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1 **Interaction of cyanobacteria with calcium facilitates the sedimentation**
2 **of microplastics in a eutrophic reservoir**

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40 Abstract

41 Low-density microplastics are frequently found in sediments of many lakes and reservoirs. The
42 processes leading to sedimentation of initially buoyant polymers are poorly understood for
43 inland waters. This study investigated the impact of biofilm formation and aggregation on the
44 density of buoyant polyethylene microplastics. Biofilm formation on polyethylene films (4 x 4 x
45 0.15 mm) was studied in a eutrophic reservoir (Bautzen, Saxony, Germany). Additionally,
46 aggregation dynamics of small PE microplastics (~85 µm) with cyanobacteria were investigated
47 in laboratory experiments. During summer phototrophic sessile cyanobacteria (*Chamaesiphon*
48 spp. and *Leptolyngbya* spp.) precipitated calcite while forming biofilms on microplastics
49 incubated in Bautzen reservoir. Subsequently the density of the biofilms led to sinking of roughly
50 10 % of the polyethylene particles within 29 days of incubation. In the laboratory experiments
51 planktonic cyanobacteria (*Microcystis* spp.) formed large and dense cell aggregates under the
52 influence of elevated Ca²⁺ concentrations. These aggregates enclosed microplastic particles
53 and led to sinking of a small portion (~0.4 %) of polyethylene microplastics. This study showed
54 that both sessile and planktonic phototrophic microorganisms mediate processes influenced by
55 calcium which facilitates densification and sinking of microplastics in freshwater reservoirs. Loss
56 of buoyancy leads to particle sedimentation and could be a prerequisite for the permanent burial
57 of microplastics within reservoir sediments.

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65 1. Introduction

66 Microplastics (MP) are frequently found in freshwater environments raising concerns about
67 distribution pathways and ecological impacts of this novel contaminant. High loadings of MP are
68 present in lake (Ballent et al., 2016) and reservoir sediments (Di and Wang, 2018), which may
69 act as permanent sinks (Corcoran et al., 2015). The largest share of MP in sediments is often
70 comprised of polyethylene (PE) which has a lower density than water (Ballent et al., 2016). Due
71 to its physical properties this polymer type is expected to stay afloat in the water column instead
72 of settling to the ground (Chubarenko et al., 2016).

73 After release into the environment, MP density, size and shape can be changed by biofouling
74 (Kaiser et al., 2017) or aggregation with natural particles and planktonic cells (Lagarde et al.,
75 2016). The term biofouling describes the attachment of microorganisms (biofilm formation) and
76 macro-organisms to submerged surfaces (Rosenhahn et al., 2010). In the oceans where
77 calcareous macro-foulers such as mussels (Kaiser et al., 2017), bryozoans (Edlin et al., 1975)
78 or barnacles (Fazey and Ryan, 2016) are commonly found on plastics, biofouling may lead to
79 sinking of buoyant polymers within weeks. The conditions in freshwater lakes differ from the
80 marine environment resulting in fouling films dominated by more soft-bodied organisms (Leiser
81 et al., 2020). Still, formation of cyanobacteria dominated biofilms can lead to the sinking of
82 buoyant polymers (Chen et al., 2019). However, the ballasting effects of cyanobacteria (ρ : 0.990
83 to 1.055 g cm^{-3}) are considered being insufficient to sink buoyant MP (Li et al., 2016). Therefore
84 it was hypothesized that the density increase originated from minerals trapped or formed inside
85 the biofilm matrix (Chen et al., 2019).

86 Sessile cyanobacteria as component of aquatic biofilms play a major role in the precipitation of
87 calcite (CaCO_3) (Jansson and Northen, 2010) and the subsequent lithification of biofilms
88 (Macintyre et al., 2000) or formation of biogenic tufa (Zippel and Neu, 2011). The finding that
89 dense biofilms ($\rho_{\text{Biofilm}} > \rho_{\text{Water}}$) may form in lakes (Chen et al., 2019) suggests that sinking of

90 buoyant MP may be facilitated by biogenic calcite precipitation. Whether the ballasting effects of
91 freshwater biofilms are derived from the microbial biomass or from minerals was investigated in
92 a field study.

93 Planktonic cyanobacteria are present in many reservoirs (Li et al., 2016) and lakes (Ortíz-
94 Caballero et al., 2019) forming extensive blooms during late summer. *Microcystis* spp. are the
95 most abundant phototrophs in Bautzen reservoir during July and August (Kamjunke et al.,
96 1997). In lake water *Microcystis* spp. are aggregating to large and sinking colonies (Chen and
97 Lüring, 2020) under the influence of dissolved Ca^{2+} (Xu et al., 2016a). These cell aggregates
98 are often exceeding 500 μm in diameter (Feng et al., 2019), possibly enclosing small inorganic
99 particles (Xu et al., 2016b). The question whether such *Microcystis* aggregates form under the
100 influence of Ca^{2+} and subsequently could enclose or sink buoyant MP was studied in lab
101 experiments.

102

103 Therefore we investigated interaction of sessile and planktonic cyanobacteria with calcium in
104 regards of their impact on the buoyancy of PE in freshwater reservoirs.

105 We hypothesized that i) biofilms form on large PE in a eutrophic reservoir leading to a loss in
106 buoyancy ii) buoyancy loss is caused by calcite precipitation iii) planktonic cyanobacteria sink
107 small PE MP through Ca^{2+} induced aggregation. These hypotheses were tested by conducting a
108 field experiment in the eutrophic Bautzen reservoir (Saxony, Germany) and laboratory batch
109 experiments with calcifying cyanobacteria.

110

111 2. Material and Methods

112

113 2.1 Study site

114 Bautzen reservoir is a freshwater body in the eastern part of Germany, providing water for
115 cooling of coal-fired power stations, fish farming, and agricultural irrigation. It is a large (5.3 km^2)

116 but rather shallow (mean depth 7.4 m) reservoir (Kasprzak et al., 2007), often experiencing
117 strong winds, which sometimes even destroy the summer stratification (Kerimoglu and Rinke,
118 2013). Bautzen reservoir is eutrophic, with extensive blooms of *Microcystis* spp. (Kamjunke et
119 al., 1997) occurring during summer which results in pH values up to 9.5 in the surface water.
120 One sediment trap (Uwitec, Austria) collecting settling matter was deployed near the deepest
121 point of the reservoir (depth: 12 m) during the year 2018 from May to December and sampled in
122 monthly intervals.

123 2.2 Plastic material, exposition, and sampling procedure

124 Squares ($4 \times 4 \times 0.15$ mm) made from low-density polyethylene (Goodfellow ET311251, ρ :
125 0.924 g cm^{-3}) were incubated near the deepest point of the reservoir within a cylindrical
126 stainless steel cage (200 particles, diameter 10 cm, length 25 cm, mesh width 3 mm) (Kettner et
127 al., 2017). The cages had a shading effect which reduced the incoming sunlight intensity by ~33
128 % (Leiser et al., 2020). Particles were sterilized by treatment with ethanol (70 % v/v, 10 min)
129 prior to the experiment. The cage was incubated from July 23 to August 21, 2019 (29 days) in
130 0.5 m depth. Profiles of oxygen concentrations, chlorophyll a, pH, temperature (multi-parameter
131 probe, Sea & Sun Technologies, Germany) were measured on both dates (Figure S1).

132

133 2.3 Biofilm characterization and sampling procedures

134 Particles were gently removed from the cage by tweezers or by flushing with reservoir water.
135 Particles for confocal laser scanning microscopy (CLSM) analysis were fixed in 4 % v/v formalin
136 solution right after removal from the cage. The other particles were stored in pre-combusted
137 ($450 \text{ }^\circ\text{C}$, 4 h) glass Petri dishes in filtered reservoir water. Additional biofilm samples were taken
138 from the inner wall of the cage and stored in reservoir water as well. These samples were used
139 to characterize the biofilms in regards of dry mass, mineral composition and elemental content.
140 The PE particles were in close proximity to or even enclosed by biofilms growing on the inner
141 cage walls. Given their visual appearance biofilms on PE were not different from biofilms from

142 the inner cage walls (Figure S2). Therefore cage walls biofilms were considered being
143 comparable to PE biofilms in regards of the above mentioned parameters. Most samples
144 (except samples for DNA extraction) were stored at 8 °C in the dark until processing. The
145 densities and volumes of three fresh biofilm sub-samples from the inner walls of the cage (300 –
146 400 mg) and three pooled sunken PE particles were analysed with pycnometers at 25 °C. Dry
147 weight (60 °C, 24 h) and ash mass (450 °C, 24 h) were determined for three individual cage
148 walls biofilm samples. The cell volume of microorganisms (biovolume) within the biofilms of ten
149 buoyant and five sinking PE-particles was analysed via CLSM and image analysis. Particles
150 were examined at five (floating) or ten (sinking) random locations resulting in a total sample size
151 of 50 for each. Calcein assay (Zippel and Neu, 2011) was used to visualize Ca carbonate
152 minerals within the biofilms of two sinking PE particles. Twenty particles for 16S amplicon
153 sequencing were carefully rinsed with DNA-free phosphate-buffered saline (pH 7.4) and stored
154 in liquid nitrogen. Calcium content within the cage walls biofilm dry mass was analysed by ICP-
155 OES (detection limit: 0.1 mg l⁻¹). X-ray diffraction was used to analyse the mineral phase of cage
156 walls biofilm dry mass.

157

158 2.4 CLSM imaging

159 Plastic particles with biofilms were prepared for CLSM as described elsewhere (Leiser et al.,
160 2020). In brief: Particles were mounted and stained (SybrGreen, calcein) in Petri dishes.
161 Imaging was done using a TCS SP5X upright microscope equipped with white laser and 63x NA
162 0.9 lens (Leica). Calcein staining (1 µg l⁻¹; 2 h, room temperature) was used to visualize divalent
163 cations such as Ca²⁺ and Ca carbonate minerals (Zippel and Neu, 2011). Calcein staining is not
164 specific for Ca²⁺ ions or calcite, and may also react with other divalent cations present in the
165 solution. The calcein stain did not bind to pristine PE particles (data not shown). PE particles
166 were stained with non-toxic iDye PolyPink following established protocols (Karakolis et al.,
167 2019). Bacteria, algae and cyanobacteria were identified via SybrGreen staining,

168 autofluorescence of chlorophyll *a* or phycobilins, respectively (Table S1 for excitation / emission
169 wavelengths). Images were visualised and projected by Imaris (Bitplane) and presented by
170 Photoshop (Adobe). An adaptation of ImageJ was used to semi-quantitatively calculate the
171 biovolumes of algae, bacteria and cyanobacteria cells (Staudt et al., 2004).

172

173 2.5 Cyanobacteria cultures and laboratory aggregation experiments

174 The effect of dissolved Ca²⁺ on the aggregation of MP with cyanobacteria was investigated
175 using non-axenic cultures of *Microcystis* sp. strain BM25 (Schwarzenberger et al., 2013) grown
176 in WC media (Guillard and Lorenzen, 1972) without vitamin solution and silicate. Pre-cultures
177 were grown on a rotary shaker (110 rpm) at room temperature and ambient daylight for 4 weeks
178 prior to the experiment. PE powder (ρ : 0.920 g cm⁻³; Alfa Aesar 9002-88-4) was sieved through
179 100 μ m and 10 μ m stainless steel sieves (Retsch, Germany) to obtain a defined size range of
180 100 – 10 μ m. The mean equivalent spherical diameter (ESD) of the sieved particles was 85 \pm 14
181 μ m (n: 60). Particles were stained with iDye PolyPink (Karakolis et al., 2019). Three different
182 Ca²⁺ concentrations (10 mg l⁻¹, 60 mg l⁻¹ and, 220 mg l⁻¹) were tested for their potential to
183 aggregate *Microcystis* sp. strain BM25.

184 Experiments were conducted in triplicates by inoculating 500 ml WC-media with 10 % v/v
185 cyanobacteria pre-culture in airtight 1-liter flasks. Cell concentrations (\sim 10⁷ cells ml⁻¹) were
186 chosen to reflect the concentration of *Microcystis* spp. in Bautzen reservoir during summer (\sim 3 -
187 5 \times 10⁷ cells ml⁻¹, data provided by the state reservoir administration of Saxony /
188 Landestalsperrenverwaltung des Freistaates Sachsen (LTV)). Right after inoculation, samples
189 for pH, Ca²⁺ and cell counts were taken. Directly afterwards pH was measured using a pH meter
190 (PP-50, Sartorius). Calcium samples were filtered (0.2 μ m) and stored at 4 °C. Cyanobacteria
191 cells were fixed in Lugol's iodine (5 % w/v iodine) until cell counting. Afterwards 10 mg (6.3 \times
192 10⁴ particles l⁻¹) of PE were added to the flasks. The cultures were then incubated at 23.5 °C
193 under constant light (70 W m⁻²) on roller incubators (10 rpm) until visible aggregates formed.

194 Depending on the Ca^{2+} concentration, aggregates formed within hours to days. Experiments
195 showing no aggregation were stopped after 7 days. Upon termination, samples for Ca^{2+} , pH and
196 cell counts were taken as described above. Aggregates formed within the flasks were
197 photographed and counted employing ImageJ cell counter plugin (Rueden et al., 2017).
198 Furthermore visible aggregates were gently removed using an inverted glass pipette. Twelve
199 aggregates per Ca^{2+} condition (4 per triplicate) were transferred into a coverwell chamber
200 (Thermo Fisher Scientific) for CLSM. Ca^{2+} and fluorescent MP within the sinking aggregates
201 were visualized via CLSM. Density measurements were conducted with 9 aggregates per Ca^{2+}
202 condition (3 per triplicate) in a temperature controlled chamber at 20 °C. Aggregates were
203 transferred to ultrapure water (20 °C) and titrated with NaI (2 g ml^{-1} , $\rho: 1.690 \text{ g cm}^{-3}$) until neutral
204 buoyancy was achieved. The density of the resulting solution was measured with pycnometers.
205 Sinking velocities of 30 individual aggregates (10 per triplicate) were determined in a column
206 (diameter 5 cm) filled with tap water (20 °C) and recorded with a camera (13 megapixel, 30 fps).
207 Afterwards the same aggregates were removed from the column and checked for their plastic
208 content under a light microscope. The ESDs of sinking aggregates were calculated from the
209 recorded images using ImageJ. Sunken biomass/aggregates remaining after aggregate
210 selection and sampling were filtered onto a stainless steel sieve (47 mm, pore size 10 μm) to
211 remove non-aggregated cells. The filters were subsequently rinsed three times with ultrapure
212 water. The biomass was dried (60 °C, 24 h) and analysed for its mineral content using XRD.
213 Cell concentrations were determined by epifluorescence microscopy after SybrGreen staining.
214 Calcium concentrations in the media were measured via ICP-OES.

215

216 2.6 X-ray diffraction

217 X-ray diffraction was performed using a PANalytical Empyrean diffractometer, equipped with a
218 Co-tube, automatic divergence slit and PIXcel 3D detector. Field samples were sieved (< 4 mm)
219 to remove PE squares and filled into 27 mm sample holders. Measurements were performed

220 from 5 to 80 °2θ with a stepsize of 0.0131 °2θ and total measurement time of 2 h 30 min. The
221 irradiated area on the sample was kept constant at 12 × 15 mm by means of respective mask
222 and the automatic divergence slit. Samples from laboratory experiments showing low mass on
223 stainless steel filters were prepared on silicon low background holders and measured under the
224 same conditions. Data were evaluated by the use of PANalytical's HighScore software and the
225 BGMN/Profex package v4.0.2 (Doebelin and Kleeberg, 2015).

226

227 2.7 DNA extraction, Illumina sequencing and bioinformatics

228 Total DNA was extracted using the DNA Power Soil Pro Kit (Qiagen) with modifications. Biofilm
229 covered plastic particles were transferred into the PowerBead Pro Tubes containing 800 µl of
230 solution CD1. The tubes were fixed horizontally to a vortex adapter and shaken for one hour for
231 mechanical disruption of bacteria. Afterwards samples were incubated for one hour with 25 µl
232 proteinase K (22 mg ml⁻¹) at 37 °C. The extraction was continued following the instructions given
233 by the supplier. Libraries, sequencing and data analysis were performed by Microsynth AG
234 (Balgach, Switzerland). To assess the bacterial diversity, the V4-V5 region of the bacterial 16S
235 rRNA gene was amplified by two-step PCR using the primer pair 515F-Y and 926R (Parada et
236 al., 2016). Libraries were sequenced using a v2 500 cycle kit and the Illumina MiSeq platform.
237 The raw data were submitted to the ENA (European Nucleotide Archive) database and were
238 assigned the BioProject ID: PRJEB38919. Standard statistical analysis and bioinformatics were
239 employed to obtain relative abundance of the bacterial phyla (S3). The OTUs assigned to
240 *Cyanobacteria* were further classified using BLAST analysis (Altschul et al., 1990) using
241 nucleotide database (nt/nr) with uncultured and environmental sample sequences excluded

242

243 2.8 Statistical analysis and programs

244 Visual MINTEQ (Version 3.1, Royal Swedish Academy of Science) was employed to calculate
245 the saturation indices of calcite (SI_{calcite}) in Bautzen reservoir for the years 2018 and 2019 using

246 the default thermodynamic database. The dataset used for modelling contained major water
247 anions, cations, pH, temperature and chlorophyll content of Bautzen reservoir and has been
248 provided by the LTV. Data normality was checked via Q-Q Plots and histograms. F-test was
249 used to test for variance homogeneity prior to conducting t-tests and ANOVA. Differences
250 between datasets were seen as statistically significant for $p < 0.05$. Akaike information criterion
251 method was used to select the best fitting multiple linear models. Residual plots were examined
252 for the validity of the linear models. Non-parametric rank-based tests and median statistics were
253 used for non-normally distributed data. R (R Core Team, 2018) was used for statistical analysis
254 and for the graphs.

255

256 3. Results & Discussion

257

258 3.1 Calcite precipitation in Bautzen reservoir

259 Summer blooms of phototrophic microorganisms accompanied by high pH values and the
260 decline of dissolved Ca^{2+} in surface water (Figure S3) were observed in 2018 and 2019. Calcium
261 made up between 0.9 % and 6.3 % of the settling matter during June-July and July-August 2018
262 (Table S2). Assuming that this Ca was present solely in the form of calcite this mineral
263 accounted for $0.24 - 2.23 \text{ g m}^{-2} \text{ d}^{-1}$ or up to 16.7 % of the total settling matter during this time.
264 Bautzen reservoir has a lower potential for calcite precipitation in the surface water compared to
265 lakes such as Baldeggersee (Luzern, Switzerland) producing $10 - 20 \text{ g calcite m}^{-2} \text{ day}^{-1}$
266 (Teranes et al., 1999) or Lake Constance (Switzerland, Germany, Austria) with $14 \text{ g m}^{-2} \text{ d}^{-1}$
267 (Stabel, 1988). Calcite precipitation is of high importance for matter flux in many eutrophic and
268 mesotrophic lakes, whereas being less intense in hyper-eutrophic and oligotrophic water bodies
269 (Koschel et al., 1983).

270

271 3.2 Field biofilms and microplastic biofouling in Bautzen reservoir

272 The incubation cage and PE-particles were covered by dense brownish biofilms after the
273 incubation period of 29 days. Whitish minerals (Figure 1a) covered the biofilms surfaces and
274 calcein stainable minerals were found in close proximity to cyanobacterial cells (Figure 1b).
275 This mineralized appearance of the biofilm was reflected by the high dry mass ($19.5 \pm 3.6 \%$, n:
276 3) and ash content ($91.2 \pm 1.9 \%$, n: 3). Mineral phases of the cage walls biofilms were
277 comprised of 98 % pure calcite and 2 % quartz as shown by XRD analysis (Figure S4). The
278 contents of major elements within the cage walls biofilms dry mass were 0.2 % Al, 31.1 % Ca,
279 0.3 % Fe, 0.4 % Mg, 0.1 % Mn, and 0.12 % Si. As Ca appeared solely in the form of calcite
280 (CaCO_3) this mineral accounted for ~78 % of the biofilms dry mass. The calcite content of
281 Bautzen reservoir biofilms was higher compared to biofilms found in Lake Velence (30 %
282 calcite) (Záray et al., 2005) and the Sanjiadian reservoir (20 – 40 %) (Tianzhi et al., 2014). Thus
283 calcite seems to be a common and major component of biofilms in lakes of different trophic
284 states and water chemistry. The respective wet density of the calcified biofilms was 1.18 g cm^{-3}
285 ± 0.012 (n: 3, 20 °C). Given the similarity of cage walls biofilms and PE biofilms (Figure S2)
286 results might be extrapolated to the MP particles. However, it cannot be excluded that biofilm
287 properties slightly differed, which should be considered while interpreting the results.

288

289 The community composition within the biofilms of buoyant and sunken PE particles was
290 examined via CLSM and 16S amplicon sequencing. CLSM analysis showed that the sunken
291 particles had significantly higher cell volumes of bacteria (0.43 to $0.007 \mu\text{m}^3 \mu\text{m}^{-2}$),
292 cyanobacteria (0.09 to $0.0002 \mu\text{m}^3 \mu\text{m}^{-2}$) and algae (0.03 to $0 \mu\text{m}^3 \mu\text{m}^{-2}$) if compared with the
293 buoyant particles (Figure 2). Cyanobacteria occurred either as filamentous colonies of
294 elongated thin cells which were found in 85 % of the analysed images (n: 50) or as colonies of
295 rounded cells found in 28 % of the images (n: 50). Classification of the 16S rRNA gene
296 sequences revealed a dominance of non-phototrophic bacteria (66 % of all sequences, Figure
297 3). Still a significant abundance (34 %) of cyanobacteria was detected within the calcifying

298 biofilms. Only 12 different bacterial phyla were found on the particles. Such low OTU richness
299 has been described previously as a common feature of microbial biofilms on MP (Amaral-Zettler
300 et al., 2020). The non-phototrophic phyla found in this study have already been described by
301 other authors to colonize MP exposed to river or lake water (Hoellein et al., 2014; Wang et al.,
302 2020; Wu et al., 2019).

303 Within the cyanobacteria Gpl were the most abundant group (26 %), followed by GpV (7 %),
304 and GpIIa (1 %) (Figure 3). Based on further classification of the sequences using BLAST
305 analysis (Altschul et al., 1990) the groups Gpl and GpV could be assigned to *Chamaesiphon*
306 spp. and *Leptolyngbya* spp. According to their morphology, the colonies of rounded cells could
307 belong to *Chamaesiphon* spp. (Kurmayer et al., 2018) whereas the filamentous colonies, which
308 were found in most of the images, resembled *Leptolyngbya* spp. (Arp et al., 2010) cells. Both
309 genera, *Chamaesiphon* spp. (Peraza Zurita et al., 2005) and *Leptolyngbya* spp. (Zippel and
310 Neu, 2011), are common members of calcifying freshwater biofilms (Arp et al., 2010). Especially
311 *Leptolyngbya* spp. has been associated with an increase of SI_{Calcite} and calcite precipitations
312 within stream biofilms (Brinkmann et al., 2015). Further, calcite grains are often found in close
313 proximity or even encrusting cyanobacterial cells (Martinez et al., 2010), which was also
314 observed in our study using calcein staining (Figure 1b). Hence it is likely that calcite was
315 precipitated by cyanobacteria leading to densification of the biofilms and subsequent sinking of
316 the PE particles. Still heterotrophic bacteria might have influenced the calcite precipitation by
317 providing nucleation sites or releasing Ca^{2+} bound to the organic biofilm matrix (López-García et
318 al., 2005). As algae and diatoms were scarce throughout, occurring in only 10 % of the image
319 datasets (n: 50), their influence on calcite precipitation might have been minor.

320 PE particle buoyancy was tested by observing their upward or downward movement in water.
321 Approximately 20 to 30 particles (10 – 15 %) lost their buoyancy at the end of the field
322 experiment. For a minor fraction of PE particles physical disturbance by the sampling procedure
323 led to a certain loss of biofilm and consequently to regaining of their buoyancy. The interior of

324 the cage was covered with biofilms entrapping and hiding some of the PE particles. As a
325 consequence the proportion of sunken particles could not be estimated precisely leading to the
326 conservative number of 20 to 30 sunken particles. The density of the sunken PE-particles was
327 1.19 g cm^{-3} at $20 \text{ }^\circ\text{C}$ (n: 1), which implies a sharp density increase compared to pristine particles
328 ($\rho: 0.924 \text{ g cm}^{-3}$). Biofilm formation has already been reported to sink buoyant MP within 18
329 days in shallow and high productive lakes (Chen et al., 2019). The authors hypothesized that
330 minerals (calcite, clays) trapped within the biofilms rather than the microbial cells induced
331 sinking of the MP (Chen et al., 2019). In the present study a biofilm volume of $2.68 \times 10^{-2} \text{ cm}^3$
332 (n: 1) was bound to the sunken MP from Bautzen reservoir of which only $\sim 1.85 \times 10^{-6} \text{ cm}^3$ (n:
333 50) was accounted for by cells. The main part of the fouling film (collected from cage walls) was
334 comprised of water (around 77 % of the weight), organic material (around 2 %) and inorganic
335 components (around 21 %) with calcite constituting 17 % of the total biofilm weight. Therefore it
336 can be assumed that the contribution of organic matter and microbial cells to the overall biofilm
337 density was minor compared to biogenic calcite.

338 In marine environments similar studies found that buoyant MP will sink within 2 to 6 weeks
339 (Fazey and Ryan, 2016; Kaiser et al., 2017) due to the development of fouling films on their
340 surfaces. Results of different studies are not easily transferable since the effect of biofouling on
341 MP density is related to particle surface to volume ratio, which is influenced by particle specific
342 size and shape (Chubarenko et al., 2016). MP films are more susceptible towards biofouling
343 than fibers or spheres (Chubarenko et al., 2016), while small particles will lose buoyancy faster
344 than large particles (Fazey and Ryan, 2016). However, the effect of biofouling has only been
345 described for large particles, yet. Given by their small size, sub-millimeter MP particles will be
346 colonized by different organisms compared to large plastics (Rogers et al., 2020). Therefore it
347 remains uncertain if findings made for larger plastics can be transferred to small MP (< 1mm).
348 Hence our finding that calcite formation reduces the buoyancy of large PE films might not be
349 extrapolated to particles smaller than 1 mm. Furthermore it should be considered that the used

350 PE films represented only a small part of the different shapes and size classes of MP found in
351 freshwater. As size and shape influence the surface to volume ratio, these parameters have to
352 be carefully taken into account when transferring the results of this study to other types of
353 particles.

354

355 3.3 Calcium, cyanobacteria and MP aggregation in lab experiments

356 Sinking aggregates of *Microcystis* spp. cells formed under the influence of 220 mg l⁻¹ and 60
357 mg l⁻¹ Ca²⁺ after < 2 h and 2 days, respectively, while no aggregates formed under 10 mg l⁻¹
358 Ca²⁺ within 7 days. Each of the experimental approaches reached pH ~ 9.7 at the end of the
359 experiment. Declining of Ca²⁺ concentration was not detected during the experiments. Calcite or
360 other mineral phases were not found within the aggregates employing XRD analysis.

361 PE particles were incorporated into the matrix (Figure S5) or attached to the outer side of the
362 aggregates. Some of the polymer particles were encrusted by calcein stainable matter (Figure
363 4). The aggregates formed in 60 mg l⁻¹ and 220 mg l⁻¹ Ca²⁺ incorporated on average 2 MP
364 particles (mean, n: 60) and subsequently transported them to the bottom of the incubation flasks
365 (Figure 5). Each flask contained approximately 65 aggregates (mean, n: 5) which in sum
366 incorporated ~130 PE particles (~0.4 % of added particles).

367 Aggregation and settling of MP with eukaryotic algae (Lagarde et al., 2016) and diatoms (Long
368 et al., 2015) have been reported before. So freshwater algae *Chlamydomonas reinhardtii*
369 formed dense aggregates (ρ : 1.2 g cm⁻³) with polypropylene MP readily sinking in culture media
370 (Lagarde et al., 2016). The aggregates in our study were denser than water (ρ : 1.1 g cm⁻³ for 60
371 mg l⁻¹ Ca²⁺; 1.05 g cm⁻³ for 220 mg l⁻¹ Ca²⁺) (Figure 5) and slightly exceeded the density ranges
372 previously reported for cyanobacterial aggregates (1.01 -1.05 g cm⁻³) (Li et al., 2016).

373 Aggregate densities might be influenced by strain specific excretion of extracellular polymeric
374 substances (EPS) (Li et al., 2016) or formation of gas vacuoles (Xu et al., 2016a). Furthermore,

375 the density of such aggregates strongly depends on other external factors such as the seasons
376 (Li et al., 2016). Under favourable environmental conditions such as high temperature, nutrient
377 loadings or light intensities (Li et al., 2016) cyanobacterial aggregates may stay afloat instead of
378 sinking to the sediments. Multiple linear regression modelling revealed that sinking speed of the
379 aggregates was dependent solely on their size, with larger colonies settling faster than smaller
380 ones. Aggregates in 60 mg l⁻¹ Ca²⁺ were significantly larger (mean: 1519 µm, n: 30) than the
381 aggregates formed in 220 mg l⁻¹ Ca²⁺ (mean 1262 µm, n: 30). Therefore they were settling
382 approximately 25 % faster than the smaller aggregates (Figure 5). Regarding the longer
383 incubation time (2 days for 60 mg l⁻¹ Ca²⁺ and 1 day for 220 mg l⁻¹ Ca²⁺) this might have been a
384 temporal effect rather than depending on the Ca²⁺ concentration. Considering the sinking
385 velocity of the aggregates (0.0036 m s⁻¹ for 60 mg l⁻¹ Ca²⁺ and 0.0029 m s⁻¹ for 220 mg l⁻¹ Ca²⁺),
386 aggregation with cyanobacteria may transport buoyant small PE to the sediment of Bautzen
387 reservoir within 34 to 42 minutes. However, this might only hold true for MP particles being
388 smaller than the enclosing cyanobacterial aggregates. Large particles (1- 5 mm), such as the
389 PE films used in the field study, are unlikely to be incorporated into the cyanobacterial
390 aggregates (~1 - 1.5 mm). Therefore the results might only be applicable for MP smaller than
391 100 µm.

392 Most likely aggregation of cyanobacterial cells was induced by the elevated Ca²⁺ concentrations
393 used in our study. Ca²⁺ ions are suspected to form bridges between the negatively charged
394 cyanobacteria or (EPS) (Xu et al., 2016a), leading to the formation of cell aggregates (Chen and
395 Lürling, 2020). Furthermore elevated Ca²⁺ concentrations can increase the production of
396 cyanobacteria EPS (Wang et al., 2011), which plays a crucial role in aggregation processes by
397 providing a sticky, flexible and robust matrix in which cells are embedded (De Oliveira et al.,
398 2020). The assumed bridging role of Ca²⁺ can be supported by the presence of calcein
399 stainable matter within the aggregates (Figure 4) connecting the cyanobacterial cells. Ca²⁺

400 concentrations $>20 \text{ mg l}^{-1}$ are required to induce aggregation of cyanobacterial cells (Chen and
401 Lüring, 2020). Bautzen reservoir has a median Ca^{2+} concentration of 35 mg l^{-1} , which lies in the
402 usual range ($10 - 120 \text{ mg l}^{-1} \text{ Ca}^{2+}$) of freshwater bodies (Wang et al., 2011). Correspondingly,
403 large cyanobacterial aggregates / colonies were observed in Bautzen reservoir during the
404 samplings in 2018 and 2019 (Figure S6). This leads to the assumption that the aggregation of
405 cyanobacteria through Ca bridging might be relevant for the fate of small MP in Bautzen
406 reservoir. The Ca^{2+} concentrations used in the lab experiments did not reflect the actual
407 concentrations found in Bautzen reservoir. Still we could show that *Microcystis* spp. aggregate
408 with MP under environmentally relevant Ca^{2+} concentrations (60 mg l^{-1}).

409

410 Field and lab experiments described two distinct processes leading to sinking of buoyant PE
411 microplastics in the context of cyanobacterial interaction with calcium. Apparently the formation
412 of biofilms sank proportionally more particles ($\sim 10 - 15 \%$) than the aggregation of small PE with
413 cyanobacteria. However, the low removal efficiency rather resulted from the low concentration
414 of cyanobacterial aggregates (mean: $130 \text{ aggregates l}^{-1}$) than from the amount of PE
415 incorporated into each of the aggregates (mean: 2). Taking into account that, during
416 cyanobacteria blooms concentrations of $3.5 \times 10^5 \text{ aggregates l}^{-1}$ can be reached (Feng et al.,
417 2019), aggregation governs a high potential for MP removal in productive lakes.

418

419

420 4. Conclusions

- 421 • A proportion of polyethylene microplastics ($\sim 10 - 15 \%$ of particles) lost its
422 buoyancy due to biofilm formation after being exposed for 29 days in a eutrophic
423 reservoir
- 424 • Biofilms were rich in calcite. Apparently the mineral had a greater effect on
425 biofilm density compared to organic matter or cells

- 426 • Cyanobacteria *Chamaesiphon* spp. and *Leptolyngbya* spp. were abundant
427 biofilm members probably facilitating calcite formation in the biofilms
428 • Planktonic cyanobacteria formed sinking aggregates with small polyethylene
429 microplastics (enclosing ~0.4 % of particles) under elevated Ca concentrations
430

431

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435

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