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# 1 **Screening Tools for the Bioconcentration Potential of monovalent organic ions in fish**

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## 10 **Abstract**

11 Currently the bioaccumulation potential of organic chemicals is assessed in a first tier approach  
12 via their octanol-water partition coefficient.. This approach has been developed for neutral  
13 chemicals and cannot work for ionizable and ionic chemicals because the latter have different  
14 sorption-mechanisms and -preferences. Thus, suitable screening tools for the bioconcentration  
15 potential of ionic and ionizable chemicals need to be developed because it cannot be expected  
16 that these chemicals are non-bioaccumulative per se. Here, we present such screening tools for  
17 monovalent ions and ionizable chemicals based on calibrated sorption models for membrane  
18 lipids, structural proteins and albumin. The molecular descriptors used for these models arise  
19 from quantum chemical calculations and are based on COSMO-RS theory. When we applied our  
20 screening tools to 1839 preselected chemicals from the REACH registration data base, we  
21 identified 187 chemicals as potentially bioconcentrating (still ignoring any kind of metabolism).  
22 Among these were carbon and sulphur based aromatic and aliphatic acids mostly with a rather  
23 high molecular surface area. We hope that this outcome will trigger further research on ion  
24 specific sorption mechanisms and lead to a re-evaluation of the bioconcentration potential of  
25 ionic chemicals.

## 27 **1 Introduction**

28 There is no generally accepted approach to estimate the bioaccumulation potential of organic ions  
29 <sup>1</sup> – despite the fact that the regulation of organic ions is a prevailing challenge.<sup>2</sup> Ionogenic organic  
30 chemicals comprise very diverse structures and chemical classes such as surfactants,  
31 pharmaceuticals, some classes of pesticides, poly- or perfluorinated acids<sup>2</sup> as well as ionic  
32 liquids.<sup>3</sup> The use of a single and easy to determine threshold value (such as a certain logarithmic  
33 octanol-water partition coefficient value), which is applied for neutral chemicals by regulation  
34 authorities,<sup>4</sup> will not suffice as a standard criterion to identify the bioaccumulation potential of  
35 charged chemicals<sup>1</sup>. Previous work focused on the description of rates of uptake and elimination  
36 (including metabolism) to describe the bioaccumulation potential of organic ions, aiming at a  
37 holistic picture.<sup>5</sup> While we agree that physiologically based pharmacokinetic modelling is highly  
38 needed, we consider the underlying physicochemical parameters, especially the equilibrium  
39 sorption coefficients to the different relevant phases, as a major uncertainty in our current  
40 knowledge. In previous work, both sorption to proteins as well as sorption to membrane lipids  
41 was estimated for organic ions via the respective octanol-water partition coefficient.<sup>5</sup> In our own  
42 work we have recently shown that this is not appropriate.<sup>6</sup> In this work we therefore develop  
43 mechanistic and semi empirical models to predict such equilibrium sorption coefficients. These  
44 can then be used to screen the bioaccumulation potential of organic ions in a first tier approach  
45 that still neglects any biotransformation or other kinetics and can thus be seen as a worst case  
46 scenario. Here, we use the newly developed predictive tools to provide such a screening of  
47 chemicals for their bioconcentration potential in fish and based on a depiction of the major  
48 sorption matrices. Analogous to the pharmacokinetic literature,<sup>4,7-9</sup> we assume the following  
49 sorption matrices in organisms to be the most relevant for organic ions: membrane lipid, muscle  
50 protein (which is our proxy for structural protein), serum albumin (which is our proxy for plasma  
51 proteins in fish) and water. We have to note that not all fish species have albumin and also there  
52 are other blood constituents that might be important sorbents for organic ions. For organic ions  
53 we assume that the sorption capacity of storage lipid (fat) can be neglected, based on the finding  
54 that ions partition into octanol only marginally as ion pairs<sup>10,11</sup> and octanol is a pretty good proxy  
55 for storage lipid,<sup>4</sup> within our general model uncertainties.

56 While our general approach is straight forward and not new<sup>4</sup> (and has been applied by us before  
57 for a few selected ionizable chemicals with available experimental data),<sup>12</sup> the challenge lies in  
58 providing all the different partition coefficients required for a broad screening. For the neutral  
59 species of ionizable chemicals the usage of poly parameter free energy relationships (pp-LFERs)  
60 is an appropriate way to obtain these data as shown in a recent review.<sup>13</sup> In general, pp-LFERs are  
61 capable of describing the equilibrium partitioning of neutral organic chemicals between a  
62 multitude of biologically relevant matrices and water as well as technical partitioning systems  
63 and water. Unfortunately, the applicability of pp-LFERs for ionic organic chemicals is still in its  
64 infancy and of rather empirical nature, limited to few chemical classes.<sup>6,14,15</sup> Thus, we investigate  
65 here, to what extent required partition coefficients can be estimated with the help of the  
66 commercial software COSMOthermX17<sup>1</sup>, which is the only predictive tool that cannot only  
67 handle neutral species but that is principally able to provide meaningful predictions for the  
68 partitioning of organic ions.<sup>6</sup> COSMOtherm is based on quantum mechanical (QM) calculations  
69 and fundamental fluid phase thermodynamics (namely the conductor like screening model for  
70 real solvents, COSMO-RS)<sup>16,17</sup> which operates with only very general fitting parameters. The  
71 COSMO-RS implementation within COSMOtherm is principally applicable to both neutral  
72 chemicals as well as ions.<sup>18</sup> For ions, it has particularly been shown to be a good model for the  
73 description of the membrane-water partition coefficient<sup>19</sup> and for ionic liquid properties.<sup>20</sup>

74 Out of the four sorption matrices, only the membrane and water itself are well-defined and are  
75 thus directly describable within COSMOtherm.<sup>19,21</sup> The other two important sorption matrices are  
76 structural proteins and plasma proteins. About 10% of the whole body mass of vertebrates is  
77 made of structural proteins, which themselves consist to about 50% of muscle proteins (e.g., actin  
78 and myosin), while the other half is mostly keratin and collagen.<sup>14</sup> In the case of blood plasma  
79 the composition of the sorbing matrix varies in different organisms and the contributions of  
80 specific proteins are not always clear. Here, we used albumin, which is expected to dominate  
81 anionic sorption in human blood, as a proxy for the plasma proteins. For structural proteins and  
82 albumin, the only chance to grasp the major characteristics of the respective sorption matrices  
83 with COSMOtherm is via fitting experimental partition coefficients of organic ions to so-called  
84 sigma moments via a multiple linear regression (MLR). The sigma moments are an output of the  
85 quantum chemical cosmo calculation for molecules and account for the solutes' interaction

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86 properties. Calibrating MLR models based on sigma moments with experimental equilibrium  
87 partitioning data works well for neutral chemicals as has been shown for a big variety of liquid-  
88 liquid partitioning systems (personal communication COSMOlogic) and is conducted in exact  
89 analogy to the pp-LFER approach, as outlined in detail below. We tested this approach both for  
90 the partitioning of organic ions between plasma protein and water and structural protein and  
91 water.

92 The major aims of this work were twofold: to develop reliable predictive sorption models, for  
93 neutral and ionic chemicals in order to describe the bioaccumulation potential of organic ions and  
94 ionizable chemicals (without metabolism); second to identify potentially bioaccumulative  
95 compounds by applying our models to a set of almost 2000 organic ions or ionizable chemicals.  
96 For the first aim we developed MLRs based on sigma moments describing the sorption to  
97 structural proteins and to albumin (for neutral and monovalent ionic chemicals, respectively). For  
98 our second aim, we combined these MLRs with the pp-LFER models for neutral species and  
99 COSMOmic for neutral and ionic chemicals and applied it to almost 2000 chemicals.

## 100 **Materials and Methods for the development of sorption model**

### 101 **Materials for the development of sorption model**

#### 102 **Temperature dependence of sorption coefficients**

103 The experimental sorption data for phospholipid membrane are available for temperatures  
104 between 20 to 37°C. The sorption differences within this temperature range are negligible, as  
105 long as the membrane is in its natural liquid crystalline state.<sup>22</sup> The data for structural and muscle  
106 proteins and albumin had been measured at 37°C because they originally aimed to describe  
107 sorption capacities in humans. Although the modeled fish has a temperature between 13 and  
108 17°C,<sup>23</sup> we expect only little influence of the temperature dependence of the sorption coefficients  
109 and regard this as one of the minor uncertainties of our model.

#### 110 **Sorption to structural (muscle) proteins**

111 Structural proteins such as muscle protein is abundant in vertebrates and of polar nature.<sup>14</sup>  
112 Analogous to previous work,<sup>14,24</sup> we assumed the experimental sorption data from water to  
113 chicken muscle to be a generally valid proxy for the partition coefficient between structural  
114 proteins and water,  $K_{\text{structural proteins/water}}$ , for both the ionic as well as the neutral species. In fact, for

115 40 neutral chemicals it has been shown in previous work that the differences in  $K_{\text{structural proteins/water}}$   
116 between chicken, fish and pig muscle proteins were small.<sup>24</sup> We used the experimental  
117 partitioning data from,<sup>14,24</sup> comprising 63 neutral chemicals, 41 anions and 10 cations (we left out  
118 those values that are only given as lower border). In order to be used in our screening model the  
119 experimental values had to be converted into volume based partitioning coefficients (multiplied  
120 with the density of muscle protein of 1.36 kg/L).<sup>25</sup> Note that this is a rather limited dataset of  
121 chemicals. Increasing predictive errors have to be expected for chemicals that do not fall into the  
122 range spanned by the calibration data. Given that there are only ten cations in the dataset, a  
123 meaningful MLR for cations is not possible (i.e., overfitting is inevitable). This gap needs to be  
124 filled by future work. For the time being it might be advisable for the screening to just use a log  
125  $K_{\text{structural proteins}}$  (cation) value of 1.5 for any cationic chemical (being the mean value of the existing  
126 experimental data).

## 127 Sorption to albumin

128 The partitioning to blood plasma is dominated by the sorption to the plasma proteins. Among  
129 these proteins serum albumin is the major sorption matrix for both neutral and ionic  
130 chemicals.<sup>26,27</sup> We rely on two consistent experimental datasets<sup>26,27</sup> for our model development.  
131 The experimental data were derived with bovine serum albumin, which is comparable to human  
132 serum albumin.<sup>27</sup> Due to the lack of reliable partitioning data for rainbow trout albumin, we use  
133 bovine serum albumin as a surrogate. Obviously, this assumption needs to be revised when new  
134 experimental values for fish plasma protein come up and as the circumstances require, a new  
135 MLR will have to be set up. In order to be used in our screening, the experimental values were  
136 converted to volume based partition coefficients (i.e., they were multiplied with the density of  
137 serum albumin, being 1.36 kg/L).<sup>25</sup>

## 138 Methods for the development of the sorption model

### 139 Calculation of $\log K_{\text{fish/water}}$

140 The partitioning of a permanently charged ionic chemical between any organism and water can  
141 be described as the additive sorption to all the sorption matrices in the body of the organism. For  
142 ions this is expressed in the following equation for the partitioning into fish:

$$143 \quad (1)$$

144 with  $f_x$  denoting the volume fractions of the respective matrices/phases and the  $K$ 's describing the

145 partition coefficients between the matrices/phases and water given in the subscripts (trivially,  
146  $K_{\text{water/water}}$  equals one and thus only  $f_{\text{water}}$  needs to be considered). For our screening approach we  
147 looked at a 1 kg rainbow trout with the following composition (volume %): storage lipid 11%,  
148 phospholipids 1.0 %, structural proteins 15.8 %, plasma proteins 0.27 %, and water 69.8 %  
149 (adapted from Nichols et al. ).<sup>28</sup> A side note to the wording used here: a ‘phase’ is per definition  
150 homogeneous like water or hexadecane. Phospholipids and albumin are highly heterogeneous,  
151 while muscle protein is probably a little less heterogeneous<sup>14</sup> – therefore we denote these latter  
152 sorption media as (sorption) matrices.

153 When we describe the bioaccumulation potential of acids and bases that are partly neutral at the  
154 investigated pH, then the partitioning of both species needs to be assessed. For the neutral  
155 species, we also consider storage lipids (triglycerides) as a major sorbing compartment in  
156 addition to membranes, structural proteins and albumin.<sup>4</sup>

157 (2)

158 The total partition coefficients of both species are then combined according to their fractionation  
159 in water that depends on the respective  $pK_a$  value.

160 (3)

161 Note again, that this model is purely based on equilibrium partitioning and does not account for  
162 any kind of metabolism and kinetics.

### 163 **Predicting $K_{x/\text{water}}$ for neutral chemicals with pp-LFERs**

164 The partitioning of neutral chemicals to the different sorption phases/matrices listed in Eq. 2 can  
165 be predicted with poly parameter free energy relationships (pp-LFERs) from the literature. In  
166 general, pp-LFER models are widely used and accepted as documented by a number of  
167 reviews.<sup>13,29,30</sup> We used the UFZ-LSER database<sup>31</sup> in order to get a maximum amount of  
168 experimentally determined solute descriptors, L (log of the hexadecane-air partition coefficient),  
169 S (dipolarity/polarizability parameter), A (solute H-bond acidity), B (solute H-bond basicity), and  
170 V (molar volume). For cases where no experimental solute descriptors were available we used the  
171 UFZ-QSPR, available free of charge from the same source. We used these solute descriptors in  
172 the following pp-LFERs from the literature to calculate  $K_{\text{membrane/water}}(\text{neutral})$ ,  
173  $K_{\text{storage lipid/water}}(\text{neutral})$ ,  $K_{\text{structural proteins/water}}(\text{neutral})$ , and  $K_{\text{albumin/water}}(\text{neutral})$  respectively<sup>22,24,26,32</sup>:

174 ; n=131, SE=0.28, T=37°C (4)

175 ; n=247, SE=0.20, T=37°C (5)

176 ; n=46, SE=0.23, T=37°C (6)

177 ; n=82, SE=0.41, T=37°C (7)

178 In addition to the pp-LFER predictions, the partitioning of neutral chemicals to structural proteins  
179 and albumin were also predicted with multi-linear regressions (MLRs) against the sigma  
180 moments of the respective chemicals (as outlined in detail below), while the partitioning of  
181 neutral chemicals to membrane was also predicted with COSMOmic. Hence, for the neutral  
182 chemicals we ended up having two predictive models (based on the same calibration data sets)  
183 one using the pp-LFER approach and one using the sigma moments derived from quantum  
184 chemical cosmo calculations (see below). We expect that all models have their shortcomings due  
185 to the finite training set, so we decided to use a consensus model for neutral chemicals, meaning  
186 that  $K_{x/water}$  of the respective sorption matrix was finally determined by the average of the two  
187 respective model results. For storage lipid we relied solely on the ppLFER Eq. 5.

### 188 **Generation of COSMOfiles**

189 Prior to the partitioning calculations with COSMOtherm (including the calculations via  
190 COSMOmic or via sigma moments) COSMOfiles of the respective chemicals were generated  
191 with quantum mechanical calculations (BP-TZVP level):<sup>33-35</sup> We used COSMOconfX16 and  
192 Turbomole version 7.1 for full energy minimization and conformer generation (up to ten  
193 conformers were generated).<sup>36</sup>

### 194 **Predicting $K_{x/water}$ of ionic and neutral chemicals via sigma moments**

195 Analogous to the pp-LFER approach which uses 5 solute descriptors (called Abraham  
196 descriptors) the interaction possibilities of a solute can also be described with five descriptors,  
197 derived from the COSMOfile of the specific chemical. In fact, it has been demonstrated that the  
198 five Abraham solute descriptors for neutral chemicals correlate well with the following five  
199 sigma moments Sig0, Sig2, Sig3, Hb\_acc3 and Hb\_don3 – all of which can be calculated with the  
200 commercial software COSMOtherm.<sup>37</sup> Given that a) these five sigma moments are also well-  
201 suited for describing partitioning for neutral chemicals via a multi-linear regression (MLR)<sup>37,38</sup>  
202 and b) the partitioning systems of structural protein and plasma protein cannot directly be

203 modelled with COSMOtherm it is an obvious choice to use the sigma moments to describe the  
204 respective partitioning systems with a MLR of the following general form:

205 (8)

206 This is done in exact analogy to the calibration of a pp-LFER equation – but unlike the pp-  
207 LFERs, sigma moments should per se be able to describe both ionic and neutral chemicals, if we  
208 also consider the additional sigma moment Sig1, which describes the charge. A big advantage of  
209 sigma moments based MLR's over other QSAR's is, that the sigma moments describe intuitively  
210 understandable physicochemical parameters, as outlined in the SI.

### 211 *Sorption to structural (muscle) proteins*

212 For  $K_{\text{structural proteins/water}}$  (ion) a tentative ppLFER had already been set up for monovalent ions by  
213 including additional descriptors accounting for the charge.<sup>14</sup> However, this pp-LFER can only  
214 account for the ionic forms of phenols, carboxylic acids, pyridines and amines and is therefore  
215 not suited for our screening purpose. Therefore we modeled  $K_{\text{structural proteins/water}}$  via the MLR based  
216 on sigma moments as discussed above (range of the sigma moments is shown in SI Table 1).

### 217 *Sorption to albumin*

218 The sorption of ions to serum albumin is partly influenced by strong steric effects,<sup>14</sup> which can  
219 only be included in a modelling approach through extensive calibration and calculation effort.<sup>25</sup>  
220 Such a model is not feasible for our screening purpose because a) it requires a very time-  
221 consuming and meticulous calculation effort and b) its domain of applicability is rather narrow.<sup>25</sup>  
222 But we can use the existing experimental data for a simplified model (which is expected to have a  
223 wider applicability domain while predicting the fitting data set less accurate) that is based on the  
224 sigma moments as discussed above. Prior to construction of this sigma-moment based model we  
225 excluded those chemicals **whose sorption behaviour to albumin is** highly influenced by steric  
226 effects, which cannot be covered by the sigma moments. Due to our previous 3D-QSAR  
227 modeling experience<sup>25</sup> we know that especially anions that have a substitution in direct vicinity to  
228 the carboxylic group are strongly influenced by steric effects (they experience a twist of the  
229 carboxyl group). Thus, we excluded these anions from the calibration dataset,<sup>27</sup> namely 2,6-  
230 dichlorobenzoic acid anion, 2-chlorobenzoic acid anion, 2-naphthalenacetic acid anion, 2-  
231 naphthoic acid anion, and naphthalene-2-sulphonate anion. We have to note here that there are

232 not enough experimental data to provide reliable rules as to which other chemicals would fall  
233 outside the application range.

### 234 **Sorption to membrane lipid**

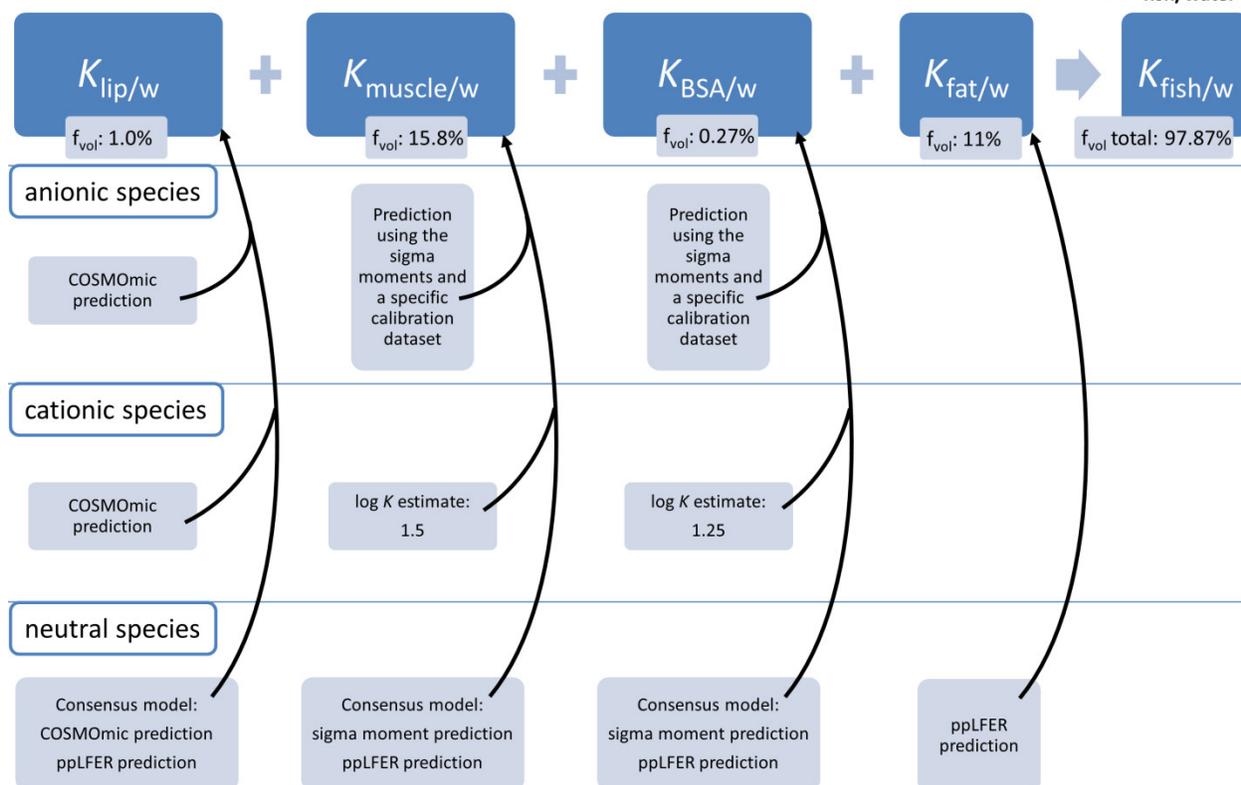
235  $K_{\text{membrane/water}}(\text{ion})$  of ionic organic chemicals can be modeled with the COSMOmic application in  
236 COSMOtherm,<sup>19</sup> which is currently the most reliable method available for this purpose<sup>6</sup> and the  
237 only prediction method that can be used for screening purposes<sup>39</sup> (in contrary to MD simulations).  
238 COSMOmic has been validated with a rather diverse dataset, including a few zwitterions and di-  
239 cations. For our screening approach, we used exactly the same calculation details as in the  
240 original COSMOmic publication: 1401 parametrization of the COSMOtherm software with an  
241 offset of 0.32 log units for the prediction of  $K_{\text{membrane/water}}$  of organic ions, using a pure DMPC  
242 membrane.<sup>19</sup>

243 For neutral chemicals the sorption to membrane lipid,  $K_{\text{membrane/water}}(\text{neutral})$  was also modeled with  
244 COSMOmic (with the same settings as used for ions), and, additionally, with the ppLFER shown  
245 in Eq. 4.

### 246 **Overall Workflow**

247 Once, all predictive models for the required partition coefficients had been set up, we were able  
248 to start the screening task. Our overall Screening workflow can be summarized as outlined in  
249 Fig. 1.

## Screening of bioconcentration potential for ions and ionic species—calculation of $\log K_{\text{fish/water}}$



250

251 **Figure 1: Workflow for our screening procedure for potentially bioaccumulative chemicals.**  
 252 **All models based on MLRs with sigma moments were newly developed in this work.**

### 253 Screening for potentially bioaccumulative chemicals

254 **For our screening for potentially bioaccumulative ions we largely investigated**  
 255 **a dataset provided by ECHA of more than 70000 non-confidential chemical**  
 256 **structures from the REACH registration database and Classification and**  
 257 **Labelling inventory.** We first filtered the dataset for those chemicals with a molecular  
 258 weight between 100 and 800, having only one  $pK_a$ . For chemicals with a  $pK_a$  between 3 and 7  
 259 both, the neutral and anionic species were considered. Here, we relied on the  $pK_a$ 's given in the  
 260 ECHA dataset, which were predicted with the ChemAxon software package. If the  $pK_a$ 's were  
 261 below 3, we only considered the anionic fraction, if the  $pK_a$ 's were above 11 we only considered  
 262 the cationic species. Also, we restricted our investigation on chemicals constructed by the atoms  
 263 H, C, N, S, O, P and halogenates. We further included some chemicals in our screening of known  
 264 environmental relevance such as perfluorinated chemicals, ionic liquids and quaternary  
 265 phosphonium cations. If adequate, we predicted the  $pK_a$  of these chemicals with JChem for Excel,  
 266 version 15.10.2600.341 (Copyright 2008-2015 ChemAxon Ltd. <https://www.chemaxon.com/>)

267 using a SMILES code as input. According to the literature, JChem performs equally well as ACD  
268 and the topological method MoKa on pKa predictions.<sup>40</sup>

## 269 **Results and Discussion**

### 270 **Models for the different sorption matrices**

#### 271 **Structural protein**

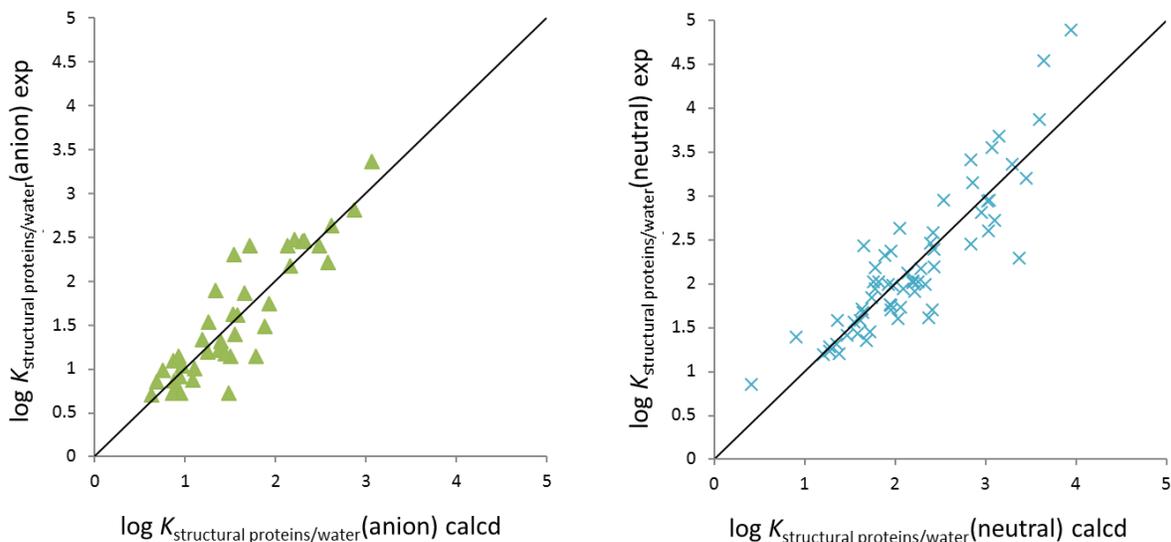
272 The Abraham solute descriptors are always positive (with the notable exception of perfluorinated  
273 chemicals and silicates), which makes the resulting pp-LFER equation instructive and easily  
274 understandable.<sup>13</sup> In contrast, Sig 1 and Sig3 can also take on negative values. This and the fact  
275 that the absolute values of the sigma moments are not normalized prohibit an easy interpretation  
276 of MLRs fitted with sigma moments as compared to pp-LFER equations.

277 In a first attempt, we fitted the experimental data of the 63 neutral chemicals, 41 anions and ten  
278 cations altogether with a MLR and obtained a promising correlation already (RMSE = 0.46, R<sup>2</sup> =  
279 0.67, SI Fig. 2). But we also assumed that differently charged chemicals might sorb to different  
280 sorption sites within the muscle proteins, so we also fitted the neutral chemicals and anions with  
281 MLRs separately. These two fits have less fitting parameters because we excluded those  
282 parameters that had a standard deviation larger than the fitted parameters themselves (resulting in  
283 three sigma moments and one constant for the anions, and four sigma moments and one constant  
284 for neutral chemicals). Additionally, the separate fits had a better statistical outcome (i.e., the  
285 RMSE was smaller and R<sup>2</sup> was higher) and are thus our first choice for screening.

286 ; R<sup>2</sup>= 0.81, RMSE=0.30, F=53, n=41 anions (9)

287 ,

288 R<sup>2</sup>= 0.78, RMSE=0.38, F=52, n=63 neutral chemicals (10)



289

290 **Figure 2: MLR based on sigma moments for structural protein (chicken muscle), left for**  
 291 **anions (3 descriptors + constant), RMSE = 0.30, R<sup>2</sup> = 0.81, right for neutral chemicals (4**  
 292 **descriptors + constant), RMSE = 0.38, R<sup>2</sup> = 0.78.**

293

294 Unfortunately, for cationic chemicals, there were only ten data points; we regard this as not  
 295 enough for a meaningful MLR. Therefore we decided to add the average  
 296  $\log K_{\text{structural proteins/water}}(\text{cation})$  value of 1.5 for cations for screening purposes as a rough estimate  
 297 (originating from the ten cations of the dataset and their  $\log K_{\text{albumin/water}}(\text{cation})$  range of 0.97 to  
 298 2.29).

299 For neutral chemicals the prediction of  $K_{\text{structural proteins/water}}(\text{neutral})$  is also possible with a pp-LFER  
 300 equation.<sup>24</sup> Analogous to the calculation of  $K_{\text{membrane/water}}(\text{neutral})$  we used a consensus model for  
 301 the neutral chemicals, averaging the outcomes of Eq.s 6 and 10.

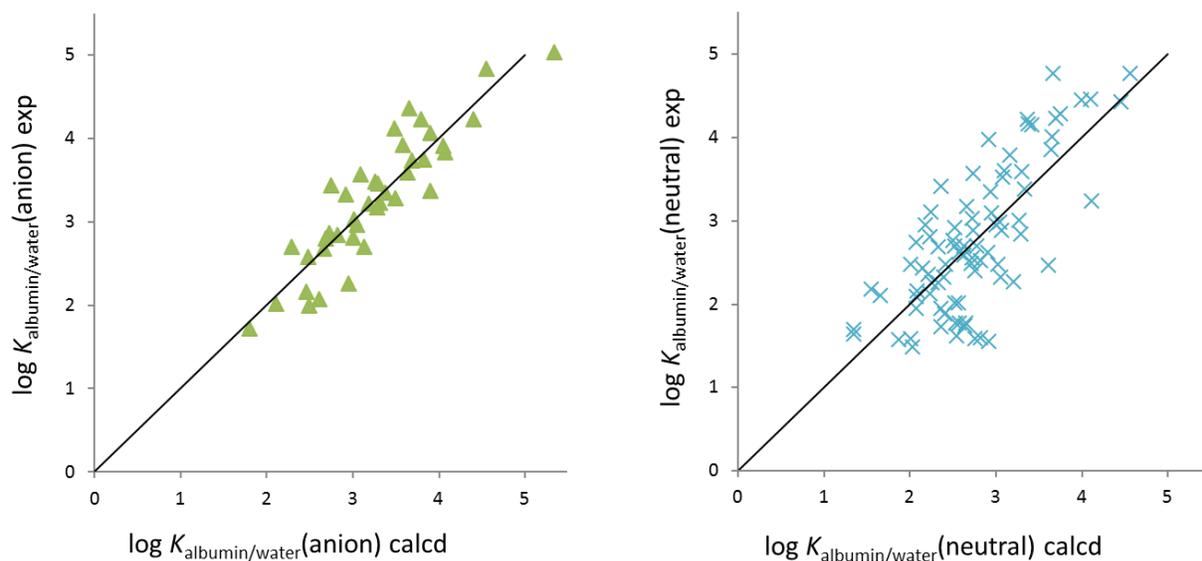
### 302 **Albumin**

303 Analogous to the structural protein, it is plausible to assume that anions and neutral chemicals  
 304 sorb to different sorption sites within the BSA protein. This can explain the rather poor fit of the  
 305 data, when the 40 anions and the 83 neutral chemicals are fitted together (SI Fig 3). The separated  
 306 fits of anions and neutral chemicals yield the following system descriptors (again leaving out  
 307 insignificant descriptors).

308 ; R<sup>2</sup>= 0.82, RMSE= 0.33, F= 39, n=40 anions (11)

309 Note that this model has to be used with caution for those anionic chemicals that are sterically  
310 hindered in vicinity to a carboxyl group, as explained above.

311 ;  $R^2= 0.56$ ,  $RMSE= 0.57$ ,  $F= 25$ ,  $n=83$  neutral chemicals (12)



312

313 **Figure 3: MLR based on sigma moments for albumin, left for anions (4 descriptors +**  
314 **constant), right for neutral chemicals (4 descriptors + constant).**

315 Again, we described the partitioning to BSA for neutral chemicals with a consensus model,  
316 averaging the results from Eq. 12 and the ppLFER Eq. 7 (SI Fig. 4).

317 As before, there are not enough data for cations to establish a MLR, so we used the average log K  
318 value of 1.25 (originating from the four cations of the dataset and their log  $K_{albumin/water}(cation)$   
319 range of 0.97 to 1.58).

### 320 **Model constraints**

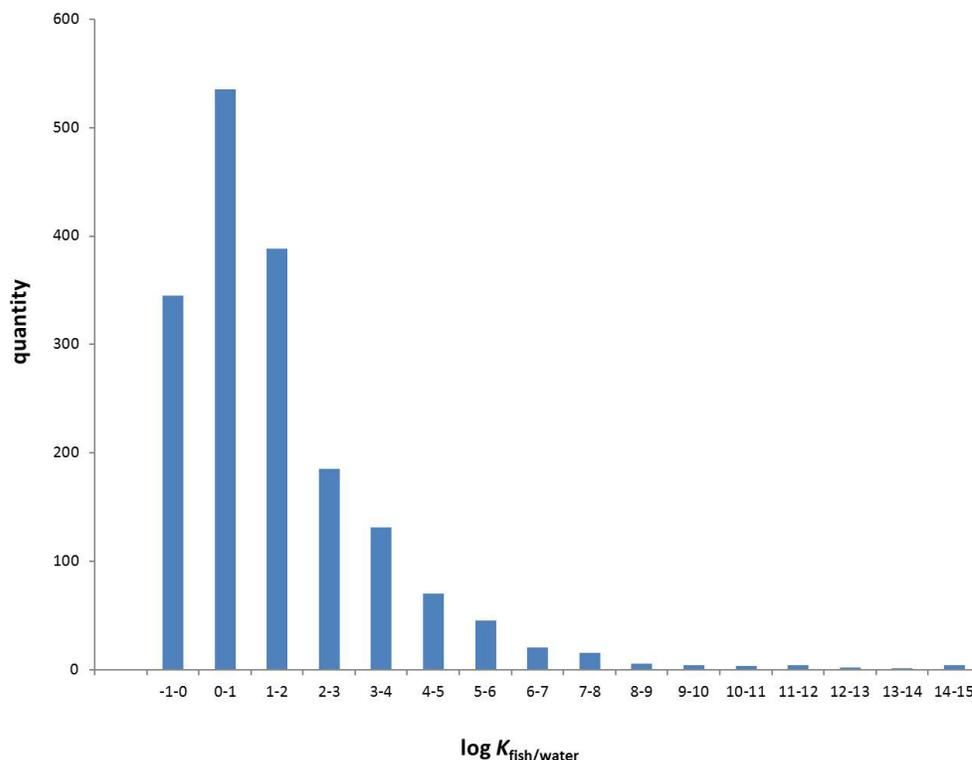
321 In order to facilitate the interpretation of the results and prevent misuse of the model, we repeat  
322 the model weaknesses in a bullet point form here.

- 323 • It is questionable whether poly- and perfluorinated chemicals are well-described with the  
324 sigma moment approach, given that van der Waals interactions are only depicted via the  
325 Sig1 (area). We therefore expect systematic deviations for perfluorinated chemicals, but  
326 due to the lack of experimental data this cannot be quantified.

- 327 • Unfortunately, also for neutral chemicals there is a lack of experimental data for  
328 perfluorinated chemicals. So also the pp-LFER based submodels for neutral chemicals  
329 can only be used with great caution for this class of chemicals
- 330 • Sorption of cations to structural proteins and plasma proteins is only roughly estimated by  
331 average values due to an insufficient number of calibration data (i.e., the sorption to  
332 serum albumin is presumably weak for cations but they sorb stronger to other plasma  
333 proteins than albumin which are not included in our screening approach due to the lack of  
334 consistent data).<sup>41</sup>
- 335 • Complex ions, i.e., ions with several ionizable groups, as well as surfactants were not part  
336 of the calibration or validation set of our models and the model performance for these  
337 chemicals/species is unknown.
- 338 • Chemicals that show a distinct steric effect in their sorption to serum albumin might not  
339 be correctly covered by our modelling approach.

340 **Screening of potentially bioaccumulative monovalent organic ions**

341



342

343 **Figure 4: Histogram of calculated  $\log K_{\text{fish/water}}$  according to Eq. 3.**

344 We screened 1839 preselected chemicals for their bioaccumulative potential, 187 (10%) of them  
 345 have predicted  $\log K_{\text{fish/water}}$  values larger than 4 (Fig 4, see SI Table 3 for the dominating sorption  
 346 phase). The molecular weight of these potentially bioaccumulative chemicals ranged between  
 347 255 and 756  $u$  thus spanning almost the entire range of the preselected values (see methods).

348

349 **Figure 5: Calculated  $\log K_{\text{fish/water}}$  (combining the contribution from the neutral and ionic species)**  
 350 **against Sig0 (area) .**

351

352  $\log K_{\text{fish/water}}$  correlates reasonably with the molecular surface area as it can be expected (Fig 5);  
 353 larger chemicals tend to be more bioaccumulative than smaller ones due to their increased  
 354 hydrophobicity. For the chemicals that possess a neutral and anionic species at a pH of 7 (acids),  
 355 the neutral species has generally the higher  $\log K_{\text{fish/water}}$  value compared to the anionic species (SI  
 356 Fig. 7). But, we also compared the pH dependent contribution of the two species to  $K_{\text{fish/water}}(\text{total})$

357 and in most of the cases the anionic species dominated the  $K_{\text{fish/water}}(\text{total})$  at pH 7 (SI Fig. 8). The  
 358 two outliers in Fig. 5 with a relatively high  $\log K_{\text{fish/water}}$  but a rather small  $\text{Sig}0$  of roughly 200 are  
 359 adamantanes, which are cubic molecules with a relatively small volume.

360 For the further discussion, we only consider the contribution and the influence of the ionic  
 361 species on the bioaccumulative potential because this is the most important contribution.

362 **Table 1: Overview of screened chemicals. Note that the sum of the chemicals in the sub-**  
 363 **groups does not always add up to the total of 1839 chemicals, because not always all**  
 364 **chemicals fit into the shown categories.**

	total	$\log K_{\text{fish/water}} < 4$		$\log K_{\text{fish/water}} > 4$	
		quantity	%	quantity	%
	<b>1839</b>	<b>1652</b>	<b>89.83%</b>	<b>187</b>	<b>10.17%</b>
aromatic	822	783	95.26%	39	4.74%
aliphatic	942	794	84.29%	148	15.71%
S based acid	409	343	83.86%	66	16.14%
C based acid	606	562	92.74%	44	7.26%
Sorbing matrix dominated by structural proteins	177	165	93.22%	12	6.78%
dominated by plasma proteins	408	354	86.76%	54	13.24%
S based acid	102	75	73.53%	27	26.47%
C based acid	223	207	92.83%	16	7.17%
aromatic	250	228	91.20%	22	8.80%
aliphatic	158	207	131.01%	32	20.25%
dominated by membrane lipids	266	171	64.29%	95	35.71%
S based acid	34	12	35.29%	22	64.71%
C based acid	33	17	51.52%	16	48.48%
aromatic	66	59	89.39%	7	10.61%
aliphatic	200	112	56.00%	88	44.00%

365

366 **Figure 6  $\log K_{\text{Fish/water}}$  ionic against the surface area of the ionic chemical. The color code**  
 367 **indicates the dominating sorption matrix that contributes for more than 60% of the total**  
 368  **$\log K$  value.**

369 Analysis of the results indicated the following general trends :

370 - aliphatic chemicals tend to be more bioaccumulative than aromatic chemicals (Tab. 1)

371 - Sorption to albumin is generally dominated by smaller chemicals while bigger molecules tend  
372 to sorb stronger to membrane lipids

373 - Structural proteins as a dominating sorption matrix (>60%) rarely leads to bioconcentration  
374 potential (Fig. 6)

375 - Albumin and membranes dominate the sorption behaviour of bioconcentrating chemicals, (Fig.  
376 6).

377 - aromatic chemicals preferably sorb to albumin while aliphatic chemicals preferably sorb to  
378 membranes (Table 1)

379 - sorption of S based acids is dominated by plasma proteins while for C based acids sorption can  
380 be dominated by both plasma proteins and membranes (Table SI 4).

381 - Sorption to plasma proteins is similar to sorption to structural proteins but higher (SI Fig. 9)

382 - Sorption to plasma proteins is considerably different to sorption to membranes but high values  
383 correlate (SI Fig. 10)

384

### 385 **Implications for the regulatory process**

386 Assessment of the bioconcentration potential of ionizable organic chemicals often assumes that  
387 ionic species do not partition into biological matrices. However, there are sufficient data in the  
388 literature to show that this general assumption does not hold. Hence, it is unacceptable to waive  
389 the bioconcentration potential of organic ions based on this assumption.. Here, we have shown  
390 that a first tier screening of ions is possible, based on molecular descriptors that came from  
391 quantum chemical calculation. It must be noted that our assessment was based on the assumption  
392 that the exposure pH value is the same as the internal pH of the fish (i.e., pH=7). If this is not the  
393 case then an ion-trap effect will occur which further increases the bioconcentration potential for  
394 organic acids <sup>42</sup>. If the pH in the exposure medium is 2 log units smaller (i.e., pH= 5) then the  
395 BCF increases by a factor 100 for acids with a pKa < 5 due to ion trapping.

396 We also point out that the presented screening was only directed towards bioaccumulation in fish.  
397 Many of the chemicals that are not expected to have a BCF potential may still have a substantial

398 bioaccumulation potential in terrestrial organisms. Chemicals that are not volatile ( $\log K_{oa} > 5$ )  
399 and that do not metabolize, possess a bioaccumulative potential in terrestrial organisms if their  
400  $\log K_{organism/water}$  exceeds 1<sup>43–46</sup>. If we take the  $K_{fish/water}$  as a proxy for a more general  $K_{organism/water}$  for  
401 all vertebrates, then a large portion of the ionizable chemicals tested here would be classified as  
402 potentially bioaccumulative in air breathing organisms. Of course, any initial screening for  
403 partition properties has to be followed by more specific testing that also considers  
404 biotransformation in the test organisms.

405

## 406 **Conflicts of interest**

407 There are no conflicts to declare

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