0.179). Random effects had a standard deviation of 0.060 for region, 0.060 for cultivar and 0.098 for farm. The right panel shows the observed values versus model prediction ( $r^2 = 0.81$ ).



Figure 4. Effect of the percentage of natural habitat and edge density on the

standardized residuals of the soybean and maize models (Figures 2 and 3 respectively).

±

## **Tables**

**Table A.1.** Data summary for quantitative variables of Table 1.

Variable class	Variable	Crop	Min value	Max value	Mean value
Environment	Latitude	Soybean	-38.59	-26.36	-33.95
		Maize	-38.03	-26.50	-33.80
	Longitude	Soybean	-65.59	-57.80	-61.52
		Maize	-65.55	-58.92	-62.36
	Environmental potential	Soybean	1	3	2.35
		Maize	1	3	2.36
Management	Nitrogen fertilization	Maize	5.50	227.66	83.09
	Phosphorus fertilization	Soybean	1.63	50.24	11.13
		Maize	2.97	73.50	20.84
	Sowing date	Soybean	10/10/18	01/21/19	11/15/18
		Maize	08/05/18	01/21/19	10/20/18
	Seed density	Soybean	22	75	40.01
		Maize	3.80	9.80	6.82
Landscape	Natural habitat	Soybean	0.06	≈100	36.90
		Maize	0.03	≈100	35.31
	Edge density	Soybean	0.00	84.93	33.55
		Maize	0.00	72.78	31.32

#### **Declaration of interests**

⊠The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

□The authors declare the following financial interests/personal relationships which may be considered as potential competing interests:

1	Effects of natural habitat composition and configuration, environment and
2	agricultural input on soybean and maize yields in Argentina
3	Matías G. Goldenberg <sup>1,2*</sup> , Alfred Burian <sup>3,4</sup> , Ralf Seppelt <sup>4,5,6</sup> , Fernanda A. Santibañez
4	Ossa <sup>1,2</sup> , Camilo E. Bagnato <sup>1</sup> , Emilio H. Satorre <sup>7,8,9</sup> , Gustavo D. Martini <sup>9</sup> , Lucas A.
5	Garibaldi <sup>1,2</sup>
6	
7	<sup>1</sup> Universidad Nacional de Río Negro, Instituto de Investigaciones en Recursos
8	Naturales, Agroecología y Desarrollo Rural, Río Negro, Argentina.
9	<sup>2</sup> Consejo Nacional de Investigaciones Científicas y Técnicas, Instituto de
10	Investigaciones en Recursos Naturales, Agroecología y Desarrollo Rural, Río Negro,
11	Argentina.
12	<sup>3</sup> Marine Ecology Department, Lurio University, Nampula, Mozambique.
13	<sup>4</sup> UFZ—Helmholtz Centre for Environmental Research, Department of Ecological
14	Modelling, Permoserstr. 15, 04318 Leipzig, Germany.
15	<sup>5</sup> Institute of Geoscience and Geography, Martin-Luther-University Halle-Wittenberg,
16	06099 Halle (Saale), Germany.
17	<sup>6</sup> German Centre for Integrative Biodiversity Research (iDiv), Halle-Jena-Leipzig,
18	Leipzig, Germany.
19	<sup>7</sup> Universidad de Buenos Aires, Facultad de Agronomía, Departamento de Producción
20	Vegetal, Cátedra de Cerealicultura, Buenos Aires, Argentina.
21	<sup>8</sup> CONICET – Universidad de Buenos Aires. Instituto de Investigaciones Fisiológicas y
22	Ecológicas Vinculadas a la Agricultura (IFEVA), Buenos Aires, Argentina.
23	<sup>9</sup> Unidad de Investigación y Desarrollo, Área de Agricultura, Asociación Argentina de
24	Consorcios Regionales de Experimentación Agrícola (AACREA), Sarmiento 1236
25	(C1041AAZ), Buenos Aires, Argentina.

26	* Corresponding author: IRNAD, Universidad Nacional de Río Negro, Anasagasti
27	1456, San Carlos de Bariloche (8400), Río Negro, Argentina. E-mail:
28	mgoldenberg@unrn.edu.ar

29

#### 30 Highlights

Agricultural input correlated negatively with the percentage of natural habitat.
Environmental and crop management variables explained spatial variation in yield.
Fields yields increased when they were recently transformed from natural habitat.
Effects of natural habitat on yield were not detected, probably due to high input.

35

#### 36 Abstract

A fundamental challenge of land use management is to sustain the production of food, 37 energy and fiber whilst preserving biodiversity and ecosystem functions. Some 38 promising solutions to current resource-use conflicts are rooted in (agro) ecological 39 intensification, which proposes that ecosystem functions provided by natural habitat can 40 41 largely replace agrochemical inputs. Here, we evaluate how natural habitat is distributed in relation to agricultural input and the environmental potential for crop production, and 42 whether natural habitat can explain the variations in yield not explained by management 43 44 and environmental factors. In our analysis, we relied on environmental and management variables from 2,858 soybean and 1,548 individual maize fields provided by a farming 45 organization in Argentina, and assessed landscape metrics of natural habitat 46 composition (percentage of natural habitat) and configuration (edge density) for each 47 48 one. We found that fields with higher fertilizer and seed input had lower percentages of natural habitat. Spatial variation in yield was well explained by environmental and 49 management variables for both soybean and maize fields, and landscape metrics showed 50

no relationship to the residuals of the models. However, fields recently transformed
from natural habitat had higher yields than those with a long history of agricultural use.
We conclude that compensatory management may mask the beneficial effects of natural
habitat to some extent, especially in fields with intensive agrochemical use. **Keywords**: Ecological intensification, agroecology, agricultural landscape, non-crop

area, conventional cropping, on-farm yield variability

58

#### 59 1. Introduction

Agricultural expansion and conventional land-use intensification have led to landscape
homogenization and global biodiversity loss (Diaz et al. 2019, Martin et al. 2019,
Seppelt et al. 2014). Decreases of biodiversity in agricultural landscapes have also
resulted in a decline in ecosystem services supporting sustainable crop production
(Tscharntke et al. 2005). Such decreases are not surprising, since conventional
intensification has largely ignored the positive role of biodiversity in crop production
systems (Seppelt et al. 2020).

67 Over the last few years, the establishment and conservation of natural habitat in 68 agricultural landscapes have been promoted under the paradigm of ecological intensification (Garibaldi et al. 2019). The presence of natural areas in agroecosystems 69 is expected to create win-win situations for biodiversity and agriculture through the 70 ecosystem functions these areas provide for crops (Garibaldi et al. 2020). Some of the 71 72 most important ecosystem functions in this context are generated by mobile species and their interactions with crops (Kremen et al., 2007). The importance of migration 73 74 between natural and agricultural habitats has been shown for multiple crop types and ecosystem functions such as pollination (Ricketts et al. 2008) and biological pest control 75

76	(Karp et al. 2018; Tscharntke et al. 2016). However, the effect of the composition and
77	configuration of natural habitat on field crop yield is still little understood.
78	Farmers are responsible for managing most of the world's populated land
79	(Ramanakutty et al. 2008). Improved knowledge of on-farm benefits of changes in the
80	composition and configuration of natural habitat on a landscape scale could have
81	worldwide environmental implications. For example, previous studies showed that both
82	soybean and maize, two of the most widely grown crops, benefit from natural habitat as
83	it promotes the natural enemies of pests (Gonzalez et al 2020, Santana Sousa et al.
84	2012). Additionally, soybean yields are also increased through more effective
85	pollination (Garibaldi et al. 2021, Monasterolos et al. 2015).
86	Our main objective was to assess the relationships between the environment,
87	land management, landscape structure and yields of soybean and maize in Argentina.
88	We targeted soybean and maize crops because, globally, 120.5 and 197.2 million ha
89	were harvested of these crops, respectively, during 2019 (FAO 2019). Argentina
90	accounted for 14% of the global soybean and 4% of the global maize production area in
91	that year (FAO 2019). To understand the relationships between landscape,
92	environmental and management variables, we first evaluated how the landscape metrics
93	(percentage of natural habitat and edge density) were distributed with regard to
94	agricultural input and environmental potential for crop production. We then assessed the
95	main drivers of crop yield and evaluated whether landscape metrics could account for
96	the spatial variation in yield that was not explained by management and environmental
97	factors. This information would indicate whether the ecosystem services provided by
98	natural habitat substantially impact yields. In total, we gathered data from 2,858
99	soybean and 1,548 maize fields.

#### 101 **2. Methods**

#### 102 2.1. Data collection

Data was collected using an extensive, standardized protocol co-developed with the 103 Regional Consortiums for Agricultural Experimentation (CREA) through the DAT 104 CREA project. Regional Consortiums for Agricultural Experimentation is a non-profit 105 civil association integrated and directed by agricultural entrepreneurs (>1,800 farms) 106 107 who meet in groups to share experiences and knowledge (https://www.crea.org.ar/). In total, the assessed area covered more than 324,000 ha of agricultural land distributed 108 across almost all the extensive grain-producing regions of Argentina (Figure 1). 109 110 Specifically, we used data from individual fields of CREA farms for the 2018-2019 111 growing season. For each soybean and maize field, we gathered data on environmental and management variables (Table 1). In addition, we used Argentina's national Crop 112 113 Data Layer 1 (INTA 2019) to quantify the landscape composition and configuration around each field in our database. We established a radius of 1,500 m as landscape size 114 since this distance covers some of the most important ecosystem functions provided by 115 natural habitat (Greenleaf & Kremen 2006) and is within the range of similar previous 116 117 studies (Martin et al. 2019). This crop data layer was divided into two categories: 118 cropped and non-cropped areas. Non-cropped areas included natural forests, grasslands and wetlands corresponding to semi-natural and natural habitats (hereafter natural 119 habitat for simplification). Land classification was carried out through the Google Earth 120 121 Engine platform (https://earthengine.google.com). For each field we then calculated the percentage of natural habitat and edge density (Table 1); i.e., the sum of the lengths of 122 123 all crop edge segments that bordered natural habitat in the landscape, divided by the total area. This analysis was implemented using the "landscapemetrics" package in R (R 124 Core Team 2020). 125

126

127 Approximate location Figure 1

- 128 Approximate location Table 1
- 129

130 2.1. Data analysis

131 2.1.1. Correlations between environmental, management and landscape variables

132 We computed Spearman's correlation to investigate associations between landscape

133 variables (percentage of natural habitat and edge density), the environmental potential

and crop management variables presented in Table 1 (i.e., nitrogen fertilization,

135 phosphorus fertilization, seed density and sowing date). The environmental potential is

136 a variable that summarizes the environmental (climatic and soil) conditions that

137 influence the yield potential of fields. Nitrogen fertilization was not considered for

138 soybeans since it is a natural nitrogen fixer, and this crop is also inoculated to promote

the biological nitrogen fixation capacity of this species (Leggett et al. 2017).

Fertilization and seed density are direct measures of intensification as they reflect the
level of input that a field crop receives. Sowing date is related to different strategies for
crop development, to take advantage of the best climatic conditions and thus maximize
yields or reduce losses.

144

145 2.1.2. Prediction of soybean and maize yields

We estimated mixed-effects models to evaluate the main drivers of crop yield, with separate models being established for soybean and maize yields. Due to the complex correlation between landscape structure and management variables (Table 2), we first implemented yield models without considering landscape metrics. We identified all the potentially relevant variables for yield prediction, which included all the environmental

and management variables shown in Table 1. Three non-nested random intercepts were 151 152 included to account for the potential confounding effects of region, crop cultivar (to account for genetic variation among cultivars) and farm (the same farm may manage 153 multiple fields). We visually determined which predictors needed transformation to 154 achieve linearity and confirmed transformation choices by comparing the Akaike 155 information criterion (AIC) values of the models with and without transformation 156 157 (Burnham et al., 2011). Phosphorus fertilization was log-transformed for soybean. To position variables on a common scale all quantitative variables were normalized. This 158 normalization involved rescaling the values of each variable so that they ranged 159 160 between 0 and 1 (min-max scaling). We also calculated pairwise correlations of 161 continuous predictors and calculated the variance inflation factor for all predictors to rule out multicollinearity. 162

163 For the model selection, we followed the Zuur et al. (2009) protocol for fitting mixed- effects models, first establishing the random structure and then the fixed effects. 164 We compared models of different complexity using the AIC. Our most complex model 165 included all fixed and random effects and the following two-way interaction effects 166 among fixed effects: previous crop x fertilization (nitrogen and phosphorus), nitrogen 167 168 fertilization x phosphorus fertilization, environmental potential x phosphorus fertilization, and environmental potential x nitrogen fertilization. Mixed-effects models 169 were fitted using the *lmer()* function from "lme4" package in R (Bates et al. 2015). To 170 select the fixed effects, the parameters of the global model were re-estimated using 171 maximum likelihood. Based on AIC, we then eliminated each interaction following a 172 173 stepwise procedure, using delta AIC > 2 as a guideline (Oddi et al. 2019, Burnham and Anderson 2002). Therefore, an interaction was considered important if it reduced the 174 AIC value of the model by at least 2 units from the value without the interaction (Oddi 175

et al. 2019). The same procedure was then carried out for non-interaction fixed-effect terms following a parsimonious criterion (Garibaldi et al. 2014). The final model parameter values were estimated using restricted maximum likelihood estimation. Model assumptions were checked by visual evaluation of the residual scatter plots (residual vs. predicted values). The conditional  $r^2$  was used as a goodness-of-fit metric and is hereafter referred to as  $r^2$ .

182

183 2.1.3. The effects of the percentage of natural habitat and edge density on models'

184 *residuals* 

185 We extracted the standardized Pearson's residuals from the final models of soybean and maize and built regression models to evaluate whether they responded to the percentage 186 of natural habitat and edge density. This was done to determine whether there was still 187 188 yield variability that could be explained by these landscape metrics after accounting for environmental and management variables through the yield model. Models (one for 189 each crop's residuals as a response variable) were fitted using the *lm()* functions of the 190 base package in R. Using AIC, we then evaluated whether the percentage of natural 191 habitat and edge density were important predictors of the residuals of soybean and 192 193 maize models. To address possible variations in the effect of the percentage of natural habitat due to landscape complexity (Tscharntke et al. 2012), the interaction between 194 the percentage of natural habitat and edge density (as a proxy for landscape complexity) 195 was also evaluated. As before, we used AIC values to identify the most parsimonious 196 models and checked model assumptions by visual evaluation of the residual scatter plots 197 198 (residual vs. predicted values).

199

200 **3. Results** 

During the 2018-2019 growing season, we collected data from 2,858 soybean and 1,548 201 202 maize fields across almost all extensive crop regions of Argentina (Figure 1). Average single field size was 70.83 ha for soybean and 63.90 ha for maize; maximum field sizes 203 were 500 and 360 ha, respectively. Landscapes with soybean fields were characterized 204 by an average of 36.9% natural habitat and an edge density of 33.6 m ha<sup>-1</sup>. Maize field 205 landscapes had an average of 35.3% natural habitat and 31.3 m ha<sup>-1</sup> edge density. The 206 soybean yield ranged between 982 kg ha<sup>-1</sup> and 5,984 kg ha<sup>-1</sup>, whilst the maize yield 207 varied between  $3,200 \text{ kg ha}^{-1}$  and  $14,300 \text{ kg ha}^{-1}$  (Figure 1). 208

209

210 3.1. Correlations between environmental, management and landscape variables

We consistently found negative correlations between conventional intensification and
natural habitats for each of the two crop types (Table 2). For soybean, phosphorus
fertilization had the highest negative correlation with the percentage of natural habitat.
Environmental potential showed a significant but weaker negative correlation. A
significant positive correlation between edge density and sowing date was detected,
although the correlation coefficient was low.

In maize fields, seed density had the highest negative correlation with the percentage of natural habitat, and sowing date also had a negative correlation with this variable. Although fertilization use had a significant negative correlation with the percentage of natural habitat, the correlation coefficient was rather low (Table 2). A strong negative correlation for maize was detected between edge density and sowing date. The negative correlation between seed density and edge density was also significant, but much lower in magnitude.

224

#### 225 Approximate location Table 2

#### 227 3.2. Prediction of soybean and maize yields

Environmental potential, phosphorus fertilization, fungicide application, previous crop, 228 and sowing date were the predictors of the fixed part of the model that best explained 229 soybean yield variability (Figure 2). The most important predictor of yield was the 230 sowing date, which had a strong negative impact. Yield decreased on average by 147 kg 231 ha<sup>-1</sup> for each week of later sowing. Between the beginning and the end of the sowing 232 period studied (103 days), yield predictions decreased by 2,169 kg ha<sup>-1</sup>. Field history 233 also affected soybean yield. Fields recently converted from natural vegetation (i.e., 234 natural area, natural habitat) had the highest yields (e.g., 260 kg ha<sup>-1</sup> more than fields 235 that had previously been used for soybean production). When phosphorus fertilization 236 increased by 20 kg ha<sup>-1</sup>, the predicted yields increased by 350 kg ha<sup>-1</sup>. Moreover, the 237 predicted yield increased by 640 kg ha<sup>-1</sup> when environmental potential went from 0 to 1 238 on the scale, and by 320 kg ha<sup>-1</sup> due to fungicide application. Of the random effects, the 239 240 region was the one that most explained yield variability (Figure 2). The model including 241 the random effects described the overall spatial variation in soybean yield with an  $r^2$ value of 0.82 and a delta AIC of 671.20 when compared with the null model. 242

243

#### 244 Approximate location Figure 2

245

The fixed part of the final model that explained yield variability in maize included environmental potential, seed density, phosphorus and nitrogen fertilization, fungicide application, previous crop type, and the interaction between previous crop type and nitrogen fertilization (Figure 3). Nitrogen fertilization was the predictor that most explained yield, the response of yield to nitrogen being much higher in fields

previously sown with maize. For example, increasing nitrogen fertilization by 20 kg ha<sup>-1</sup> 251 led to a yield increase of 1,186 kg ha<sup>-1</sup> in fields where maize had previously been sown, 252 but the increase was only 205 kg ha<sup>-1</sup> in fields that had recently been converted from 253 natural areas. This stronger response to fertilization was accompanied by lower 254 expected yields when no fertilizer was applied (decrease of model intercept by 744 kg 255 ha<sup>-1</sup> in fields when maize was followed by maize). In contrast, a previous plantation of 256 a different grain or service crop had a positive effect on yield. Furthermore, yield 257 increased by 1,466 kg ha<sup>-1</sup> when environmental potential changed from 0 to 1 on the 258 scale. Increasing seed density from 3.8 seeds m<sup>-2</sup> to 9.8 seeds m<sup>-2</sup> raised yields from 259 7,929 to 10,015 kg ha<sup>-1</sup>, and fungicide application increased yields by 744 kg ha<sup>-1</sup>. The 260 random effect of the farm was the one that most explained yield variability in maize 261 (Figure 3). This model (including the random effects) described the spatial variation in 262 maize yield with an  $r^2$  value of 0.81 and a delta AIC value of 424.97 when compared 263 with the null model. 264

265

#### 266 Approximate location Figure 3

267

268 3.3. The effects of the percentage of natural habitat and edge density on models'

269 *residuals* 

The standardized residuals of neither the soybean model nor the maize model were significantly related to the percentage of natural habitat or edge density (Figure 4). For soybean, the model that included natural habitat, edge density and their interaction had 4.75 more AIC units than the null model. In the case of maize, the model that included natural habitat, edge density and their interaction had an AIC that was 2.32 units higher than the null model. 276

#### 277 Approximate location Figure 4

278

#### 279 **4. Discussion**

In this study, we gathered data from hundreds of fields in Argentina to assess the 280 relationships between the environment, land management, landscape structure and 281 282 yields. Due to the complex correlation structure between environmental, management and landscape variables, we implemented a three-step approach. We found that fields 283 with greater agricultural input negatively correlated with the percentage of natural 284 285 habitat. Land with higher environmental potential for grain production (i.e., field productivity) correlated negatively with the percentage of natural habitat in soybean. 286 287 Spatial variation in yield was well explained by environmental and management 288 variables for both soybean and maize fields. Neither percentage of natural habitat nor edge density could explain the variation in crop yield that was not described by 289 290 environmental and management variables in the datasets analyzed in this study.

291

292 4.1. Correlations between environmental, management and landscape variables

We found that fields with greater agricultural input negatively correlated with the percentage of natural habitat. Our results agree with previous findings which show that conventional agricultural intensification associated with higher levels of agricultural input and intensified crop sequences are currently co-occurring in landscapes with few areas of natural habitat in Argentina (Satorre and Andrade 2021). In our study, this process is especially evident in the negative correlations between the percentage of natural habitat and phosphorus fertilization in soybean and seed density in maize fields.

Across soybean fields, land with higher environmental potential for grain 300 301 production correlated negatively with the percentage of natural habitat. This pattern is expected because agriculture has been expanding in Argentina since the late 1980s, due 302 to modern technology (no-tillage techniques and genetically modified crops), climate 303 change (increase in warm-period rainfalls), and market conditions (global increase in 304 soybean demand) (Baldi and Paruelo 2008, Satorre 2005). Mixed cattle grazing-305 306 cropping systems were replaced by continuous cropping, and an increase in field sizes led to landscape homogenization and fewer natural habitats in most new productive 307 agricultural areas (Medan et al. 2011). 308

309 In some regions, a high percentage of the natural habitat measured in this study was possibly related to lowlands where cropping has lower yields due to soil and 310 311 weather limitations. Therefore, cattle raising is still the main activity in these areas (Cid 312 et al 2011) since agriculture is risky and limited to scattered productive areas. On the other hand, in the north of the country, natural habitat was represented by forested land 313 that was often only recently cleared for agriculture (Volante et al. 2016). In this case, 314 due to the few environmental limitations (i.e., soil and weather) for crop production, 315 large areas of natural habitat are more likely to be converted for agricultural use despite 316 317 the limitations imposed by the forest regulatory framework, which has proven to be insufficient for protecting these areas (Vallejos et al. 2021). 318

These different regions are important drivers of the negative correlations between the sowing date and landscape metrics of maize. This crop is now being sown late in new, less productive areas, where fields tend to be large (Satorre and Andrade 2020), since late sowing avoids summer drought and prevents yield variability at these northern latitudes (Satorre, et al. 2021).

324

#### 325 4.2. Prediction of soybean and maize yields

326 In the soybean yield model, the sowing date was the predictor with the highest impact on yield. Our results confirm the results of other studies performed in different regions 327 of Argentina, which found that the sowing date is a key variable in explaining yield 328 variability (Madias et al. 2021, Vitantonio-Mazzini et al. 2021, Di Mauro et al. 2018). 329 In fact, we found that late sowing led to an average decrease of 21 kg ha<sup>-1</sup> d<sup>-1</sup>, which 330 331 lies within the range of yield losses found for different regions of Argentina (Madias et al. 2021, Vitantonio-Mazzini et al. 2021) and other parts of the world (Rattaliano et al. 332 2017). This negative effect of late sowing is related to the environmental conditions, 333 334 which the crop experiences during critical periods of its cycle (Satorre et al. 2003). Soybean development is regulated by temperature and photoperiod, and their 335 interactions (Constable and Rose 1988). In consequence, the shorter days experienced 336 337 by later-sown soybean crops cause the plants to flower more rapidly (Lawn and Byth, 1973), shortening the vegetative period and positioning shifting the most important 338 yield determination periods in less favorable, less productive conditions (Satorre et al. 339 340 2003).

341 Our paper demonstrates that phosphorus fertilization, fungicide application and 342 previous crop are also key management variables that explain soybean yields, in agreement with previous findings in Argentina (Di Mauro et al. 2018). Environmental 343 variables were also addressed, with the field environmental potential variable in the 344 345 fixed part of the model and the region as a random effect. Removing region as a random effect from the model reduced the AIC value by 80.49 units, indicating the importance 346 347 of environmental and management conditions at a regional level on crop yield. In the maize model, nitrogen fertilization was the predictor with the highest 348 impact on yield; furthermore, the response to nitrogen addition was around 5 times 349

higher in fields where maize had previously been grown than in those that had 350 351 previously been natural areas. Varvel and Peterson (1990) also found this enhanced response of yield to nitrogen in maize-maize rotation compared with other rotations. 352 353 This effect arises due to differences in nutrient immobilization as a response to different previous crops (Kramberger et al. 2009). Maize as a previous crop has high 354 immobilization rates which means that added nitrogen, contributes a larger part of the 355 356 total available nitrogen to the crop. Although maize-maize rotations show strong yield responses to nitrogen fertilization, yields are the lowest under low levels of fertilization 357 (Figure 3). The previous crop can also increase soil nitrogen availability by 358 359 symbiotically incorporating nitrogen into the soil, or by mineralization of the soil nitrogen, depending on the previous species (Kramberger et al. 2009). 360 Seed density also had an important impact on maize yield. Seed density has a 361 362 direct association with stand density, which is known to be another important management decision that affects maize yield (Satorre et al. 2021, Gambin et al. 2016, 363 Hernandez et al., 2014). Although yield response to stand density usually follows a non-364 linear response (Sarlangue et al. 2007), we found a linear response, which suggests that 365 our field data included only the linear part of the yield response to this variable. 366 367 Environmental potential and fungicide application had a similar positive effect on maize as it did on soybean yield, in accordance with previous studies (Vitantonio-Mazzini et 368 369 al. 2020).

370

371 *4.3.* The effect of the percentage of natural habitat and edge density on models'

372 residuals

373 Ecological intensification proposes that the conservation of natural habitats which

374 provide ecosystem functions in agricultural landscapes will diminish external inputs and

favor more environmentally friendly agriculture (Garibaldi et al. 2019). We analyzed
whether landscape variables could explain the yield variability not explained by
environmental and crop management variables. In this study, we did not find evidence
for a substantial impact of landscape structure on yields, contrasting with earlier
findings of positive relationships in maize (Yang et al. 2019, Santana Sousa et al. 2011)
and soybean yields (Gonzalez et al. 2020, Monasterolos et al. 2015).

381 The positive effects of natural habitats on crop yield mainly arise from natural pest control and pollination as key ecosystem services (Alexandridis et al. 2021, 382 Garibaldi et al. 2021). In our study, we investigated commercial farms, which depend 383 384 on the intensive use of pesticides and transgenic Bt maize hybrids with resistance to Lepidoptera. This high, consistent input is reflected in transgenic Bt maize covering 385  $\sim$ 98% of total maize sown area (argenbio.com.ar) and the preventive use of insecticides 386 387 (Butinof et al. 2014). Such high investments in pest control are likely to mask the ecosystem service pest control provided by natural habitats in these field crops 388 (Costamagna et al. 2008). 389

Pollination is another important ecosystem service provided by natural habitat 390 which is relevant for soybean (Garibaldi et al. 2021), but does not affect the wind-391 392 pollinated maize plants. Soybean has an intermediate dependency on pollinators: reductions of 10% to slightly less than 40% have been found when comparing 393 experiments with and without animal pollinators (Klein et al. 2007, Chacoff et al. 2010). 394 A previous study found a mean increase of 21% in soybean yield when comparing open 395 396 versus exclosure treatments (Garibaldi et al. 2021). Although there is increasing scientific evidence to support soybean entomophile pollination, this has traditionally 397 been neglected by farmers in the study area (e.g., soybean farmers do not usually place 398 hives in or near their fields). In our study, we found that natural habitat, which provides 399

important nesting grounds and food sources during non-crop flowering periods for
pollinators, did not increase soybean yields. A possible explanation for this could be
that the general level of pollinator densities could have been sufficiently high, since in
our study 70% of soybean fields were surrounded by at least 20% of natural habitat. A
proportion of 20% of natural habitat is often suggested as an important threshold to
saturate the requirements of most crops for the provision of pollination and other
supporting ecosystem services (Garibaldi et al. 2020).

We would like to point out that our classification of natural habitat did not consider differences in the quality (i.e., plant diversity) of natural habitat, and that for simplicity we assessed only a single landscape size (i.e., the radius around fields). Both landscape scale (Le Provost et al. 2021) and habitat quality could be very important for pollination and ecosystem services provision in general (Kremen et al. 2007, Liere et al. 2015).

It should also be noted that yield responses to natural habitat are complex and 413 may have not only positive but also negative effects. For example, natural habitat can 414 compete for resources, which can have substantial negative effects on crop yields 415 (Zhang et al. 2007) and outbalance the positive impacts, such as the decrease in 416 417 numbers of herbicide-resistant weeds in the presence of natural habitats, which was observed in our study area (Alexandridis et al. 2022, Garibaldi et al. 2022). Resulting 418 net neutral responses might be a common result, especially when some key ecosystem 419 420 services are masked, as expected in our study.

421 Conservation of natural habitat in private agricultural landscapes is considered 422 for several different reasons. Legal frameworks can impose natural habitat conservation 423 in productive land (Garibaldi et al. 2020), but the recognition of its effects on crop yield 424 could provide a strong additional motivation for farmers to contribute to natural habitat

conservation. Through this study, we propose that in high-input dependent cropping 425 systems such as soybean and maize in Argentina, farm gross income (USD ha<sup>-1</sup>) 426 response to landscape composition and configuration should be studied in depth. Direct 427 costs are sensitive to ecosystem functions since much of the cost responds to dynamics 428 between natural habitat and crop performance (e.g., plagues/natural enemies, pesticide 429 use). The high agricultural input (e.g., intense use of pesticide) could mask the benefits 430 provided by natural habitat. However, the economic cost of these agricultural practices 431 could lead to lower farm gross income once production costs are considered (Zou et al. 432 2020). For example, if the frequency and amount of pesticide used diminishes with 433 434 greater quantities of natural habitat and/or edge density, the increase of natural habitat may help to substantially reduce farming costs while maintain yields resulting in more 435 environmentally friendly production systems. 436

437

#### 438 **5.** Conclusions

In this paper, we covered almost all major field-crop regions of Argentina, where 439 soybean and maize production represents one of the country's main sources of income. 440 Agricultural input was negatively correlated with natural habitat. Environmental and 441 442 management variables explained yield variability in both crops, and yield models considering these effects satisfactorily described the large spatial variation of yield in 443 the study regions. Neither percentage of natural habitat nor edge density could account 444 for the variation in crop yield that was not explained by environmental and management 445 variables in our datasets. As compensatory management probably masks to some extent 446 the beneficial effects of natural habitat in terms of yields, we recommend that future 447 studies focus on agricultural costs (USD ha<sup>-1</sup>). This could help to determine whether the 448

449 interaction between crop performance and the ecosystem functions provided by natural

450 habitat could have a beneficial influence on them.

451

## 452 Declaration of Competing Interest

453 The authors declare that they have no known competing financial interests or personal

454 relationships that could have appeared to influence the work reported in this paper.

455

- 456 Appendix A. Supporting information
- 457 See Table A.1.

458

#### 459 Acknowledgments

460 We thank Dr. Facundo Oddi for previous discussions, Dr. Juan Aguero for reading the

461 manuscript in previous versions and CREA research and development unit for valuable

462 feedback during the research process. Our thanks also go to two anonymous reviewers

that made valuable comments on the manuscript, and to the Consejo Nacional de

464 Investigaciones Científicas y Técnicas - Argentina, and the Asociación Argentina de

465 Consorcios Regionales de Experimentación Agrícola for co-funding a scholarship to the

466 first author.

## **References**

468	Alexandridis, N., Marion, G., Chaplin-Kramer, R., Dainese, M., Ekroos, J., Grab, H.,
469	Jonsson, M., Karp, D.S., Meyer, C., O'Rourke, M.E., Pontarp, M., Poveda, K.,
470	Seppelt, R., Smith, H.G., Walters, R.J., Clough, Y., Martin, E.A. 2022.
471	Archetype models upscale understanding of natural pest control response to
472	land-use change. Ecol. Appl.e2696. https://doi.org/10.1002/eap.2696
473	Alexandridis, N., Marion, G., Chaplin Kramer, R., Dainese, M., Ekroos, J., Grab, H.,
474	Jonsson, M., Karp, D.S., Meyer, C., O'Rourke, M.E., Pontarp, M., Poveda, K.,
475	Seppelt, R., Smith, H.G., Martin, E.A., Clough, Y. 2021. Models of natural pest
476	control: Towards predictions across agricultural landscapes. Biol.
477	Control. 163: 104761. 10.1016/j.biocontrol.2021.104761
478	Andrade, J.F., Satorre, E.H., 2015. Single and double crop systems in the Argentine
479	Pampas: environmental determinants of annual grain yield. Field Crop Res. 177:
480	137-147. https://doi.org/10.1016/j.fcr.2015.03.008.
481	Baldi, G., Paruelo, J.M. 2008. Land-use and land cover dynamics in South American
482	temperate grasslands. Ecol Soc 13:6
483	Bates, D., Mächler, M., Bolker, B., Walker, S. 2015. Fitting Linear Mixed-Effects
484	Models Using lme4. J Stat Softw. 67, 1–48.
485	https://doi.org/10.18637/jss.v067.i01.
486	Burnham, K.P., Anderson, D.R., Huyvaert, K.P., 2011. AIC model selection and
487	multimodel inference in behavioral ecology: Some background, observations,
488	and comparisons. Behav. Ecol. Sociobiol. 65, 23-35.
489	https://doi.org/10.1007/s00265-010-1029-6
490	Burnham, K.P., Anderson, D.R. 2002. Model selection and multimodel inference: a
491	practical information-theoretic approach. Springer, New York

492	Butinof. M., Fernández, R., Lantieri, M.J. Stimolo, M.I., Blanco, M., Machado, A.L.
493	Franchini,, G., Gieco, M., Portilla, M., Eandi, M., Sastre, A., Diaz, M.P. 2014.
494	Pesticides and agricultural work environments in Argentina. In: Larramendy M,
495	Soloneski S, editores. Pesticides - toxic aspects. InTech. pp. 105-34
496	Chacoff, N.P., Morales, C.L., Garibaldi, L.A., Ashworth, L., Aizen, M.A., 2010.
497	Pollinator dependence of Argentinean agriculture: current status and temporal
498	analysis. Am J Plant Sci Biotechnol. 3: 106–116.
499	Cid, M.S., Grecco, R.C.F., Oesterheld, M., Paruelo, J.M., Cibils, A.F., Brizuela, M.A.,
500	2011. Grass-fed beef production systems of Argentina's flooding pampas:
501	understanding ecosystem heterogeneity to improve livestock production.
502	Outlook Agric. 40: 181–189. https://doi.org/10.5367/oa.2011.0040
503	Constable, G.A., Rose, I.A.1988. Variability of Soybean Phenology Response to
504	Temperature, Daylength and Rate of Change in Daylength. Field Crops
505	Research. 18:57-69
506	Costamagna, A. C., D. A. Landis, and M. J. Brewer 2008. The role of natural enemy
507	guilds in Aphis glycines suppression. Biological Control 45: 368-379.
508	Díaz, S., Settele, J., Brondízio, E.S., Ngo, H.T., Agard, J., Arneth, A., Balvanera, P.
509	2019. Pervasive Human-Driven Decline of Life on Earth Points to the Need for
510	Transformative Change. Science 366 (1327): eaax3100.
511	https://doi.org/10.1126/science.aaw3100
512	Di Mauro, G., Cipriotti, P.A., Gallo, S., Rotundo, J.L., 2018. Environmental and
513	management variables explain soybean yield gap variability in Central
514	Argentina. Eur J Agron. 99: 186–194.

- Egli, L., Schröter, M., Scherber, C., Tscharntke, T., Seppelt. R. 2020. Crop asynchrony
  stabilizes food production. Nature 588 (7837): E7–12.
- 517 https://doi.org/10.1038/s41586-020-2965-6.
- 518 FAO. 2019. https://www.fao.org/faostat/en/#home. Accessed in November 2021.
- 519 Gambini, B.L., Coyos, T., Di Mauro, G., Borrás, L., Garibaldi, L.A. 2016 Exploring
- 520 genotype, management, and environmental variables influencing grain yield of
- 521 late-sown maize in central Argentina. Agri Sys. 146: 11-19.

522 <u>https://doi.org/10.1016/j.eja.2017.09.013</u>

- 523 Garibaldi, L., Goldenberg, Burian, A., Santibañez, F., Satorre, E.H., Martini, G.D.,
- Seppelt, R. 2022. Smaller agricultural fields, more edges, and natural habitats
  reduce herbicide-resistant weeds. [Manuscript submitted for publication]
- 526 Garibaldi, L., Oddi, F., Miguez, F., Bartomeus, I., Orr, M., Jobbagy, E., Kremen, C.,
- 527 Schulte, L., Hughes, A., Bagnato, C., Abramson, G., Brodgewater, P., Gomez
- 528 Carella, D., Díaz, S., Dicks, L., Ellis E., Goldenberg, M.G., Huaylla, C.,
- 529 Kuperman, M., Locke, H., Mehrabi, Z., Santibañez, F., Chao-Dong, Z. 2020.
- 530 Working landscapes need more than 20% native hábitat. Cons Lett. 14:
- 531 e12773. <u>https://doi.org/10.1111/conl.12773</u>
- 532 Garibaldi, L.A., Schulte, L.A., Nabaes Jodar, D.N., Gomez Carella, D.S. Kremen, C.
- 533 2021. Time to integrate pollinator science into soybean production. Trends Ecol.
  534 Evol. 36: 573–575
- 535 Garibaldi, L.A., Peréz-Mendez, N., Garratt, M.P.D. Gemmill-Herren, B., Miguez, F.E.
- 536 Dicks, L.V. 2019. Policies for Ecological Intensification of Crop Production.
- 537 Trends Ecol Evol. 34, 4

538	Garibaldi, L.A., Casas, C., Biganzoli, F. 2014. Datos jerárquicos en Ciencias
539	Ambientales: ejemplos prácticos y análisis de modelos jerárquicos en lenguaje
540	R. 1a ed Bariloche: el autor. pp.242. ISBN 978-987-33-6434-1
541	Gonzalez, E., Landis, D.A., Knapp, M., Valladares, G. 2020. Forest cover and proximity
542	decrease herbivory and increase crop yield via enhanced natural enemies in soybean
543	fields. J App Ecol. 57,11.
544	Hernandez, F., Amelong, A., Borras, L., 2014. Genotypic differences among
545	Argentinean maize hybrids in yield response to stand density. Agron. J. 106,
546	2316–2324
547	INTA 2019 Mapa Nacional de Cultivos campaña 2018/2019
548	Karp, D.S., Chaplin-kramer, R., Meehan, T.D., Martin, E.A., Declerck, F., Grab, H.,
549	Zubair, M., Avelino, J., Batáry, P., Baveco, J.M., Bianchi, F.J.J.A., Carrière, Y.,
550	Carvalheiro, L.G., Cayuela, L., Centrella, M., Cetkovi, A., Charles, D., Chabert,
551	A., Costamagna, A.C., Mora, A. De, Kraker, J. De, Desneux, N., Diehl, E.,
552	Diekötter, T., Dormann, C.F., Eckberg, J.O., Entling, M.H., Fiedler, D., Franck, P.,
553	Veen, F.J.F. Van, Frank, T., Gagic, V., Garratt, M.P.D., Getachew, A., Gonthier,
554	D.J., Goodell, P.B., Graziosi, I., Groves, R.L., Gurr, G.M., Hajian-forooshani, Z.,
555	Heimpel, G.E., Herrmann, J.D., Huseth, A.S., Inclán, D.J., Ingrao, A.J., Iv, P.,
556	Jacot, K., Johnson, G.A., Jones, L., Kaiser, M., Kaser, J.M., Keasar, T., Kim, T.N.,
557	Kishinevsky, M., Landis, D.A., Lavandero, B., Lavigne, C., Le, A., Lemessa, D.,
558	Madeira, F., Mader, V., Marie, A., Marini, L., Martinez, E., Molina-montenegro,
559	M.A., Neal, M.E.O., Opatovsky, I., Ortiz-martinez, S., Nash, M., Östman, Ö.,
560	Ouin, A., Pak, D., Paredes, D., Parsa, S., Parry, H., Peterson, J.A., Petit, S.,
561	Philpott, S.M., Plantegenest, M., Ple, M., Pluess, T., Pons, X., Potts, S.G., Pywell,
562	R.F., Takada, M.B., Taki, H., Tamburini, G., Thomson, L.J., Tricault, Y. 2018.

- 563 Crop pests and predators exhibit inconsistent responses to surrounding landscape
- 564 composition. PNAS. 115 (33) E7863-E7870
- 565 <u>https://doi.org/10.1073/pnas.1800042115</u>
- 566 Klein, A.M., Vaissière, B.E., Cane, J.H., Steffan-Dewenter, I., Cunnigham, S.A.,
- 567 Kremen, C., Tscharntke, T., 2007.Importance of pollinators in changing landscapes
- for world crops. Proceedings of the Royal Society, 274: 303–313.
- 569 Kramberger, B., Gselman, A., Janzekovic, M., Kaligaric, M., Bracko, B. 2009. Effects
- 570 of cover crops on soil mineral nitrogen and on yield and nitrogen content of maize.
- 571 Eur. J. Agron. 31, 103-109.
- 572 Kremen, C., Williams, N.M., Aizen, M.A., Gemmill-Herren, B., LeBuhn, G., Minckley,
- 573 R., Packer, L., Potts, S.G., Roulston, T., Steffan-Dewenter, I., Vázquez, D.P.,
- 574 Winfree, R., Adams, L., Crone, E.E., Greenleaf, S.S., Keitt, T.H., Klein, A.-M.,
- 575 Regetz, J., Ricketts, T.H. 2007. Pollination and other ecosystem services
- 576 produced by mobile organisms: a conceptual framework for the effects of land-
- 577 use change. Ecol Lett. 10: 299-314. <u>https://doi.org/10.1111/j.1461-</u>
- **578** <u>0248.2007.01018.x</u>
- Lawn, R.J., Byth, D.E., 1973. Response of soya beans to planting date in South-eastern
- 580 Queensland. I. Influence of photoperiod and temperature on phasic
- developmental patterns. Aust J Agric Res. 24: 67-80.
- 582 Leggett, M., Diaz-Zorita, M. Koivunen, M., Bowman, R. Pesek, R. Stevenson,
- 583 C., Leister, T. 2017. Soybean response to inoculation with Bradyrhizobium
- japonicum in the United States and Argentina. Agron J. 109, 1031-1038.
- 585 <u>https://doi.org/10.2134/agronj2016.04.0214</u>
- Le Provost, G., Thiele, J., Westphal, C., Penone, C., Allan, E., Neyret, M., Van Der
- 587 Plas, F., Ayasse, M., Bardgett, R.D., Birkhofer, K., Boch, S., Bonkowski, M.,

588	Buscot, F., Feldhaar, H., Gaulton, R., Goldmann, K., Gossner, MM., Klaus, VH.,
589	Kleinebecker, T., Krauss, J., Renner, S., Scherreiks, P., Sikorski, J., Baulechner,
590	D., Blüthgen, N., Bolliger, R., Börschig, C., Busch, V., Chisté, M., Fiore-Donno,
591	A.M., Fischer, M., Arndt, H., Hoelzel, N., John, K., Jung, K., Lange, M.,
592	Marzini, C., Overmann, J., Paŝalić, E., Perović, D.J., Prati, D., Schäfer, D.,
593	Schöning, I., Schrumpf, M., Sonnemann, I., Steffan-Dewenter, I., Tschapka, M.,
594	Türke, M., Vogt, J., Wehner, K., Weiner, C., Weisser, W., Wells, K., Werner
595	,M., Wolters, V., Wubet, T., Wurst, S., Zaitsev, A.S., Manning, P. 2021.
596	Contrasting responses of above-and belowground diversity to multiple
597	components of land-use intensity. Nat Commun 12:1-13
598	Madias, A., Di Mauro, G., Vitantonio-Mazzini, L.N., Gambin, B.L., Borrás, L.
599	Environment quality, sowing date, and genotype determine soybean yields in the
600	Argentinean Gran Chaco. Eur J Agron. 2021, 123:126217.
601	https://doi.org/10.1016/j.eja.2020.126217
602	Martin, E.A., Dainese, M., Clough, Y., Báldi, A., Bommarco, R., Gagic, V., Garratt,
603	M.P., Holzschuh, A., Kleijn, D., Kovács-Hostyánszki, A., Marini, L., Potts,
604	S.G., Smith, H.G., Al Hassan, D., Albrecht, M., Andersson, G.K., Asís, J.D.,
605	Aviron, S., Balzan, M.V., Baños-Picón, L., Bartomeus, I., Batáry, P., Burel, F.,
606	Caballero-López, B., Concepción, E.D., Coudrain, V., Dänhardt, J., Diaz, M.,
607	Diekötter, T., Dormann, C.F., Duflot, R., Entling, M.H., Farwig, N., Fischer, C.,
608	Frank, T., Garibaldi, L.A., Hermann, J., Herzog, F., Inclán, D., Jacot, K., Jauker,
609	F., Jeanneret, P., Kaiser, M., Krauss, J., Le Féon, V., Marshall, J., Moonen, A
610	C., Moreno, G., Riedinger, V., Rundlöf, M., Rusch, A., Scheper, J., Schneider,
611	G., Schüepp, C., Stutz, S., Sutter, L., Tamburini, G., Thies, C., Tormos, J.,
612	Tscharntke, T., Tschumi, M., Uzman, D., Wagner, C., Zubair-Anjum, M.,

613	Steffan-Dewenter, I. 2019. The interplay of landscape composition and
614	configuration: new pathways to manage functional biodiversity and
615	agroecosystem services across Europe. Ecol Lett, 22: 1083-
616	1094. <u>https://doi.org/10.1111/ele.13265</u>
617	Medan, D., Torretta, J. P., Hodara, K., Fuente, E. B., Montaldo, N. H. 2011. Effects of
618	agriculture expansion and intensification on the vertebrate and invertebrate
619	diversity in the Pampas of Argentina Biodivers. Conserv. 20: 3077–100.
620	Biodivers Conserv. 20:3077–3100
621	Monasterolo, M., Musicante, M. L., Valladares, G. R., Salvo, A. 2015. Soybean crops
622	may benefit from forest pollinators. Agri Ecosys Envio. 202, 217–222.
623	https://doi.org/10.1016/j.agee.2015.01.012
624	Nosetto, M.D., Jobbágy, E.G., Brizuela, A.B., Jackson, R.B. 2012. The hydrologic
625	consequences of land cover change in central Argentina. Agr Ecosys Environ.
626	154, 2-11. https://doi.org//10.1016/j.agee.2011.01.008
627	Oddi, FJ, Miguez, FE, Ghermandi, L, Bianchi, LO, Garibaldi, LA. 2019. A nonlinear
628	mixed-effects modeling approach for ecological data: Using temporal dynamics
629	of vegetation moisture as an example. Ecol Evol.
630	9: 10225-10240. <u>https://doi.org/10.1002/ece3.5543</u>
631	Rattalino Edreira, J.I., Mourtzinis, S., Conley, S.P., Roth, A.C., Ciampitti, I.A., Licht,
632	M.A., Kandel, H., Kyveryga, P.M., Lindsey, L.E., Mueller, D.S., Naeve, S.L.,
633	Nafziger, E., Specht, J.E., Stanley, J., Staton, M.J., Grassini, P., 2017. Assessing
634	causes of yield gaps in agricultural areas with diversity in climate and soils. Agr.
635	Forest Meteorol. 247, 170–180.

636	Ramankutty, N., Evan, A. T., Monfreda, C., Foley, J. A. 2008. Farming the planet: 1.
637	Geographic distribution of global agricultural lands in the year 2000. Global
638	Biogeochem. Cycles. 22, GB1003. https://doi.org/10.1029/2007GB002952
639	R Core Team. 2020. R: a language and environment for statistical computing. R
640	Foundation for Statistical Computing, Vienna, Austria. Available from
641	http://www.r-project.org.
642	Ricketts, T.H., Regetz, J., Steffan-Dewenter, I., Cunningham, S.A., Kremen, C.,
643	Bogdanski, A., Gemmill-Herren, B., Greenleaf, S.S., Klein, A.M., Mayfield,
644	M.M., Morandin, L.A., Ochieng', A. and Viana, B.F. 2008. Landscape effects
645	on crop pollination services: are there general patterns? Ecol Lett. 11, 499-515.
646	https://doi.org/10.1111/j.1461-0248.2008.01157.x
647	Santana Sousa, E.H., Costa Batista Matos, M., Almeida, R.S., Vieira Teodoro, A. 2011.
648	Forest Fragments' Contribution to the Natural Biological Control of Spodoptera
649	frugiperda Smith (Lepidoptera: Noctuidae) in Maize. Braz arch biol technol. 54,
650	755-760. https://doi.org/10.1590/S1516-89132011000400015
651	Sarlangue, T., F.H. Andrade, P.A. Calviño, Purcell, L.C. 2007. Why do maize hybrids
652	respond differently to variations in plant density? Agron. J. 99:984–991.
653	https://doi.org/10.2134/agronj2006.0205
654	Satorre, E.H., Andrade, F.H. 2021. Cambios productivos y tecnológicos de la
655	agricultura extensiva argentina en los últimos quince años. Ciencia Hoy. 29
656	(173): 19-27
657	Satorre, E.H., Tronfi, E., Costamagna, C., Iturrez, T., Arinci, A. 2021. Factores
658	ambientales y de manejo determinantes del rendimiento de maíz en la Región
659	Centro-Norte de la provincia de Córdoba. Revista Agronomía & Ambiente, Rev.
660	Facultad de Agronomía UBA, 41 (2): 100-113

- 661 Satorre, E.H. 2005. Cambios tecnológicos en la agricultura argentina actual. Ciencia
  662 Hoy Vol. Temático (2): 17-20.
- Satorre, E.H. 2003. El Libro de la Soja. Servicios y Marketing Agropecuario, Buenos
  Aires, Argentina, pp. 264.
- 665 Seppelt, R., Arndt, C., Beckmann, M., Martin, E.A., Hertel, T.W., 2020. Deciphering
- the Biodiversity Production Mutualism in the Global Food Security Debate.
- 667 Trends Ecol. Evol. 35, 1011–1020. https://doi.org/10.1016/j.tree.2020.06.012
- 668 Seppelt, R., Manceur, A.M., Liu, J. Fenichel, E.P., Klotz. S. 2014. Synchronized peak-
- rate years of global resources use". Ecol Soc. 19 (4): art50.
- 670 https://doi.org/10.5751/ES-07039-190450.
- 671 Tscharntke, T., Karp, D.S., Chaplin-kramer, R., Batáry, P., Declerck, F., Gratton, C.,
- Hunt, L., Ives, A., Jonsson, M., Larsen, A., Martin, E.A., Martínez-salinas, A.,
- 673 Meehan, T.D., Rourke, M.O., Poveda, K., Rosenheim, J.A., Rusch, A., Schellhorn,
- N., Wanger, T.C., Wratten, S., Zhang, W., 2016. When natural habitat fails to
- enhance biological pest control Five hypotheses. BIOC.
- 676 https://doi.org/10.1016/j.biocon.2016.10.001
- Tscharntke, T., Klnein, A.M., Kruess, A., Steffan-Dewenter, I., Thies, C..2005.
- 678 Landscape perspectives on agricultural intensification and biodiversity –
- ecosystem service management. Ecol Lett. 8: 857–874.
- 680 https://doi.org/10.1111/j.1461-0248.2005.00782.x
- Vallejos, M.N., Camba Sans, G.H., Aguiar, S., Mastrángelo, M.E., Paruelo, J.M. 2021.
- The law is spider's web: An assessment of illegal deforestation in the Argentine
- 683 Dry Chaco ten years after the enactment of the "Forest Law". Env Develop.
- 684 38:100611. https://doi.org/10.1016/j.envdev.2021.100611.

686	E., Borr'as, L. 2021. Sowing date, genotype choice, and water environment
687	control soybean yields in central Argentina. Crop Sci. 212: 82-94.
688	https://doi.org/10.1002/csc2.20315.
689	Vitantonio-Mazzini, L.N., Borrás, L., Garibaldi, L.A, Pérez, D.H., Gallo, S., Gambin,
690	B.L. 2020. Management options for reducing maize yield gaps in contrasting
691	sowing dates. Field Crops Res. 251: 107779.
692	https://doi.org/10.1016/j.fcr.2020.107779
693	Volante, J.N., Mosciaro, M.J., Gavier-Pizarro, G.I., Paruelo, J.M. 2016. Agricultural
694	expansion in the Semiarid Chaco: Poorly selectivecontagious advance. Land Use
695	Pol. 55:154-165. http://dx.doi.org/10.1016/j.landusepol.2016.03.025 0264-8377/
696	Yang, L., Xu, L., Liu, B., Zhang, Q., Pan, Y., Li, Q., Li, H., Lu, Y. 2019. Non-crop
697	habitats promote the abundance of predatory ladybeetles in maize fields in the
698	agricultural landscape of northern China. Agric Ecosyst Environ. 277, 44-52.
699	https://doi.org/10.1016/j.agee.2019.03.008.
700	Zhang, W., Ricketts, T.H., Kremen, C., Carney, K., Swinton, S.M. 2007. Ecosystem
701	services and dis-services to agriculture. Ecol Econ. 64, 253-260
702	Zou, Y., de Kraker, J., Bianchi, F.J.J.A., Xiao, H., Huang, J., Deng, X., Hou, L., van der
703	Werf, W. 2020. Do diverse landscapes provide for effective natural pest control
704	in subtropical rice? J. Appl. Ecol. 57:170-180. <u>https://doi.org/10.1111/1365-</u>
705	<u>2664.13520</u>
706	Zuur, A.F., Ieno, E.N., Walker, N., Saveliev, A.A., Smith, G. M. 2009. Mixed effects
707	models and extensions in ecology with R. New York: Springer.
708	

Vitantonio-Mazzini, L., Gomez, D., Di Mauro, G., Gambín, B.L., Iglesias, R., Jobbagy,