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#### 20 Abstract

21 Land-use intensification and climate change are main threats to the abundance and 22 diversity of soil macrofauna. However, little is known about their biomass in response to these concurrent drivers. Here, we investigated the biomass responses of soil 23 24 macrofauna along a land-use intensity gradient of five land-use regimes (i.e., from 25 extensively-used grassland to conventional cropland) under two climate scenarios 26 (ambient *vs.* future). We found that land-use intensification (but not climate change) 27 significantly reduced soil macrofauna biomass at the community rather than individual 28 level. Further, the community structure of soil macrofauna based on total biomass data 29 varied with land-use type (i.e., grasslands vs. croplands). Collectively, our findings suggest that land-use intensification can negatively shift the community biomass 30 31 patterns of soil macrofauna consistently under both ambient and future climates in 32 agroecosystems.

33

34 Keywords: Agroecosystem; Body mass; Climate change; Community biomass;
35 Intensive land use; Soil invertebrate

36

#### 37 Introduction

Soil macrofauna account for the majority of soil fauna biomass across many terrestrial
ecosystems, and contribute substantially to ecosystem functions (Gongalsky, 2021). For
example, 'litter transformers' (e.g., Julida, Isopoda) fragment coarse organic debris and
promote microbial decomposition (David, 2014). 'Ecosystem engineers' (e.g.,

42 earthworms) incorporate large amounts of organic fragments into the mineral soil
43 horizon to facilitate nutrient cycling and primary production (Blouin et al., 2013) and
44 form the environment for other species in the soil (Eisenhauer, 2010).

Larger soil fauna may be more affected by environmental changes than smaller fauna represented by *r*-strategists (e.g., Briones, 2014; Tsiafouli et al., 2015; Phillips et al., 2019). However, soil macrofauna attract much less attention, compared with other groups of soil organisms (e.g., microbes, microfauna, and mesofauna) (Basturk et al., 2021). Additionally, previous macrofauna studies focused more on the responses of their abundance and diversity to global changes (Franco et al., 2016; Phillips et al., 2019; Yin et al., 2019), but less on their biomass responses.

52 Body mass, as one of the most important physio-morphological traits, largely 53 determines the food intake and metabolic rate of invertebrates (Eklöf et al., 2017). 54 Moreover, the resource utilization and allocation capacity of a community is often reflected by the total biomass of co-occurring species within that community (Post and 55 56 Pedersen, 2008). Changes in biomass pattern of macrofauna community may directly 57 influence the energy fluxes in soil food webs, therefore biomass may be even more 58 important to ecosystem functioning than their abundance and diversity (Basturk et al., 59 2021).

As soil fauna are thermosensitive, their biomass patterns could be shifted by climate change (Xu et al., 2017). For instance, Vestergård et al. (2015) found that the total biomass of mesofauna communities was reduced by warming, especially when combined with drought. Warming speeds up individual metabolism (Scheffers et al., 64 2016), resulting body size/mass reduction of soil mesofauna (Yin et al., 2020).
65 Additionally, land-use intensification have been reported to cause abundance and
66 diversity loss of soil fauna (especially for macrofauna, Postma-Blaauw et al., 2012),
67 and simultaneously reduce the complexity of their community structure (Decaëns et al.
1994; Tsiafouli et al. 2015). But less is known about how these two global change
69 drivers individually and collectively influence soil macrofauna biomass, at the both
70 individual and community levels.

Here, we address this question in the framework of the 'Global Change 71 72 Experimental Facility' (GCEF) in Germany. Specifically, we performed a two-year 73 study in a full-factorial combination of climate (ambient vs. future, i.e., increased 74 temperature and altered precipitations) and land use (from low to high land-use intensity: extensively-used meadow  $\rightarrow$  extensively-used pasture  $\rightarrow$  intensively-used meadow  $\rightarrow$ 75 organic cropland  $\rightarrow$  conventional cropland). We tested two hypotheses: (1) climate 76 77 change and land-use intensification will reduce soil macrofauna biomass at the both 78 individual and community levels; (2) interactions between climate and land use will be 79 significant on soil macrofauna biomass. More precisely, land-use intensification will 80 exacerbate the detrimental effects of climate change on soil macrofauna biomass; whilst 81 extensive land use will alleviate these climate change effects.

82

### 83 Materials and Methods

Our study was conducted at the 'Global Change Experimental Facility' (GCEF) in Bad Lauchstädt, Germany (51° 23' 30" N, 11° 52' 49" E, 116 m a.s.l.). This study site is characterized by a subcontinental climate with a mean annual temperature and precipitation of 8.9°C and 498 mm, respectively. The soil type is Haplic Chernozem, the contents of total carbon and nitrogen within the upper soil (15 cm) varied between 1.71–2.09% and 0.15–0.18%, respectively (Schädler et al., 2019).

90 The GCEF was established in late 2012, and consists total 50 plots arranged into 91 10 blocks (Fig. S1A). Half of these plots (25/50) are under five ambient climate blocks, 92 and the other half (25/50) are subjected to five future climate blocks (i.e., increased 93 temperature by ~0.55 °C, and changed precipitation patterns with ~20% reduction in 94 summer and ~10% increment in spring/autumn), which is a projection of the climate of 95 Central Germany for the years of 2070-2100 based on several models, i.e., COSMO-96 CLM (Rockel et al., 2008), REMO (Jacob and Podzun, 1997), and RCAO (Döscher et 97 al., 2002). Using a split-plot design, each five land-use regimes are randomly arranged 98 into either ambient or future climate block (Fig. S1B). Descriptions of these five land-99 use regimes are provided in Figure S2.

We sampled totally 200 soil cores (Ø 16 cm, 10 cm depth) during a two-year study period for autumn season: in 26.10.2015 and 25.10.2016. Specifically, two soil cores per plot were taken and gathered into one sample at each sampling period to extract soil macrofauna using a Kempson extraction method (Kempson et al., 1963). Using a digital microscope (VHX-600, Keyence Corp., Osaka, Japan), the extracted macrofauna (body length > 3 mm) were determined to some specific taxa, i.e., Araneae, Chilopoda,

106 Coleoptera, Diplura, Diptera, Formicidae, Gastropoda, Isopoda, Julida, Lumbricidae, 107 Psocoptera, Symphyla, and Thysanoptera. For all taxa, we divided the number of 108 individuals in each sample by the surface area  $(0.04 \text{ m}^2)$  of the two soil cores to calculate 109 density, and measured the body size (length,  $\mu$ m) of each individual. The body mass 110 (M,  $\mu$ g) of each taxon was calculated by a body size assessment method based on taxon-111 specific formulas (Table S1). For each plot, the total biomass and mean body mass of 112 macrofauna communities were represented by mean  $\pm$  se.

113 Linear mixed models (LMMs) were conducted using the 'lme4' package (Bates et 114 al., 2011) to assess the effects of climate (as mainplot factor level; ambient vs. future), 115 land use (as subplot factor level; extensively-used meadow vs. extensively-used pasture 116 vs. intensively-used meadow vs. organic farming vs. conventional farming), date (2015 117 autumn and 2016 autumn), and their interactions on the total biomass and mean body 118 mass of soil macrofauna. Land use was nested into mainplot, and mainplot was served 119 as a random effect. Data was tested for assumptions of normality and homogeneity of 120 variances using the Shapiro-Wilk and Levene's tests. The post-hoc comparisons of means among groups were applied using the 'emmeans' package (Lenth et al., 2019). 121 122 To visualize how climate, land use, and their interaction influence community structure 123 of soil macrofauna (based on total biomass data), permutational multivariate analysis 124 of variance (PERMANOVA) and pairwise comparison were conducted using 'adonis' 125 and 'pairwise.adonis' functions with the 'bray' method and 999 permutations fitted to 126 non-metric multidimensional scaling (NMDS) ordination in R 'vegan' package (Martinez Arbizu, 2020; Oksanen et al., 2013). All statistical analyses were conducted 127 128 using R version 4.0.3 (R Core Team, 2021).

129

### 130 **Results and Discussion**

In this study, climate change did not significantly affect the biomass patterns of soil macrofauna (Table S2). The same experiment showed that climate change significantly reduced the total biomass of soil mesofauna (Yin et al., 2020). However, the total biomass of soil macrofauna remained unchanged under future climate (Fig. 1A; Table S2), possibly because soil macrofauna are more resistant to the variations in climate than soil mesofauna (Bokhorst et al., 2012). Therefore, the responses of macrofauna biomass to climate change may take longer to manifest.

Additionally, we found no evidence of body mass reduction in soil macrofauna under future climate (Fig. 1 B; Table S2), despite climate change has been reported to shrink the body size and mass of many organisms (Gardner et al., 2011). However, soil fauna body mass has been reported to decrease from mesic to arid conditions; the aridity is an environmental filter especially for large-bodied fauna (Andriuzzi et al., 2020). We sampled in two consecutive autumn years (2015/2016), the wetter autumn conditions in our future climate plots may mitigate the detrimental warming effects.

Detrimental effects of land-use intensification are prevalent in soil biota, especially in large-bodied fauna (Postma-Blaauw et al., 2010). We found that land-use intensification significantly reduced the total biomass but not the mean body mass of soil macrofauna across all plots and dates (Fig. 1A-B; Table S2). These findings indicate that agricultural intensification threatens macrofauna biomass at the community but not individual level. 151 Further, the significant effects of land use on macrofauna total biomass were 152 taxon-specific (Table S3). For example, the total biomass of Coleoptera and Chilopoda 153 was significantly higher in extensively-used pasture than that in croplands (Fig. 1C-D); whereas the total biomass of Julida in intensively-used meadow and Diplura in 154 155 extensively-used meadow was significantly higher than that in croplands, respectively 156 (Fig. 1E-F). In general, grazing (along with livestock trampling) may cause soil 157 compaction and reduction in porosity, as well as the formation of local anoxic conditions (Schrama et al., 2013), and these effects are expected to be detrimental to 158 159 soil macrofauna. In turn, livestock dung is an attractive food resource for Coleoptera 160 (Galante et al., 1995), and low to moderate grazing intensity may facilitate their total biomass (Tonelli et al., 2018). Further, agricultural extensification can increase the 161 162 complexity of soil food webs, and the mean body mass of soil fauna (Postma-Blaauw 163 et al., 2010), but may also cause the increased predation (Flohre et al., 2011). As the top predators of soil food webs, Chilopoda are highly mobile and prefer to dwell at the 164 soil surface when the soil porosity is low and when hypoxic conditions are prevalent. 165 166 In our pastures, therefore, the increase of total biomass of Coleoptera may accordingly increase the total biomass of predators (e.g., Chilopoda) through a bottom-up trophic 167 168 cascade (Wu et al., 2011).

In the context of land-use intensification, the reduced total biomass of soil fauna
might be due to their decreased density (Yin et al., 2020). Indeed, a generally positive
correlation was detected between total biomass of soil macrofauna and their density;
Strikingly, this pattern was significant only in croplands, but not in grasslands (Fig. 2A).

173 By contrast, there was no significant correlation between their mean body mass and 174 density (Fig. 2B). Furthermore, the community structure of soil macrofauna (based on 175 total biomass data) was significantly affected by land-use (but not climate) treatment (Fig. 2C; Table S4A), and these significant effects were driven by the two land-use 176 types (i.e., grasslands vs. croplands) (Table S4B). As partially supported by other 177 178 studies (Guan et al., 2021; Yin et al., 2019), land-use type conversion is the foremost 179 driving force of shifts in the community composition or food web structure of soil fauna. Collectively, our results partially support our 1<sup>st</sup> hypothesis, showing that land-use 180 intensification shifted the biomass patterns of soil macrofauna with a significant 181 182 reduction in total biomass; however, we did not find any evidence of significant interactions between climate and land use on soil macrofauna biomass to support our 183 2<sup>nd</sup> hypothesis, suggesting that the effects of land use were consistent under the both 184 185 ambient and future climate scenarios.

186

#### 187 Conclusion

Land-use intensification was a more severe immediate threat than climate change to soil macrofauna biomass. Our findings provide strong experimental evidence that landuse change from cropland to grassland could shift the community structure of soil macrofauna from a higher to a lower total biomass community. Despite no significant effects of climate change on soil macrofauna biomass were found approximately three years after the establishment of the GCEF experimental platform, long-term observation is needed to determine if the current response pattern will be consistent or 195 change over time. Furthermore, these detrimental effects of land-use intensification on 196 total biomass of soil macrofauna may have far-reaching consequences for the provision 197 of ecosystem functions and energy fluxes, e.g., litter decomposition, nutrient turnover, 198 and productivity of agricultural ecosystems (Taylor et al., 2010). In order to better 199 understand and predict the potential effects of soil fauna biomass loss and community 200 changes on related ecosystem functions in this changing world, we call for future 201 research to explore these functional links and elucidate the mechanisms.

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339

#### 340 Conflict of Interest

341 The authors declare that they have no known competing financial interests or personal

342 relationships that could have appeared to influence the work reported in this paper.

343

## 344 Authors Contributions

345 MS is part of the GCEF steering committee that developed the experimental platform;

- 346 MS and NE conceived this study; RY and MS conducted the field and lab work; R.Y.
- analyzed the data and wrote the first draft; P.K. offered advices on writing structure.
- 348 All authors greatly contributed to revising of this paper.

#### 350 **Figure legends**

351 Fig. 1 Interaction effects of climate and land use on total biomass (A) and mean body 352 mass (**B**) of soil macrofauna, as well as total biomass of Coleoptera (**C**), Chilopoda (**D**), Julida (E), and Diplura (F). Boxplots show the mean (the solid dot), the median (the 353 horizontal line), the first and third quartile (the rectangle), and the  $1.5 \times$  interquartile 354 355 range (the whiskers). Jitter points around the boxplot represent sample numbers for 356 each climate (ambient climate in blue, future climate in red). The histogram on the upper right (A) shows the effects of land use on total biomass of soil macrofauna (mean 357 358  $\pm$  se). Different lowercase letters represent significant differences among land-use 359 regimes at P < 0.05 based on the post-hoc Tukey's HSD tests. Abbreviation: EM (in lawn green): extensively-used meadow, EP (in olive drab): extensively-used pasture, 360 361 IM (in lime green): intensively-used meadow, OF (in yellow): organic farming, and CF 362 (in orange): conventional farming.

364 Fig. 2 Linear correlations between density and total biomass of soil macrofauna (A), as well as between density and mean body mass of soil macrofauna (B). Shown are the 365 366 fitted regression lines with 95% confidence intervals, and dots for all individual samples 367 for extensively-used meadow (EM, in lawn green), extensively-used pasture (EP, in olive drab), intensively-used meadow (IM, in lime green), organic farming (OF, in 368 yellow), and conventional farming (CF, in orange). When P < 0.05,  $R = 0.2 \sim 0.4$  and 369  $0.7 \sim 0.9$  represents low and high correlation, respectively. Effects of climate and land 370 use on community structure of soil macrofauna based on total biomass data (C). Non-371 16

372	metric multidimensional scaling (NMDS) plot based on Bray-Curtis distances
373	visualizes community differences between climate scenarios represented by dots
374	(ambient), and triangles (future), and among land-use regimes (EM, with lawn green
375	symbols; EP, with olive drab symbols; IM, with lime green symbols; OF, with yellow
376	symbols; CF, with orange symbols). Ellipses represent 95% confidence estimates for
377	mean NMDS scores of five land-use regimes.

Fig. 1







B

Total biomass (mg/m<sup>2</sup>)

F

 $p/m^2$ 





NMDS1

#### **Supplementary materials**



**Fig. S1** Global Change Experimental Facility (GCEF). (**A**) Aerial image of GCEF setup: total 50 plots (24 m × 16 m = 384 m<sup>2</sup> per plot) are arranged into 10 blocks (i.e., 5 plots/block), of which 5 blocks for ambient climate, the other 5 blocks for future climate. Picture: Tricklabor/Service Drone, copyrights: UFZ. (**B**) Climate and land use treatment settings. Climate treatments as main-plot factor have two levels: ambient climate (the control) *vs*. future climate (i.e., the increased temperature by ~0.55 °C, and the altered precipitation by ~20% reduction in summer and ~10% increment in spring/autumn). Picture: Andrè Künzelmann, copyrights: UFZ. Land-use treatments as

sub-plot factor have five land-use regimes, i.e., EM = extensive-used meadow; EP = extensive-used pasture; IM = intensive-used meadow; OF = organic farming; CF = conventional farming. Each these five land-use regimes are randomly nested into a block.

**Fig. S2** Detailed description of five land-use regimes at the Global Change Experimental Facility (GCEF).



Taxa	Formula	$\mathbf{B}_0$	<b>B</b> <sub>1</sub>	Reference
Araneae	$\log M = B_0 + B_1 \times \log L$	-2.42	1.84	Mercer et al., 2001
Chilopoda	$M=e^{\;(B_0+B_1\timesln(L))}$	-3.29	2.10	Ganihar, 1997
Coleoptera	$M=e^{\;(B_{0}+B_{1}\timesln(L))}$	-3.25	2.49	Sample et al., 1993
Diplura	$M=e^{\;(B_{0}+B_{1}\timesln(L))}$	-3.43	2.59	Ganihar, 1997
Diptera	$M=e^{\;(B_{0}+B_{1}\timesln(L))}$	-3.14	2.59	Ganihar, 1997
Formicidae	$M=e^{(B_0+B_1\times ln(L))}$	-3.14	2.34	Ganihar, 1997
Gastropoda	$M=e^{\;(B_0+B_1\timesln(L))}$	-2.75	1.59	Wardhaugh, 2013
Isopoda	$M=e^{\;(B_{0}+B_{1}\timesln(L))}$	-4.81	3.44	Wardhaugh, 2013
Julida	$M=e^{\;(B_{0}+B_{1}\timesln(L))}$	-4.59	2.54	Gowing and Recher, 1984
Lumbricidae	$\logM=B_0+B_1\!\times\log L$	0.93	1.09	Mercer et al., 2001
Psocoptera	$M=e^{\;(B_{0}+B_{1}\timesln(L))}$	-3.07	2.30	Ganihar, 1997
Symphyla	$M=e^{\;(B_0+B_1\timesln(L))}$	-3.07	2.30	Ganihar, 1997
Thysanoptera	$M=e^{\;(B_0+B_1\timesln(L))}$	-3.07	2.30	Ganihar, 1997

**Table S1** Taxon-specific formulas for the estimation of body mass  $(M, \mu g)$  for different soil macrofauna taxa based on the body size  $(L, \mu m)$  spectrum method.

**Table S2** Results (*F*-values) of linear mixed effects models (Type III ANOVA withSatterthwaite's method) testing the effects of climate, land use, date, and theirinteractions on total biomass and mean body mass of soil macrofauna community. *F*-value with \*\* (P < 0.01) representing significant effects is indicated in bold font.

Traatmanta	T	otal biomas	SS	N	Mean body mass			
Treatments	$Df_{(num:den)}$	<i>F</i> -value	<i>P</i> -value	$Df_{(num:den)}$	<i>F</i> -value	<i>P</i> -value		
Climate (C)	1:80	1.09	0.30	1:62	0.37	0.54		
Land use (L)	4:80	3.60	< 0.01**	4:62	0.46	0.76		
Date (D)	1:80	3.24	0.07	1:62	0.05	0.82		
$\mathbf{C} \times \mathbf{L}$	4:80	1.04	0.39	4:62	1.30	028		
$\mathbf{C} \times \mathbf{D}$	1:80	0.03	0.87	1:62	2.34	0.13		
$L \times D$	4:80	0.86	0.49	4:62	1.18	0.33		
$C \times L \times D$	4:80	0.44	0.78	4:62	0.75	0.56		

**Table S3** Results (*F*-values) of linear mixed effects models (Type III ANOVA with Satterthwaite's method) testing the effects of climate, land use, date, and their interactions on total biomass of Lumbricidae, Araneae, Coleoptera, Chilopoda, Julida, Diptera, Isopoda, Formicidae, Gastropoda, Diplura, Symphyla, Psocoptera, and Thysanoptera across two samplings. *F*-values with \*\*\* (P < 0.001), \*\* (P < 0.01), \* (P < 0.05) representing significant effects are indicated in bold font.

Treatments	Df	Araneae	Chilopoda	Coleoptera	Diplura	Diptera	Formicidae	Gastropoda
Climate (C)	1,80	1.16	1.26	0.003	0.00	0.48	1.78	1.69
Land use (L)	4,80	1.09	4.54**	6.41***	4.29**	1.18	1.01	1.33
Date (D)	1,80	1.14	21.45***	0.84	8.19**	0.33	1.85	3.70
$C \times L$	4,80	1.01	1.49	1.53	0.78	0.36	0.85	0.74
$\mathbf{C} \times \mathbf{D}$	1,80	0.33	3.04	0.47	0.83	0.49	1.79	1.69
$L \times D$	4,80	1.23	3.95	0.69	1.54	0.91	1.04	1.33
$C \times L \times D$	4,80	1.36	1.51	1.32	1.40	0.82	0.99	0.75
Treatments	Df	Isopoda	Julida	Lumbricidae	Psocoptera	Symphyla	Thysanoptera	
Climate (C)	1,80	0.13	0.07	0.49	1.02	0.59	0.03	
Land use (L)	4,80	1.18	3.30*	1.94	0.60	1.97	1.42	

Date (D)	1,80	2.76	8.93**	1.53	0.43	0.34	13.23***
$C \times L$	4,80	0.99	0.04	1.17	0.64	0.09	1.02
$\mathbf{C} \times \mathbf{D}$	1,80	0.08	0.11	0.27	1.44	2.40	0.08
$L \times D$	4,80	1.14	2.83	0.81	0.98	1.00	1.34
$C \times L \times D$	4,80	1.11	0.13	1.22	0.74	0.44	1.10

**Table S4** Results ( $R^2$ -values) of permutational multivariate analysis of variance (PERMANOVA) based on Bray-Curtis distances testing the effects of climate and land use, and their interaction on community structure of soil macrofauna using total biomass data (**A**). Pairwise comparisons based on Bray-Curtis dissimilarities showing differences between land-use regimes (**B**).  $R^2$ -values with \*\*\* (P < 0.001), \*\* (P < 0.01) representing significant effects are indicated in bold font.

(A) PERMANOVA results						
Treatments	Df	$R^2$ -value	<i>P</i> -value			
Climate (C)	1	0.03	0.16			
Land use (L)	4	0.25	< 0.001***			
$C \times L$	4	0.06	0.52			
( <b>B</b> ) Pairwise comparis	ons betwe	en land-use regimes				
Comparisons	Df	$R^2$	<i>P</i> -value			
EM vs. EP	1	0.05	0.45			
EM vs. IM	1	0.14	0.88			
EM vs. OF	1	0.24	0.002**			
EM vs. CF	1	0.20	0.006**			
EP vs. IM	1	0.06	0.37			
EP vs. OF	1	0.33	< 0.001***			
EP vs. CF	1	0.26	0.003**			
IM vs. OF	1	0.29	< 0.001***			
IM vs. CF	1	0.24	0.003**			
OF vs. CF	1	0.03	0.46			