

This is the accepted manuscript version of the contribution published as:

Sartori Jeunon Gontijo, E., Monteiro, A.S.C., Tonello, P.S., Roeser, H.M.P., **Friese, K.**, Rosa, A.H. (2022):

Analyses of colloidal, truly dissolved, and DGT-labile metal species and phosphorus in mining area surrounded by tailing dams using self-organising maps

Chemosphere **303, Part 2** , art. 135003

The publisher's version is available at:

<http://dx.doi.org/10.1016/j.chemosphere.2022.135003>

1 **Analyses of colloidal, truly dissolved, and DGT-labile metal species and**
2 **phosphorus in mining area surrounded by tailing dams using self-**
3 **organising maps**

4 Erik Sartori Jeunon Gontijo^{a, *}, Adnivia Santos Costa Monteiro^b, Paulo Sérgio Tonello^a,
5 Hubert Mathias Peter Roeser^c, Kurt Friese^d, André Henrique Rosa^{a, *}

6
7 *a - Institute of Science and Technology, São Paulo State University (UNESP), Av. Três de*
8 *Março, 511, Alto da Boa Vista – 18087-180 – Sorocaba- SP – Brazil, (+55 15) 3238-3414*

9 *b - Federal University of Sergipe (UFS), Campus São Cristóvão, Av. Marechal Rondon, s/n,*
10 *Jardim Rosa Elze – 49100-000 – São Cristóvão- SE– Brazil*

11 *c - Federal University of Ouro Preto (UFOP), Campus Universitário, Morro do Cruzeiro*
12 *354000-000, Ouro Preto-MG, Brazil*

13 *d - Department of Lake Research, Helmholtz Centre for Environmental Research – UFZ,*
14 *Brueckstr. 3a, 39114, Magdeburg, Germany*

15

16

17

18

19

20

21

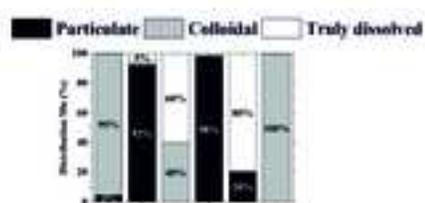
22 *corresponding authors: Erik S.J. Gontijo (sartori_jg@hotmail.com) and André H. Rosa

23 (andre.rosa@unesp.br)

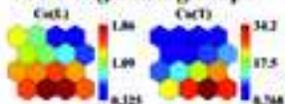
24

Highlights

- Values for trace metals and P were high before the Fundão dam failure in 2015
- P was mainly in the truly dissolved and DGT-labile fractions
- Increase of colloidal Fe reduced P in the DGT-labile fraction
- Most of total dissolved Cu was found in the colloidal fraction in Conceição River
- Differences between dry and wet seasons for Co, Fe and Ni detected by SO-Maps



Self-organising maps

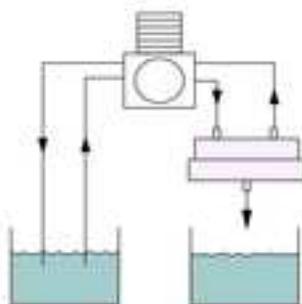


Bioavailability / Toxicity / Mobility / Fate / Scenario before accident

- DGT-labile
- Total
- Particulate
- Total dissolved
- Colloidal
- Truly dissolved



in-situ



Filtration / ultrafiltration

Water samples

- Cr
- Cu
- Co
- Mn
- Ni
- Pb
- Zn
- P

Upper Doce River catchment



- Mining area
- Sampling site
- Tailings
- River or stream

**Mining area!!
Dam burst!! Tailings!!
Tailings dam!!**

1 **Analyses of colloidal, truly dissolved, and DGT-labile metal species and**
2 **phosphorus in mining area surrounded by tailing dams using self-**
3 **organising maps**

4 Erik Sartori Jeunon Gontijo^{a, *}, Adnivia Santos Costa Monteiro^b, Paulo Sérgio Tonello^a,
5 Hubert Mathias Peter Roeser^c, Kurt Friese^d, André Henrique Rosa^{a, *}

6
7 *a - Institute of Science and Technology, São Paulo State University (UNESP), Av. Três de*
8 *Março, 511, Alto da Boa Vista – 18087-180 – Sorocaba- SP – Brazil, (+55 15) 3238-3414*

9 *b - Federal University of Sergipe (UFS), Campus São Cristóvão, Av. Marechal Rondon, s/n,*
10 *Jardim Rosa Elze – 49100-000 – São Cristóvão- SE– Brazil*

11 *c - Federal University of Ouro Preto (UFOP), Campus Universitário, Morro do Cruzeiro*
12 *354000-000, Ouro Preto-MG, Brazil*

13 *d - Department of Lake Research, Helmholtz Centre for Environmental Research – UFZ,*
14 *Brueckstr. 3a, 39114, Magdeburg, Germany*

15

16

17

18

19

20 *corresponding authors: Erik S.J. Gontijo (sartori_jg@hotmail.com) and André H. Rosa

21 (andre.rosa@unesp.br)

22

23 **Abstract**

24 The knowledge of size-distribution and lability of metals and nutrients in freshwater systems is
25 important for estimation of the ecological effects of mining. However, it is still limited in
26 several mining areas such as the Quadrilátero Ferrífero (Brazil) which was severely polluted by
27 the collapse of the Fundão tailings dam in November 2015. In this study, results of an
28 investigation from 2014 using a neural network named self-organising map (SO-Map) into the
29 conditions of selected trace metals that are of particular importance to mining areas (Cr, Cu,
30 Co, Mn, Ni, Pb, Zn) are presented. Additionally, P was considered by its high importance as a
31 nutrient and sites later affected by the dam burst were also included by chance. Water samples
32 were collected at six sites in dry and rainy seasons and filtered and ultrafiltered for
33 determination of total dissolved ($< 0.45 \mu\text{m}$) and truly dissolved ($< 1 \text{ kDa}$) fractions. Diffusive
34 gradients in thin films (DGT) devices were deployed *in situ* for determination of the DGT-labile
35 fraction. All data were analysed using SO-Map and Spearman's rank correlation. Phosphorus
36 in the Carmo River occurred mainly in the truly dissolved and DGT-labile fractions. The higher
37 amounts of this element in the river water (up to $263 \mu\text{g L}^{-1}$ of total P) might be related to
38 untreated sewage discharge. Moreover, the concentrations of other trace metals (Mn, Cu, Co,
39 Ni, Zn) were high, even under the "natural" conditions (before the dam failure) due to natural
40 and anthropogenic factors such as local lithology and mining.

41

42 **Keywords:** size fractionation, diffusive gradients in thin films, ultrafiltration, Kohonen neural
43 network, mining impacts

44

1. Introduction

The Quadrilátero Ferrífero is a well-known mineral province located in the state of Minas Gerais, in southeast of Brazil (Deschamps et al., 2002; Varcjao et al., 2011). The gold and iron mining has caused several negative impacts in this region, including soil erosion and contamination of water, sediments, and soils (Deschamps et al., 2002; Roeser and Roeser, 2010; Varcjao et al., 2011; Gontijo et al., 2016). The construction of several tailings dams for storing wastes from mining activities has also become problematic, particularly by leaking of mine wastes and the collapse of tailings dams (Rodrigues et al., 2014; Santamarina et al., 2019). The accident near the city of Mariana in November 2015 (collapse of the Fundão tailings dam) for instance released more than 40 million m³ of mining waste into the Upper Doce River catchment (Carmo et al., 2017; Vergilio et al., 2020).

Since the behaviour and fate of metals and metalloids in rivers are dependent on their chemical form, it is of high interest to study size-distribution and lability of metals and nutrients in aquatic systems. Metals and nutrients may be adsorbed in suspended particles, incorporated in living organisms, complexed with organic or inorganic ligands, or appear as free (hydrated) ions. These different forms have different mobility, toxicity and bioavailability in the environment (Buffle, 1991). As a result, the sole determination of total and even total dissolved concentrations of a given element is not sufficient to predict its effects on aquatic systems (Morel, 1983).

An important approach that can be used to better investigate trace metal speciation and mobility of metals and metalloids is size fractionation of water samples. Filtration and ultrafiltration (UF) are examples of techniques generally used for fractionating a given element into particulate, total dissolved, colloidal, and truly dissolved fractions, according to the used operational definition (Singhal et al., 2006; Gontijo et al., 2017a). While the species in the particulate fraction have low mobility and bioavailability, species in the truly dissolved fraction are more mobile and potentially bioavailable (Allen and Hansen, 1996; Wen et al., 1997).

71 The technique of diffusive gradients in thin films (DGT) can be used complementarily
72 to UF. It is able of detect *in situ* labile species (including free species and DGT-labile
73 complexes) which may be more bioavailable in water systems (Forsberg et al., 2006; Tonello
74 et al., 2007; Rougerie et al., 2021). DGT devices are deployed in aqueous environments where
75 metal ions diffuse through a diffusive boundary layer, a gel layer and accumulate in the binding
76 layer of the sampler. The concentration of the metals accumulated in the binding layer during
77 the deployment time is then measured and used to calculate the metal flux through the gel layer
78 and the amount of labile species in the water (Davison and Zhang, 1994). These labile species
79 are basically the ones able to dissociate and reach equilibrium along the diffusion medium
80 during the deployment time (van Leeuwen and Jansen, 2005; Garmo et al., 2006; Sans-Duñó et
81 al., 2021). Then, DGT is considered a dynamic technique since it represents the flux of metal
82 species to the device (van Leeuwen et al., 2005).

83 The investigation of water quality and metal fractions requires the measurement of many
84 variables that generates a huge amount of data. The analysis of these data and the complex
85 relationship among variables requires the use of exploratory tools with good visualisation
86 capabilities (Çinar and Merdun, 2009; Gontijo et al., 2021). The self-organising map (SO-Map),
87 also known as Kohonen neural network, is an example of such tools. It is a type of neural
88 network used for clustering and visualisation that converts high-dimensional data in a low-
89 dimensional space, typically a two-dimensional grid of nodes (neurons) forming a rectangular
90 map (Vesanto, 1999; Kohonen, 2001; Brereton, 2012).

91 This investigation has the objective to determine the distribution Cr, Cu, Co, Mn, Ni,
92 Pb, Zn and P in the fractions total, total dissolved, truly dissolved, and DGT-labile in samples
93 from rivers and streams in an area impacted by iron and gold mining using SO-Map. Sampling
94 was carried out before the collapse of the Fundão tailings dam and included sites in the affected
95 area. In this sense, our study can be a basis of comparison for the assessment of the
96 contaminating effect of the dam failure. The presence of several tailings dams in the catchment

97 also requires continuous monitoring of truly dissolved and DGT-labile metal and nutrient
 98 species for mobility and toxicity prediction. This type of study is still scarce in such impacted
 99 areas and may contribute to understand the effect on mining activities on metal distribution,
 100 mobility, and lability.

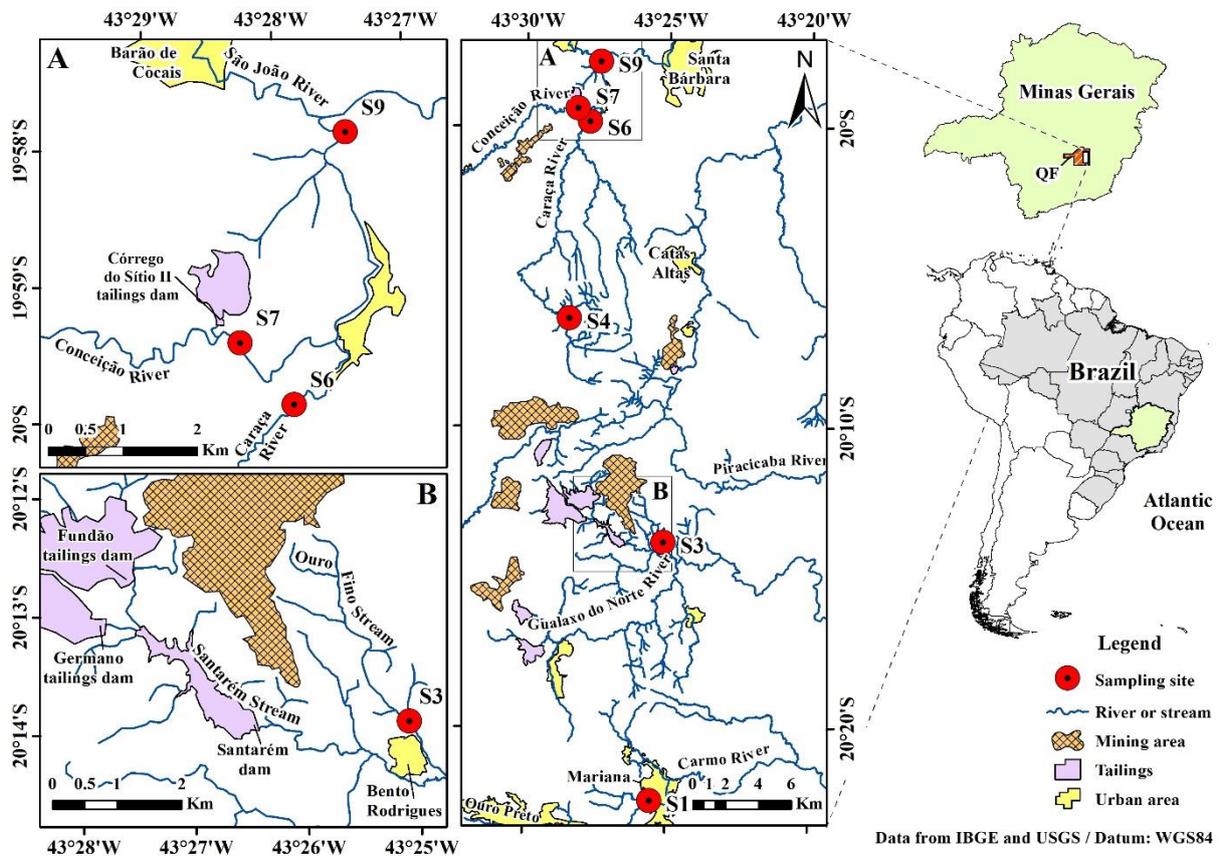
101

102 2. Material and methods

103 2.1. Study site and sampling

104 Six sampling sites (S1, S3, S4, S6, S7 and S9, Fig. 1) were investigated in Upper Doce
 105 River catchment in rainy and dry season (February and October, respectively) of 2014.
 106 Conductivity ($\mu\text{S cm}^{-1}$), total dissolved solids (TDS, mg L^{-1}), pH and temperature ($^{\circ}\text{C}$) were
 107 measured *in situ* using a multi-parameter probe Ultrameter II (Myron L Company).

108



109

110 **Fig. 1** Sampling sites (S1, S3, S4, S6, S7 and S9), mining areas and tailing dams in Upper Doce
 111 River catchment (in Quadrilátero Ferrífero, QF). Mining areas and tailing dams were delineated

112 using cloud-free Landsat 8 imagens taken from path 217 and row 74 on 11th October and 12th
113 November 2015. The images were acquired from the U.S. Geological Survey (USGS) Earth
114 Explorer webpage (<https://earthexplorer.usgs.gov/>, accessed in December 2020). All
115 delineations were performed using the software ArcGis 10.5 (ESRI, USA).

116

117 The sampling sites were originally selected to investigate the As, Al and Fe distribution
118 in rivers from Quadrilátero Ferrífero and the influence of dissolved organic matter on it (Gontijo
119 et al., 2016). However, the presence of several dams in the Doce River catchment and the dam
120 collapse in 2015 nearby the city of Mariana makes the investigation of the distribution of the
121 trace metals Cr, Cu, Co, Mn, Ni, Pb and Zn in rivers near the dams remarkably important. The
122 selected sites included analysis of DGT-labile species, except S3 which was selected because
123 it was directly affected by the collapse of Fundão dam in 2015 [Fig. A.1, Appendix A,
124 Supplementary Information (SI)]. While S1 and S3 are located in areas affected by the dam-
125 burst disaster, sites S4, S6, S7 and S9 are located in rivers or streams surrounded by mining
126 areas and tailings dams (Fig. 1) but were not affected by this disaster. Site S4 is a pristine area
127 located in a private protected place; Private Reserve of the Natural Heritage, RPPN.

128 Freshwater samples were collected from all sites using precleaned plastic bottles (2 × 1
129 L for each site). The samples were subsequently transported to laboratory for determination of
130 the elements in the total, total dissolved (< 0.45 µm) and truly dissolved (< 1 kDa) fractions.
131 Amber glass bottles (60 mL) were used to store samples to analyse the content of dissolved
132 organic carbon (DOC) at each site. Samples for DOC were filtered through 0.45 µm membranes
133 and preserved with phosphoric acid (H₃PO₄). All samples were stored at 4°C until analysis.

134

135 **2.2. Filtration, ultrafiltration and acid digestion**

136 300 mL of the samples were filtered through 0.45 µm cellulose nitrate membranes for
137 determination of the total dissolved fraction of the analysed elements. The truly dissolved

138 fraction was analysed by further ultrafiltration (UF) of 150 mL through 1 kDa regenerated
139 cellulose membranes (Millipore). The UF was carried out using a Teflon home-made tangential
140 UF system coupled to a Gilson Minipuls[®] 3 peristaltic pump, similar as described in Monteiro
141 et al. (2016). The system and membranes were cleaned using NaOH (0.1 mol L⁻¹) and HCl
142 (< 0.01 mol L⁻¹) solutions followed by rinsing with ultrapure water (18.2 MΩ, Millipore) before
143 each filtration.

144 Acid digestion of the fractions total (not filtered, 150 mL), total dissolved (< 0.45 μm,
145 150 mL) and truly dissolved (< 1 kDa, 150 mL) followed by a six-fold preconcentration (four-
146 fold for the truly dissolved fraction) was performed to detect elements in lower concentrations.
147 The digestion was carried out on a hot plate based on the method 3005A of the U.S.
148 Environmental Protection Agency (US EPA, 1992). Determination of the elements in all
149 digested fractions were performed by inductively coupled plasma optical emission
150 spectrometry (ICP-OES, Agilent Technologies 700 Series). Limits of detection (LoD) and
151 quantification (LoQ) for the elements analysed by ICP-OES are given in the SI section (Table
152 A.1, Appendix A). Blanks of ultrapure water were used to correct measured concentrations.
153 Particulate fraction (> 0.45 μm) and colloidal fraction (< 0.45 μm and > 1 kDa) were calculated
154 by the difference between total and total dissolved fraction and by the difference between total
155 dissolved and truly dissolved fraction, respectively. All experiments were performed in
156 duplicate. A summary of all fractions cited in this investigation is presented within the
157 supplement (Table A.2, SI).

158 The recovery for all elements measured (from recovery tests for validation) were in the
159 range of 80-120% as shown in Table A.3 (Appendix A, SI).

160

161 **2.3. Diffusive gradients in thin films (DGT)**

162 Diffusive gradients in thin films (DGT) units were assembled using piston-type devices
163 acquired from DGT research (Lancaster, UK) for determination of DGT-labile species. Each

164 device contained a 0.45 μm cellulose nitrate membrane, a diffusive gel (agarose crosslinked
165 polyacrylamide) layer, and a binding gel layer. Chelex-100 gel was used as binding gel layer for
166 determination of labile Cd, Cr, Cu, Co, Fe, Mn, Ni, Pb and Zn. Ferrihydrite gel was used as
167 binding gel for determination of labile P. The surface area of the piston-like units exposed to
168 the water was 3.14 cm^2 and the thickness of the of the diffusive gel and the cellulose nitrate
169 membrane were 0.8 mm and 0.13 mm, respectively.

170 Three DGT units containing chelex-100 gel and three units containing ferrihydrite were
171 deployed at each sampling site. The units were retrieved after 4 to 9 days (cf. Table A.4,
172 Appendix A, SI), rinsed with ultrapure water, stored in plastic bags, and transported to the
173 laboratory. Water temperature was measured *in situ* during deployment and retrieval at each
174 site for determination of the diffusion coefficients of all analysed elements. In laboratory, the
175 DGT units were opened, and the chelex-100 gel layers were eluted with 2 mL of HNO_3 (1 mol
176 L^{-1}). The ferrihydrite gels were eluted with 2 mL of concentrated HCl for 24 h in a shaker and
177 then diluted with ultrapure water.

178 The concentration of the analysed elements (C_e , $\mu\text{g L}^{-1}$) after elution was analysed with
179 ICP-OES (Agilent Technologies 700 Series). The accumulated mass of each analyte (M , ng)
180 was subsequently calculated using the following equation (Zhang and Davison, 1995):

$$181 \quad M = C_e(V_e + V_g)/E_f \quad (1)$$

182 where V_e is the volume of the eluent (mL), V_g is the volume of the binding gel (mL) and E_f is
183 the elution factor (cf. Table A.4, Appendix A). The DGT-labile concentration of each analyte
184 (C_a , $\mu\text{g L}^{-1}$) was then calculated using the following equation (Zhang and Davison, 1995):

$$185 \quad C_a = M\Delta g/DtA \quad (2)$$

186 where Δg is the thickness of the diffusive gel and cellulose nitrate filter (cm), A is the exposed
187 area to water in the river/stream (cm^2), t is the deployment time (s), and D is the diffusion
188 coefficient ($\text{cm}^2 \text{s}^{-1}$). The diffusion coefficients of each element analysed are available from
189 DGT research (<https://www.dgtresearch.com>, accessed in October 2020; Table A.5, Appendix

190 A, SI). The coefficients were corrected using the average temperature measured *in situ* during
191 deployment and retrieval of the DGT units (Table A.6, Appendix A, SI).

192

193 **2.4. Self-organising maps and Spearman's rank correlation**

194 The relationships among samples, variables and between samples and variables were
195 investigated using the SO-Map approach. The SO-Map consists of neurons in a two-
196 dimensional grid. Each input sample in this technique is represented by a vector whose elements
197 correspond to the concentrations of the four fractions (total, total dissolved, truly dissolved, and
198 DGT-labile) of each investigated element (Cd, Cr, Cu, Co, Fe, Mn, Ni, P, Pb and Zn) and the
199 physicochemical parameters (conductivity, TDS, pH and temperature) measured at each
200 sampling site. The neurons in the output map have the same dimension (each one represented
201 by a component plane) of the input vectors. The input and output layers are connected by weight
202 vectors. These vectors are initialised with small random numbers in the analysis. Then, the
203 Euclidean distance between an input vector and the weight vector of each output neuron is
204 calculated. The neuron with the smallest distance (best matching unit, BMU) from the input
205 vector is then selected and the weight vector of the BMU and its neighbours are updated (Çinar
206 and Merdun, 2009; Asan and Ercan, 2012; Brereton, 2012).

207 The data were normalised before the analysis using z-score transformation for
208 converting the variables to a common scale with mean zero and standard deviation one.
209 Different map architectures (2×2 to 6×6) were tested to find the map with the most
210 informative distribution of the samples. The optimal number of neurons should not be too high
211 or too low because the samples would be respectively too far apart or too close, decreasing the
212 possibility to extract information (Garcia et al., 2007). The software Matlab 2017b
213 (MathWorks, Natick, MA) and the SOM toolbox 2.1 (Alhoniemi et al., 2000) were used to
214 perform the SO-Map analysis.

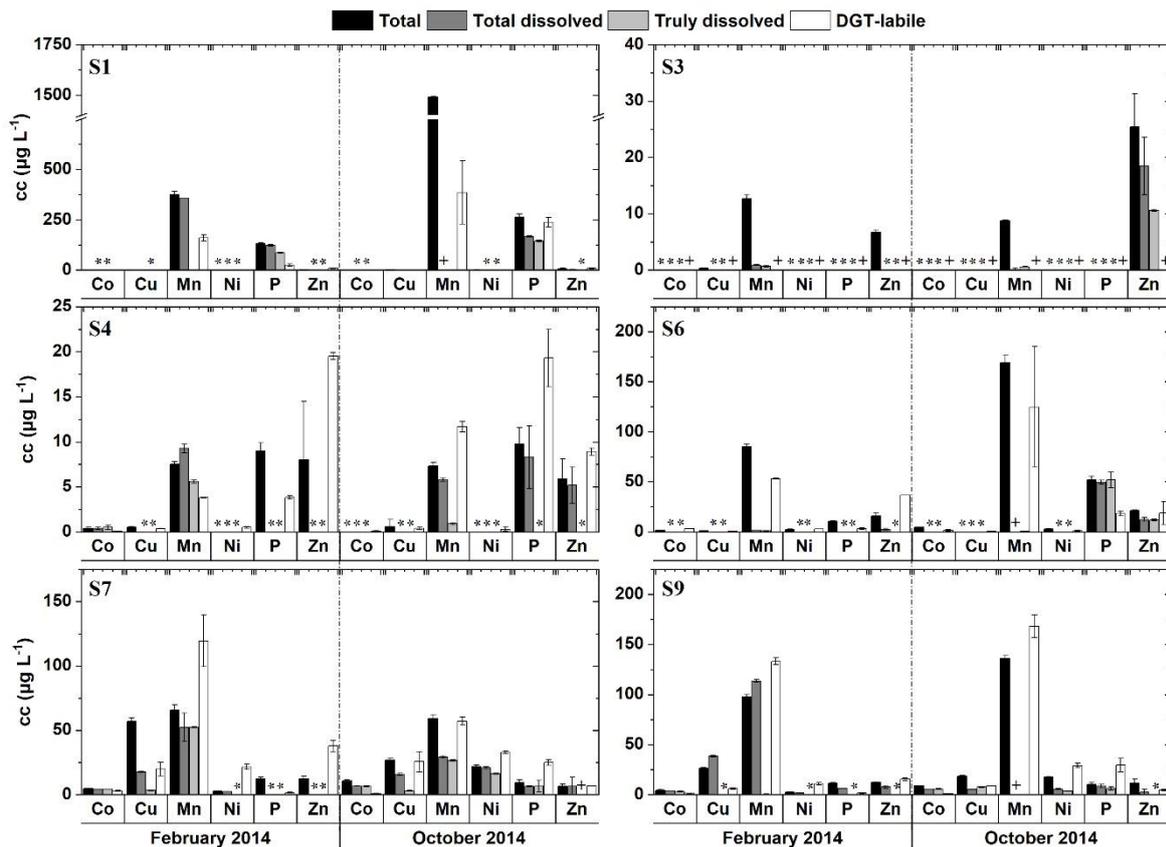
215 The correlation between the variables was investigated using Spearman's rank
216 correlation coefficients in the R software. Moderate or strong correlations were expressed by
217 coefficients (ρ) higher than 0.40 (positive correlation) or lower than -0.40 (negative
218 correlation). To run SO-Map and correlation analyses, concentration values of elements
219 measured by ICP-OES below the limit of detection were replaced by these values. The file (.txt)
220 used for running the SO-Map analysis is available in the SI.

221

222 **3. Results**

223 Sampling site S1 showed the highest concentrations of Mn and P, especially in the dry
224 season (Fig. 2). This site also recorded the highest conductivity ($292 \mu\text{S cm}^{-1}$), pH (7.2), TDS
225 (194 mg L^{-1}) and the highest concentration of DOC (4.0 mg L^{-1}) in the dry season (Table A.7,
226 Appendix A, SI). In contrast, lowest values of conductivity, TDS, and Mn and P concentrations
227 were observed at S3 and S4. Sites S7 and S9 presented the highest concentrations of Co, Cu
228 and Ni, while S3 showed the highest amounts of Zn. Cd, Cr, and Pb were not detected in most
229 samples (cf. Table A.8, Appendix A). Small amounts (up to $5.7 \mu\text{g L}^{-1}$) of total Cr and Pb were
230 detected at sites S1, S6, and S9. DGT-labile Cd ($0.12 \pm 0.05 \mu\text{g L}^{-1}$) and Cr ($0.13 \pm 0.08 \mu\text{g L}^{-1}$)
231 were also detected at site S7 in the rainy season (Fig. 2, Table A.8 within SI).

232

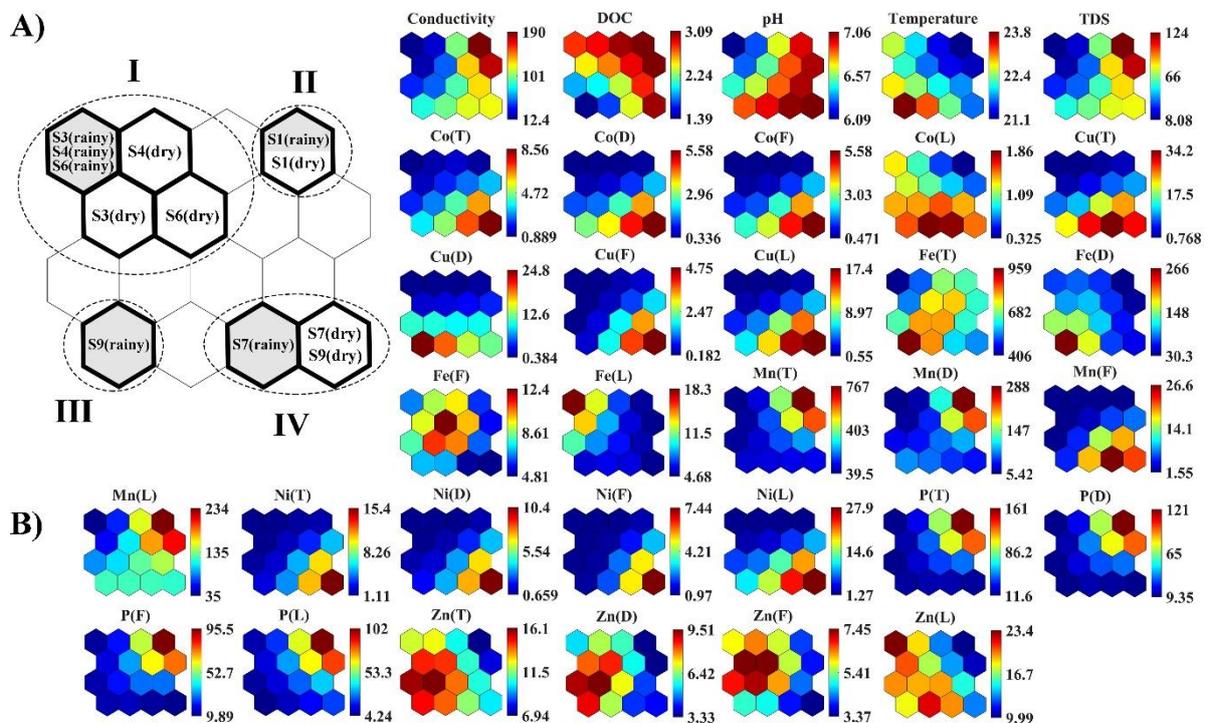


233
 234 **Fig. 2** Concentration (cc) of Co, Cu, Mn, Ni, P and Zn in the total, total dissolved, truly
 235 dissolved, and DGT-labile fractions from the six sampling sites (S1, S3, S4, S6, S7 and S9) in
 236 Upper Doce River catchment. Not detected and not determined values are depicted as * and +,
 237 respectively. These results are available in the Appendix B, supplementary information (SI),
 238 tables B.1-B.2.

239
 240 Regarding the different fractions (Fig. A.2, Appendix A), Co and P occurred
 241 predominantly in the truly dissolved fraction (< 1 kDa) in most samples. Mn was predominant
 242 in the colloidal fraction at S1 and S9 in the rainy season and at S4 in the dry season, while it
 243 was predominant in the truly dissolved fraction at S4 and S7 in the rainy season. Cu occurred
 244 mainly in the particulate fraction at sites S1 and S9 in the dry season and at S7 in the rainy
 245 season. The highest concentrations of DGT-labile Ni occurred at sites S7 and S9 (33.0 ± 0.9 and
 246 $29.1 \pm 2.2 \mu\text{g L}^{-1}$, Fig 2) and of DGT-labile Cu at S7 ($25.8 \pm 7.8 \mu\text{g L}^{-1}$). Most of total dissolved
 247 P was found in the DGT-labile fraction (Fig. 2) and the same behaviour was observed for Mn,
 248 except at S1 (rainy season) where only about half of the total dissolved Mn was DGT-labile.

249 The samples were divided by SO-Map analysis in four (I-IV) distinct groups that are
 250 circled in Fig. 3 A. These groups represent samples that have similar properties since they are
 251 in the same or neighbouring neurons (hexagonal units). The larger the distance between two
 252 samples in the maps, the more dissimilar from each other they are (Garcia et al., 2007). The
 253 map architecture 4×4 was selected because it gives the best distribution of samples (most
 254 informative) in the SO-Map analysis. Behind the map of samples (Fig. 3 A), there are
 255 component planes (map of variables, Fig. 3 B) that correspond to the variables used to create
 256 the maps (Brereton, 2012).

257



258

259 **Fig. 3** Self-organising maps: A) Map of samples, in which samples from six sites (S1, S3, S4,
 260 S6, S7 and S9) were clustered in four distinct groups (I-IV) as circled in the map. The
 261 sampling season (rainy or dry) when water was collected is indicated in parenthesis and by
 262 grey (rainy season) and white (dry season) tones in the neurons (hexagonal units). B) Maps of
 263 variables (component planes): the bars beside each map indicate the intensity of each variable
 264 (red is high intensity and blue is low intensity). The fractions total, total dissolved, truly
 265 dissolved, and DGT-labile are indicated by the letters T, D, F and L (in brackets),
 266 respectively. The used Fe data were derived from Gontijo et al. (2016).

267

268 Group I in the map includes samples with lower conductivity, pH, TDS and lower
269 content of Co, Cu, Mn, Ni, and P (blue colours in the maps of these variables at the
270 corresponding position of group I, Fig. 3 B). The total and total dissolved Fe content in group I
271 is lower than in other groups, but it has samples with the highest truly dissolved and DGT-labile
272 Fe contents (see red colours for these variables at the position of group I, Fig. 3). Group II is
273 formed by S1 samples which exhibit the highest conductivity, DOC, TDS and highest Mn and
274 P contents. Water temperature and concentrations of all other measured elements were low in
275 this second group. Group IV (samples from S7 and S9) had samples with higher amounts of Cu
276 and Ni. The sample taken at S9 in the rainy season formed a separated group out from other
277 samples at S9 because of its higher concentrations of total and total dissolved Fe and Zn. A
278 summary of the most relevant variables for each group is presented in the supplementary section
279 (Table A.9, Appendix A).

280 The analysis of the map of variables (Fig. 3 B) can also indicate correlations between
281 variables. These correlations are revealed when there are similar or opposite patterns at the
282 same positions in the neurons of a pair of maps of variables (Vesanto, 1999). Neurons with
283 higher intensity for total dissolved Fe (red tones, Fig. 3 B) for instance have lower intensity for
284 P fractions (blue tones), indicating inverse relationship between these P-fractions and total
285 dissolved Fe. In fact, a negative correlation (-0.55 , $p=0.10$; Table A.10, Appendix A) was
286 confirmed between total dissolved Fe and DGT-labile P. A positive relationship is observed
287 among the variables conductivity, DOC, TDS, total Mn, total dissolved Mn, DGT-labile Mn,
288 total P, total dissolved P, truly dissolved P, and DGT-labile P, since all these variables presented
289 red tones in the corresponding neurons in the maps (note that the maps of these variables are
290 similar to each other, Fig. 3 B). Significant ($p < 0.05$) positive correlations were confirmed for
291 example between total P and conductivity/TDS (both 0.75), total Mn and conductivity/TDS
292 (0.78 and 0.79, respectively) and DOC and labile P (0.65). Similar patterns in the maps can also

293 be observed between Co, Cu and Ni fractions, confirmed by positive correlations (up to 0.95)
294 between total Ni and Co.

295

296 **4. Discussion**

297 **4.1. Species distribution, potential bioavailability and seasonality**

298 The high P contents (up to $263 \pm 16 \mu\text{g L}^{-1}$) in samples from Carmo River (site S1,
299 samples from group II, Fig. 3) are probably related to untreated urban sewage effluent
300 discharge, which is a problem already reported in the literature (Santos et al., 2019). The higher
301 conductivity, DOC and TDS also support this assumption because these variables are also
302 related to sewage inputs in rivers (Daniel et al., 2002). P concentration was lower during the
303 rainy season (Fig. 2), probably because of dilution of the sewage effluent by higher river flow
304 as a result of increased precipitation in the summer period (cf. Fig. A.3, Fig. A.4, Appendix A).

305 P appeared predominantly in the truly dissolved fraction and is mostly DGT-labile at all
306 sampling sites. These DGT-labile species include species that dissociate and/or desorb in the
307 medium, diffusing through the gel layer and accumulating in the binding gel during the
308 deployment time (Zhang and Davison, 1995; Zhang and Davison, 2015). Although different
309 organisms have different uptake mechanisms, the labile species measured by DGT are
310 considered potentially bioavailable (Zhang and Davison, 2015). Therefore, higher
311 concentrations of DGT-labile and potentially bioavailable P at S1 and S6 may cause
312 cyanobacteria blooms which have been already recorded downstream in Doce River catchment
313 and correlated to total P, particularly in the dry season (Jardim et al., 2014). The data also
314 indicated that the presence of total dissolved Fe may decrease the amount of DGT-labile P
315 (negative correlation), probably because P tends to bind effectively to Fe colloids (e.g. Fe-
316 oxyhydroxides and Fe-organic matter) that are not labile and may be not readily bioavailable to
317 biota (Eisenreich and Armstrong, 1980; Baken et al., 2014; Baken et al., 2016; Gontijo et al.,

318 2016). The positive correlation (0.68) between total Fe and total P may also be related to the P
319 binding of Fe.

320 The highest measured Mn content was also detected at S1, where it may be derived from
321 manganese-bearing rocks in the banded iron formation of the Quadrilatero Ferrífero (Dorr,
322 1969; Costa et al., 2003). Mobilisation from old gold and manganese exploitation may have
323 contributed to increase the amounts of Mn in Carmo River (Costa et al., 2003). Mn appeared
324 predominantly in the particulate and colloidal fractions, except in S7 samples. This colloidal
325 Mn may appear as Mn oxides or bound to humic substances (HS), Fe-rich colloids and other
326 ligands in the aquatic system (Huangfu et al., 2013; Li et al., 2019).

327 The sampling sites S3, S4 and S6 (group I, Fig. 3) showed the best water quality by
328 lowest concentrations of all elements analysed, except Zn at S3 and S6. Measurements of DOC
329 at the retrieval time of DGT devices at S4 showed a DOC increase from 4.7 ± 0.1 to 9.0 ± 0.1 mg
330 L^{-1} (Table A.13, Appendix A) after several days of strong rainfall (72.8 mm between 15 and
331 19th February 2014, Table A.11; see also Table A.12, Appendix A). This increase of DOC was
332 associated with a drop in pH down to 4.4 and it is related to the presence of HS (Gontijo et al.,
333 2016; Gontijo et al., 2017b). Higher contents of HS at S4 after rainfalls may impact the species
334 distribution in the rivers from Upper Doce River catchment because HS stabilise for instance
335 Fe in the colloidal fraction, thereby increasing its mobility (Ritter et al., 2006). A greater
336 presence of Fe complexed to HS can also affect the distribution of P in the catchment, moving
337 it from the truly dissolved to the colloidal fraction by binding to the complexed Fe (Ritter et al.,
338 2006; Gontijo et al., 2017b; Saeed et al., 2018).

339 The samples from Conceição River (particularly at S9) are the ones (groups III and IV,
340 Fig. 3) where seasonality was detected using SO-Map analysis (samples are in different groups).
341 This difference was caused especially by higher Fe and lower Ni and Co content in the rainy
342 season than in the dry season. The higher Fe content may be attributed to the leaching of soils
343 due to rainfall during the rainy season (cf. annual precipitation in the area in Fig. A.3, Appendix

344 A). Conversely, Ni and Co are diluted in the rainy season because of higher water flow of the
345 river. Ni may be derived from ultramafic rocks or laterite ores present in the area (Terezinha
346 Costa et al., 2006). Co may also appear in laterite ores as substitute of Fe and its high correlation
347 with Ni (0.95, $p < 0.05$) supports the natural origin (Ribeiro et al., 2020). Most of the total
348 dissolved Zn was DGT-labile and consequently potentially bioavailable. The DGT-lability of
349 both, Cu and Zn decreased from S7 to S9 in the Conceição River. This may be connected to the
350 input of HS from Caraça River catchment (Gontijo et al., 2016) that can bind these elements,
351 reducing their lability (McKnight et al., 1983; Chakraborty and Chakrabarti, 2008).

352 The trace-metals Cd, Cr, and Pb were not detected in most samples. Cr and Pb were
353 found in samples from Carmo and Conceição rivers mostly in the total fraction which may be
354 related to resuspension from sediments where they have already been recorded (Deschamps et
355 al., 2002; Terezinha Costa et al., 2006; Marques et al., 2019). DGT-labile Cd was found at S7
356 ($0.12 \pm 0.05 \mu\text{g L}^{-1}$ Cd), probably due to the pre-concentration capabilities of DGT (Zhang and
357 Davison, 1995).

358 Table A.14 (Appendix A) presents the total dissolved concentrations of the analysed
359 elements (without digestion) at the retrieval time in October/2014. It shows that the
360 concentrations of most of the detected elements were quite similar to the ones from the
361 deployment time (Fig. 2) in the dry season, except for Cu at S7. This difference reflects the
362 dynamic environment with presence of several processes such as the ones described above in a
363 way that aquatic systems are hardly at chemical equilibrium (van Leeuwen et al., 2005).

364

365 **4.2. Legislation and possible effects of the dam collapse**

366 The Brazilian National Environmental Council (CONAMA), in its Resolution 357/2005
367 (Brazil, 2005), established regulations and guidelines for the classification of water bodies
368 according to their intended use. There are several classes in which water quality standards are
369 assigned and must be preserved. The waters from Doce River catchment are classified as Class

370 2, in which the use is intended for protection of aquatic life, water supply after conventional
371 treatment, recreation, fishing and irrigation (IGAM, 2015). Several parameters in the samples
372 from Upper Doce River catchment were above the limits of class 2, especially in Carmo and
373 Conceição rivers. Total P and Mn concentrations at S1 for instance were much higher (up to
374 $263 \mu\text{g L}^{-1}$ P and up to $1491 \mu\text{g L}^{-1}$ Mn) than the standards for class 2 ($100 \mu\text{g L}^{-1}$ for both P
375 and Mn, Table A.15, Appendix A). Besides total Mn, samples from Conceição River (at S7 and
376 S9) had also higher values for total dissolved Cu (max. $38.5 \mu\text{g L}^{-1}$ Cu) than the limit by
377 legislation ($9 \mu\text{g L}^{-1}$ Cu). These results shows that the water quality in the studied area was
378 already impaired by natural and human factors, even before the dam breach at Mariana in 2015.

379 From literature it is known that the dam burst in the investigated area has initially caused
380 an increase in the amount of suspended particulate matter (SPM) and turbidity in the river
381 waters. That derived from the sludge and mud and was composed mainly of Fe and Si oxides
382 (Pires et al., 2003; IBAMA, 2015; IGAM, 2015; Fraga et al., 2021). This material was
383 transported downstream and stored in riverbeds and flood plains after most of the tailings from
384 the accident was settled. According to Hatje et al. (2017), heavy rainfall may contribute to
385 further erosion and remobilisation of the deposited material in the future. Such events may
386 change the distribution and bioavailability of some elements in the affected systems. P for
387 example was mostly present in the truly dissolved fraction but it will probably bind to increasing
388 contents of Fe-rich particulate matter, thereby transferring it from the truly dissolved to the
389 particulate fraction (Smith and Longmore, 1980). This change of distribution subsequently will
390 affect the mobility and lability of the elements in the water column and may also impact for
391 example the levels of chlorophyll-a and trophic status in some reservoirs along Doce River
392 (Coimbra et al., 2020).

393 Comparing total and total dissolved metal contents analysed after the accident (Hatje et
394 al., 2017; IGAM, 2018; Silva et al., 2018; Foesch et al., 2020, Table A16, Appendix A) with
395 the results from our study, higher Cr, Cu, Fe, Pb, and Mn contents were observed in Carmo

396 River and Ouro Fino Stream after the accident. The concentration of total Mn in Carmo River
397 for instance was much higher after the accident (up to 13400 $\mu\text{g L}^{-1}$ Mn in 2015; IGAM, 2018)
398 than before (up to 1491 $\mu\text{g L}^{-1}$ Mn in this study). The total dissolved Mn in Ouro Fino Stream
399 was also higher after the accident (117 $\mu\text{g L}^{-1}$ Mn, Hatje et al. (2017)) than before (up to 0.9 μg
400 L^{-1} Mn in this study). Total Pb which was not detected in Carmo River in this investigation
401 reached values above 40 $\mu\text{g L}^{-1}$ in 2015 and 2016 (IGAM, 2018; Foesch et al., 2020), exceeding
402 the Brazilian limit of Pb for class 2 freshwaters (Table A.15, Appendix A). Total dissolved Fe
403 and total Cu contents in Carmo River also presented an expressive increase, from maximum of
404 9 $\mu\text{g L}^{-1}$ Fe and 2 $\mu\text{g L}^{-1}$ Cu in this study to 2620 $\mu\text{g L}^{-1}$ Fe and 20 $\mu\text{g L}^{-1}$ Cu after the accident
405 (Hatje et al., 2017; IGAM, 2018; Foesch et al., 2020), respectively. However, there is still no
406 information regarding species distribution and DGT-lability in the water column after the dam
407 collapse in the area for comparison and impact evaluation.

408

409 **5. Conclusions**

410 Both techniques (DGT and ultrafiltration) were complementary used to study the
411 distribution and DGT-lability of P and some trace elements within the upper Rio Doce
412 catchment affected by mining (e.g., aluminium, iron, gold and manganese). Our results show
413 that even under "natural" background conditions before the dam failure of Mariana in 2015, the
414 element concentrations of some trace metals and phosphorus were in part above the Brazilian
415 limits for freshwater class 2 due to natural and anthropogenic factors such as local lithology
416 (Mn, Cu, Ni, Co), mining (Cu, Zn, Pb) and untreated sewage discharge (P, DOC).

417 Carmo River exhibited the highest P concentrations, which occurred predominantly in
418 the truly dissolved and DGT-labile fractions, indicating that it is mobile and potentially
419 bioavailable. The data indicated that Fe may have decreased the amount of DGT-labile P by
420 adsorption processes to increase the particulate amount of P. Cd was detected only in the DGT-
421 labile fraction but not in the fractions measured by filtration and ultrafiltration because of the

422 pre-concentration capabilities of DGT. The DGT-labile amount of Cu and Zn decreased over
423 the investigated stretch of the Conceição River, indicating the presence of other ligands (e.g.,
424 HS) and/or particles (> 0.45 µm) in the water phase which is supported by a contemporaneous
425 increase of the colloidal and/or particulate concentrations. Effects of seasonality (dry vs. rainy
426 seasons) were detected only in Conceição River and was attributed especially to Fe, Ni and Co.
427 Concentrations of Ni and Co were lower in the wet season than in the dry season due to the
428 dilution effect. In contrast, the concentrations of Fe increased in the rainy season, presumably
429 due to increased erosion of lateritic soils.

430 The dam burst in the catchment with the subsequent increase in particulate material
431 deposited downstream in the riverbed and floodplains has the potential to affect the aquatic life
432 in the catchment. To assess the ecological impacts and fate of the dam failure of Mariana in
433 2015, a continuous chemical monitoring including the distribution and DGT-lability of trace-
434 metals and nutrients would be mandatory. Our study provides a basis for such monitoring and
435 allows comparison of post-accident chemical exposure with "quasi-natural" background
436 concentration ranges of the situation before the dam failure.

437

438 **Acknowledgments**

439 We are thankful to the Graduate Program in Environmental Engineering (ProAmb) from the
440 Federal University of Ouro Preto (UFOP) that provided infrastructure for some analysis and the
441 field trips. We grateful acknowledge the help of Ms. Deyse Reis, Mr. Edilson Oliveira and Ms.
442 Laura Nascimento for their support in field and laboratory.

443

444 **Declarations**

445 **Ethics approval and consent to participate**

446 Not applicable

447 **Consent for publication**

448 Not applicable

449 **Research data**

450 All data generated or analysed during this study are included in this published article [and its
451 supplementary information files]. They can also be accessed at
452 <https://data.mendeley.com/datasets/6bwcgxsrkx/2>.

453

454 **Funding**

455 This research was supported by the São Paulo Research Foundation (FAPESP, grant numbers
456 12/17727-8, 13/14122-0 and 19/06800-5), the German Academic Exchange Service (DAAD,
457 grant number DAAD- ID 57414997) and the Brazilian Coordination for the Improvement of
458 Higher Education (CAPES, grant numbers BEX10452/14-1 via CAPES/DAAD scholarship
459 program and 99999.008107/2015- 07).

460

461 **Supplementary Information (SI)**

462 Supplementary files to this article including appendices A and B can be found online.

463

464 **References**

- 465 Alhoniemi, E., Himberg, J., Parhankangas, J., Vesanto, J., 2000. SOM Toolbox.
- 466 Allen, H.E., Hansen, D.J., 1996. The importance of trace metal speciation to water quality
467 criteria. *Water Environ. Res* 68, 42-54. <https://doi.org/10.2175/106143096X127307>.
- 468 Asan, U., Ercan, S., 2012. An Introduction to Self-Organizing Maps. in: Kahraman, C. (Ed.).
469 Computational Intelligence Systems in Industrial Engineering: With Recent Theory and
470 Applications. Atlantis Press, Paris, pp. 295-315.
- 471 Baken, S., Moens, C., van der Grift, B., Smolders, E., 2016. Phosphate binding by natural
472 iron-rich colloids in streams. *Water Res.* 98, 326-333.
473 <https://doi.org/10.1016/j.watres.2016.04.032>.

474 Baken, S., Nawara, S., Van Moorlehem, C., Smolders, E., 2014. Iron colloids reduce the
475 bioavailability of phosphorus to the green alga *Raphidocelis subcapitata*. *Water Res.* 59,
476 198-206. <https://doi.org/10.1016/j.watres.2014.04.010>.

477 Brereton, R.G., 2012. Self organising maps for visualising and modelling. *Chemistry Central*
478 *Journal* 6, S1. <https://doi.org/10.1186/1752-153X-6-S2-S1>.

479 Buffle, J., 1991. *Complexation Reactions in Aquatic Systems: An Analytical Approach*. Ellis
480 Horwood Ltd.

481 Carmo, F.F.d., Kamino, L.H.Y., Junior, R.T., Campos, I.C.d., Carmo, F.F.d., Silvino, G.,
482 Castro, K.J.d.S.X.d., Mauro, M.L., Rodrigues, N.U.A., Miranda, M.P.d.S., Pinto,
483 C.E.F., 2017. Fundão tailings dam failures: the environment tragedy of the largest
484 technological disaster of Brazilian mining in global context. *Perspectives in Ecology*
485 *and Conservation* 15, 145-151. <https://doi.org/10.1016/j.pecon.2017.06.002>.

486 Chakraborty, P., Chakrabarti, C.L., 2008. Competition from Cu(II), Zn(II) and Cd(II) in Pb(II)
487 Binding to Suwannee River Fulvic Acid. *Water, Air, and Soil Pollution* 195, 63-71.
488 <https://doi.org/10.1007/s11270-008-9727-7>.

489 Çinar, O., Merdun, H., 2009. Application of an unsupervised artificial neural network
490 technique to multivariant surface water quality data. *Ecological Research* 24, 163-173.
491 <https://doi.org/10.1007/s11284-008-0495-z>.

492 Coimbra, K.T.O., Alcântara, E., de Souza Filho, C.R., 2020. Satellite evidence for pervasive
493 water eutrophication in the Doce River reservoirs following the collapse of the Fundao
494 dam in Brazil. *Environmental Pollution*, 116014.
495 <https://doi.org/10.1016/j.envpol.2020.116014>.

496 Costa, A.T., Nalini, H.A., de Lena, J.C., Friese, K., Mages, M., 2003. Surface water quality
497 and sediment geochemistry in the Gualaxo do Norte basin, eastern Quadrilátero
498 Ferrífero, Minas Gerais, Brazil. *Environ. Geol.* 45, 226-235.
499 <https://doi.org/10.1007/s00254-003-0870-6>.

500 Daniel, M.H.B., Montebelo, A.A., Bernardes, M.C., Ometto, J.P.H.B., Camargo, P.B.d.,
501 Krusche, A.V., Ballester, M.V., Victoria, R.L., Martinelli, L.A., 2002. Effects of Urban
502 Sewage on Dissolved Oxygen, Dissolved Inorganic and Organic Carbon, and Electrical
503 Conductivity of Small Streams along a Gradient of Urbanization in the Piracicaba River
504 Basin. *Water, Air, and Soil Pollution* 136, 189-206.
505 <https://doi.org/10.1023/A:1015287708170>.

506 Davison, W., Zhang, H., 1994. In situ speciation measurements of trace components in natural
507 waters using thin-film gels. *Nature* 367, 546-548. <https://doi.org/10.1038/367546a0>.

508 Deschamps, E., Ciminelli, V., Lange, F., Matschullat, J., Raue, B., Schmidt, H., 2002. Soil
509 and sediment geochemistry of the iron quadrangle, Brazil the case of arsenic. *Journal of*
510 *Soils and Sediments* 2, 216-222. <https://doi.org/10.1007/bf02991043>.

511 Dorr, J.V.N., 1969. Physiographic, Stratigraphic and Structural Development of the
512 Quadrilátero Ferrífero, Minas Gerais, Brazil: Regional Geology of the Quadrilátero
513 Ferrífero, Minas Gerais, Brazil. Regional geology of the Quadrilatero Ferrifero, Minas
514 Gerais, Brazil, 110.

515 Eisenreich, S.J., Armstrong, D.E., 1980. Association of organic matter, iron and inorganic
516 phosphorus in lake waters. *Environment International* 3, 485-490.
517 [https://doi.org/10.1016/0160-4120\(80\)90156-7](https://doi.org/10.1016/0160-4120(80)90156-7).

518 Foesch, M.D.S., Francelino, M.R., Rocha, P.A., Gomes, A.R.L., 2020. River Water
519 Contamination Resulting from the Mariana Disaster, Brazil. *Floresta e Ambiente* 27.

520 Forsberg, J., Dahlgvist, R., Gelting-Nystrom, J., Ingri, J., 2006. Trace metal speciation in
521 brackish water using diffusive gradients in thin films and ultrafiltration: Comparison of
522 techniques. *Environ. Sci. Technol.* 40, 3901-3905. <https://doi.org/10.1021/es0600781>.

523 Fraga, M.d.S., da Silva, D.D., Reis, G.B., Guedes, H.A.S., Elesbon, A.A.A., 2021. Temporal
524 and spatial trend analysis of surface water quality in the Doce River basin, Minas

525 Gerais, Brazil. *Environment, Development and Sustainability*.
526 <https://doi.org/10.1007/s10668-020-01160-8>.

527 Garcia, J.S., Da Silva, G.A., Arruda, M.A.Z., Poppi, R.J., 2007. Application of Kohonen
528 neural network to exploratory analyses of synchrotron radiation x-ray fluorescence
529 measurements of sunflower metalloproteins. *X-Ray Spectrom.* 36, 122-129.
530 <https://doi.org/10.1002/xrs.950>.

531 Garmo, Ø.A., Lehto, N.J., Zhang, H., Davison, W., Røyset, O., Steinnes, E., 2006. Dynamic
532 Aspects of DGT as Demonstrated by Experiments with Lanthanide Complexes of a
533 Multidentate Ligand. *Environ. Sci. Technol.* 40, 4754-4760. 10.1021/es060674v.

534 Gontijo, E.S.J., Herzsprung, P., Lechtenfeld, O.J., de Castro Bueno, C., A.C. Barth, J., Rosa,
535 A.H., Friese, K., 2021. Multi-proxy approach involving ultrahigh resolution mass
536 spectrometry and self-organising maps to investigate the origin and quality of
537 sedimentary organic matter across a subtropical reservoir. *Org. Geochem.* 151, 104165.
538 <https://doi.org/10.1016/j.orggeochem.2020.104165>.

539 Gontijo, E.S.J., Monteiro, A.S.C., Rosa, A.H., 2017a. Metal and Metalloid Speciation in
540 Aquatic Environments: Concepts, Techniques and Applications. *Revista Virtual de*
541 *Química* 9, 1910-1929. <https://doi.org/10.21577/1984-6835.20170112>.

542 Gontijo, E.S.J., Watanabe, C.H., Monteiro, A.S.C., da Silva, G.A., Roeser, H.M.P., Rosa,
543 A.H., Friese, K., 2017b. Effects of Fe(III) and quality of humic substances on As(V)
544 distribution in freshwater: Use of ultrafiltration and Kohonen neural network.
545 *Chemosphere* 188, 208-217. <https://doi.org/10.1016/j.chemosphere.2017.08.143>.

546 Gontijo, E.S.J., Watanabe, C.H., Monteiro, A.S.C., Tonello, P.S., da Silva, G.A., Friese, K.,
547 Roeser, H.M.P., Rosa, A.H., 2016. Distribution and bioavailability of arsenic in natural
548 waters of a mining area studied by ultrafiltration and diffusive gradients in thin films.
549 *Chemosphere* 164, 290-298. <http://dx.doi.org/10.1016/j.chemosphere.2016.08.107>.

550 Hatje, V., Pedreira, R.M.A., de Rezende, C.E., Schettini, C.A.F., de Souza, G.C., Marin, D.C.,
551 Hackspacher, P.C., 2017. The environmental impacts of one of the largest tailing dam
552 failures worldwide. *Scientific Reports* 7, 10706. [https://doi.org/10.1038/s41598-017-](https://doi.org/10.1038/s41598-017-11143-x)
553 [11143-x](https://doi.org/10.1038/s41598-017-11143-x).

554 Huangfu, X., Jiang, J., Ma, J., Liu, Y., Yang, J., 2013. Aggregation Kinetics of Manganese
555 Dioxide Colloids in Aqueous Solution: Influence of Humic Substances and
556 Biomacromolecules. *Environ. Sci. Technol.* 47, 10285-10292.
557 <https://doi.org/10.1021/es4003247>.

558 IBAMA, 2015. *Lauto Técnico Preliminar: Impactos Ambientais Decorrentes do Desastre*
559 *Envolvendo o Rompimento da Barragem de Fundão, em Mariana, Minas Gerais.*, 38 p.

560 IGAM, 2015. *Acompanhamento da Qualidade das Águas do Rio Doce Após o Rompimento*
561 *da Barragem da Samarco no distrito de Bento Rodrigues – Mariana/MG.*
562 *Monitoramento da Qualidade das Águas Superficiais do Rio Doce no Estado de Minas*
563 *Gerais. Minas Gerais, Brazil, 66 p.*

564 IGAM, 2018. *Acompanhamento da Qualidade das Águas do Rio Doce Após o Rompimento*
565 *da Barragem da Samarco no distrito de Bento Rodrigues – Mariana/MG.*
566 *Monitoramento da Qualidade das Águas Superficiais do Rio Doce no Estado de Minas*
567 *Gerais. Minas Gerais, Brazil.*

568 Jardim, F.A., Sperling, E.v., Jardim, B.F.d.M., Almeida, K.C.d.B., 2014. Fatores
569 determinantes das florações de cianobactérias na água do Rio Doce, Minas Gerais,
570 Brasil. *Engenharia Sanitaria e Ambiental* 19, 207-218.

571 Kohonen, T., 2001. *Self-Organizing Maps.* Springer Berlin Heidelberg.

572 Li, Q., Xie, L., Jiang, Y., Fortner, J.D., Yu, K., Liao, P., Liu, C., 2019. Formation and stability
573 of NOM-Mn(III) colloids in aquatic environments. *Water Res.* 149, 190-201.
574 <https://doi.org/10.1016/j.watres.2018.10.094>.

575 Marques, L.d.S., Reis, D.A.d., Nascimento, L.P.d., Oliveira, E.G., Santiago, A.d.F., Roeser,
576 H.M.P., 2019. Mobility of metals in river sediments from a watershed in the Iron
577 Quadrangle, Brazil. *Geochimica Brasiliensis* 33, 273-285.
578 <https://doi.org/10.21715/GB2358-2812.2019333273>.

579 McKnight, D.M., Feder, G.L., Thurman, E.M., Wershaw, R.L., Westall, J.C., 1983.
580 Complexation of copper by aquatic humic substances from different environments. *Sci.*
581 *Total Environ.* 28, 65-76. [https://doi.org/10.1016/S0048-9697\(83\)80008-4](https://doi.org/10.1016/S0048-9697(83)80008-4).

582 Monteiro, A.S.C., Parat, C., Rosa, A.H., Pinheiro, J.P., 2016. Towards field trace metal
583 speciation using electroanalytical techniques and tangential ultrafiltration. *Talanta* 152,
584 112-118. <http://dx.doi.org/10.1016/j.talanta.2016.01.053>.

585 Morel, F.M.M., 1983. *Principles of Aquatic Chemistry*. Wiley, New York, NY, USA.

586 Pires, J.M.M., Lena, J.C.d., Machado, C.C., Pereira, R.S., 2003. Potencial poluidor de resíduo
587 sólido da Samarco Mineração: estudo de caso da barragem de Germano. *Revista Árvore*
588 27, 393-397.

589 Ribeiro, P.P.M., de Souza, L.C.M., Neumann, R., dos Santos, I.D., Dutra, A.J.B., 2020.
590 Nickel and cobalt losses from laterite ore after the sulfation-roasting-leaching
591 processing. *Journal of Materials Research and Technology* 9, 12404-12415.
592 <https://doi.org/10.1016/j.jmrt.2020.08.082>.

593 Ritter, K., Aiken, G.R., Ranville, J.F., Bauer, M., Macalady, D.L., 2006. Evidence for the
594 aquatic binding of arsenate by natural organic matter-suspended Fe(III). *Environ. Sci.*
595 *Technol.* 40, 5380-5387. <https://doi.org/10.1021/es0519334>.

596 Rodrigues, A.S.d.L., Malafaia, G., Costa, A.T., Nalini Júnior, H.A., 2014. Iron ore mining
597 promotes iron enrichment in sediments of the Gualaxo do Norte River basin, Minas
598 Gerais State, Brazil. *Environmental Earth Sciences* 71, 4177-4186.
599 <https://doi.org/10.1007/s12665-013-2808-y>.

600 Roeser, H.M.P., Roeser, P.A., 2010. O Quadrilátero Ferrífero - MG, Brasil: Aspectos sobre
601 sua história, seus recursos minerais e problemas ambientais relacionados. *Geonomos* 18,
602 34-37.

603 Rougerie, J., Martins de Barros, R., Buzier, R., Devillers, D., Fondanèche, P., Lissalde, S.,
604 Leblanc, J., Saüt, M., Rebillard, J.-P., Mazzella, N., Guibaud, G., 2021. Diffusive
605 gradients in thin films (DGT): A suitable tool for metals/metalloids monitoring in
606 continental waterbodies at the large network scale. *Sci. Total Environ.* 754, 142147.
607 <https://doi.org/10.1016/j.scitotenv.2020.142147>.

608 Saeed, H., Hartland, A., Lehto, N.J., Baalousha, M., Sikder, M., Sandwell, D., Mucalo, M.,
609 Hamilton, D.P., 2018. Regulation of phosphorus bioavailability by iron nanoparticles in
610 a monomictic lake. *Scientific reports* 8, 17736-17736. [https://doi.org/10.1038/s41598-](https://doi.org/10.1038/s41598-018-36103-x)
611 [018-36103-x](https://doi.org/10.1038/s41598-018-36103-x).

612 Sans-Duñó, J., Cecilia, J., Galceran, J., Puy, J., 2021. Availability of metals to DGT devices
613 with different configurations. The case of sequential Ni complexation. *Sci. Total*
614 *Environ.* 779, 146277. <https://doi.org/10.1016/j.scitotenv.2021.146277>.

615 Santamarina, J.C., Torres-Cruz, L.A., Bachus, R.C., 2019. Why coal ash and tailings dam
616 disasters occur. *Science* 364, 526-528. <https://doi.org/10.1126/science.aax1927>.

617 Santos, O.S.H., Avellar, F.C., Alves, M., Trindade, R.C., Menezes, M.B., Ferreira, M.C.,
618 França, G.S., Cordeiro, J., Sobreira, F.G., Yoshida, I.M., Moura, P.M., Baptista, M.B.,
619 Scotti, M.R., 2019. Understanding the Environmental Impact of a Mine Dam Rupture in
620 Brazil: Prospects for Remediation. *Journal of Environmental Quality* 48, 439-449.
621 <https://doi.org/10.2134/jeq2018.04.0168>.

622 Silva, D.d.C., Bellato, C.R., Marques Neto, J.d.O., Fontes, M.P.F., 2018. Trace Elements in
623 River Waters and Sediments Before and After a Mining Dam Breach (Bento Rodrigues,
624 Brazil). *Química Nova* 41, 857-866.

625 Singhal, R.K., Preetha, J., Karpe, R., Tirumalesh, K., Kumar, S.C., Hegde, A.G., 2006. The
626 use of ultra filtration in trace metal speciation studies in sea water. *Environment*
627 *International* 32, 224-228. <https://doi.org/10.1016/j.envint.2005.08.015>.

628 Smith, J.D., Longmore, A.R., 1980. Behaviour of phosphate in estuarine water. *Nature* 287,
629 532-534. <https://doi.org/10.1038/287532a0>.

630 Terezinha Costa, A., Arias Nalini Jr., H., de Tarso Amorim Castro, P., Carvalho de Lena, J.,
631 Morgenstern, P., Friese, K., 2006. Sediment contamination in floodplains and alluvial
632 terraces as an historical record of gold exploitation in the Carmo River basin, Southeast
633 Quadrilátero Ferrífero, Minas Gerais, Brazil. *Acta Hydroch. Hydrob.* 34, 245-256.
634 <https://doi.org/10.1002/ahch.200400625>.

635 Tonello, P.S., Rosa, A.H., Abreu Jr, C.H., Menegário, A.A., 2007. Use of diffusive gradients
636 in thin films and tangential flow ultrafiltration for fractionation of Al(III) and Cu(II) in
637 organic-rich river waters. *Anal. Chim. Acta* 598, 162-168.
638 <http://dx.doi.org/10.1016/j.aca.2007.07.013>.

639 US EPA, 1992. Method 3005A: Acid Digestion of Waters for Total Recoverable or Dissolved
640 Metals for Analysis by Flaa or ICP Spectroscopy. 5 p.

641 van Leeuwen, H.P., Jansen, S., 2005. Dynamic aspects of metal speciation by competitive
642 ligand exchange–adsorptive stripping voltammetry (CLE–AdSV). *Journal of*
643 *Electroanalytical Chemistry* 579, 337-342.
644 <https://doi.org/10.1016/j.jelechem.2005.03.006>.

645 van Leeuwen, H.P., Town, R.M., Buffle, J., Cleven, R.F.M.J., Davison, W., Puy, J., van
646 Riemsdijk, W.H., Sigg, L., 2005. Dynamic Speciation Analysis and Bioavailability of
647 Metals in Aquatic Systems. *Environ. Sci. Technol.* 39, 8545-8556.
648 <https://doi.org/10.1021/es050404x>.

649 Varcjao, E.V.V., Bellato, C.R., Fontes, M.P.F., Mello, J.W.V., 2011. Arsenic and trace metals
650 in river water and sediments from the southeast portion of the Iron Quadrangle, Brazil.
651 Environ. Monit. Assess. 172, 631-642. <https://doi.org/10.1007/s10661-010-1361-3>.

652 Vergilio, C.d.S., Lacerda, D., Oliveira, B.C.V.d., Sartori, E., Campos, G.M., Pereira, A.L.d.S.,
653 Aguiar, D.B.d., Souza, T.d.S., Almeida, M.G.d., Thompson, F., Rezende, C.E.d., 2020.
654 Metal concentrations and biological effects from one of the largest mining disasters in
655 the world (Brumadinho, Minas Gerais, Brazil). Scientific Reports 10, 5936.
656 <https://doi.org/10.1038/s41598-020-62700-w>.

657 Vesanto, J., 1999. SOM-based data visualization methods. Intelligent Data Analysis 3, 111-
658 126. [https://doi.org/10.1016/S1088-467X\(99\)00013-X](https://doi.org/10.1016/S1088-467X(99)00013-X).

659 Wen, L.-S., Santschi, P.H., Gill, G.A., Paternostro, C.L., Lehman, R.D., 1997. Colloidal and
660 Particulate Silver in River and Estuarine Waters of Texas. Environ. Sci. Technol. 31,
661 723-731. <https://doi.org/10.1021/es9603057>.

662 Zhang, H., Davison, W., 1995. Performance Characteristics of Diffusion Gradients in Thin
663 Films for the in Situ Measurement of Trace Metals in Aqueous Solution. Anal. Chem.
664 67, 3391-3400. <https://doi.org/10.1021/ac00115a005>.

665 Zhang, H., Davison, W., 2015. Use of diffusive gradients in thin-films for studies of chemical
666 speciation and bioavailability. Environmental chemistry (Online) 12, 85-101.
667 <https://doi.org/10.1071/EN14105>.

668

Figure captions

Fig. 1 Sampling sites (S1, S3, S4, S6, S7 and S9), mining areas and tailing dams in Upper Doce River catchment (in Quadrilátero Ferrífero, QF). Mining areas and tailing dams were delineated using cloud-free Landsat 8 imagens taken from path 217 and row 74 on 11th October and 12th November 2015. The images were acquired from the U.S. Geological Survey (USGS) Earth Explorer webpage (<https://earthexplorer.usgs.gov/>, accessed in December 2020). All delineations were performed using the software ArcGis 10.5 (ESRI, USA).

Fig. 2 Concentration (cc) of Co, Cu, Mn, Ni, P and Zn in the total, total dissolved, truly dissolved, and DGT-labile fractions from the six sampling sites (S1, S3, S4, S6, S7 and S9) in Upper Doce River catchment. Not detected and not determined values are depicted as * and +, respectively. These results are available in the Appendix B, supplementary information (SI), tables B.1-B.2.

Fig. 3 Self-organising maps: A) Map of samples, in which samples from six sites (S1, S3, S4, S6, S7 and S9) were clustered in four distinct groups (I-IV) as circled in the map. The sampling season (rainy or dry) when water was collected is indicated in parenthesis and by grey (rainy season) and white (dry season) tones in the neurons (hexagonal units). B) Maps of variables (component planes): the bars beside each map indicate the intensity of each variable (red is high intensity and blue is low intensity). The fractions total, total dissolved, truly dissolved, and DGT-labile are indicated by the letters T, D, F and L (in brackets), respectively. The used Fe data were derived from Gontijo et al. (2016).

