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Costs of implementing agricultural soil protection policies – Insights from two German cases

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Abstract

Transaction costs (TCs) are often claimed to be a key determinant of how policies are actually implemented on the ground and what effect they ultimately deliver on soil quality and functions. Focussing on agriculture-related soil protection policies in Eastern Germany, we analyse data from key informant interviews in two case study areas (Brandenburg and Saxony-Anhalt) in order to provide new evidence that TCs do indeed matter for policy implementation. We systematically map TCs that occur at the policy implementation and operation stages and their drivers. Our data showed that in addition to TCs for ‘information management’ and ‘coordination’, existing frameworks need to be extended to explicitly consider TCs for ‘enforcement’. Results illustrate that there is a broad range of TCs that are due to the complexity of soils and their management, property rights assignment and administrative processes. To some extent TCs in one policy arena can be reduced; however, often they are only superseded in place and time and, moreover, there are trade-offs between different kinds of TCs. The paper emphasises that every assessment of effective policy implementation requires a specification of TCs and over what time-frame they occur.

Keywords

Transaction cost economics, soil conservation, agricultural policies, policy evaluation, Brandenburg, Saxony-Anhalt

1 Introduction

Soil degradation as a result of intensive agriculture continues to be a serious issue in Europe and worldwide (EC, 2006; Boardman & Poesen, 2006; Banwart, 2011). Several agricultural and environmental policies at EU and national level try to address soil conservation aspects, such as agri-environmental schemes (AES), the Nitrates Directive, Cross Compliance Regulation, as well as national Soil Protection Acts and Nature Conservation Schemes (Louwagie, Gay & Sammeth, 2011). However, all these policies have in common that they address soil as a by-product with limited effect on the ground (Prager, Hagemann, Schuler & Heyn, 2011; Louwagie et al., 2011). There is a need to better understand why existing policy tools are not effective in addressing soil degradation.

38 Previous studies of decision making processes in other policy fields (Schleyer & Theesfeld, 2011;
39 McCann, 2013; Alexandrescu, Martinat, Klusáček & Bartke, 2014) suggest that an analysis of the
40 institutional framework can shed light on the challenges of implementing policies with a soil
41 conservation focus. Indeed, Dobbs (2012) regards institutional analysis as “key for fostering
42 agricultural sustainability”. Transaction costs (TCs) are a central concept in institutional analysis, but
43 as Birner & Wittmer (2009) note, TCs are seldom empirically measured in analyses. The
44 acknowledgement of transaction costs, especially in the design of policy instruments, “enables the
45 analyst to bring in practical issues that are normally ignored” (McCann, 2013, p. 260). Moreover, TCs
46 are not only important for setting-up policies but also for running them (Vatn, 2010), and they can
47 contribute to understanding and informing policy processes. Accounting for TCs can further help
48 “evaluate current policies in order to improve their effectiveness” (McCann, Colby, Easter, Kasterine
49 & Kuperan, 2005, p. 528). Rørstad, Vatn & Kvakkestad (2007, p. 1) emphasize that “the cost of
50 managing a policy may be as important for efficiency as the cost of producing the goods and
51 services”. Understanding where, when and why TCs occur is important in order to reduce the
52 administrative burden, bureaucratic workload and costs for the implementing authorities, as well as
53 for farmers and other land managers.

54 So far, the academic literature has mainly covered TCs that occur for the implementation of agri-
55 environmental schemes (AES) (*inter alia* Beckmann, Eggers & Mettepenningen, 2009; Matzdorf, Pierr
56 & Sattler, 2003; Falconer & Whitby, 1999). For national soil protection legislation or other European
57 policies that influence soil protection, such as the Nitrates Directive or the Cross Compliance
58 Regulation, such analyses do not exist. In the current literature, there is also a lack of understanding
59 of the diversity of TCs, as well as the substantial impact that some of these TCs can have on timely
60 and adequate implementation of policies.

61 Policy characteristics, specific policy content and the way that policies are implemented, are key to
62 policy success (Louwagie et al., 2011). TCs are important for all policy stages as they influence
63 people’s behaviour in operating the policy and responding to it, its efficiency, and ultimately what
64 results and impacts a policy has ‘on-the-ground’. Their neglect during policy design processes could
65 negatively influence the implementation of, for example, AES (Falconer & Saunders, 2002).

66 The contribution of this paper is twofold. First, it addresses the gap in scientific literature on the
67 institutional dimension of ex-ante policy assessments in the specific field of soil conservation, with a
68 particular focus on transactions costs. We make a theoretical contribution by extending existing
69 analytical frameworks to allow the systematic mapping of TCs that arise at the policy implementation
70 and operation stage, including their drivers. Second, the paper explores reasons for the lack of
71 effectiveness of existing policies in reducing agricultural soil degradation. Better understanding the
72 drivers for TCs (what exactly creates these costs) allows for specific recommendations as to how
73 certain TCs can be reduced, to increase the efficiency of policy implementation and operation. As
74 such, the paper makes a contribution to enhance evidence-based policy making that is transferable
75 to other policy fields.

76

77 **2 Research approach**

78 **2.1 Conceptual considerations**

79 Transaction cost economics (TCE) dates back to Coase (1937, 1960) and Williamson (1985, 1999) and
80 has been applied to many topics, including natural resource management. This paper focuses on
81 public TCs that occur in political decision making, policy implementation and monitoring. To some,
82 TCE might imply that actual costs are being measured in quantitative terms and optimised. However,
83 TCs are seldom measured in monetary terms, because they arise during or as a result of policy
84 processes and are either not measurable or they are subsumed in general agency expenditures
85 (Ostrom, 1992). In our understanding of TCs, we follow Vatn (2010, p. 1246) and ask “how costly it is
86 to coordinate actions that are interrelated”?

87 In the context of natural resource governance, TCE has been applied in several studies, in particular
88 with respect to AES (see Beckmann et al., 2009 for an overview). Mettepenningen, Beckmann &
89 Eggers (2011) investigate determinants of public TCs and how costs can be reduced in the case of
90 AES. Existing frameworks for the analysis of TCs especially for environmental policy making (e.g.
91 Garrick, Whitten & Coggan, 2013; Hagedorn, Arzt & Peters, 2002; Ostrom, 1998 and more detailed
92 Krutilla & Krause, 2011; Birner & Wittmer, 2004) already provide a set of criteria for the analysis of
93 aspects relevant to TCs. Most of them are summarized by Paavola (2002, p. 97): “Gaining
94 information, conducting negotiations, making collective decisions, encoding collective choices into
95 institutional arrangements and rules, and enforcing these institutional arrangements are all costly
96 efforts.”

97 Krutilla & Krause (2011) look at TCs for environmental policy making by focusing on the transactions
98 between the regulator, as a representative of society, and those who are regulated. These authors
99 acknowledge that especially in the implementation and enforcement phase several kinds of TCs
100 occur, such as costs of information gathering, administrative costs, and political costs due to
101 stakeholders seeking to influence the design of regulations and guidelines. Paavola & Adger (2005)
102 further differentiate sources of costs of information which are: i) limited cognitive capacity; ii) self-
103 interested agents who do not disclose their preferences; iii) learning processes regarding attributes
104 of environmental resources that take place over a long period of time; iv) adjustments taking time,
105 and requiring learning and resources; and v) information often being scattered or not accessible for
106 different actor groups due to a lack of authority.

107 Birner & Wittmer (2004) present different types of TCs arising from decision making and
108 implementation. Because the step of transposing European policies into national policies and
109 instruments involves decision making - especially for AES - we include these transaction costs in our
110 analysis. For policy design Birner & Wittmer (2004, p. 669) distinguish i) costs of acquiring
111 information, “including scientific and indigenous knowledge on natural resources and information on
112 preferences in case of conflicting goals”, and ii) costs of coordination including the organisation of
113 events and conflict settlement. Key aspects of policy implementation and operation that impact on
114 TCs are incentives for compliance, asymmetrical information, measurability of the outcome, use of
115 social control and problems that arise from non-compliance (Birner & Wittmer, 2004). From these
116 theoretical assumptions we derive that TCs that occur during policy implementation can be
117 categorised into ‘TCs for information management’ and ‘TCs for coordination’.

118 **2.2 Empirical foundations and analysis methods**

119 Based on the theoretical background of TCE, this paper investigates the specific issues of policy
 120 implementation and enforcement in two German federal states – so-called Länder. The focus is on
 121 policies that target soil conservation to secure soil as the basic resource for agricultural activities. For
 122 Germany, these policies include 1) the Federal Soil Protection Act, 2) regulations and directives that
 123 farmers need to comply with under the Cross Compliance Regulation (e.g. the Nitrates Directive and
 124 its transposition), and 3) agri-environmental schemes (AES) targeting soil degradation under the
 125 Rural Development Programmes of individual states (Prager et al., 2011).

126 The Soil Protection Act and the AES were already mature policies during the time of the study. The
 127 Soil Protection Act came into force in 1998 and is a ‘command and control’ measure prescribing good
 128 practice approaches. AES are even older, with the basic idea of the programme stemming from the
 129 MacSharry reform in 1992. Although AES are a European programme they are individually designed
 130 at Länder level, with the design requiring concurrent processes at several levels (ministry, local
 131 offices, land managers) and several stakeholder groups (administrators, NGOs, farmers).

132 The Cross Compliance Regulation came into force in 2003. It introduces requirements such as the
 133 Good Agricultural and Ecological Conditions (GAEC), but it also compiles a range of established
 134 regulations such as the Nitrate Directive. Although the data underlying this paper are from 2003/4
 135 and 2008, they provide a solid basis for analysis, because the policies have not substantially changed
 136 since that time, only the contents of the AES went into a new phase (2008-2013).

137 The two Länder Brandenburg and Saxony-Anhalt are used as case studies (Yin, 2009) in order to (1)
 138 gather empirical evidence regarding the influence of TCs, and (2) complement the aforementioned
 139 frameworks to reflect the empirical evidence. The objective is to present data that show where in the
 140 institutional structure TCs occur, when and why.

141 The empirical data for the analysis were gathered through a total of 43 guided key informant
 142 interviews (based on Gilchrist & Miller, 1999) conducted in Brandenburg (BB) (2008) and Saxony-
 143 Anhalt (ST) (2004/2005) (Figure 1). Although data are from two states and from different time
 144 periods, results are still comparable as the implementation of soil policy and administrative
 145 structures in these two neighbouring states are similar.

146 *Figure 1 to be placed here (Map of case study areas)*

147 All interviews were conducted face-to-face, except two telephone interviews. Interviewees were
 148 chosen with the aim to represent the broad range of actors involved in soil policy implementation
 149 and operation (Table 1). They included government (G) staff of the Ministry of Agriculture and
 150 Environment and of local agricultural offices, and non-governmental (NG) interest group
 151 representatives such as the farmers union, Friends of the Earth and the Organic Farming Association.
 152 The aim of the interviews was to learn about processes in design and implementation of agricultural
 153 soil protection policies including the role of TCs as one key aspect. Respondents were not asked
 154 directly for TCs, but instead asked to give their views on the perceived policy implementation and
 155 operation, which provided the basis for deriving a wide picture of TCs. The interview guides
 156 (available from the authors on request) contained open questions on the design and implementation
 157 process, roles of actors involved, communication patterns and coordination between actors, as well
 158 as involvement in decision-making processes related to policy implementation and
 159 operationalization.

160

<i>Table 1 to be placed here</i>

161 The interviews are labelled according to the states (either BB or ST), the type of interviewee (G or
 162 NG) and a number, e.g. BB-G-03 is the third governmental interviewee in Brandenburg. The empirical
 163 material illustrating the role of TCs in soil protection policy implementation was analysed as follows:
 164 Interview notes were analysed qualitatively based on robust textual analysis from interview
 165 transcripts. Statements were systematically analysed by coding them to a list of TCs that was derived
 166 from the literature, in particular the frameworks introduced in the previous section. Based on the
 167 results of this exercise, the coding frame was revised.

168

169 **3 Transaction costs in policy implementation**

170 This section presents empirical evidence for different TCs and their determinants. The results of the
 171 analysis are presented in the following in three categories. These categories reflect the theoretical
 172 assumptions discussed in Section 2.1. The analysis of our empirical data suggested that the
 173 distinction of TCs for ‘information management’ and ‘coordination’ was not sufficient, hence we
 174 introduced an additional category of TCs called ‘enforcement’. As we will discuss in Section 4, the
 175 distinction of three categories is easier to explain to decision makers and makes it more simple for
 176 them to relate their experiences.

177 We extracted influencing factors (determinants) of TCs from the interviews and sorted them into the
 178 three categories. In the order of presentation, we labelled each factor with a capital letter to ease
 179 later discussion and references. For each factor, we discuss whether its impact increases or reduces
 180 the extent of TCs. Table 2 summarises the conclusions by type of TC and the expected impact
 181 directions. This approach to presentation aims to enhance the clarity of the paper, but we
 182 acknowledge that other categorisation of the empirical information would have been possible (and
 183 would be equally valid).

184 **3.1 Enforcement**

185 It is impossible to have policies which require compliance with standards without enforcement. Our
 186 empirical material provides interesting evidence to support this claim. In particular, the roles of
 187 enforcement and control measures, and of taking evidence versus social control are discussed.

188 Detecting non-compliance and imposing penalties requires resources (Lehmann, Schleyer, Wätzold &
 189 Wüstemann, 2009). In this article, compliance is understood as farmers adhering to the rules
 190 previously outlined, e.g. farmers apply the rules of the GAEC standards and do not remove
 191 hedgerows from their fields. The environmental effectiveness of the measures is not a subject of the
 192 study. Governmental interviewees present themselves as being convinced that the control system is
 193 the basis of successful policies, and ensures enforcement of regulations (BB-G-5) and that “farmers
 194 are much more willing to cooperate when they hear of Cross Compliance checks, because they fear a
 195 reduction of their subsidies” (BB-G-3). If the institutional framework provides incentives for the
 196 target audience of a policy to comply, which includes informal rules such as peer pressure and social
 197 control, then costs for enforcing the policy are lower. However, administrative staff need to dedicate
 198 time for on-site checks, as well as checking records, which translates into TCs. We surmise that an
 199 increased number of checks are associated with increasing TCs (A).

200 While social control and peer-pressure can reduce TCs, the possibility of social control is very limited
201 in particular with regard to AES, because in most cases external actors are not aware of which
202 schemes a farmer has signed up for, if any. They would not know what the prescriptions are with
203 which the farmer should comply. So, if a farmer received payments for conversion to organic farming
204 and hence only carries out a minimum amount of field maintenance, this might not be immediately
205 obvious as he could also participate in a non-tillage scheme (ST-G-12). Therefore, social control and
206 peer-pressure can reduce or increase TCs **(B)**. Some interviewees view penalties as an effective
207 enforcement mechanism, yet in areas such as soil conservation, where both determining the
208 outcome of a conservation action, as well as detecting and producing evidence for offences is
209 difficult, TCs can be highly significant. It is generally difficult to prove what exactly a farmer has done
210 wrong, or what was due to weather conditions (e.g. in erosion events). Some environmental offences
211 are reported by citizens (BB-G-7), but this has also disadvantages, as one interviewee explained:

212 “People file a notification based on their own personal views which then starts up this
213 control and sanctioning machinery that is not always necessary.” (BB-G-1)

214 Hence: to detect non-compliance and to produce evidence for offences tends to increase TCs **(C)**.

215 Furthermore, interviews repeatedly revealed that one crucial determinant of ensuring enforcement
216 is the personnel capacity of administrations and related organisations. The personnel capacity for
217 gathering and analysing necessary baseline data and drawing conclusions on the state of the soil is
218 considered to be decreasing. For example, in the Brandenburg case there is only one member of staff
219 at the district level responsible for the tasks related to soil conservation. When the federal soil
220 protection law was passed, no additional staff or funding was made available for its implementation
221 (BB-G-2). In addition, staff with a soil science background in research institutes are increasingly rare
222 (BB-G-2, BB-G-9) but were previously an important source of data and expertise for agencies.
223 Interviewees complain that budget cuts both in Brandenburg and Saxony-Anhalt have led to merged
224 offices and departments and reductions in staff in administrative units at state, regional and local
225 level. Qualified personnel tend to reduce TCs **(D)**, nevertheless, gathering and analysing data tends to
226 remain an inevitable TCs increasing factor **(E)**. In this context, it was stated that handling and
227 processing applications and contracts, carrying out on-site checks, and other tasks are time-
228 consuming and costly activities for an authority. Therefore, processing applications and contracts **(G)**,
229 as well as on-site checks both tend to increase TCs **(F)**.

230 Some interviewees complained that there is little room for communication between agency staff and
231 farmers. Agency staff claim that advice on AES is not their responsibility, but should be provided by
232 the privately-organised advisory services (ST-G-12). However, the uptake of schemes and innovative
233 conservation measures will increase with the availability and quality of advice to farmers (BB-NG-9).
234 More interest in schemes would lead to more applications and queries. Hence, it might be in the
235 interest of some personnel not to promote AES (ST-G-13). Promoting AES and advising on changes to
236 legislation and regulations also requires training for agricultural advisors, yet in the long run, more
237 efficient soil conservation could be ensured. Providing advice and instructions comes with a cost that
238 might be outweighed by the benefits, thus it could increase or reduce TCs **(H)**.

239 In summary, this section outlined the determinants of administrative costs faced by the
240 implementing authorities. Due to the characteristics and range of soil processes, outcomes are
241 difficult to monitor and non-compliance is difficult to prove, which increases TCs **(C)**. Timely, strict
242 enforcement and better data provision would raise the TCs **(A, C)**, because it requires increasing the
243 personnel capacity and gathering of data **(E, F, G)**. However, a crucial determinant is the personnel

244 capacity in the long term for enforcement by providing evidence and assistance (**D**). We find that TCs
 245 are not only a burden for the actors involved, but need to be understood from the perspective of
 246 administrators and farmers, with regard to investment into capacity (**D**) and community (**B**)
 247 development, e.g. in conjunction with advice (**H**) – an aspect that will recur in the following sections.

248 **3.2 Information management**

249 Information is essential for policy implementation. Information includes biophysical data, technical
 250 information, knowledge of communication channels and administrative procedures, as well as
 251 practical information about farmer support needs. Different actors have different kinds and levels of
 252 detailed information available, at different points in time and in different forms – in other words,
 253 actors have imperfect, i.e. asymmetric or incomplete, information. In practice, asymmetric
 254 information is a concern for all actors. Asymmetries can have different reasons; either that actors
 255 withhold information to be in a more favourable situation than others, or actors do not communicate
 256 the information they have, because the information is not required or regarded as unnecessary in the
 257 specific situation. In both case studies, information asymmetry was found between different actor
 258 groups and within actor groups, i.e. within one authority. Some staff may be less motivated than
 259 others to exchange information.

260 There is evidence that information asymmetry increases TCs (**I**). In order to address the issue of
 261 incomplete information, an investment in (scientific) research would be needed. Our evidence shows
 262 that the perception of the role of science and scientific organisations differs between actor groups.
 263 Regional authority staff access and use scientific knowledge provided by research organisations in
 264 Brandenburg (BB-G-2; BB-G-9), but non-governmental actors often lack access to this information
 265 (ST-NG-02). An (external) contribution of scientific knowledge tends to decrease TC (**J**).

266 The contact with interest groups is important for the ministry because they represent the main
 267 channels for information on AES (ST-G-08). To tap into external knowledge, the ministry uses several
 268 mechanisms such as organising consultations, requesting reports from the local offices, and informal
 269 knowledge transfer. Some interviewees are convinced of the usefulness of consultations while others
 270 think they waste time and resources. NGOs in Saxony-Anhalt asserted that there are “definitely not
 271 too many meetings” (ST-NG-02), whilst ministry staff found there are too many (ST-G-17).

272 The broader the audience at meetings, the more opinions can be gathered and negotiated, but again
 273 with a trade-off in time. Therefore, these meetings involve TCs not only in terms of facilitation and
 274 time for such an intense exchange but, as Crase, O’Keefe & Dollery (2013) show in their case study
 275 on consultation for the reallocation of water in Australia, might even result in higher costs for
 276 taxpayers afterwards. Hence consultations can increase or decrease TCs (**K**). Requesting and
 277 analysing reports seems to be less effective and tends to increase TCs (**L**).

278 Working groups are a platform for exchange of information between a broad range of actors. Some
 279 of the actors are much in favour: “A working group would make a lot of sense. We have been
 280 suggesting it for a while” (ST-NG-06). In particular, for the localised adaptation of soil conservation
 281 measures and discussion of problems, a working group of farmers, local authorities, all levels of
 282 administration, and ministry officials is seen as useful (BB-G-9). But interviewees also acknowledge
 283 the extensive effort needed to organise a working group and convincing people to become involved,
 284 because they are asked to contribute over and above their usual workload (BB-G-8). Others voice the
 285 opinion that a permanent working group is not the best use of people’s time and prefer ad-hoc
 286 organised meetings to exchange information. The willingness to talk, but also the ability to

287 contribute one's own opinion and be heard, is seen as essential. Working groups tend to decrease
288 TCs **(M)**.

289 A stumbling block on the road to efficiency appears in segregated organisation, e.g. although the
290 agriculture and environment sectors are combined in one ministry in Saxony-Anhalt, departments
291 typically dealt with separate sets of stakeholders so that AES were informed by agricultural interest
292 groups only (Prager & Nagel, 2008) (cf. **(S)** below). Moreover, coordination and communication
293 becomes more costly due the practice of rotating administrative staff with the objective of building
294 administrators' skill base and reducing the risk of individuals taking advantage or misusing their
295 network and contacts. Many interviewees raised this issue and commented that this practice brings
296 friction and loss of continuity, and ultimately knowledge, indicating that staff rotation tends to
297 increase TCs **(N)**.

298 To ensure information is distributed, the ministry in Saxony-Anhalt organises training for technical
299 staff at the local agricultural offices in order to familiarise them with new procedures, directives,
300 funding rules and applications. These events are generally useful for information transfer and
301 coordination between the various administrative units but there can be an information overload and
302 saturation: "If there are too many training events, we don't attend anymore" (ST-G-01) – so training
303 can either decrease or increase TCs **(O)**.

304 Agricultural authorities organise information events for farmers, for example after substantial
305 changes to AES are implemented, i.e. such changes have knock-on effects in terms of costs of
306 organising events and distributing information. TCs in the process of information distribution are
307 increased by uncertainty regarding what actions are required. This might be due to the desire of
308 decision makers to decrease costs during a negotiation phase, but this increases TCs in the
309 implementation and enforcement phase. Interviews provided evidence that these problems exist for
310 many regulations. Information events to disseminate information to farmers can increase or
311 decrease TCs **(P)**.

312 The empirical data provide evidence that the way in which information is distributed is crucial with
313 respect to successful communication. A Brandenburg interviewee explained that an option for
314 reducing costs and still reaching a large share of farmers is for an authority representative to speak at
315 an event organised by an interest group, or by using farmer associations to distribute information
316 (BB-G-10). The farmer association would know how best to get the message across to their
317 constituency. Utilizing existing platforms for information transfer can decrease TCs **(Q)**. There is
318 awareness among governmental interviewees that using official communication channels may hinder
319 information transfer: "The process is slowed down and some units, who feel they might be impacted,
320 will, for example, hold back a report. In this case I go to the interest groups, who take it straight up to
321 the ministry" (ST-G-03). Informal communication is important for decision making procedures: "The
322 main lines of our everyday business are decided in an informal setting" (ST-NG-02). The main reason
323 for using informal channels is inefficiencies in the official channels, for example the requirement for
324 duplication of communication by post and email. Many interviewees admitted that they regularly
325 email directly to the relevant person "because we need the response quickly, [and] following the
326 official communication channels takes too long" (ST-G-01). In the face of costly official
327 communication procedures, staff seek to reduce TCs, mainly by shortening the time needed to get
328 access to information or to pass it on. Utilizing informal communication and knowledge transfer can
329 decrease TCs **(R)**.

330 To summarize, information is needed for the implementation of a policy. Different stakeholders hold
 331 different bits of information and gathering all this information from different sources requires
 332 resources. However, the actual level of TCs is dependent on how information is distributed, how
 333 much effort people put into actively transferring the message and the use of the most efficient
 334 channels. It will not be optimal for policy implementation to always minimise TCs as gathering
 335 information and facilitating communication is crucial, and there is also a productive component
 336 inherent in the costs. Therefore, the focus should be on balancing the costs and benefits of gathering
 337 and sharing information efficiently for policy implementation and operation – involving all relevant
 338 actors.

339 **3.3 Coordination**

340 In addition to information-related costs, coordination costs matter. By coordination we understand
 341 processes and actions such as administering data, drafting and checking contracts, on-farm checks,
 342 issuing payments, and managing conflicts.

343 In line with increased information requirements in segregated organisations, costs for coordination
 344 are likely to be higher in policy areas that require coordination between different authorities (e.g.
 345 agriculture and nature conservation) as compared to areas that can be addressed within one
 346 department only. In Saxony-Anhalt, for example, the ministry comprises a division for agriculture and
 347 a division for nature conservation. Both should be involved in designing AES, but cooperation is
 348 found to be limited. TCs tend to be higher in policy areas organised in a segregated manner **(S)**.

349 Interviewees indicate that increasing coordination costs occur as a result of involving relevant actors
 350 **(T)**, but also for managing conflicts that might occur when interests differ **(U)**. Consultation is a
 351 statutory requirement, but the ministry has discretion regarding whom to involve and to what
 352 extent. Similar to working groups, consultations are a tool to coordinate different interests and
 353 gather information but take time and effort to organise (cf. **K** above). The underlying assumption is
 354 that consultations will create support and legitimacy for a policy as well as help distributing
 355 information, e.g. through the representatives to the interest group members. Generating 'buy-in' is
 356 also expected to increase uptake of AES.

357 Environmental groups in Saxony-Anhalt find it increasingly difficult to contribute to consultations, as
 358 AES and soil management are not their core business and they lack the capacity to keep up with the
 359 changing legislation and funding regulations, which in turn limits their ability to provide competent
 360 input when consulted (ST-NG-04). There is also an issue around neglecting consultation outputs in
 361 further decision making (BB-G-5). If input is perceived to be disregarded in decision making, actors
 362 will be less motivated to contribute or become involved. With decreased motivation to participate,
 363 the costs for getting people involved would increase. Therefore, less motivated actors can increase
 364 TCs **(V)**.

365 Birner and Wittmer (2004, p. 669) identified the settlement of conflicts as a TC together with
 366 'resources spent on meetings' and 'costs arising from delayed decisions': "These costs are obviously
 367 influenced by the number of different actors or interest groups involved in a particular governance
 368 structure, and by the prevailing conflicts of interest between them." In Saxony-Anhalt, a ministry
 369 representative claimed that the ministry's task is directing and leading, but instead they have to
 370 become involved in technical problems, such as administering and checking the field blocks that are
 371 the basic unit for calculating AES payments (ST-G-15). There are also latent conflicts between
 372 different levels of administration, and between administration and NGOs/farm advisors. For

373 example, “the ministry sees the local office only as the administrator, not as a partner with technical
 374 expertise” (ST-G-12). TCs can decrease if the own role perceived by actors was a clear and important
 375 one (**W**).

376 Conflicts may also arise when the purpose of a scheme is perceived differently by different
 377 stakeholders. Some might want to see environmental outcomes on the ground, while others are
 378 focused solely on reducing the risk of complaints and law suits, and making a scheme operational and
 379 efficient from an administrative point of view (ST-G-12, ST-NG-01). Others (especially the farming
 380 lobby) said they prefer to maximise payments to farmers with only minor changes to farming
 381 operations. Detecting and managing latent and open conflicts is increasing TCs (**U**), but understood
 382 to be a key task. In particular, the clarification and operationalization of policy instruments can
 383 increase TCs (**X**), but seems to be inevitable for effective policy implementation and operation.

384 This section showed that TCs for coordinating processes are sometimes perceived to be very high.
 385 Not only do different actors have to be integrated in the process, which is often an obstacle for non-
 386 governmental actors, but also discussions are time-consuming. On the other hand, an open
 387 discussion may lead to an enhanced understanding of positions and an outcome that is supported by
 388 all actors, thereby being of crucial importance for successful policy implementation. Due to differing
 389 interests and positions, conflict resolution is a necessary step in this process. Increased TCs at this
 390 stage are assumed to be a good investment in delivering an effective policy later on.

391

392 **4. Discussion and Conclusions**

393 In the previous sections, the results were presented according to three categories of TC:
 394 ‘enforcement’, ‘information sharing’ and ‘coordination’. The first TC category ‘enforcement’ is driven
 395 by several determinants such as the measurability of outcomes that are regarded as a main task for
 396 the administrations under consideration. Soil conservation is regarded as what Birner & Wittmer
 397 (2004, p. 673) call ‘care-intensive transactions’ which are “activities that are difficult to monitor
 398 because they involve carefulness, watchfulness, and diligence and, therefore, leave ample room for
 399 shirking – or even sabotage”. This attribute is crucial for agricultural land use as policies suffer from
 400 the fact that soil degradation is difficult and costly to measure, especially in the case of erosion and
 401 compaction. Soil organic matter by contrast is measurable to some extent; however, in all cases the
 402 cause-effect relation is often hard to detect (cf. Towers et al., 2006).

403 The second category ‘information management’ involves TCs for acquiring information and
 404 distributing information. AES almost certainly require more coordination efforts than other policies
 405 because of the active integration of a range of actors which leads to the third category
 406 ‘coordination’. On the one hand, this ensures that the policy is targeted to the specific local needs,
 407 but on the other hand, coordination becomes a central determinant for successful implementation.
 408 Conflict management is an additional factor that requires coordination and also communication. It
 409 has not been explicitly mentioned in existing frameworks but it has a great influence on
 410 implementation and operation processes.

411

<i>Table 2 to be placed here</i>

412

413 Capturing such a wide range of TCs was facilitated by collecting data through personal interviews not
 414 directly asking for TCs, but instead collecting rich descriptions of how actors experienced policy
 415 implementation and operation, an approach that we argue has provided the basis for deriving a
 416 more complete picture of TCs. Table 2 illustrates the types of TCs and the determinants that can
 417 increase or decrease these costs – according to the perception of our interviewees. For example, a
 418 downward facing arrow should be read as likely reducing TCs, e.g. ‘Informal communication and
 419 knowledge transfer’ is a factor that reduces the cost of ‘Information management’. The following
 420 sections summarise what we conclude from these findings with regard to the diversity of TCs, their
 421 drivers, and displacement of TCs over time and place.

422 **4.1 Diversity in transaction costs and their drivers**

423 The data indicate a diverse set of TC drivers. Individual determinants may increase or decrease TCs
 424 depending on circumstance. The analysis points to a number of trade-offs that precludes the
 425 depiction of unambiguous costs. For example, the more information that can be gathered by
 426 involving more stakeholders in the consultations (**K, P**), the better the basis for the resulting decision
 427 (because diverse actor groups contribute different types of knowledge, Widmark & Sandstrom
 428 (2012)) and the greater the support for the subsequent decision. However, more meetings and
 429 consultations also increase costs. In addition, involving more stakeholders increases the diversity of
 430 views and interests and may increase or decrease costs for conflict management (**T, U**).

431 TCs of agencies for information distribution could be reduced if all relevant advice was provided
 432 through advisory services (**Q**). However, farmers tend not to demand advice that does not
 433 immediately increase profits such as agri-environmental or soil conservation advice, hence this
 434 advice would have to be subsidised to increase its uptake.

435 The more legislation that needs implementation and the higher the requirements of monitoring and
 436 data, the more TCs occur (**A, C, E, F, G, L, X**). If more staff were available or existing personnel had
 437 better capacity to check enforcement, we could assume improved policy implementation (**D, X**);
 438 however, this would come with an increase in staff costs and higher costs for coordination between
 439 staff and organisations (**N, O, S**). Informal networks and trust (e.g. between senior managers and
 440 staff, between agency staff and non-governmental stakeholders) can facilitate the use of informal
 441 (direct) communication channels, which can increase information flow and reduce TCs (**Q, R**).
 442 However, if too much information is passed through informal channels, there will be a risk of a loss of
 443 accountability and information overload. Control intensity (and related TCs (**E, G**)) can be reduced by
 444 providing adequate advice which would invest in awareness-raising for soil functions and methods of
 445 soil protection (**H**).

446 The analysis shows that no classification of TCs is clear cut. Using existing typologies of TCs (e.g.
 447 McCann et al., 2005) would have created similar ambiguities caused by overlap between types. There
 448 can be several (perhaps even unintended) benefits from spending resources on one transaction. For
 449 example, the costs for a consultation event may support information distribution, information
 450 acquisition, enforcement and conflict resolution at the same time. Due to the overlap in benefits of
 451 one event, it may turn out that defining specific transactions is preferable to the chronology of the
 452 policy cycle suggested by McCann et al. (2005).

453 A different context could create large differences in costs, depending on the actors, their interests
 454 and the organisational, political and institutional structures within which they are embedded, as well

455 as the natural environment. Care-intensity regarding soil-related transactions is high. This underlines
456 the influence of asset specificity on TCs in assessing policy implementation.

457 **4.2 Displacement of transaction costs over time and place**

458 Based on the empirical data and established framework, we can identify opportunities to reduce TCs
459 in policy implementation for responsible authorities, but these are likely to create an increase in
460 costs elsewhere (Coggan et al., 2010), for NGOs, advisors, farmers, other departments, and – not
461 least – the environment. Hence, discussions regarding reducing TCs must be about optimising and
462 not minimising TCs in all of the three categories established in our framework (Table 2). Furthermore
463 the asset specificity must be taken into account (Coggan, Buitelaar, Whitten & Bennett 2013), which
464 is especially high when it comes to soil functions. If more information distribution is undertaken by
465 advisory services, it reduces the authorities' costs but requires subsidies for soil-related advice. This
466 means the risk of merely displacing costs makes the identification of true savings for TCs overall
467 difficult, if not impossible.

468 While it is possible to measure TCs empirically if a study is limited to public administration (Mann,
469 2000), or to landowner's environmental management in a well-defined geographical area (Mburu,
470 Birner & Zeller, 2003), TC measurement is much more challenging if a broader perspective is
471 adopted. A related issue is the displacement of TCs in time (Kuperan, Abdullah, Pomeroy, Genio &
472 Salamanca, 2008). If we suggest that cooperation and exchange between relevant actors at the local
473 and regional scale lead to better targeted soil conservation measures in AES and regionally relevant
474 provisions in a state soil protection legislation, this would increase TCs for information distribution
475 and coordination in the short term but would reduce them in the medium to long term, coupled with
476 increased societal benefits due to better protection of soils. Mutual learning processes from
477 cooperation and exchange potentially decrease TCs in the long term, assuming that they reduce
478 information collection costs, as well as later monitoring and enforcement costs (Falconer, Dupraz, &
479 Whitby, 2001).

480 **4.3 Conclusions**

481 In contrast to typical economic approaches, our aim was not to identify 'the most efficient' of all
482 possible governance structures, instead, we emphasise the difficulties and trade-offs that have to be
483 taken into account, if the factors that influence TCs in policy implementation are to be analysed and
484 measured comprehensively. The TCs categories and the list of determinants that we identified for
485 the implementation phase of soil conservation policies (Table 2) are comprehensive for this phase
486 and administrative actors. Similar costs can be expected for other areas across Germany, countries
487 with a similar administrative structure and policies which try to address wicked problems relating to
488 the natural environment.

489 At the same time, we stress the importance of gathering data and information on TCs *in-situ*.
490 Following predefined categories such as the ones suggested by McCann et al. (2005) or Coggan et al.
491 (2010) would have led us to miss the TCs caused by internal conflicts in the administration, barriers
492 to information flow, and costs associated with false alerts regarding breaches of soil conservation
493 policy.

494 Some of the decisions affecting TCs are made earlier in the policy cycle (policy formulation and
495 decision-making phase) such as the sometimes ill-defined requirements for actions farmers have to
496 take. As a consequence such aspects have to be taken up in the policy design process (see e.g.
497 McCann 2013). Our analysis outlines these TCs together with their drivers and suggests that policy

498 makers should take them into account to avoid reducing policy effectiveness at the implementation
499 stage. In addition, we emphasise that it is necessary to clarify whose TCs are being measured, over
500 what time frame and how they might be influenced by pre-existing policies in order to avoid double
501 counting and displacement effects. This finding resonates with McCann et al.'s (2005) emphasis on
502 the importance of considering all policy stages for a complete efficiency analysis. Moreover, in line
503 with the general conclusion that an enhanced uptake of sustainable land management in policy
504 making needs to go hand-in-hand with high-level politicians' awareness for the topic (cf. Seiler,
505 Bartke, Mienert & Schwarze, 2009), we underline that due to their critical role, the consideration of
506 TCs should be mandatory and more explicit in policy formulation processes.

507 Due to the costly nature of comprehensive TCs measurements ("collecting detailed data on
508 transaction costs is difficult and costly" McCann et al. (2005, p. 532)) across the lifetime of a policy
509 and the caveats associated with double counting, boundaries, definitions, implicit/ explicit costs,
510 displacement and time effects, we suggest it is more useful to take account of estimations of TCs.
511 Estimates would potentially provide a more accurate picture of reality – rather than being precise but
512 (often) inaccurate (Mayer, 1993) – and allow sufficient information for improved decisions on
513 preferable policies. Such estimation would need to involve 'insiders' with knowledge about the policy
514 area (administrators, policy makers, NGOs, land managers), be informed by an understanding of why
515 respective TCs occur, and then allow for adjustments which increase effectiveness rather than taking
516 cost reduction (efficiency) as the overruling goal, since even the most efficient policy may completely
517 miss its objective.

518 Finally, our analysis highlights that TCs are not only constraints but long term investments, i.e.
519 investments into information for sound and effective implementation procedures, and that
520 stakeholder participation is a key requirement. To achieve more effective policy implementation, the
521 existence of TCs has to be acknowledged by decision makers, and both policies and governance
522 structures have to be designed with the aim of reducing these TCs.

523 Our results add empirical grounding to future research that aims to identify TCs and their
524 implications. Future studies should explore if there is merit in separating out fixed TCs and variable
525 TCs. The examples from the empirical studies provide insights where the potential for improving
526 implementation procedures lies. This has to be taken into account in the policy design process where
527 the foundations for the implementation process are established.

528

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538 **References**

- 539 Alexandrescu, F., Martinat, S., Klusáček, P. & Bartke, S. (2014). The Path from Passivity towards
540 Entrepreneurship: Public Sector Actors in Brownfield Regeneration Processes in Central and Eastern
541 Europe. *Organization & Environment* 27(2), 181-201.
- 542 Banwart, S. (2011). Save our soils. *Nature*, 474(7350), 151-152.
- 543 Beckmann, V., Eggers, J. & Mettepenningen, E. (2009). Deciding how to decide on agri-environmental
544 schemes: the political economy of subsidiarity, decentralisation and participation in the European
545 Union. *Journal of Environmental Planning and Management* 52(5), 689-716.
- 546 Birner, R. & Wittmer, H. (2004). On the 'efficient boundaries of the state': the contribution of
547 transaction-costs economics to the analysis of decentralization and devolution in natural resource
548 management. *Environment and Planning C: Government and Policy* 22(5), 667-685.
- 549 Birner, R. & Wittmer, H. (2009). Making Environmental Administration More Effective: A Contribution
550 from New Institutional Economics. In V. Beckmann & M. Padmanabhan (Eds.), *Institutions and*
551 *Sustainability*. Berlin & Heidelberg: Springer Verlag, 153-173.
- 552 Boardman, J. & Poesen, J. (Eds.) (2006). *Soil Erosion in Europe*. Wiley: Chichester.
- 553 Coase, R. (1937). The Nature of the Firm. *Economica* 4(16), 386-405.
- 554 Coase, R. (1960). The problem of social cost. *Journal of Law and Economics* 3, 1-44.
- 555 Coggan, A., Whitten, S.M. & Bennett, J. (2010). Influences of transaction costs in environmental
556 policy. *Ecological Economics* 69, 1777-1784.
- 557 Coggan, A., Buitelaar, E., Whitten, S. & Bennett, J. (2013). Factors that influence transaction costs in
558 development offsets: Who bears what and why? *Ecological Economics* 88, 222-231.
- 559 Crase, L. & O'Keefe & Dollery, B. (2013). Talk is cheap, or is it? The cost of consulting about uncertain
560 reallocation of water in the Murray-Darling Basin, Australia. *Ecological Economics* 88, 206-213.
- 561 Dobbs, T. L. (2012). The enduring importance of understanding institutions and values. *International*
562 *Journal of Agricultural Sustainability* 19(3), 201-203.
- 563 European Commission (2006). Thematic Strategy for Soil Protection. Vol. COM (2006) 231.
- 564 Falconer, K. & Whitby, M. (1999). *Transactions and Administrative costs in Countryside Stewardship*
565 *Policies: An Investigation for Eight European Member States*. Centre for Rural Economy Research
566 Report. Retrieved from: <http://www.ncl.ac.uk/cre/publish/pdfs/rr99.1a.pdf>.
- 567 Falconer, K., Dupraz, P. & Whitby, M. (2001). An Investigation of Policy Administrative Costs Using
568 Panel Data for the English Environmentally Sensitive Areas. *Journal of Agricultural Economics* 52, 83-
569 103.
- 570 Falconer, K. & Saunders, C. (2002). Transaction costs for SSSIs and policy design. *Land Use Policy*
571 19(2), 157-166.
- 572 Garrick, D., Whiten, S.M. & Coggan, A. (2013). Understanding the evolution and performance of
573 water markets and allocation policy: A transaction costs analysis framework. *Ecological Economics*
574 88, 195-205.

- 575 Gilchrist, V.J. & Williams, R.L. (1999). Key Informant Interviews. In B.F. Crabtree & W.L. Miller (Eds.),
576 *Doing Qualitative Research*. Thousand Oaks: Sage Publications, 71-88.
- 577 Hagedorn, K., Arzt, K. & Peters, U. (2002). Institutional Arrangements for Environmental Co-
578 operatives: a Conceptual Framework. In K. Hagedorn (Ed.), *Environmental Cooperation and*
579 *Institutional Change: Theories and Policies for European Agriculture*. Cheltenham: Edward Elgar, 3-25.
- 580 Krutilla, K. & Krause, R. (2011). Transaction Costs and Environmental Policy: An Assessment
581 Framework and Literature Review. *International Review of Environmental and Resource Economics*
582 4(3-4), 261-354.
- 583 Kuperan, K., Abdullah, N.M.R., Pomeroy, R.S., Genio, E.L. & Salamanca, A.M. (2008). Measuring
584 transaction costs of fisheries co-management. *Coastal Management* 36(3), 225-240.
- 585 Lehmann, P., Schleyer, Ch., Wätzold & F. Wüstemann, H. (2009). Promoting Multifunctionality of
586 Agriculture: An Economic Analysis of New Approaches in Germany. *Journal of Environmental Policy*
587 *and Planning* 11(4), 315-332.
- 588 Louwagie, G., Gay, S.H., Sammeth, F. & Ratering, T. (2011). The Potential of European Union Policies
589 to Address Soil Degradation in Agriculture. *Land Degradation and Development* 22(1), 5-17.
- 590 Mann, S. (2000). Transaktionskosten landwirtschaftlicher Investitionsförderung: Ein komparativer
591 Ansatz [Transaction costs of agricultural investment promotion: A comparative analysis].
592 *Agrarwirtschaft [German Journal of Agricultural Economics]* 94(7), 259–269.
- 593 Matzdorf, B., Piorr, A. & Sattler, C. (2003). *Halbzeitbewertung des Plans zur Entwicklung des*
594 *ländlichen Raums gemäß VO (EG) Nr. 1257/1999 des Landes Brandenburg [Interim Evaluation of the*
595 *Rural Development Plan according to the Council Regulation (EC) No 1257/1999 of the Land*
596 *Brandenburg]*. Retrieved from: http://z2.zalf.de/oa/Bericht_61.pdf (10/11/2012).
- 597 Mayer, T. (1993). *Truth Versus Precision in Economics (Advances in Economic Methodology)*.
598 Aldershot: Edward Elgar.
- 599 Mburu, J., Birner, R. & Zeller, M. (2003). Relative importance and determinants of landowners'
600 transaction costs in collaborative wildlife management in Kenya: an empirical analysis. *Ecological*
601 *Economics* 45, 59-73.
- 602 McCann, L., Colby, B., Easter, K.W., Kasterine, A. & Kuperan, K.V. (2005). Transaction cost
603 measurement for evaluating environmental policies. *Ecological Economics* 52(4): 527-542.
- 604 McCann, L. (2013). Transaction costs and environmental policy design. *Ecological Economics* 88, 253-
605 262.
- 606 Mettepenningen, E., Beckmann, V. & Eggers, J. (2011). Public transaction costs of agri-environmental
607 schemes and their determinants – Analysing stakeholders' involvement and perceptions. *Ecological*
608 *Economics* 70, 641-650.
- 609 Ostrom, E. (1992). *Crafting Institutions for Self-Governing Irrigation Systems*. San Francisco: ICS Press.
- 610 Ostrom, E. (1998). The Institutional Analysis and Development Approach. In E. T. Loehman and D. M.
611 Kilgour (Eds), *Designing Institutions for Environmental and Resource Management*. Cheltenham:
612 Edward Elgar, 68-90.

- 613 Paavola, J. (2002). Rethinking the choice and performance of environmental policies. In: D. W.
 614 Bromley & J. Paavola, J. (Eds.), *Economics, Ethics, and Environmental Policy: Contested Choices*.
 615 Malden: Blackwell, 87-102.
- 616 Paavola, J. & Adger, W.N. (2005). Institutional ecological economics. *Ecological Economics* 53, 353-
 617 368.
- 618 Prager, K. & Nagel, U.J. (2005). *Kommunikationsbeziehungen der Ämter für Landwirtschaft und*
 619 *Flurneuordnung im Bereich der Agrarumweltprogramme - Fallstudie Sachsen-Anhalt [Communication*
 620 *relationships for agri-environmental schemes between the Department of Agriculture and*
 621 *reorganisation of land – a case study from Saxony-Anhalt]*. SUTRA-Working Paper No 7.
- 622 Prager, K., Hagemann, N., Schuler, J. & Heyn, N. (2011). Incentives and enforcement: the institutional
 623 design and policy mix for soil conservation in Brandenburg (Germany). *Land Degradation and*
 624 *Development* 22(1), 111-123.
- 625 Rørstad, P.K., Vatn, A. & Kvakkestad, V. (2007). Why do transaction costs of agricultural policies vary?
 626 *Agricultural Economics* 36(1), 1-11.
- 627 Schleyer, C. & Theesfeld, I. (2011). Agricultural and Environmental Policies from an Institutional
 628 Economics Perspective: a Method for Ex-ante Policy Assessment. *Agrarwirtschaft [German Journal of*
 629 *Agricultural Economics]* 60(3), 189-199.
- 630 Seiler, P., Bartke, S., Mienert, M. & Schwarze, R. (2009). Forschung für die Nachhaltigkeit Research
 631 for Sustainability – Nachhaltigkeitsbewertungsverfahren in der Flächenpolitik stärken, *GAIA –*
 632 *Ecological Perspectives for Science and Society* 18(4), 334–336.
- 633 Towers, W., Grieve, I.C., Hudson, G., Campbell, C.D., Lilly, A., Davidson, D.A., Bacon, J.R., Langan, S.J.
 634 & Hopkins, D.A., (2006). Scotland's soil resource – current state and threats. *Environmental Research*
 635 *Report 2006/01*. The Macaulay Institute & University of Stirling.
- 636 Vatn, A. (2010). An institutional analysis of payments for environmental services. *Ecological*
 637 *Economics* 69, 1245–1252.
- 638 Widmark, C. & Sandstrom, C. (2012). Transaction Costs of Institutional Change in Multiple-Use
 639 Commons: The Case of Consultations Between Forestry and Reindeer Husbandry in Northern
 640 Sweden. *Journal of Environmental Policy & Planning* 14(4), 428-449,
- 641 Williamson, O.E. (1985). *The Economic Institutions of Capitalism*. New York: Macmillan.
- 642 Williamson, O.E. (1999). Public and private bureaucracies: a transaction cost economics perspectives.
 643 *Journal of Law, Economics, and Organization* 15(1), 306-342.
- 644 Yin, R. K. (2009). *Case Study Research. Design and Methods*. Los Angeles: Sage.
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646 **Figure 1:** Map of case study areas



647
648

649 **Table 1:** Interviews conducted in Brandenburg and Saxony-Anhalt

Land	Governmental stakeholders (G)	Non-governmental stakeholders (NG)	Total
Brandenburg (BB)	10	10	20
Saxony-Anhalt (ST)	17	6	23
Total	27	16	43

650 Source: Own compilation

651

652 **Table 2:** (Transaction) Cost categories and their influencing factors in agricultural soil protection
653 policy implementation and operation in Brandenburg and Saxony-Anhalt

(Transaction) Cost type	Influencing factors	Impact*
Enforcement	<ul style="list-style-type: none"> • Personal capacity (D) • Social control & peer pressure(B) • Providing advice and instructions(H) • Checking records (A) • Detecting and producing evidence of offences(C) • Gathering and analysing baseline data(E) • Processing applications and contracts(F) • On-site / Field checks (G) 	<p>↓</p> <p>↕</p> <p>↕</p> <p>↗</p> <p>↗</p> <p>↗</p> <p>↗</p> <p>↗</p>
Information management	<ul style="list-style-type: none"> • Contribution of scientific knowledge(J) • Ad hoc working groups (M) • Information transfer trough existing platforms (Q) • Informal communication and knowledge transfer (R) • Consultations (K) • Training for technical staff (O) • Information events for farmers (P) • Asymmetric information (I) • Requesting reports (L) • Staff rotation (N) 	<p>↓</p> <p>↓</p> <p>↓</p> <p>↓</p> <p>↕</p> <p>↕</p> <p>↕</p> <p>↕</p> <p>↗</p> <p>↗</p> <p>↗</p>
Coordination	<ul style="list-style-type: none"> • Actor perceives own role to be clear and important (W) • Higher segregation of policy organisation (S) • High number of actors involved (T) • Detecting and managing latent and open conflicts (U) • Less motivated actors (V) • Clarification and operationalization of policy instruments (X) 	<p>↓</p> <p>↗</p> <p>↗</p> <p>↗</p> <p>↗</p> <p>↗</p>

654 Source: Own compilation

655 *Legend for impact: ↓ = factor tends to reduce (transaction) costs, ↕ = reducing or increasing
656 (transaction) cost possible, ↗ = factor tends to increase (transaction) costs.

657 Source: Own compilation