

**This is the author's draft version of the contribution published as:**

**Groeneveld, J.,** Berger, U., Henschke, N., Pakhomov, E.A., Reiss, C.S., Meyer, B. (2020):  
Blooms of a key grazer in the Southern Ocean – an individual-based model of *Salpa*  
*thompsoni*  
*Prog. Oceanogr.* **185** , art. 102339

**The publisher's version including minor corrections in the text and  
used equations is available at:**

<https://doi.org/10.1016/j.pocean.2020.102339>

**Blooms of a key grazer in the Southern Ocean – an individual-based model of *Salpa thompsoni***

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1 Abstract:

2 The Southern Ocean near the Western Antarctic Peninsula (WAP) is strongly affected by  
3 climate change resulting in warmer air temperature, accompanied with reduced sea ice  
4 coverage, increased sea water temperature and potential changes in the abundances of two  
5 key grazer species *Salpa thompsoni* (salp) and *Euphausia superba* (Antarctic krill). While  
6 salp abundance is hypothesized to increase, krill abundance is hypothesized to decline with  
7 dramatic consequences for the entire food web of the Southern Ocean. A better understanding  
8 of the biotic interaction between krill and salps and their population dynamics is thus crucial.  
9 However, the life cycle of salps is complicated and barely understood. Therefore, we have  
10 developed an individual-based model describing the whole life cycle to better understand the  
11 population dynamics of salps and the conditions for blooms. The model has been used to  
12 explore if and under what conditions the empirical pattern of large variability in observed  
13 salp abundances at the WAP, generated by the long-term data of the US Antarctic Marine  
14 Living Resources Program (AMLR) can emerge from a small seeding population. The model  
15 reproduced this empirical pattern if daily growth rates of oozoids were higher than previously  
16 reported for the WAP (mean growth rate for oozoids  $\sim 1 \text{ mm d}^{-1}$ ) and if growth rates of  
17 blastozooids were lower (mean growth rate  $\sim 0.2 \text{ mm d}^{-1}$ ). The model suggests that a  
18 prerequisite for local salp blooms requires a small founding population in early spring. With  
19 climate change it has been suggested that more frequent and earlier transport of salps into the  
20 WAP or winter survival will occur. Hence, the risk of salp blooms in the WAP is likely to  
21 substantially increase. These findings highlight the importance for an improved quantitative  
22 understanding of how primary production and the southward advection of salps will be  
23 impacted by climate change.

24

25 Key words: Individual-based model, simulation, *Salpa thompsoni*, climate change, Western  
26 Antarctic Peninsula

27

28 Introduction:

29 The pelagic tunicate, *Salpa thompsoni* (salp hereafter), and the Antarctic krill, *Euphausia*  
30 *superba* (krill hereafter), are key grazers in the Southern Ocean and important elements of the  
31 marine Antarctic food web (Atkinson et al. 2004). Historically salps and krill occupied  
32 different areas of the Southern Ocean. While salps were more abundant in the food poorer  
33 and warmer waters north of the Antarctic Polar Front, krill were more abundant in colder and  
34 more productive areas south of the Antarctic Polar Front (Foxton 1966, Pakhomov and Hunt  
35 2017). With rising sea water temperatures and changes in the sea ice dynamics in the  
36 Southern Ocean, however, salps may expand their range farther south and interact more often  
37 with krill (Atkinson et al. 2004). The fact that salps have a sexual and an asexual  
38 reproduction cycle that allows them to reproduce explosively results occasionally in very  
39 high salp densities also called blooms (e.g. mean densities up to 3.5 individuals m<sup>-3</sup> and  
40 higher values in recent years > 18 individuals m<sup>-3</sup> are reported; Loeb and Santora 2012). Salp  
41 blooms may severely reduce the food availability for krill by consuming a substantial amount  
42 of the primary production (Dubischar and Bathmann 1997). In addition, salps can consume  
43 young life stages of krill directly and may therefore increase the mortality rates on krill  
44 populations (Huntley et al. 1989). The potential increase of competition between krill and  
45 salps may have substantial impact on the krill-dependent Antarctic food web. In contrast to  
46 salps, krill are a preferred prey for many marine mammals and seabirds, and are a significant  
47 commercial fishery (Nicol and Foster 2016). Although the contribution of salps as a food  
48 resource is not well understood (Pakhomov et al., 2002, Henschke et al. 2016) an increase in  
49 salp densities may put the krill-dependent Antarctic food web at risk.

50

51 Currently, the population dynamics of salps, particularly in a changing environment, are not  
52 well understood and the development of population models for salps and pelagic tunicates  
53 have just begun (Henschke et al. 2015, Henschke et al. 2018). Modelling the population  
54 dynamics of salps needs to reflect the large variability of salp abundance in space and time. If  
55 environmental conditions are suitable, oozoids show a massive asexual reproduction by  
56 releasing chains of genetically identical blastozoids (Fig. 1). In contrast, each blastozoid  
57 can produce one oozoid embryo during the sexual part of the reproductive cycle. Earlier  
58 studies have reported the reproductive cycle (both sexual and asexual parts) lasts between 9  
59 and 12 months, suggesting a single annual production period (Foxton 1966, Loeb and Santora  
60 2012). However, recently Pakhomov and Hunt (2017), using an eulerian study, showed that  
61 salps were capable of completing two reproductive cycles per year. If this were possible, the

62 per capita growth rate would increase significantly given that one single oozoid may release  
63 under optimal conditions more than 800 blastozooids (Daponte et al. 2001). Clearly,  
64 resolving the limits of the salp life history is critical to projecting future impacts on the  
65 pelagic ecosystem of the Southern Ocean.

66

67 To improve our understanding of the population dynamics and salp bloom risk, we have  
68 developed an individual-based population model for salps. In contrast to previous pelagic  
69 tunicate models, our model aims to reproduce the broad distribution of observed salp  
70 abundances at the AMLR sites (US Antarctic Marine Living Resource Program) ranging  
71 from 0 to 44.000 individuals  $1000\text{ m}^{-3}$  (Fig. 2a) instead of modelling the average density  
72 (Henschke et al. 2018). The US Antarctic Marine Living Resources Program (AMLR) at the  
73 WAP offers a unique long-term data set of salp and krill observations and chlorophyll *a*  
74 measurements among other variables (Fig. 2). It covers a large spatial extent from the  
75 Brainsfield Strait to the Elephant Island area (see Reiss et al. 2008 and Loeb et al. 2010 for  
76 more details and a map of the survey grid) between 1996 and 2011. Correlating chlorophyll *a*  
77 and salp counts is challenging because it is often impossible to disentangle positive effects of  
78 food on growth and negative effects on food provision due to grazing. The observational data  
79 suggest (Fig. 2c) that high salp densities limit chlorophyll *a* availability although there is  
80 large uncertainty. The data also show that salp blooms cannot be found where the summer  
81 water temperatures are low (below  $-0.5^{\circ}\text{ C}$ , Fig. 2d). The individual-based model includes the  
82 feedback between primary production and salp abundance to limit salp population growth as  
83 well as individual variability to better understand the emergence of salp bloom years. We  
84 calibrated the model using two population level patterns: 1) the intra-annual distribution  
85 pattern of salps 2) and the observed large variability of population densities in space and time  
86 at the AMLR sites.

87 From the simulations we observed the population densities, the number of completed life  
88 cycles in a season, the oozoid vs. blastozooid ratio, the time when a founding population  
89 migrated successfully into the simulation area and the individual daily growth rate in length.

90

91 Methods:

92 We calibrated the model to reproduce the variability in salp abundance data from the US  
93 Antarctic Marine Living Resource (AMLR) program run by the National Oceanic and  
94 Atmospheric Administration (NOAA). During the AMLR program, abundance data of krill  
95 and salps as well as environmental data (e.g. chlorophyll *a*) were annually collected around  
96 the Western Antarctic Peninsula (WAP) between 1996 and 2011 (see Loeb et al. 2010 for  
97 details). Surveys were conducted during the austral summer (January-March). We compared  
98 our model results with the distribution of all salp abundance observations which ranges over  
99 four orders of magnitude (Fig. 2a). We differentiate two model scenarios related to food  
100 availability: 1) Constant food scenario: Each year the same peak chlorophyll *a* concentration  
101 will be reached if no grazing would occur (mean of the observed chlorophyll *a* concentrations  
102 from the AMLR sites). 2) Variable food scenario: The observed chlorophyll *a* concentrations  
103 (Fig. 2 b) are used to derive a lognormal distribution that is used in the model to determine  
104 the peak chlorophyll *a* concentration in summer (see sub-model “Chlorophyll dynamics” for  
105 more details).

106

107 Model description:

108 The model description follows the Overview, Design and Details protocol suggested by  
109 Grimm et al. (2006, 2010).

110

111 *Purpose*

112 The purpose of the individual-based model is to describe and project the population dynamics  
113 of salps in the Antarctic Peninsula regions of the Southern Ocean. Specifically, we quantify  
114 the patterns of intra-annual abundance distribution, the inter-annual fluctuations, the number  
115 of reproductive cycles per year and the ratio between the oozoid stage and the blastozoid  
116 stage. In this study we focus on the local population dynamics on similar spatial resolution as  
117 empirical samples have been taken (several 1000 m<sup>3</sup>) in order to understand and predict the  
118 enormous variability of salp abundances in space and time in the context of environmental  
119 changes as observed in nature. We include three sources of variability which we hypothesize  
120 that could lead to the emergence of the observed abundance distributions: 1) demographic  
121 stochasticity, 2) variability in food availability and 3) variability in the initial seeding  
122 population, i.e. whether and when a small seeding population has been transported into the  
123 simulated arena. Furthermore, the model should reflect the measured distribution of daily  
124 length growth and therefore a pattern on the individual level. Since the model will be

125 eventually linked to a krill model in the near future, the long-term purpose of the salp model  
126 is also to develop a simulation tool for assessing the potential interaction of the two key  
127 grazers of the Southern Ocean.

128

### 129 *Entities, state variables, and scales*

130 The model has four entities: patches, and three kinds of salp individuals: oozoids, chains of  
131 female blastozoids and male blastozoids. Patches have the state variable *chla* (density of  
132 chlorophyll *a* measured in mg chlorophyll *a* m<sup>-3</sup>) and the state variable *pchla* (total amount of  
133 chlorophyll *a* in one patch measured in mg chlorophyll *a*) as a proxy for the available food  
134 for salps. Oozoids, chains and male blastozoids have some state variables in common such  
135 as: length *l*, age *a*, number of individuals *n* (which is the number of blastozoids in one chain  
136 and always one for oozoids and male blastozoids), and the accumulated number of days  
137 without sufficient food, *t<sub>starv</sub>*. In addition, blastozoids are sex specific, whereas female  
138 blastozoids form chains. The temporal resolution in the model is one day. The simulation  
139 runs over several years to compile time series. In this paper, the overall run time was 300 000  
140 days. The spatial resolution is 16 cubic meters and the simulation arena is a layer of 13 × 13  
141 = 169 patches resulting in a simulated volume of 2,704 m<sup>-3</sup>. Thus, we assume that each patch,  
142 which is the NetLogo term for the smallest spatial resolution in the model, is representing a  
143 volume of 16 cubic meters like a cube with edge length of roughly 2.5 m. The amount of food  
144 in one patch *pchla* is therefore  $pchla = 16 \text{ m}^3 \times chla \text{ mg chlorophyll } a \text{ m}^{-3}$ . The simulated  
145 world is a torus, i.e. individuals that leave the simulation arena at one edge will reappear from  
146 the opposite edge. This is a technical procedure to avoid edge effects and does not aim to  
147 represent the real transport processes. Underlying these periodic boundary conditions is the  
148 assumption that the simulation arena is part of a larger spatial domain with the same  
149 environmental conditions and salp densities. The model has been implemented in NetLogo  
150 version 6.0.3 (Wilensky 1999).

151

### 152 *Process Overview and Scheduling*

153 An overview of all processes in the model can be found in Figure 3. Each daily time step  
154 begins with all salps exploiting the resources at their location (patch). Next, the daily growth  
155 in body length is determined for each salp followed by an asexual or sexual reproduction  
156 event if the individual has reached a particular length. Then, the mortality of salps is applied,  
157 followed by updating primary production for each patch. If salp abundance is zero in the  
158 simulation arena, some immigration of salps from outside is allowed with an immigration

159 probability. To address the uncertainty of the immigration probability we have run the model  
160 for different immigration probabilities ranging from 0.001 to 1. Finally, each salp moves  
161 randomly to one of its neighboring patches representing a random walk. All procedures are  
162 repeated every time step for each salp. During the course of the year primary production and  
163 the vonBertalanffy growth rate in body length changes whereas all other parameters remain  
164 constant during the year (e.g. mortality).

165

#### 166 *Design principles*

##### 167 Basic principles:

168 Growth in body length depends on the available food, temperature, and the current body  
169 length. If food is not limited, the growth follows a vonBertalanffy growth curve as in  
170 Dynamic Energy Budget models (Jager et al. 2013). Apart from that, we model the following  
171 fundamental processes feeding, death, birth and immigration. We do not explicitly model  
172 emigration. We model population dynamics on a local scale matching the resolution of single  
173 empirical observations to better understand under which conditions the large variability of  
174 observed salp densities emerges on the observational resolution, while the overall extent of  
175 the AMLR survey is in the order of hundred thousand km<sup>2</sup> (Reiss et al. 2008).

176

##### 177 Emergence:

178 The main patterns of intra-annual abundances, inter-annual fluctuations in abundance, body  
179 length growth rate distributions, and number of reproductive cycles will emerge from local  
180 path dependent indirect interactions between salps through their resource.

181

##### 182 Sensing:

183 Salps sense the amount of food in the patch they are located in, but only move based on  
184 random movements, and only a single patch per day.

185

##### 186 Interaction:

187 Salps interact indirectly via the food resource.

188

##### 189 Observations:

190 During the simulation, the abundances of salps are accumulated for each month. For each  
191 time step the ratio between blastozooids and oozoids is calculated and recorded (only if  
192 oozoids are around to avoid division by zero). Also, the daily body length growth rates of

193 oozoids and blastozooids are stored in a list, but only for the first 5000 days to reduce  
194 memory allocation. The peak abundance during a year is also stored. Daily abundances are  
195 stored during the summer months (January to March) over all years to compare this  
196 distribution with the empirical abundance distribution (see Fig. 6). Finally, the number of full  
197 reproductive cycles is stored for each year and the day of the year when migration into the  
198 simulation arena occurred.

199

200 Initialization:

201 The simulation starts during midsummer (January 1), with a daily time step. Since most  
202 simulations run for 300000 consecutive days the choice of the starting date does not affect the  
203 results. At the initiation of the simulation the state variable *chla* is initialized for all patches  
204 with a value of 0.5 mg chlorophyll *a* m<sup>-3</sup>. Two oozoids are created with length 2 cm assuming  
205 that these oozoids have not released a chain yet. We initialize the model with oozoids since  
206 there is evidence that they are longer lived (up to 24 months, Loeb and Santora 2012) than  
207 blastozooids (up to 7 months, Loeb and Santora 2012) and that oozoids may overwinter at  
208 larger depths (Loeb and Santora 2012). We do not explicitly model fertilization and assume  
209 that single male blastozooids will be present to fertilize the first cohort of released chains.

210

211 *Submodels*

212 Growth

213 For each salp (each female blastozooid in a chain is modelled as an individual) in the  
214 simulated arena the growth is determined for each daily time step. Salps located in the same  
215 patch compete for the resource. Potential food uptake (equation 1) of individual salps, *ing*, is  
216 proportional to their surface, *l*<sup>2</sup>, and food availability, *f*, (with *chla*: chlorophyll *a* density in a  
217 grid cell in mg chlorophyll *a* m<sup>-3</sup>, and *K* the half saturation constant) following a Holling type  
218 II functionality (equation 2) and a factor, *g*, describing the maximum area specific  
219 assimilation rate (see Table 1 for an overview of all parameters and their elasticities):

$$220 \quad ing = gf \times l^2 \quad (1)$$

$$221 \quad f = \frac{chla}{chla+K} \quad (2)$$

222 Parameters (*g* = 0.0025 mg chlorophyll *a* cm<sup>-2</sup> and *K* = 0.2 mg chlorophyll *a* m<sup>-3</sup>) have been  
223 parameterized that modelled ingestion rates are in good agreement with ingestion rates  
224 reported in the literature (von Harbou 2009, p. 108 and p. 205, Fig. S3). As potentially many  
225 individuals compete for the available resource we first calculate the overall food demand in a

226 given patch,  $F_{dem}$ . If the food demand exceeds the available food,  $F_{supply}$ , each food uptake is  
227 reduced by  $F_{supply}/F_{dem}$  to avoid consuming more resources than are available and to model  
228 the interaction as scramble competition in contrast to contest competition (Begon et al. 2006).  
229

230 The vonBertalanffy growth rate is modelled temperature dependent following the Arrhenius  
231 relation:

$$232 \quad r_B(T) = r_{B,ref} \times \left( \frac{T_A}{T_1} - \frac{T_A}{T} \right) \quad (3)$$

233 Where,  $T$ , is the temperature in Kelvin,  $T_1$ , is the reference temperature where  $r_B(T_1) = r_{B,ref}$ ,  
234  $T_A = 8000$  K at  $T_1 = 275$  K is the Arrhenius temperature which is good agreement with  
235 reported  $Q_{10} = 2.8$  (Iguchi and Ikeda 2004). Temperature varies between  $2^\circ\text{C}$  and  $-2^\circ\text{C}$  from  
236 summer to winter (see section ‘primary production’ for more details). The parameter  $r_{B,ref}$  is  
237 different for oozoids and blastozooids ( $r_{B,ref,oozoids} = \text{d}^{-1}$ ,  $r_{B,ref,blastozooids} = 0.0155 \text{ d}^{-1}$ ).

238 Parameter values are calibrated pattern oriented (see section “parameter calibration” for  
239 further details). The growth in body length is modelled following the vonBertalanffy growth  
240 curve (equation 4) with functional response,  $f$ , given in equation 2. This functional response  
241 may be adjusted if the demand is higher than the supply by the factor  $F_{supply}/F_{demand}$  as  
242 outlined above. The only inputs are the current body length,  $l$ , and the density of chlorophyll  
243  $a$  in the patch. The parameter,  $l_{max}$ , gives the asymptotic body length that is assumed to be  $l_{max}$   
244  $= 14$  cm for oozoids and  $l_{max} = 5$  cm for blastozooids (for individual blastozooids not for  
245 chains) for both sexes.

246

$$247 \quad \frac{\Delta l}{\Delta t} = r_B(T) \times (f \times l_{max} - l) \quad (4)$$

248

249 If the increase in length based on equation 3 is positive the length of the individual will be  
250 updated. If it is negative the length of the individual will not change, but the state variable,  
251  $t_{starv}$ , will be increased by one. Finally, the food in the patch,  $pchla$ , will be reduced by each  
252 individual by,  $ing$ , given in equation 1. If the  $chla$  density ( $chla = pchla / 16 \text{ m}^{-3}$ ) in a patch  
253 drops to zero it will be set to  $chla = 0.005$  mg chlorophyll  $\text{a m}^{-3}$ . This reflects some small  
254 inflow from outside of organic matter without explicitly modelling a computational  
255 expensive diffusion process.

256

257 Reproduction:

258 The reproductive cycle of salps is relatively complex. Oozoids asexually produce multiple  
259 chains of blastozooids that can lead to massive recruitment events. In the model we assume  
260 that oozoids start reproduction when they have reached a body length of  $l = 6$  cm based on  
261 empirically estimates of first block releases that ranges from 55 mm to 70 mm (Foxton 1966,  
262 von Harbou 2009). Foxton (1966) concluded from his empirical work that four chains with  
263 750-800 aggregates could be released. This is in accordance with the empirical findings of  
264 Daponte et al. (2001) who suggested that under optimal conditions more than 800 aggregates  
265 could be released in up to five chains by one oozoid based on specimen ranging from 13 mm  
266 to 95 mm in size. Given this empirical evidence we have implemented a length dependence  
267 reproduction process such that each oozoid releases their first chain with 150 blastozooids  
268 when it has grown to 6 cm or larger. Thus, the second chain with 180 blastozooids will be  
269 released if oozoid length is 7 cm, the third chain with 210 blastozooids will be released if  
270 oozoid length is 8 cm, and the final fourth chain with 240 blastozooids will be released if the  
271 oozoid length is 9 cm. After the release of the fourth chain the oozoid dies in the model  
272 assuming that the reproductive inactive oozoid will have a high mortality and to reduce  
273 computational operations. Not all oozoids release four chains, since some will die before or  
274 never reach the needed body length. In the model the process of chain release will take more  
275 than four weeks from the first release to the final chain release. Chains consist of female  
276 blastozooids. We assume that all female blastozooids will be fertilized. Thus, each female  
277 blastozooid can give birth to one oozoid by sexual reproduction. This is implemented in a  
278 way that each female blastozooid is releasing one oozoid once it has reached the body length  
279 of  $l = 2.5$  cm following the observations by Foxton (1966) who found that most embryos are  
280 released by blastozooids in the size class of 25-30 mm. We assume that 70% of all embryos  
281 survive based on estimated proportions of failed embryos observed across various latitudes  
282 and temperatures (Henschke et al. 2018, Henschke et al. 2019). After releasing the embryo,  
283 the blastozooid changes sex from female to male and fertilizes other female blastozooids  
284 which we do not explicitly model. At this stage we assume that the chains break up and that  
285 male blastozooids act as individuals.

286

## 287 Mortality

288 Apart from embryo mortality described in the reproduction paragraph, mortality can occur in  
289 three different ways in the model namely by predation, age, and starvation. Predation is  
290 implemented as a daily mortality,  $m = 0.025$ . There are no direct measurements of mortality  
291 for *Salpa thompsoni*, but length dependent extrapolations for daily mortality based on

292 experimental work for the smaller salp species *Thalia democratica* have been suggested  
293 (Henschke et al. 2015). Together with considerations of negative density dependence daily  
294 mortality rate,  $m \sim 0.03 \text{ d}^{-1}$ , have been used in a previous salp model (Henschke et al. 2018).  
295 Age related mortality is introduced to avoid immortality and is prescribed and set at 500 days,  
296 mostly due to a lack of information regarding maximum life span. This prescribed maximum  
297 age-related mortality is not a significant source of mortality. Oozoids also die if they have  
298 released four chains. Finally, individuals also die if they have not met their food demands for  
299 a number of days specified by the parameter,  $t_{starv} = 30$  days. Whether food demand is met or  
300 not is determined by the vonBertalanffy growth equation (1), if the calculated growth  
301 increment is negative we assume that the food demand has not been fulfilled. Parameter  
302 values have been chosen that the seasonal abundance patterns (observed abundances in  
303 summer and very low abundances in winter) are well represented and that they are in the  
304 ranges of other modelling studies (Henschke et al. 2018). With the given parameterization  
305 simulated local abundances go to zero during winter.

306

307 Primary production:

308 To allow for an interaction between foraging salps and primary production we have included  
309 a patch-based conceptual primary production that follows a logistic equation (5) where the  
310 growth rate  $r_{algae}$  depends on the time of the year (6). As in other population models the  
311 conceptual primary production model is based on a logistic equation and the parameters have  
312 been calibrated that the observed chlorophyll *a* content matches the observations (e.g.  
313 Ryabov et al. 2017). The change in *chl a* in equation 5 depends on the carrying capacity  $K =$   
314  $0.675 \text{ mg chlorophyll } a \text{ m}^{-3}$ , the maximum growth rate,  $r_{algae,max} = 0.25 \text{ day}^{-1}$ , and the decay  
315 rate,  $d = 0.05 \text{ day}^{-1}$ . The carrying capacity has been chosen such as that the peak chlorophyll  
316 *a* content in the model matches the average observed chlorophyll *a* content  $chl a = 0.54 \text{ mg}$   
317 chlorophyll *a*  $\text{m}^{-3}$ . In the scenario where we base our assumption on the chlorophyll  
318 availability of the AMLR data we draw the expected *chl a* peak abundance,  $N^*$ , from a  
319 lognormal distribution with  $\text{meanlog} = 3.83$  and  $\text{sdlog} = 0.58$  derived from a maximum  
320 likelihood estimate of the measured AMLR chlorophyll data presented in Figure 2 b.  
321 Carrying capacity is then determined by  $K = N \times (1 - d/r_{algae,max})$ . Doing so the peak *chl a*  
322 content will match  $N^*$  in the absence of salp predation. The seasonal fluctuation of the  
323 growth rate  $r_{algae}$  is described in equation 6. The maximum growth rate  $r_{algae,max} = 0.25 \text{ day}^{-1}$   
324 and,  $vd$  (in units of days), parameterizes the time before new year's day when primary  
325 production is at its maximum.

326

$$327 \quad \frac{\Delta chla}{\Delta t} = r_{algae}(t) \times chla \times \left(1 - \frac{chla}{K}\right) - d \times chla \quad (5)$$

328

$$329 \quad r_{algae}(t) = r_{algae,max} \times \left(0.5 \times \cos\left(\frac{t+vd}{365} \times 360\right) + 0.5\right) \quad (6)$$

330

331 Temperature is modelled in a similar conceptual way to reflect intra-annual changes in the  
332 simulation arena with a summer temperature of 2°C (see Fig. 2d for the distribution of  
333 observed temperatures between -1.6°C and 4.3°C and a mean of 1.6°C) and a winter low of -  
334 2°C.

$$335 \quad T(t) = \cos\left(\frac{t}{365} \times 360\right) \times 2 + 273 \quad (6)$$

336 Temperature, T, is measured in Kelvin to be consistent with the Arrhenius relation in  
337 equation 3 and annual temperature varies between its maximum and minimum for four  
338 degrees Kelvin.

339

#### 340 Immigration

341 During winter the number of salps is getting low and due to demographic stochasticity and  
342 starvation local extinction usually occurs in the model. If that happens we model stochastic  
343 migration into the local model arena. Therefore, each time step we draw a random number  
344 from a uniform distribution [0,1). If the random number is below the immigration probability  
345 ( $p_i = 0.0085$ ) we allow  $n_l = 10$  oozoids (body length  $l_l = 3$  cm) to migrate into the model  
346 arena. The migration probability has been chosen after systematic variation of this parameter  
347 and comparing with the observed data (see Fig. 6). The chosen immigration probability  
348 results in migration events every 118 days on average. The size of the seed population of 10  
349 has been chosen to be able to investigate the question whether local salp blooms can be  
350 caused by small initial populations that undergo explosive population growth. The elasticity  
351 analyses (Table 1) shows that the results do not change substantially if the number of  
352 immigrants is slightly varied. The immigrating individuals are randomly distributed in space.

353

#### 354 Movement

355 Movement of salps within the model domain is implemented as a random walk to allow salps  
356 to distribute themselves in space. Therefore, all salps are randomly moved each time step  
357 (day) to one of their eight neighboring patches.

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## Input

In the variable food scenario, we read in a time series of chlorophyll *a* values generated from a lognormal distribution that was derived from the observed data (meanlog = 3.83 and sdlog = 0.58).

## Parameter calibration and Elasticity

The model has been calibrated to reproduce the intra- and inter-annual abundance patterns on the population level. Mortality related parameters have been chosen that the intra-annual pattern of very low abundances during winter and variable abundances in summer could be met. Suitable ranges of the vonBertalanffy growth rates have been determined using the BehaviorSearch tool using genetic algorithms (BehaviorSearch v1.10, Forrest Stonedahl & Uri Wilensky 2010) with an earlier model version to match simulated peak abundances and the observed abundance data. Final calibration of the vonBertalanffy growth rate for oozoids,  $rb_{\text{oozoid}}$ , has been done by choosing a value that matches the median of the observed salp abundances well. The Elasticities of the parameters and their values are given in Table 1. Elasticities report the relative change in a response variable if parameters change. Therefore, we changed all parameters one at a time for  $\pm 10\%$ . Elasticities are computed as  $E = (y(p_{\text{max}}) - y(p_{\text{min}})) / y(p_{\text{ref}}) / 0.2$ , with  $p_{\text{min}}$  is the parameter decreased by 10%,  $p_{\text{max}}$  is the parameter increased by 10% and  $p_{\text{ref}}$  is the reference value. The response  $y$  has been averaged over 10 simulations. We use the maximum abundance during the simulation and the median of seasonal peak abundances as the response variables (see Table 1). The simulation results are most sensitive to changes in the parameters describing the growth rates in body length (oozoid vonBertalanffy growth rate  $rb_{\text{ref,oozoid}}$ , blastozoid vonBertalanffy growth rate  $rb_{\text{ref,blastozoid}}$ , see Table 1), the functional response (half saturation constant  $K$ , see Table 1), and the primary production (Rate of primary production  $r_{\text{algae,max}}$ ). This highlights the importance of an improved understanding of the physiological processes and primary production especially in a changing environment where empirically measured relationships may change. In addition, changes in daily mortality also change the simulation results substantially (daily mortality  $m$ , see Table 1). Immigration probability ( $p_i$ ) does not affect the simulated peak abundance, but the median peak abundance. This finding reflects that the maximum abundance in the model is capped by the food resource.

392 Results:

393 The explicit spatial nature of the model shows how strongly the abundance estimates and  
394 their variability depend on the resolution of the observations. A snapshot of the model is  
395 presented in Figure 4 for very high salp densities where the mean abundance is 20 salps  $\text{m}^{-3}$   
396 on the extent of the simulation arena (2704  $\text{m}^{-3}$ ). However, the variability of salps abundances  
397 on the patch level is between 0 (2% of all patches are empty) and 9920 individuals per patch  
398 (= 62 individuals  $\text{m}^{-3}$ ). The fact that female blastozooids are aggregated in chains intensifies  
399 this spatial variation. This resolution dependent variability can be shown in a similar way for  
400 the chlorophyll *a* content. The mean concentration is 0.56 mg chlorophyll *a*  $\text{m}^{-3}$  in the  
401 snapshot (Fig. 3b), but ranges from 0.14 mg chlorophyll *a*  $\text{m}^{-3}$  to 0.86 mg chlorophyll *a*  $\text{m}^{-3}$  in  
402 exploited and unexploited patches, respectively.

403

404 The model is able to reproduce empirical patterns such as the ratio between oozoids and  
405 blastozooids (both sexes) (Fig. 5a). Since each oozoid can produce several hundred  
406 blastozooids during asexual reproduction the blastozooid to oozoid ratio increase as high as  
407 5000 in the constant food scenario and to 19174 in variable food scenario. However, the  
408 median value for the blastozooid to oozoid ratio is 43 and 4 for the constant food scenario  
409 and the variable food scenario matching observed ratios between 4 and 100 (Pakhomov and  
410 Hunt 2017). In the constant food environment daily growth in length of blastozooids is also  
411 very variable ranging from 0  $\text{mm d}^{-1}$  to 0.5  $\text{mm d}^{-1}$  with a mean of 0.2  $\text{mm d}^{-1}$ . Daily growth  
412 of oozoids is substantially larger varying from 0  $\text{mm d}^{-1}$  to 2.3  $\text{mm d}^{-1}$  with an average of 1  
413  $\text{mm d}^{-1}$ . In the food variable scenario the growths rates are more variable for oozoids (0 – 2.8  
414  $\text{mm d}^{-1}$ , average: 1.2  $\text{mm d}^{-1}$ ) and blastozooids (0 – 0.6  $\text{mm d}^{-1}$ , average: 0.2  $\text{mm d}^{-1}$ ). The  
415 simulated intra-annual abundance distribution is consistent with overall expectation that salps  
416 are mostly present in the Southern Ocean during austral summer, as found in observations  
417 (Foxton 1966, Atkinson et al. 2017). In the model, the number of reproductive cycles (Fig.  
418 5d) ranges from 0 (reproductive failure) to 2.5. Simulated salp abundances during the  
419 summer months vary over four orders of magnitude, similar to the observations in the AMLR  
420 data. The distribution of salp abundances depends on the immigration probability (Fig. 6a). In  
421 the constant food scenario the observed maximum abundances cannot be reproduced (Fig.  
422 6a). For the variable food scenario the maximum observed abundances can be simulated and  
423 the results are less sensitive to changes in the immigration probability matching the observed  
424 distribution well (Fig. 6b). Simulated peak abundances vary substantially between years (see  
425 Fig. S2) and the resulting peak abundance in a season depends very much on the time when

426 the seeding population migrates into the local model arena (Fig. 7a). Only if the seed  
427 population migrates into the domain during early spring the very high abundances can be  
428 achieved the following summer (Fig. 7a). If we define a salp bloom as abundances larger or  
429 equal to the 95% quantile of observed salp abundances ( $q = 1.4 \text{ salps m}^{-3}$ ) we can simulate  
430 the bloom risk for different immigration probabilities (Fig. 7b). Bloom risk as defined above  
431 (more than  $1.4 \text{ salps m}^{-3}$ ) increases with immigration probability, however, the risk rises more  
432 slowly in the variable food scenario and converges at a lower level. In the constant food  
433 scenario, the bloom risk grows up to 0.8 with immigration probability = 1 (not shown in the  
434 graph).  
435

436 Discussion:

437 We have developed the first individual-based model for salps that is able to resolve  
438 physiological and demographic processes to better understand the population dynamics of  
439 salps and its potential for their explosive population growth. The model reproduces empirical  
440 patterns on several organizational levels. Starting from the individual level modelling daily  
441 growth rates, our model results (variable food scenario: growth rates for oozoids: mean = 1.2  
442  $\text{mm d}^{-1}$ ,  $\text{sd} = 1 \text{ mm d}^{-1}$  and  $\text{max} = 2.8 \text{ mm d}^{-1}$ ) are consistent with the latest empirical  
443 observations (Pakhomov and Hunt 2017) reporting  $2.83 \pm 0.42 \text{ mm d}^{-1}$  for oozoid daily  
444 growth and cohort-modelling results ( $0.7 - 2 \text{ mm d}^{-1}$ , Henschke et al. 2018). It is worth noting  
445 that growth rates in the simulation were measured all year round and often included zero  
446 growth due to possible severe competition or low productivity during winter. On the  
447 observational scale the grazing impact of salps is often described as moderate (e.g. Pakhomov  
448 et al. 2002) although it has been stated that salps can consume more than 100% of the  
449 primary production (Dubischar and Bathmann 1997). On small spatial and temporal  
450 resolution salps may be food limited negatively affecting their growth.

451

452 Older estimates of daily growth rates based on cohort analyses by Loeb and Santora (2012)  
453 were substantially lower for oozoids (growth rates for oozoids: mean  $0.23 \text{ mm d}^{-1} \pm 0.04 \text{ mm}$   
454  $\text{d}^{-1}$ ) and higher for blastozooids ( $0.4 \text{ mm d}^{-1} \pm \text{mm d}^{-1}$ ). The lower estimates have profound  
455 consequences for the estimates of one full reproduction cycle duration suggested to be 9 – 12  
456 months (Foxton 1966, Loeb and Santora 2012). However, if we use parameterizations  
457 resulting in daily growth rates comparable with those reported by Loeb and Santora (2012)  
458 the salp peak abundances could never exceed  $0.5 \text{ individuals m}^{-3}$  even if the most positive  
459 assumptions for migration (immigration probability 100%) and food availability (variable  
460 food scenario) were used (see Fig. S1). This is in line with a previous modelling study where  
461 a scenario based on the growth rate estimates by Loeb and Santora (2012) resulted in non  
462 viable salp populations (Henschke et al. 2018). Assuming low growth rates salp blooms in  
463 our model would require a massive immigration or potential accumulation of oozoids over at  
464 least two years, which was in fact suggested by Loeb and Santora (2012). Indeed, if a  
465 substantial fraction of the salp population would survive beyond summer/fall, it could act as  
466 the starting ground for a salp bloom in the next year. This could not be possible in our model  
467 because in our simulation salp abundances during winter get very low. Thus, salp blooms  
468 only occurred if a successful migration of a seed population started in early spring. Whether  
469 the early year migration could potentially be interpreted as salps surviving from the previous

470 season at greater depths or suitable adjacent habitats still requires confirmation. Interestingly  
471 the simulated median of the peak abundance for the slow growth ( $0.03 \text{ individuals m}^{-3}$ ) is in  
472 the same order of magnitude as the observed median of  $0.06 \text{ individuals m}^{-3}$  from the AMLR  
473 sites. This highlights the importance of probabilistic risk assessments instead of comparing  
474 statistics such as the median. Another mechanism that could lead to massive salp blooms is  
475 the aggregation of salps from a large spatial domain by physical forces. To clarify the  
476 importance of this process model of ocean currents and turbulences and individual-based  
477 models need to be coupled (Dorman et al. 2011).

478

479 Our simulations highlight the variability in the population of salps that are caused by two  
480 sources of uncertainty: environmental stochasticity (variability in the primary production) and  
481 demographic stochasticity caused by random death events and immigration of small seeding  
482 populations. Our results show that if maximum primary production does not change from  
483 year to year, the abundance of salps is mainly driven by the immigration (or population  
484 survival during overwintering) of small seeding populations. With global change intrusion of  
485 warm waters transporting salps southward may become more frequent potentially resulting in  
486 more salp blooms (Loeb et al. 1997, Pakhomov et al. 2002). To better understand changes in  
487 southward transportation of salps coupling of advection models and salp models is needed on  
488 larger spatial scales similar to the modelling of krill larvae transport and distribution (Mori et  
489 al. 2019). The increase in the risk of salp blooms, however, is modified by other  
490 environmental factors, which can be seen in our simulations when maximum primary  
491 production is drawn from a lognormal distribution. In that case, salp blooms may be locally  
492 prevented in years of lower primary production. This highlights the importance of forcing  
493 models with values drawn from the empirical distributions on the relevant scale for  
494 demographic and physiological processes, rather than average values over large spatial and  
495 temporal scales. Improved environmental models are needed that can predict how temporal  
496 and spatial variability of chlorophyll *a* will vary with global climate change to project the  
497 population dynamics and the probability of salp blooms. To upscale local results to  
498 biogeographical regions is challenging since crucial information such as primary productivity  
499 and sea ice coverage will be only available on spatial resolutions that are not necessarily  
500 matching the spatial resolution of the relevant demographic and physiological processes  
501 (Levin 1992, Chave 2013). In particular, the non-linear relationship between food availability  
502 and duration of sea ice cover on growth and reproduction will require suitable transfer  
503 functions to link the observation and process scales (Bloeschl and Sivapalan 1995).

504 Therefore, the linkage of improved ecosystem models with higher spatial resolution and  
505 spatial explicit ecological models will be highly beneficial especially if the biotic interaction  
506 between krill and salps are to be considered (Elith and Leathwick 2009, Schurr et al. 2012).

507

508 Conclusions:

509 With an individual-based local salp model we have explored how the huge variability in  
510 observed salp abundances can be better understood. We have found that fluctuations in  
511 environmental conditions together with demographic processes such as the migration of small  
512 seed populations can cause variability in the observed range on the local scale. For large salp  
513 blooms successful immigration in early spring is required. This may require that oozoids have  
514 successfully overwintered in the area (Loeb and Santora 2012) and that sea ice retreat started  
515 early which is assumed to favor salp blooms (Atkinson et al. 2004). However, the  
516 overwintering hypothesis is difficult to support since no salp monitoring campaigns have  
517 been conducted during winter in the AMLR area yet (Atkinson et al. 2017). It is suggested  
518 that the migration of salps together with the intrusion of warm waters or/and salp winter non-  
519 extinction events will become more probable (Loeb et al. 1997, Pakhomov et al. 2002). The  
520 increase in the risk of salp blooms due to migration (or/and winter survival) will still be  
521 buffered to some extent by environmental stochasticity of primary productivity. These  
522 findings highlight the importance of considering the large variability in the population  
523 dynamics of salps together with heterogeneous environmental conditions in space and time to  
524 understand the risk of salp blooms in the Southern Ocean.

525

526

527 Acknowledgements: This work was supported by the PEKRIS BMBF Project 03F0746B  
528 awarded to BM and JG (The PERformance of KRill vs. Salps to withstand in a warming  
529 Southern Ocean). We also thank the DAAD Project 57386792 for funding travelling for BM  
530 and JG to partially support this work. We gratefully acknowledge the support of the German  
531 Centre for Integrative Biodiversity Research (iDiv) Halle-Jena-Leipzig, funded by the  
532 German Research Foundation (FZT 118).

533

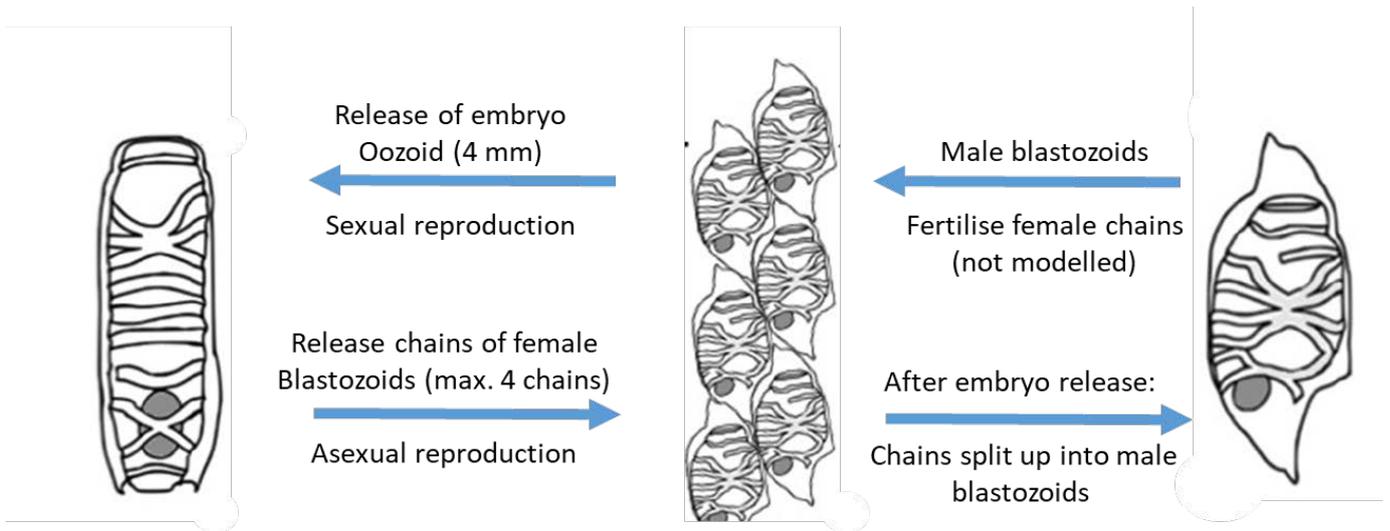
534 Table 1: Overview of parameters, reference values and their elasticities. Elasticities are  
535 computed as  $E = ((y(p_{max}) - y(p_{min})) / y(p_{ref})) / 0.2$ , where  $p_{min}$  is the parameter decreased by  
536 10%,  $p_{max}$  is the parameter increased by 10% and  $p_{ref}$  is the reference value. The response  $y$   
537 has been averaged over 10 simulations. We use the maximum abundance during the  
538 simulation to determine  $E_{peak}$  and the median of seasonal peak abundances to determine  
539  $E_{median}$ .

540

Parameter	Parameter	Submodels	Value	Units	Elasticity $E_{peak}$	Elasticity $E_{median}$
$g$	Assimilation rate	Growth	0.0025	mg chl a / cm <sup>2</sup>	-0.02	-0.27
$rb_{ref,oozoid}$	Oozoid vonBertalanffy growth rate	Growth	0.024	d <sup>-1</sup>	2.20	3.80
$rb_{ref,blastozoid}$	Blastozoid vonBertalanffy growth rate	Growth	0.0155	d <sup>-1</sup>	2.0	3.40
$K$	Half saturation constant	Growth	0.2	mg chl a m <sup>-3</sup>	-2.6	-5.50
$t_{surv}$	Starvation	Death	30	days	0.3	0.5
$m$	Daily mortality	Death	0.025	-	-3.1	-3.60
$p_e$	Embryo survival	Reproduction	0.7	-	0.7	0.5
$vd$	Vegetation delay	Primary production	45	days	-0.1	0.1
$r_{algae,max}$	Rate of primary production	Primary production	0.25	d <sup>-1</sup>	0.9	2.9
$d$	Chl a decay	Primary production	0.05	d <sup>-1</sup>	-0.5	-2.60
$p_i$	Immigration probability	Immigration	0.0085	-	-0.2	2.70
$n_i$	Number of immigrants	Immigration	10	-	0.60	0.60
$l_i$	Size of immigrants	Immigration	3	cm	0.5	1.1

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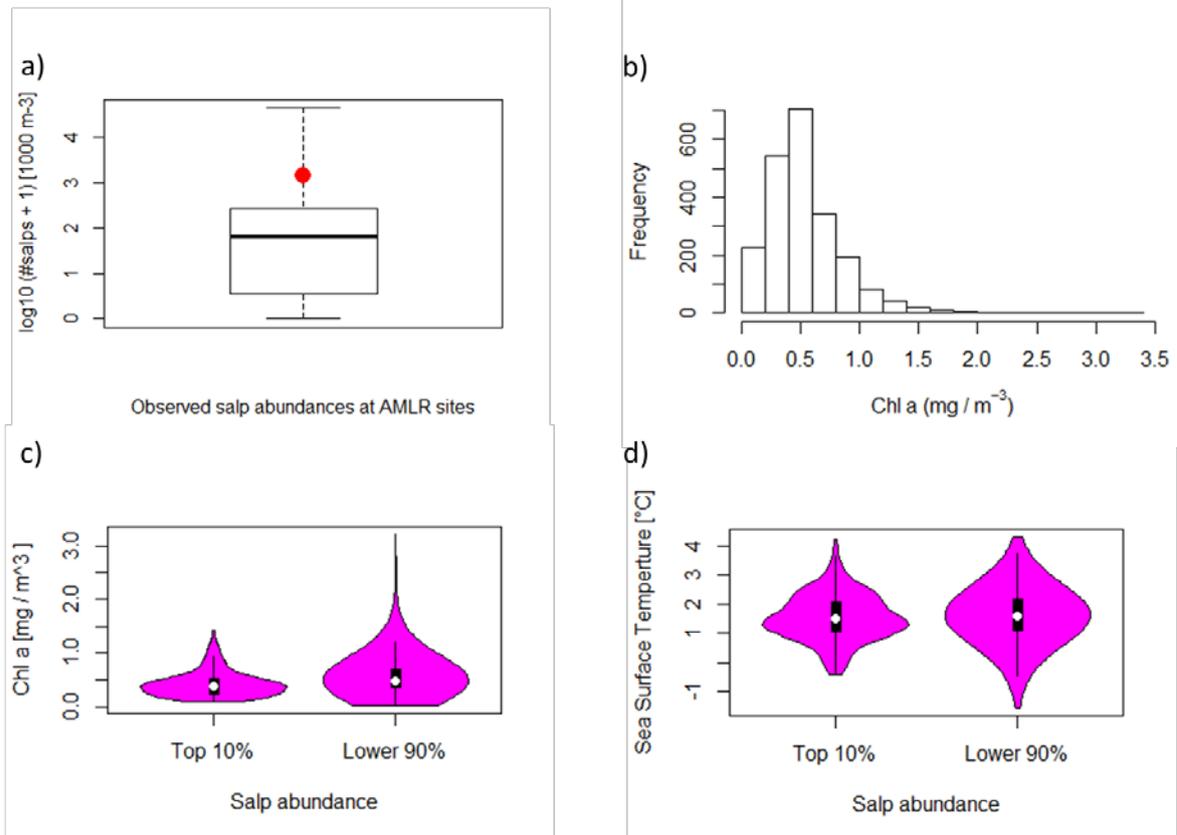


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544 Figure 1: Modified from Henschke et al. 2018. The life cycle of *Salpa thompsoni*. Oozoid  
 545 individuals release chains of female blastozoids. Each female blastozoid can release one  
 546 oozoid if fertilized. After embryo release the female blastozoid changes sex into a male  
 547 blastozoid that potentially fertilizes other female blastozoids. We refer to one cycle of  
 548 sexual and asexual reproduction as one regeneration cycle.

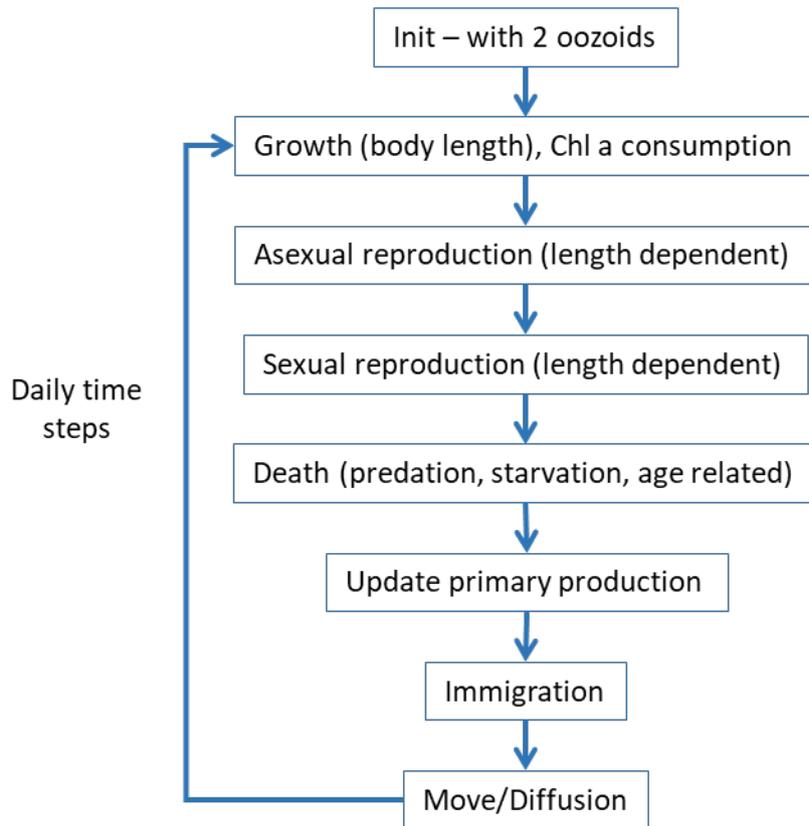
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551 Figure 2: Overview of the long term data set of the Antarctic Marine Living Research (AMLR)  
 552 by NOAA that has been used to constrain the model. A) Distribution of the observed salp  
 553 densities. The distribution is dominated by observations with no salps observed. If salps  
 554 have been observed, the number varies over four orders of magnitude with a maximum  
 555 density of 44,241 salps in 1000 m<sup>3</sup> of water. The red point indicates the 95% quantile of  
 556 observed densities  $q = 1,440$  salps 1000 m<sup>-3</sup> (the box presents the range of the second and  
 557 third quartiles, the line in the box is the median and whiskers show the full range of the data  
 558 as long as the distance of the extreme points does not exceed 1.5 times the interquartile  
 559 range from the box). B) Measured chlorophyll *a* also varies substantially. C-D Here we  
 560 present C) chlorophyll *a* and D) sea surface temperature distributions for the highest 10%  
 561 salp abundances and the remainder (lower 90%). It can be seen that salp abundances are  
 562 controlled by low temperatures and that high densities of salps have not been observed for  
 563 very high chlorophyll *a* content (the area in the violin plot C and D indicate the frequency of  
 564 the observed value).  
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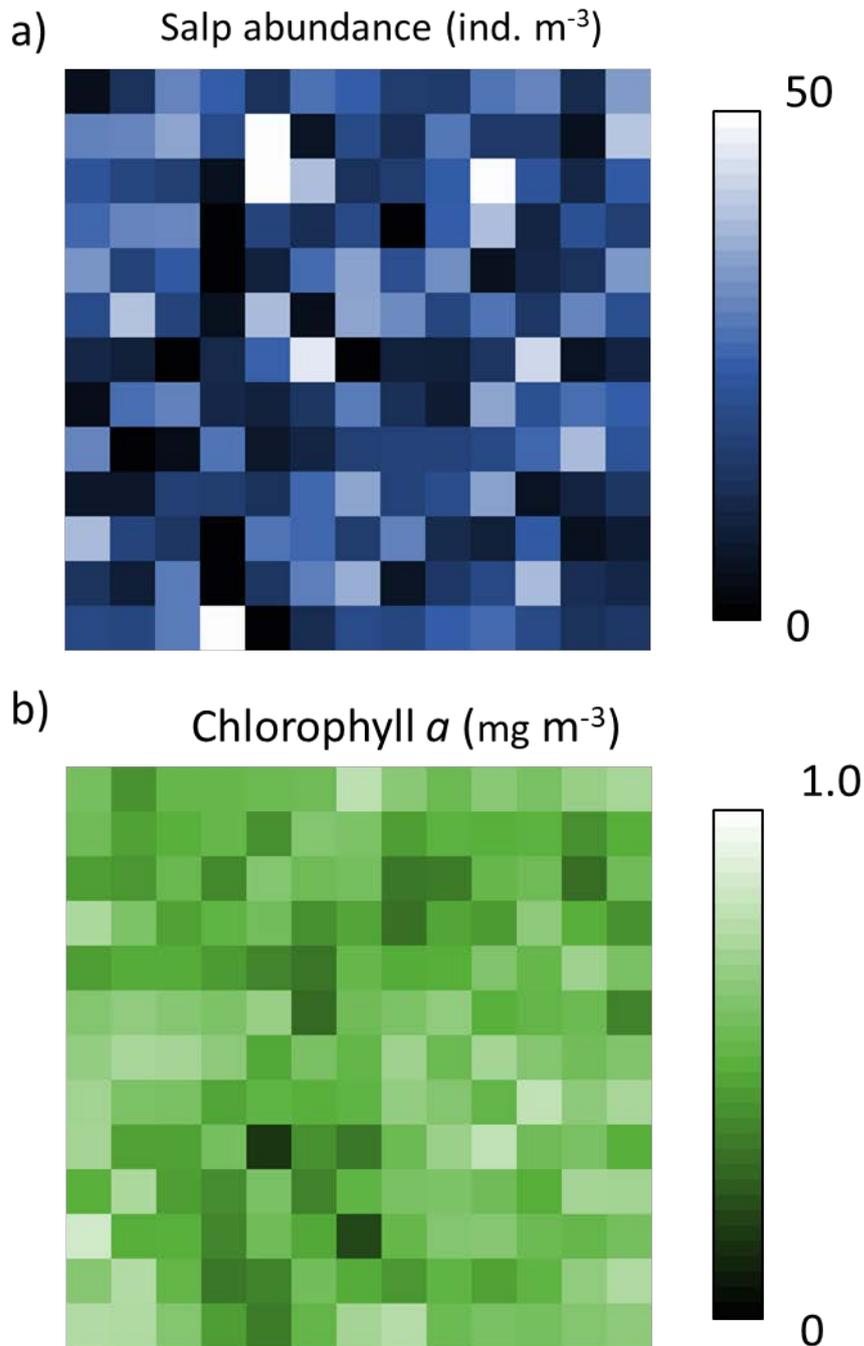
572 Figure 3: Flow Chart of the model. Each day the following processes are executed: each salp  
573 potentially grows in body length and reduces chlorophyll *a* availability in its patch, if body  
574 length exceeds a certain threshold oozoids will reproduce asexually and blastozooids will  
575 reproduce sexually, mortality is determined, the chlorophyll *a* concentration in each patch is  
576 updated (Update primary production), immigration is modelled, and the spatial distribution  
577 of salps is updated (Move/Diffusion). See the section “Submodels” for further details.

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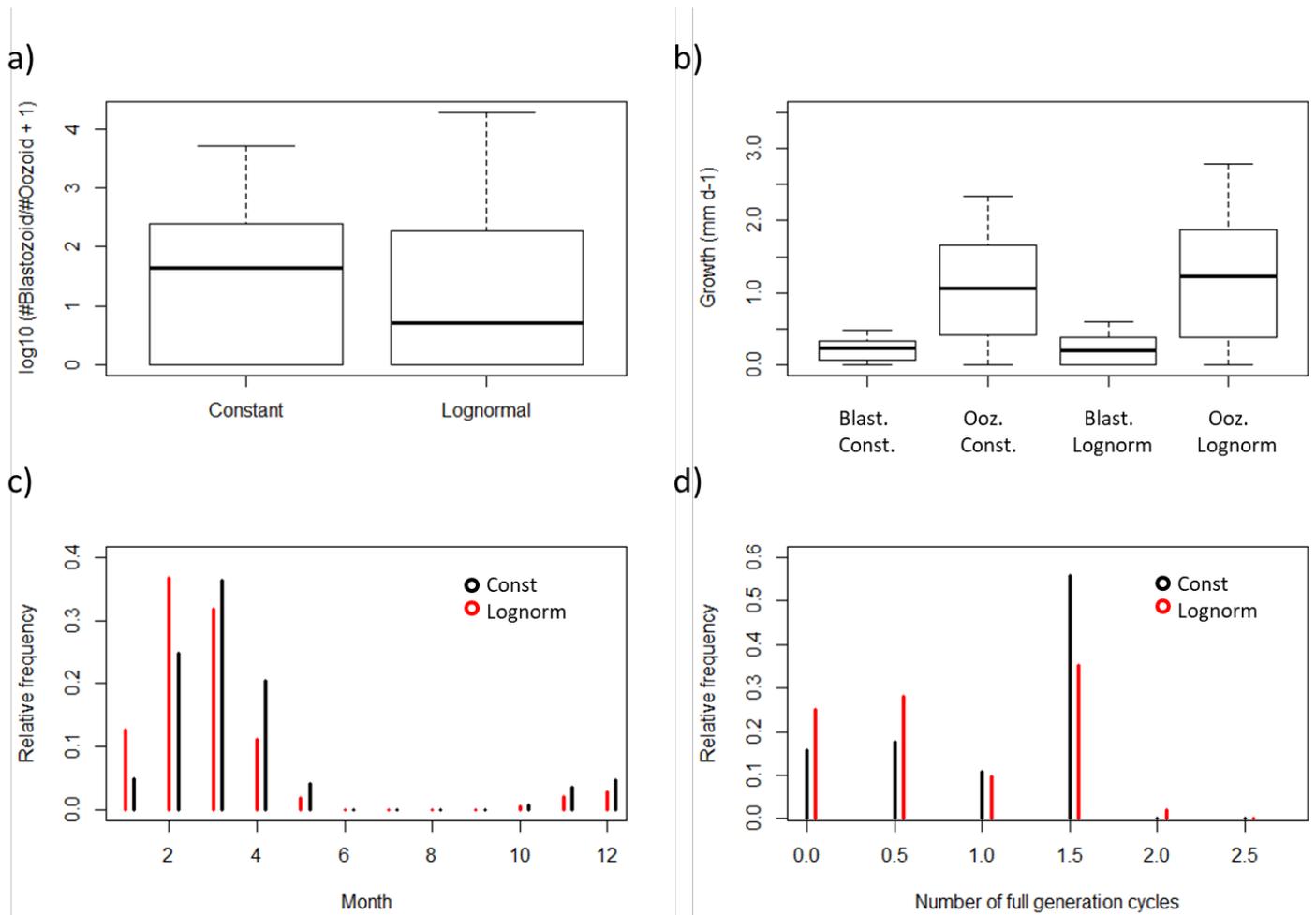


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584 Figure 4: Typical state of the model during a salp bloom ( $N = 54,400$ ). Spatial resolution is 16  
 585 m<sup>-3</sup> and spatial extent is 13 x 13 patches. While the average salp density is 20 salps m<sup>-3</sup> the  
 586 density on the resolution of the simulation varies between 0 and 62 salps m<sup>-3</sup>. The mean  
 587 chlorophyll *a* content is 0.56 mg chlorophyll *a* m<sup>-3</sup>. On the grid resolution chlorophyll *a*  
 588 content varies between 0.14 and 0.86 mg chlorophyll *a* m<sup>-3</sup>.

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593 Figure 5: Diagnostic patterns produced by the model. A) On average more blastozoids than

594 oozoids are present in the system (the median ratios between blastozoids and oozoids are 43

595 for the constant chlorophyll *a* scenario and 4 for the variable food (lognormal) scenario). b)

596 Daily growth rates covering the range reported in the literature. c) Intra-annual distributions

597 are in line with the fact that salps are only expected in the Antarctic Summer around the

598 Western Antarctic Peninsula. d) The distribution of fully completed life cycles ranges from 0

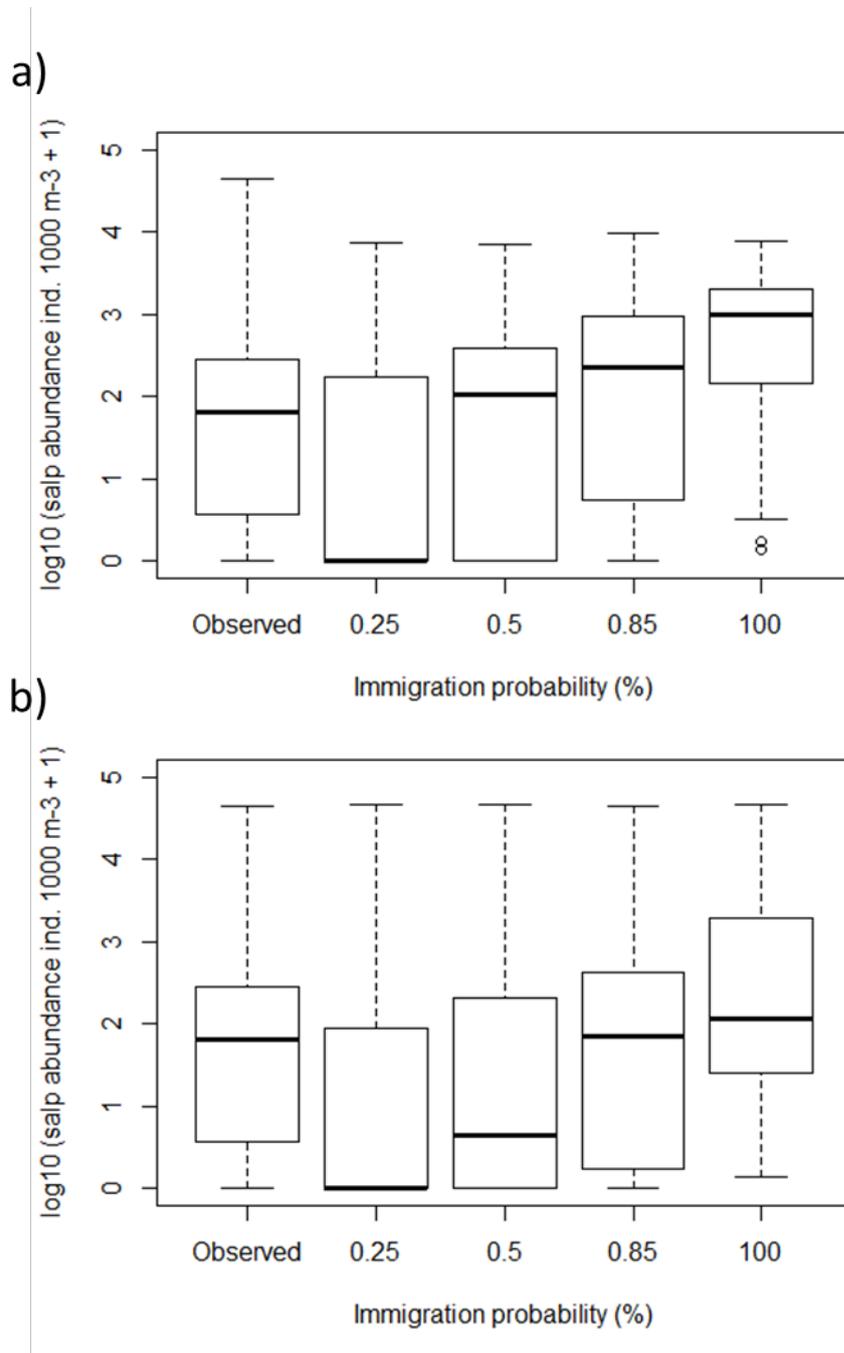
599 to 2.5.

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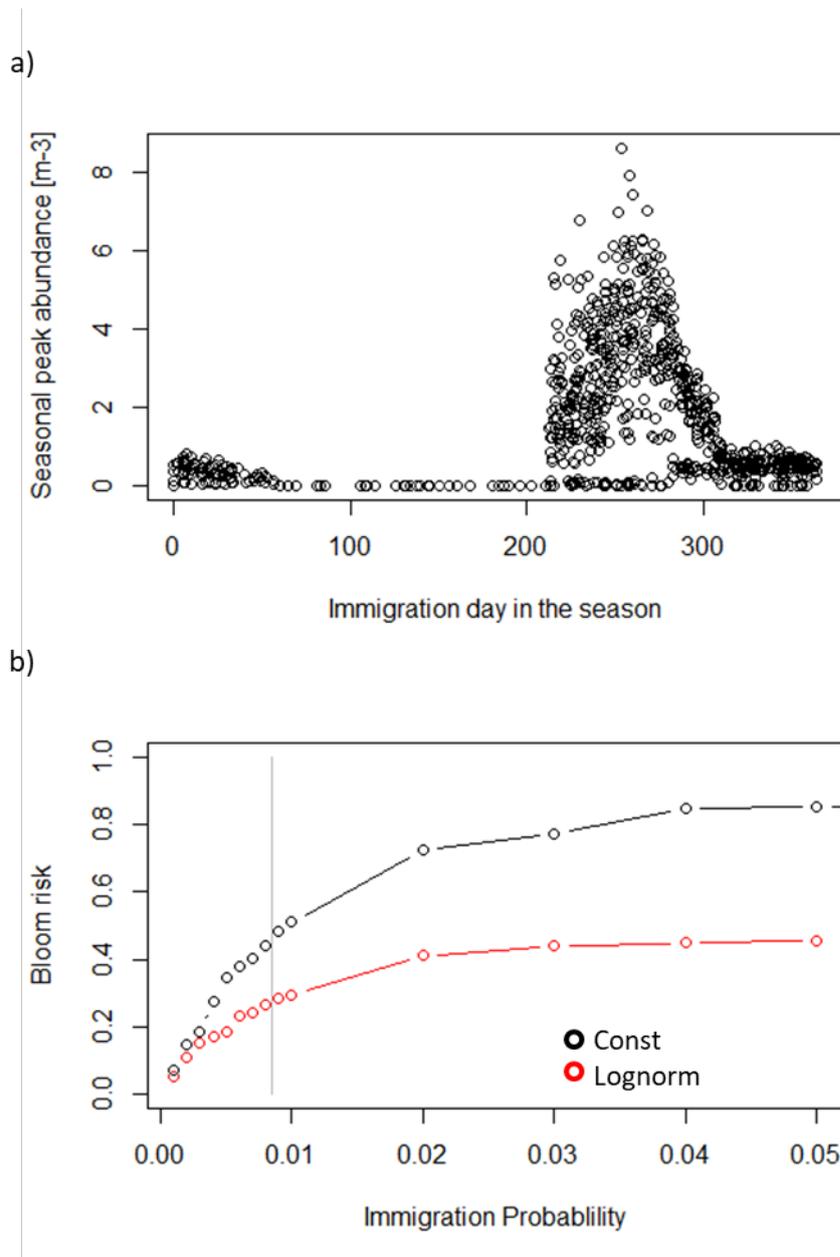
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605 Figure 6: Comparing distributions of observed salp abundances and simulated daily summer  
606 month abundances (January – March) for 800 simulated years. The simulated abundances  
607 strongly depend on the immigration probability. a) In the scenario with constant food  
608 availability the maximum observed salp density cannot be reached. b) Only in the variable  
609 chlorophyll *a* scenario (lognormal) observed maximum abundances can be achieved.

610



611

612 Figure 7: a) Maximum peak abundances in the season strongly depend on the timing when a  
 613 seeding population has successfully migrated into the local area. Only if migration happens in  
 614 early spring massive blooms can be expected in the following summer (simulations have  
 615 been made with the constant food scenario). b) The risk for a salp bloom (here defined as the  
 616 95% quantile of observed salp densities which is  $1.4 \text{ salps m}^{-3}$ ) is higher for the constant  
 617 chlorophyll *a* scenario (black line) and can partly be buffered by heterogeneities in the  
 618 chlorophyll *a* content (red line). The gray line indicates the immigration probability that  
 619 resulted in the best match with the observed salp abundance distributions.

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621

622 References

623 Atkinson, A., et al. (2017). KRILLBASE: a circumpolar database of Antarctic krill and salp  
624 numerical densities, 1926-2016. *Earth System Science Data*, 9(1), 193-210.

625

626 Atkinson, A., Siegel, V., Pakhomov, E., Rothery, P. (2004). Long-term decline in krill stock  
627 and increase in salps within the Southern Ocean. *Nature*, 432(7013), 100-103.

628

629 Begon, M., Townsend, C.R., Harper, J.L. (2006). *Ecology: From Individuals to Ecosystems*.  
630 USA, Blackwell Publishing Ltd.

631

632 Blöschl, G. and M. Sivapalan (1995). SCALE ISSUES IN HYDROLOGICAL MODELING  
633 - A REVIEW. *Hydrological Processes*, 9(3-4), 251-290.

634

635 Chave, J. (2013). The problem of pattern and scale in ecology: what have we learned in  
636 20years? *Ecology Letters*, 16, 4-16.

637

638 Daponte, M. C., Capitanio, F.L., Esnal, G.B. (2001). A mechanism for swarming in the  
639 tunicate *Salpa thompsoni* (Foxton, 1961). *Antarctic Science*, 13(3), 240-245.

640

641 Dorman, J. G., Powell, T.M., Sydeman, W.J., Bograd, S.J. (2011). Advection and starvation  
642 cause krill (*Euphausia pacifica*) decreases in 2005 Northern California coastal populations:  
643 Implications from a model study. *Geophysical Research Letters*, 38. DOI:  
644 10.1029/2010GL046245.

645

646 Dubischar, C. D. and U. V. Bathmann (1997). Grazing impact of copepods and salps on  
647 phytoplankton in the Atlantic sector of the Southern Ocean. *Deep-Sea Research Part II-*  
648 *Topical Studies in Oceanography*, 44(1-2), 415-433.

649

650 Elith, J. and J. R. Leathwick (2009). Species Distribution Models: Ecological Explanation  
651 and Prediction Across Space and Time. *Annual Review of Ecology Evolution and*  
652 *Systematics*, 40, 677-697.

653

654 Foxton, P. (1966). The distribution and life history of *Salpa thompsoni* Foxton with  
655 observations on a related species, *Salpa gerlachei* Foxton. *Discovery Reports*, 34, 1-116.

656

657 Grimm, V., et al. (2006). A standard protocol for describing individual-based and agent-  
658 based models. *Ecological Modelling*, 198(1-2), 115-126.

659

660 Grimm, V., Berger, U., DeAngelis, D.L., Polhill, J.G., Giske, J., Railsback, S.F. (2010). The  
661 ODD protocol A review and first update. *Ecological Modelling*, 221(23), 2760-2768.

662 Henschke, N., Smith, J.A., Everett, J.D. Suthers, I.M. (2015). Population drivers of *Thalia*  
663 *democratica* swarm: insights from population modelling. *Journal of Plankton Research*, 37,  
664 1074-1087.

665 Henschke, N., Everett, J.D., Richardson, A.J., Suthers, I.M. (2016). Rethinking the Role of  
666 Salps in the Ocean. *Trends in Ecology & Evolution*, 31(9), 720-733.

667

668 Henschke, N., Pakhomov, E.A., Groeneveld, J., Meyer, B. (2018). Modelling the life cycle of  
669 *Salpa thompsoni*. *Ecological Modelling*, 387, 17-26.

670

671 Henschke, N., Smith, J.A., Everett, J.D., Suthers, I.M. (2015). Population drivers of a *Thalia*  
672 *democratica* swarm: insights from population modelling. *Journal of Plankton Research*,  
673 37(5), 1074-1087.

674 Henschke, N. and Pakhomov, E.A. (2019). Latitudinal variations in *Salpa thompsoni*  
675 reproductive fitness. *Limnology and Oceanography*, 64, 575-584.

676

677 Huntley, M. E., Sykes, P.F., Marin, V. (1989). BIOMETRY AND TROPHODYNAMICS OF  
678 *SALPA-THOMPSONI FOXTON* (TUNICATA, THALIACEA) NEAR THE ANTARCTIC  
679 PENINSULA IN AUSTRAL SUMMER, 1983-1984. *Polar Biology*, 10(1), 59-70.

680

681 Iguchi, N. and Ikeda, T. (2004). Metabolism and elemental composition of aggregate and  
682 solitary forms of *Salpa thompsoni* (Tunicata: Thaliacea) in waters off the Antarctic Peninsula  
683 during austral summer 1999. *Journal of Plankton Research*, 26, 1025-1037.

684

685 Jager, T., Martin, B., Zimmer, E.I. (2013). DEBkiss or the quest for the simplest generic  
686 model of animal life history. *Journal of Theoretical Biology*, 328, 9-18.

687

688 Levin, S. A. (1992). THE PROBLEM OF PATTERN AND SCALE IN ECOLOGY.  
689 *Ecology*, 73(6), 1943-1967.

690

691 Loeb, V., Siegel, V., Holm-Hansen, O., Hewitt, R. Fraser, W., Trivelpiece, W., Trivelpiece,  
692 S. (1997). Effects of sea-ice extent and krill or salp dominance on the Antarctic food web.  
693 Nature, 387(6636), 897-900.

694

695 Loeb, V., Hofmann, E.E., Klinck, J.M. and Holm-Hansen, O. (2010) Hydrographic control of  
696 the marine ecosystem in the South Shetland-Elephant Island and Bransfield Strait region, 57,  
697 519 – 542.

698

699 Loeb, V. J. and J. A. Santora (2012). Population dynamics of *Salpa thompsoni* near the  
700 Antarctic Peninsula: Growth rates and interannual variations in reproductive activity (1993-  
701 2009). Progress in Oceanography, 96(1), 93-107.

702

703 Mori, M., Corney, S.P., Melbourne-Thomas, J., Klocker, A., Kawaguchi, S., Constable, A.  
704 and Summer, M. (2019) Modelling dispersal of juvenile krill released from the Antarctic ice  
705 edge: Ecosystem implications of ocean movement. Journal of Marine Systems, 189, 50-61.

706

707 Nicol, S. and J. Foster (2016). The Fishery for Antarctic Krill: Its Current Status and  
708 Management Regime. In: V. Siegel (Ed.), Biology and Ecology of Antarctic Krill (pp. 387-  
709 421). Springer International Publishing.

710

711 Pakhomov, E. A., Froneman, W., Perissinotto, R. (2002). Salp/krill interactions in the  
712 Southern Ocean: spatial segregation and implications for the carbon flux. Deep-Sea Research  
713 Part II-Topical Studies in Oceanography, 49(9-10), 1881-1907.

714

715 Pakhomov, E. A. and B. P. V. Hunt (2017). Trans-Atlantic variability in ecology of the  
716 pelagic tunicate *Salpa thompsoni* near the Antarctic Polar Front. Deep-Sea Research Part II-  
717 Topical Studies in Oceanography, 138, 126-140.

718

719 Reiss, C.S., Cossio, A.M., Loeb, V., Demer, D.A. (2008). Variations in the biomass of  
720 Antarctic krill (*Euphausia superba*) around the South Shetland Islands, 1996-2006. Ices  
721 Journal of Marine Science, 65(4), 497-508.

722

723 Ryabov, A.B., de Roos, A.M., Meyer, B., Kawaguchi, S., Blasius, B. (2017). Competition-  
724 induced starvation drives large-scale population cycles in Antarctic krill. Nature Ecology &  
725 Evolution, 1(7). doi: 10.1038/s41559-017-0177

726

727 Schurr, F. M., et al. (2012). How to understand species' niches and range dynamics: a  
728 demographic research agenda for biogeography. *Journal of Biogeography*, 39(12), 2146-  
729 2162.

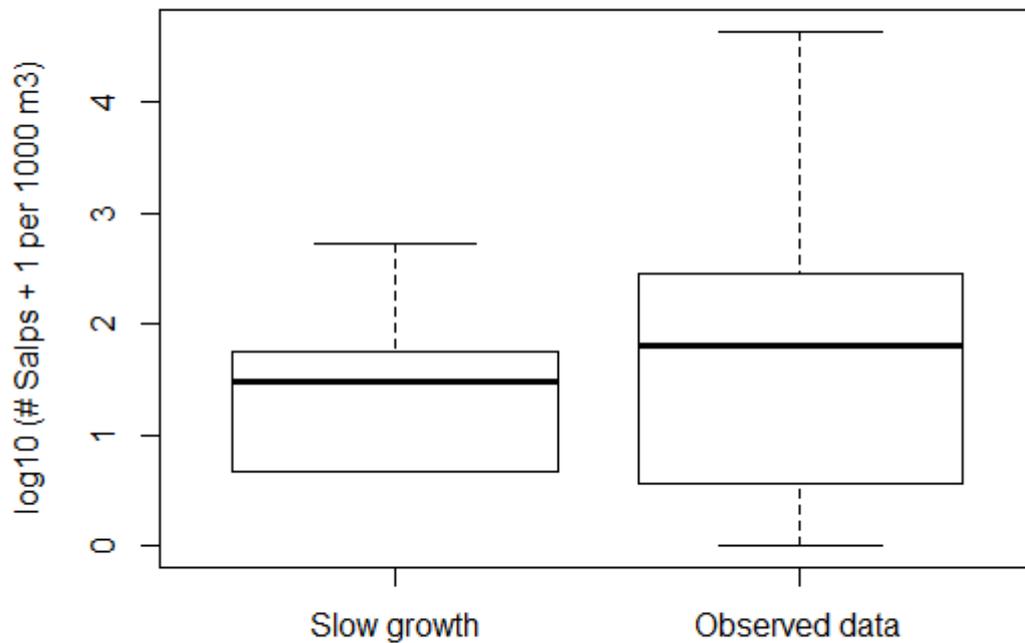
730

731 Stonedahl, F. and Wilensky, U. (2010). BehaviorSearch [computer software]. Center for  
732 Connected Learning and Computer Based Modeling, Northwestern University, Evanston, IL.  
733 available online: <http://www.behaviorsearch.org>

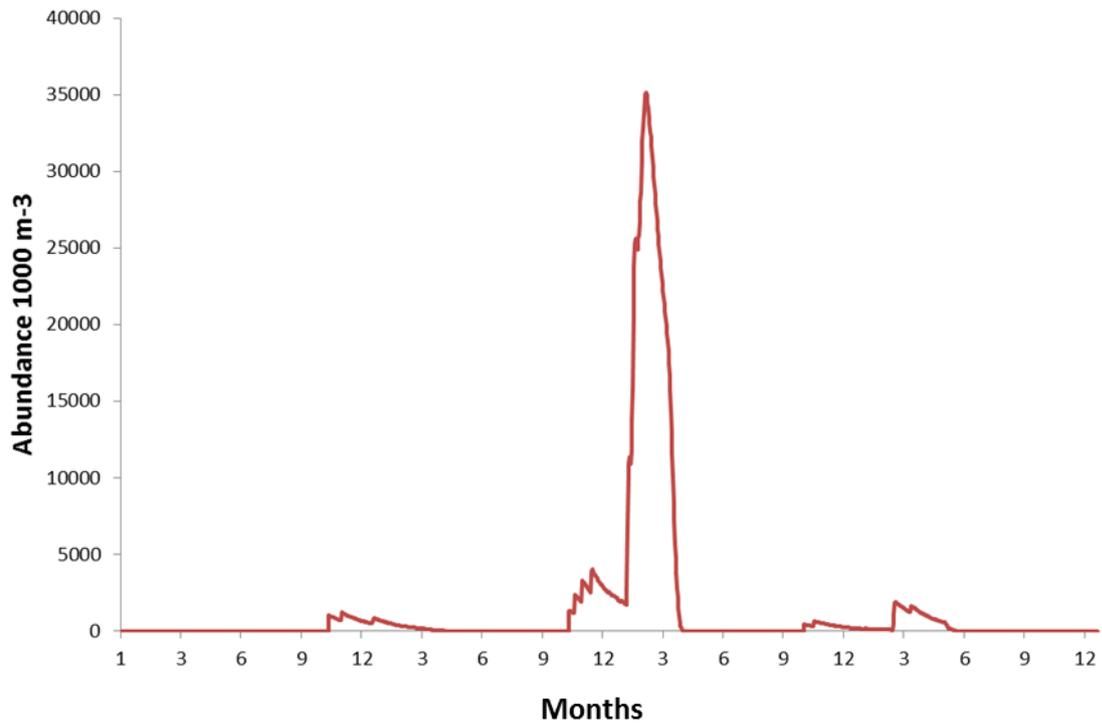
734

735 von Harbou, L. (2009). Trophodynamics of Salps in the Atlantic Southern Ocean University  
736 of Bremen, University of Bremen. Ph.D.

737 Appendix:  
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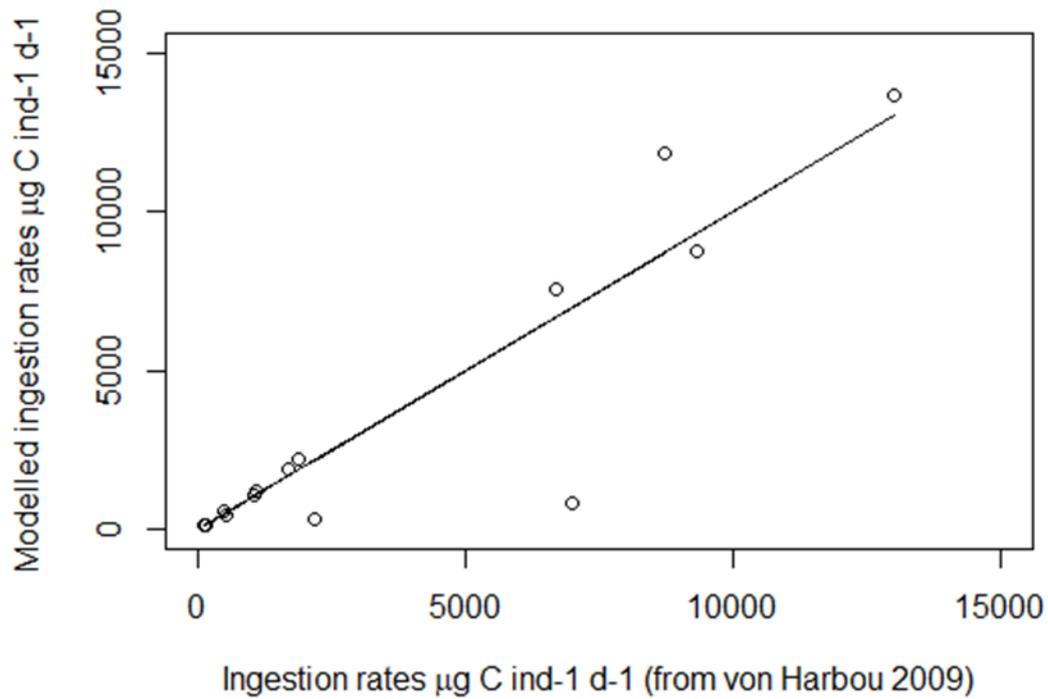


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745 Figure S1: Simulated seasonal peak abundances using the lower growth rates reported by  
746 Loeb and Santora (2012). With slow growth the simulated abundance never exceeds 0.54  
747 individuals  $m^{-3}$  compared to more than 40 individuals  $m^{-3}$  in the AMLR data set, although the  
748 highest possible migration probability and the variable food scenario have been used.  
749 Nevertheless, the medians of both distributions are in the same order of magnitude. This  
750 highlights the importance of considering the variability in population abundances in addition  
751 to the median or the average abundance for risk assessment.  
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Figure S2: Exemplary time series for four years for the variable food scenario. To have a salp bloom successful establishment has to occur in early spring and the food availability needs to be sufficient. Peak chlorophyll *a* levels have been 0.2 mg m<sup>-3</sup>, 0.31 mg m<sup>-3</sup>, 0.79 mg m<sup>-3</sup> (year of the abundance peak), 0.39 mg m<sup>-3</sup> and finally 1.23 mg m<sup>-3</sup> in the last summer. During the last summer no salps have been migrated into the area.



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763 Figure S3: Modelled ingestion rates and ingestion rates reported in von Harbou 2009 (p. 108  
 764 and p. 205). To compare ingestion rates a C:Chla ratio of 60 has been assumed which is also  
 765 in the range of reported values (von Harbou 2009). For data points reported as low  
 766 chlorophyll we assumed  $0.75 \text{ mg m}^{-3} \text{ Chl-}a$  and  $2 \text{ mg m}^{-3} \text{ Chl-}a$  respectively for the high  
 767 chlorophyll scenario and  $1.8 \text{ mg m}^{-3} \text{ chl}a$  were explicitly stated (Table 4 page 204 von  
 768 Harbou 2009).

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