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2 **Equilibrium Biopartitioning of Organic Anions – a Case Study for**

3 **Humans and Fish**

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13

14 **Abstract**

15 In this work we combine partition coefficients between water and membrane lipid, storage lipids, the
16 plasma protein albumin as well as structural protein with the tissue dependent fraction of the respective
17 phases in order to obtain a clearer picture on the relevance of various biological tissues for the
18 bioaccumulation of 31 organic anions. Most of the partition coefficients are based on experimental data,
19 supplemented by some predicted ones. The data suggest that the plasma protein, albumin, will be the major
20 sorption matrix in mammals. Only small fractions of the studied chemicals will occur freely dissolved in an
21 organism. For the investigated acids with $pK_a < 5$, partitioning is dominated by the ionic species rather than
22 the corresponding neutral species. Bioconcentration in fish is not expected to occur for many of these acids
23 unless pH in the aqueous environment is low or specific sorption mechanisms are relevant. In contrast,
24 biomagnification in terrestrial mammals would be expected for most organic anions if they are not
25 sufficiently metabolized. We conclude that sorption is important for the toxicokinetics of ionizable organic
26 chemicals and the dominating sorbing matrices are quite different from those for neutral species.

27

28 **Keywords: bioaccumulation, organic anions, partitioning, bioconcentration, risk assesement,**
29 **chemical regulation**

30 1. Introduction

31 Among the many organic chemicals that have to be regulated there are many that are ionizable (i.e., have
32 at least one ionic species depending on the pH). In the regulation procedure these ionizable chemicals need
33 to be assessed –among others– for their bioaccumulation potential and their toxicity. For these tasks it is
34 fundamental that we understand the biopartitioning of ionic species. For neutral organic chemicals our
35 understanding of biopartitioning has reached a rather high level(Endo et al., 2013) but it is still in its infancy
36 for ionic organic species. In recent papers we have investigated the partitioning of organic ionic species to
37 distinct biological matrices such as phospholipids, structural proteins and albumin.(Bittermann et al., 2014;
38 Henneberger et al., 2016a, b) Here, we combine this information in order to shed some light on the partition
39 behavior of some typical organic acid anions in organisms. Cations were excluded from this study due to
40 limited data availability. Armitage et al. (Armitage et al., 2013) have presented a toxicokinetic model for the
41 bioaccumulation of ionizable chemicals in fish. Their model had a focus on kinetic processes and was built
42 on the assumption that partition coefficients of ionic species can be estimated by subtracting an empirically
43 found constant from the logarithmic partition coefficients of the corresponding neutral species. Similar
44 approaches have also been used by others in models for the bioconcentration of ionizable chemicals in fish
45 (Erickson et al., 2006; Fu et al., 2009). This approach goes back to a work of Escher and Schwarzenbach on
46 membrane partitioning of phenolic acids (Escher and Schwarzenbach, 1996). However, in a later work
47 Escher and Sigg presented data revealing that correction by a single constant does not really work when
48 different compound classes need to be covered (Escher and Sigg, 2004). Our own work also does not
49 support this approach (Bittermann et al., 2014). In the pharmaceutical literature there is also the attempt to
50 relate anionic partition coefficients to the K_{ow} of the neutral species based on an empirical relationship
51 calibrated with a small number of experimental data (Rodgers and Rowland, 2006). For our work here we
52 concluded that it was important to use reliable, mostly experimental partition coefficients rather than less
53 reliable estimated values in order to analyse the equilibrium partition preferences of ions in organisms. The
54 specific goal of this work was to answer three questions: 1) What is the relative importance of different
55 sorption matrices (proteins, phospholipids, triglycerides) for the partitioning of acidic chemicals in an
56 organism (here a typical human is used as an example)? 2) Can we infer a distinct organ-specific

57 partitioning of acidic chemicals from the available data using composition and size of organs that are typical
58 for humans? 3) What is the expected bioaccumulative behavior of non-metabolized, acidic chemicals in fish
59 (here a rainbow trout) at a physiological pH of 7.4 and at various external pH values. For all questions, we
60 start with considering the ionic species alone and then discuss the partitioning of the total compound (i.e.,
61 neutral and ionic species).

62

63 2. **Methods**

64 2.1 *General Approach*

65 In our recent work we were able to generate equilibrium sorption data for about 40 organic anions in
66 structural proteins, albumin and phospholipids.(Bittermann et al., 2014; Henneberger et al., 2016a, b) These
67 data were supplemented with data from the literature and will serve as the basis for the work presented here.
68 It was not possible though to base the work solely on experimental data because there is only limited overlap
69 between the investigated ionic chemicals from the different datasets. We therefore had to resort to estimation
70 methods for ionic partition coefficients in order to fill data gaps. Available estimation methods for ionic
71 partitioning to albumin and structural proteins have a rather narrow application domain (Henneberger et al.,
72 2016b; Linden et al., 2017) and were avoided. Thus, we decided to use only those ions for which we had
73 experimental values for their partitioning from water to albumin and to structural protein. For four of these
74 chemicals we also had experimental phospholipid-water partition coefficients (4-n-octylbenzenesulfonate,
75 bromoxynil, ibuprofen and pentachlorophenol). For the remaining ions, the phospholipid-water partition
76 coefficients were derived from a quantum chemically based approach, namely COSMOmic, which we
77 expect to possess a very wide application range (Bittermann et al., 2014). This procedure left us with 31
78 organic acid anions representing pharmaceuticals, pesticides and other chemicals. Most of these chemicals
79 are carboxylic acids (n=23), particularly benzoic acids (n=8) and naphthoic acids (n=7).

80 For the partitioning of the neutral species we largely rely on pp-LFER predictions as outlined in earlier
81 work (Endo and Goss, 2011; Endo et al., 2012; Geisler et al., 2012). Both, experimental partition
82 coefficients as well as those from the pp-LFER equations are for 37°C.

83 This work is limited to the equilibrium partitioning of ions although kinetic uptake and elimination
84 processes may strongly affect the actual concentrations in organisms. In fact, kinetics may be more
85 important for ionic species than for neutral ones because ions are known to have a much lower membrane
86 permeability (Saparov et al., 2006) and are expected to be more susceptible to active transport processes
87 across membranes (Roth et al., 2012). Nevertheless, equilibrium partitioning must be understood first before
88 more complex kinetically controlled scenarios can be investigated.

89 All calculations here assume that partitioning within an organism can be treated as additive with respect
90 to the different sorption matrices of an organ or organism.. It is also assumed that partition coefficients that
91 have been measured with isolated proteins and lipids are representative of *in vivo* partitioning. These are
92 standard assumptions in pharmaco- and toxicokinetic modeling and at least for neutral organic chemicals
93 they are supported by some experimental evidence (Endo et al., 2013) For plasma-muscle partitioning of
94 ions we have some qualitative indication that this approach should work as well (see SI Figure S7 in
95 (Henneberger et al., 2016b)). Again following pharmacokinetic practice we assume that water, storage
96 lipids, phospholipids, structural proteins and albumin are the relevant sorbing matrices for neutral organic
97 species. For ionic species we follow the same approach but ignore storage lipids (triglycerides). This is
98 supported by experimental evidence showing that partitioning of ions to organic solvents is orders of
99 magnitude smaller than partitioning to phospholipids (Escher and Sigg, 2004). Neglect of storage lipids as
100 sorbing compartment for ions is further supported by energetic considerations: a sorbing organic matrix
101 likely needs ionic (or zwitterionic) functional groups in order to efficiently compete with the strong water
102 dipole for organic ions because of the rather attractive ion-dipole interactions that occur in water. While
103 proteins and phospholipids possess such ionic functions, storage lipids do not. A comprehensive
104 investigation of ionic partitioning in organisms also needs to include ion-trap effects that arise from pH
105 differences within the body and/or between an organism and surrounding water (Fu et al., 2009;
106 Neuwoehner and Escher, 2011).

107 *2.2 Specific Calculations*

108 Table 1 shows the fractional composition for a total human and for the major organs that we have used for
109 our calculations.

110

111 **Table 1**

112 Composition of a woman (BW = 60 kg, H = 163 cm, BMI = 22.6 kg/m²). Blood volumes within the organs
 113 were estimated with Krogh’s cylinder model (Krogh, 1922) and subtracted from the organ volumes reported
 114 in (Willmann et al., 2007). Composition of the organs was taken from (Schmitt, 2008a; Schmitt, 2008b)
 115 Interstitial volume in the organs and albumin concentration therein was taken from (Schmitt, 2008b) and
 116 (Ellmerer et al., 2000) respectively. All values are given in mL.

| | total | volume of | volume of | volume of | volume of | volume of | volume |
|--------------|---------------------|---------------------|------------------|-------------------|------------------|-------------------|-----------------|
| | volume of | phospholipid | storage | structural | water | in | of |
| | (without | s | lipids | protein | | interstiti | al |
| | capillaries) | | | | | al | space of |
| | | | | | | | organs |
| Adipose | 22076 | 42.5 | 15177 | 990 | 5845 | 20.66 | |
| Brain | 1311 | 89.1 | 56.9 | 106 | 1059 | 0.04 | |
| Gut | 1223 | 21.9 | 48.7 | 151 | 1000 | 0.83 | |
| Heart | 343 | 14.8 | 13.7 | 49.1 | 265 | 0.32 | |
| Kidneys | 427 | 12.1 | 4.3 | 57.3 | 352 | 0.59 | |
| Liver | 1843 | 73.3 | 30.4 | 275 | 1463 | 2.21 | |
| Lung | 1034 | 7 | 7.3 | 39.3 | 977 | 2.75 | |
| Muscle | 19114 | 79.1 | 76 | 2946 | 15997 | 16.26 | |
| Skin | 3516 | 25.7 | 231 | 751.4 | 2501 | 6.33 | |
| Spleen | 231 | 2.1 | 0.9 | 34.6 | 193 | 0.31 | |
| Gonads | 12 | 0.3 | 0 | 1.3 | 10 | 0.01 | |
| Blood | 4800 | 16.3 | 15.8 | 624 | 3986 | 158 | |
| total | 55929 | 384 | 15662 | 6025 | 33649 | 208.32 | |

117

118

119 This total composition from Table 1 together with the corresponding partition coefficients listed in the
 120 Supporting information (SI Table S1) allows the calculation of the relative sorption capacity for each ionic
 121 species, with the following formulas:

122

(1)

123

(2)

124 (3)

125 (4)

126
127 where $f_{i \text{ matrix}}$ is the fraction of the total amount of a chemical species i that resides in 'matrix' at
128 equilibrium, V is the volume of the respective biological matrix (see Table 1) and $K_{i,xy}$ is the equilibrium
129 partition coefficient between matrix x and y (in L/L). When needed these were calculated from weight
130 based partition coefficients in SI Table S1 using density of 1.36 kg/L for all proteins. Subscripts are used as
131 follows: w – water, alb –albumin, prot – structural proteins, p.lip - phospholipids. Partition coefficients
132 between two matrices are calculated via the thermodynamic cycle using the known matrix-water partition
133 coefficients. The relative concentrations in the matrices at equilibrium are directly given by these partition
134 coefficients.

135 The partitioning of the total chemical (ionic plus neutral species) is obtained by applying above formulas
136 to the neutral species as well (including storage lipids as additional sorbing matrix) and then weighing the
137 results for both species according to their fractionation in water that is calculated from the Henderson and
138 Hasselbalch equation:

139 (5)

140 with α_{acid} being the fraction of the neutral acid and $(1- \alpha_{i \text{ acid}})$ as the fraction of the acid anion. The pK_a values
141 used for this calculation are listed in SI Table S2.

143 3. Results & discussion

144 3.1 Matrix specific partitioning

145 For a typical human adult with a composition as shown in Table 1, the expected relative sorption
146 capacity of different matrices for each anion is shown in Figure 1.

147
148 **Fig. 1.** Relative sorption capacity (calculated as product of partition coefficients and relative volume of the
149 matrix) of different matrices for the investigated anions, sorted according to chemical classes (all data refer
150 to the anionic species although the names of the neutral chemicals are shown).

151

152 Almost all anions are predominately sorbed to albumin, with structural proteins and phospholipids as the
153 next relevant sorbing matrices. Despite its dominating contribution to the volume of a human, water plays
154 only a minor role for hosting these chemicals. Besides the relative distribution of the total internal amount of
155 a chemical it is also interesting to look at the relative concentrations in various matrices (see SI Table S3). In
156 this case, albumin is even more dominating than for the mass distribution because albumin has the highest
157 loadings while contributing little to the total volume of humans.

158 Figure 2 shows the expected relative sorption capacity for the neutral species of the tested organic acids.
159 Storage lipids are typically expected to dominate the partitioning of neutral organic chemicals. It is therefore
160 interesting to see that the picture is somewhat different for these rather polar neutral molecules. For all of
161 them, storage lipids, phospholipids and structural proteins contribute in the same order of magnitude to the
162 overall accumulation in the organism.

163

164 **Fig. 2.** Relative sorption capacity (calculated as product of partition coefficients and relative volume of the
165 matrix) for the neutral species of the tested organic acids based on estimated partition coefficients (listed in
166 SI Table S4) and matrix volumes (Table 1).

167

168 The distribution of the sum of both molecular species of each chemical at pH 7.4 is given in SI Table S5.
169 It differs little from the situation shown for the ionic species alone in Figure 1 because the anionic fraction in
170 water of all chemicals studied here amounts to more than 99% of the total chemical at the physiological pH
171 of 7.4. Sorption of the neutral species is not sufficiently extreme to compensate for this fractionation.

172 *3.2 Organ specific partitioning*

173 Among all matrices investigated above, albumin has shown the highest affinity for organic anions. Most of
174 the albumin (ca. 75%) is located in the blood and so blood is estimated to carry the highest load of organic
175 anions as compared to the organs (see Figure 3 and SI Table S6). Differences in the calculated mass
176 loadings of different organs are mostly due to the different volume of these organs. Hence, muscles are the
177 next important compartment for storing organic anions followed by the skin and adipose. This picture does

178 not change much if both species of each chemical are considered together (i.e., anionic and neutral species),
179 as already discussed above: at pH 7.4 the neutral species make up less than 1 % of the total aqueous
180 concentration and their respective partition coefficients are not high enough to let neutral species dominate
181 the overall partitioning of the investigated chemicals. We have to note here that in reality there might be
182 organ specific sorption processes for ions that are not considered in our modeling approach because too little
183 systematic knowledge about such processes is available. A prominent example for such a specific binding is
184 the binding of long chain perfluorinated acids to fatty acid binding proteins in the liver (Luebker et al., 2002;
185 Woodcroft et al., 2010; Zhang et al., 2013).

186

187

188

189

190 **Fig. 3.** Sorption capacity (calculated as product of partition coefficients and relative volume of the organs)
191 of different organs of a human (see Table 1 for characteristics) for various organic anions (although the
192 names of the neutral species were used in the graphic all data apply to the anionic species).

193

194 *3.3 Bioconcentration in fish*

195 So far, our work was only concerned with partitioning within organisms. For the bioaccumulation
196 perspective, partitioning between the environment and the organism is important. The relevant metric for
197 regulation is currently the bioconcentration factor (BCF) which is defined as the steady-state concentration
198 in fish over the concentration in the water it is exposed to. This BCF should not exceed the value of 2000 for
199 a fish with 5 % lipids. While standardization of total lipid content is useful for neutral hydrophobic
200 chemicals it is irrelevant for ionic chemicals because their sorption is not dominated by storage lipids but by
201 proteins and to some extent by phospholipids as shown above. For our calculations here, we looked at a 1 kg
202 rainbow trout with the following composition in volume %: storage lipids 11%, phospholipids 1.0 %,
203 structural proteins 15.8 %, blood proteins 0.27 %, water 69.8 % (adapted from (Nichols et al., 1990)). For
204 our calculations we assumed that the partition coefficient for bovine serum albumin can also be used for fish

205 blood transport protein although we have no further evidence for this assumption. We note that this is a
206 major uncertainty in our model approach. Partition equilibrium can be seen as a worst case estimation for
207 bioconcentration because it does not account for any metabolism. Here, we use the organism-water partition
208 coefficients of the anionic and neutral species as an indicator of bioaccumulation trends of different
209 ionizable organics that could be used in a lower tier screening of chemicals in the regulatory process. Figure
210 4 shows how the different sorption matrices contribute to the overall distribution between rainbow trout and
211 water for the neutral and the respective anionic species (the indicated relative contributions refers to the non-
212 logarithmic K -value).

213

214

215 **Fig. 4.** Logarithmic partition coefficients for the neutral and the anionic species between water and whole
216 fish (rainbow trout). The color coding indicates how different sorption matrices contribute to this overall
217 partitioning (note that the indicated relative contributions refer to the non-logarithmic K -value).

218

219 For most chemicals (74 %) the neutral species exhibit stronger sorption from water into fish than their
220 corresponding ionic species. In those cases where we find the ionic species to possess a higher affinity for
221 fish than the neutral species this is always due to an exceptionally high partitioning to albumin. The latter is
222 likely due to a combination of a favorable three dimensional structure and favorable ionic/and or H-bond
223 interactions (Linden et al., 2017). The neutral species of these chemicals will have almost the same 3D
224 structure but not the same ionic/and or H-bond interactions. Our partition coefficient to albumin for neutral
225 species are all predicted with a model that does not account for 3D effects. We can thus not be sure whether
226 these values are really correct or whether –for these chemicals- the neutral form would have actually have a
227 much higher affinity to albumin so that the neutral species would still dominate in the total affinity to fish.

228 The overall sorption of the investigated chemicals at physiological pH is always dominated by the ionic
229 species, simply because all studied chemicals are > 99% ionized at pH 7.4, while the equilibrium partition
230 coefficients of the ions are less than 100 times smaller than that of their corresponding neutral species. Thus,
231 for the chemicals studied here the organism/water partition coefficients of the ionic species can also be used

232 as an approximation of the total partition coefficients for the sum of both species (anionic and neutral) for
233 each chemical as long as the pH value in the exposure medium is identical to the physiological pH.

234 If ionizable chemicals partition between two aqueous compartments that are separated by a membrane,
235 which is only permeable to the neutral species, then partition equilibrium is reached when the concentration
236 of the neutral species is the same on both sides of the membrane. If, however, the pH value on both sides of
237 the membrane differs then the equilibrium concentration of the ionized species is different on both sides of
238 the membrane and hence also the total concentration of the ionizable chemical (Neuwoehner and Escher,
239 2011) This so-called ion-trap effect will increase the bioaccumulation of acids if the pH in the exposure
240 water is smaller than the internal pH in the organism. The ion-trap effect can easily be calculated for the
241 worst-case assumption that the ions have no membrane permeability at all:

$$(6)$$

243 Hence, at pH 6.4 in the exposure medium there would be an increase in the sorption coefficients of most
244 acids by one log unit for an organism that keeps up an internal pH of 7.4. In this case two of the investigated
245 chemicals would slightly exceed the BCF threshold of 2000 (assuming that no metabolism takes place):
246 pentachlorophenol and 4-n-octylbenzenesulfonate. At pH 5.4 which is the lower end of what one would
247 typically expect for natural waters more than 50% of the discussed anions would exceed the BCF of 2000 if
248 no metabolism occurs.

249 For biomagnification in terrestrial organisms the acceptable thresholds for equilibrium partitioning are
250 much lower than for bioconcentration. For chemicals that are not volatile ($\log K_{oa} > 5$) and that do not
251 metabolize, $\log K_{organism/water}$ would have to be smaller than 1 in order to be assessed as non-bioaccumulative
252 based on purely physico-chemical reasoning (Czub and McLachlan, 2004; Armitage and Gobas, 2007; Kelly
253 et al., 2007; Goss et al., 2013). The partition coefficients shown in Figure 4 for fish can also be taken as an
254 approximation for terrestrial vertebrates because the overall composition does not vary greatly between fish
255 and other vertebrates (compare data from Table 1 and (Nichols et al., 1990)). One can see that most
256 chemicals exceed this threshold of $\log K_{organism/water} > 1$. Thus, bioaccumulation for acidic chemicals might be
257 more of an issue for terrestrial organisms than for aquatic species. Another effect that might contribute to the

258 bioaccumulation of ionizable chemicals, which is not relevant for hydrophobic neutral molecules, is active
259 transport across membranes with carrier proteins (Ng and Hungerbühler, 2014).

260 The discussion in this work gives a first idea of certain trends when it comes to the biopartitioning of
261 organic acids. For a complete picture, more processes will have to be considered in a quantitative way (e.g.,
262 active transport, ion-trap effects within organisms in different organelles, metabolism, electric fields across
263 membranes). In addition to the organic acids investigated here, organic bases, zwitterions and permanent
264 ions also have to be considered.

265 The bioaccumulation metrics BCF and BMF are defined for steady state situations in which the sum of all
266 kinetic uptake processes matches the sum of all kinetic clearance processes. For chemicals with low or no
267 metabolism these accumulation metrics are close to what one can expect from equilibrium partitioning,
268 which is what we discussed here. However, the more important metabolism becomes the more these metrics
269 become kinetically controlled which means that also uptake kinetics and thus membrane permeability
270 become an issue (Armitage et al., 2013).

271 We are just at the beginning of a sound mechanistic understanding of the toxicokinetics of organic ions in
272 organisms and there is a long way ahead. We hope that this work helps to raise awareness for the existing
273 knowledge gap and that more work in this direction can be triggered.

274

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- 281 Armitage, J.M., Arnot, J.A., Wania, F., Mackay, D., 2013. Development and evaluation of a mechanistic
282 bioconcentration model for ionogenic organic chemicals in fish. *Environ. Toxicol. Chem.* 32, 1-14.
- 283 Armitage, J.M., Gobas, F.A.P.C., 2007. A terrestrial food-chain bioaccumulation model for POPs. *Environ. Sci. Technol.*
284 41, 4019-4025.
- 285 Bittermann, K., Spycher, S., Endo, S., Pohler, L., Huniar, U., Goss, K.-U., Klamt, A., 2014. Prediction of phospholipid-
286 water partition coefficients of ionic organic chemicals using the mechanistic model COSMOmic. *J. Phys. Chem. B* 118,
287 14833-14842.
- 288 Czub, G., McLachlan, M.S., 2004. Bioaccumulation potential of persistent organic chemicals in humans. *Environ. Sci.*
289 *Technol.* 38, 2406-2412.
- 290 Ellmerer, M., Schaupp, L., Brunner, G.A., Sendlhofer, G., Wutte, A., Wach, P., Pieber, T.R., 2000. Measurement of
291 interstitial albumin in human skeletal muscle and adipose tissue by open-flow microperfusion. *American Journal of*
292 *Physiology-Endocrinology and Metabolism* 278, E352-E356.
- 293 Endo, S., Bauerfeind, J., Goss, K.U., 2012. Partitioning of neutral organic compounds to structural proteins. *Environ.*
294 *Sci. Technol.* 46, 12697-12703.
- 295 Endo, S., Brown, T.N., Goss, K.-U., 2013. A General Model for Estimating Partition Coefficients to Organisms and their
296 Tissues using the Biological Compositions and Polyparameter Linear Free Energy Relationships. *Environ. Sci. Technol.*
297 47, 6630-6639.
- 298 Endo, S., Goss, K.U., 2011. Serum Albumin Binding of Structurally Diverse Neutral Organic Compounds: Data and
299 Models *Chem. Res. Toxicol.* 24, 2293-2301.
- 300 Erickson, R.J., McKim, J.M., Lien, G.J., Hoffman, A.D., Batterman, S.L., 2006. Uptake and elimination of ionizable
301 organic chemicals at fish gills: I. Model formulation, parameterization, and behavior. *Environ. Toxicol. Chem.* 25,
302 1512-1521.
- 303 Escher, B.I., Schwarzenbach, R.P., 1996. Partitioning of Substituted Phenols in Liposome-Water, Biomembrane-
304 Water, and Octanol-Water Systems. *Environ. Sci. Technol.* 30, 260-270.
- 305 Escher, B.I., Sigg, L., 2004. Chemical Speciation of Organics and of Metals at Biological Interphases. in: van Leeuwen,
306 H.P., Köster, W. (Eds.). *Physicochemical Kinetics and Transport at Biointerfaces*. Wiley & Sons, pp. 205-242.
- 307 Fu, W.J., Franco, A., Trapp, S., 2009. Methods For Estimating The Bioconcentration Factor Of Ionizable Organic
308 Chemicals. *Environ. Toxicol. Chem.* 28, 1372-1379.
- 309 Geisler, A., Endo, S., Goss, K.-U., 2012. Partitioning of Organic Chemicals to Storage Lipids: Elucidating the
310 Dependence on Fatty Acid Composition and Temperature. *Environ. Sci. Technol.* 46, 9519-9524.
- 311 Goss, K.U., Brown, T.N., Endo, S., 2013. Elimination half-life as a metric for the bioaccumulation potential of
312 chemicals in aquatic and terrestrial food chains. *Environ. Toxic. Chem.* 32, 1663-1671.
- 313 Henneberger, L., Goss, K.U., Endo, S., 2016a. Equilibrium Sorption of Structurally Diverse Organic Ions to Bovine
314 Serum Albumin. *Environ. Sci. Tech.* 50, 5119-5126.
- 315 Henneberger, L., Goss, K.U., Endo, S., 2016b. Partitioning of Organic Ions to Muscle Protein: Experimental Data,
316 Modeling, and Implications for in Vivo Distribution of Organic Ions. *Environ. Sci. Tech.* 50, 7029-7036.
- 317 Kelly, B.C., Ikononou, M.G., Blair, J.D., Morin, A.E., Gobas, F.A.P.C., 2007. Food web-specific biomagnification of
318 persistent organic pollutants. *Science* 317, 236-239.
- 319 Krogh, A., 1922. *The anatomy and physiology of capillaries*. Yale University Press, New Haven.
- 320 Linden, L., Goss, K.U., Endo, S., 2017. 3D-QSAR predictions for bovine serum albumin-water partition coefficients of
321 organic anions using quantum mechanically based descriptors. *Environ. Sci.: Processes Impacts* 19, 261-269.
- 322 Luebker, D.J., Hansen, K.J., Bass, N.M., Butenhoff, J.L., Seacat, A.M., 2002. Interactions of fluorochemicals with rat
323 liver fatty acid-binding protein. *Toxicology* 176, 175-185.
- 324 Neuwoehner, J., Escher, B.I., 2011. The pH-dependent toxicity of basic pharmaceuticals in the green algae
325 *Scenedesmus vacuolatus* can be explained with a toxicokinetic ion-trapping model. *Aquatic Toxicology* 101, 266-275.
- 326 Ng, C.A., Hungerbühler, K., 2014. Bioaccumulation of Perfluorinated Alkyl Acids: Observations and Models.
327 *Environmental Science & Technology* 48, 4637-4648.
- 328 Nichols, J.W., McKim, J.M., Andersen, M.E., Gargas, M.L., Clewell, H.J., Erickson, R.J., 1990. A physiologically based
329 toxicokinetic model for the uptake and disposition of waterborne organic-chemicals in fish. *Toxicology and Applied*
330 *Pharmacology* 106, 433-447.
- 331 Rodgers, T., Rowland, M., 2006. Physiologically based pharmacokinetic modelling 2: Predicting the tissue distribution
332 of acids, very weak bases, neutrals and zwitterions. *Journal Of Pharmaceutical Sciences* 95, 1238-1257.

333 Roth, M., Obaidat, A., Hagenbuch, B., 2012. OATPs, OATs and OCTs: the organic anion and cation transporters of the
334 SLCO and SLC22A gene superfamilies. *British Journal of Pharmacology* 165, 1260-1287.
335 Saparov, S.M., Antonenko, Y.N., Pohl, P., 2006. A new model of weak acid permeation through membranes revisited:
336 Does Overton still rule? *Biophysical Journal* 90, L86-L88.
337 Schmitt, W., 2008a. Corrigendum to "General approach for the calculation of tissue to plasma partition coefficients".
338 *Toxicology In Vitro* 22, 1666.
339 Schmitt, W., 2008b. General approach for the calculation of tissue to plasma partition coefficients. *Toxicology in*
340 *Vitro* 22, 457.
341 Willmann, S., Hohn, K., Edginton, A., Sevestre, M., Solodenko, J., Weiss, W., Lippert, J., Schmitt, W., 2007.
342 Development of a physiology-based whole-body population model for assessing the influence of individual variability
343 on the pharmacokinetics of drugs. *Journal Of Pharmacokinetics And Pharmacodynamics* 34, 401-431.
344 Woodcroft, M.W., Ellis, D.A., Rafferty, S.P., Burns, D.C., March, R.E., Stock, N.L., Trumpour, K.S., Yee, J., Munro, K.,
345 2010. Experimental Characterization Of The Mechanism Of Perfluorocarboxylic Acids' Liver Protein Bioaccumulation:
346 The Key Role Of The Neutral Species. *Environ. Toxicol. Chem.* 29, 1669-1677.
347 Zhang, L.Y., Ren, X.M., Guo, L.H., 2013. Structure-Based Investigation on the Interaction of Perfluorinated
348 Compounds with Human Liver Fatty Acid Binding Protein. *Environmental Science & Technology* 47, 11293-11301.
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