

# Genetic richness affects trait variation but not community productivity in a tree diversity experiment

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### Summary

• Biodiversity–ecosystem functioning experiments found that productivity generally increases with species richness, but less is known about effects of within-species genetic richness and potential interactions between the two. While functional differences between species can explain species richness effects, empirical evidence regarding functional differences between genotypes within species and potential consequences for productivity is largely lacking.

• We therefore measured within- and among-species variation in functional traits and growth and determined stand-level tree biomass in a large forest experiment factorially manipulating species and genetic richness in subtropical China.

• Within-species variation across genetic seed families, in addition to variation across species, explained a substantial amount of trait variation. Furthermore, trait responses to species and genetic richness varied significantly within and between species. Multivariate trait variation was larger among individuals from species mixtures than those from species monocultures, but similar among individuals from genetically diverse vs genetically uniform monocultures. Correspondingly, species but not genetic richness had a positive effect on stand-level tree biomass.

• We argue that identifying functional diversity within and among species in forest communities is necessary to separate effects of species and genetic diversity on tree growth and community productivity.

### Introduction

Biodiversity–ecosystem functioning (BEF) research has repeatedly demonstrated that increases in plant species richness can lead to increases in plant productivity (Tilman *et al.*, 1996; Hector *et al.*, 1999; Liang *et al.*, 2016; Huang *et al.*, 2018) and other ecosystem functions (Balvanera *et al.*, 2006; Isbell *et al.*, 2011). Plant species can be different in their use of resources, their resistance to natural enemies or their responses to abiotic conditions. Differences in these ecological functions can lead to differences in productivity and fitness over a range of environmental conditions. Based on niche theory, increasing species richness is hypothesized to increase plant productivity as a result of the increased diversity of ecological functions within communities (McGill *et al.*, 2006; Turnbull *et al.*, 2016).

Identifying the ecologically relevant functions of species is challenging, but the use of physiological and morphological traits to measure, describe and predict functions of individuals, species and ecosystems has become common practice in ecology (Wright *et al.*, 2004; Bruelheide *et al.*, 2018; Kattge *et al.*, 2020). In addition, it is commonly assumed that differences in ecological functions would be related to and identified by differences in trait values among species (McGill *et al.*, 2006). However, general relationships between traits and their functions are still elusive as a result of, among other things, the dependency on abiotic and biotic environmental conditions and the use of species mean values instead of individual values (Shipley *et al.*, 2016; Yang *et al.*, 2018). The assumption that higher functional trait diversity at the community level has a positive effect on community productivity has been tested and sometimes confirmed by relationships between community-level trait diversity and community productivity (Kröber *et al.*, 2015; Tobner *et al.*, 2016; Wagg *et al.*, 2017; Roscher *et al.*, 2018b).

Assuming that community trait diversity only depends on species mean trait values ignores trait variation within and trait overlap between species and therefore overestimates species

differences in niche and ecological functionality (Violle et al., 2012). The degree of intraspecific trait variation is determined by genetic differences and plastic responses to local environmental conditions. Although intraspecific trait variation is often smaller than interspecific trait variation (Albert et al., 2010; Hulshof & Swenson, 2010; Poorter et al., 2018), intraspecific trait variation can still affect community functioning (Hughes et al., 2008; Bolnick et al., 2011; Siefert et al., 2015; Des Roches et al., 2018). In addition species that coexist in mixtures are also expected to further minimize overlap to avoid strong interspecific competition (Kraft et al., 2008; Zuppinger-Dingley et al., 2014; Turnbull et al., 2016). This so-called niche differentiation by species in diverse communities could result from plastic or genetic shifts in species functional trait values (Zuppinger-Dingley et al., 2014; Lipowsky et al., 2015; Roscher et al., 2015) by which total variation among individuals from mixtures increases compared with individuals from monocultures.

While most designed BEF experiments focus on effects of species diversity (reviewed in Tilman et al., 2014; Verheyen et al., 2016; Weisser et al., 2017), it is conceivable that within-species genetic or even epigenetic diversity also affects plant productivity and other ecosystem functions (Schmid, 1994; Crutsinger et al., 2006; Kotowska et al., 2010; Latzel et al., 2013; Prieto et al., 2015; Hahn et al., 2017; Zeng et al., 2017). While positive effects of within-species diversity on plant productivity have been observed (Booth & Grime, 2003; Hughes & Stachowicz, 2004; Crutsinger et al., 2006; Kotowska et al., 2010; Cook-Patton et al., 2011; Drummond & Vellend, 2012), to date explicit empirical tests of whether this is a result of functional differences between genotypes have, to our knowledge, only been done for the model species Arabidopsis thaliana (Wuest & Niklaus, 2018) and rice (Montazeaud et al., 2018). Especially in tree communities, the degree of trait variation created by within-species genetic diversity has not yet been determined. In addition, there is a principal lack of information about the effects of genetic diversity on tree productivity (Fischer et al., 2017; Hahn et al., 2017; Zeng et al., 2017). Most studies of tree communities looked at effects of genetic diversity on the interaction between plants and arthropod communities (e.g. diversity and abundance of insects) (Barton et al., 2015; Campos-Navarrete et al., 2015; Cao et al., 2018; Barantal et al., 2019). Overall, identifying functional trait variation among groups of tree individuals of known genetic relationship (here seed families; individuals from the same seed family share the same mother) within and between species in forest communities would be the first step towards identifying ecologically relevant biological processes that underlie effects of species and genetic diversity on tree growth and community productivity.

In this study, we evaluated functional trait variation among species and among different seed families within species across two degrees of species richness (one vs four species) factorially crossed with two degrees of genetic richness (one vs four seed families per species) in experimental tree communities 8 yr after establishment. First, we studied how much of the observed trait variation was explained by species and seed-family identity, species and genetic richness and the interactions between them. Second, we determined to what extent species and seed families were distinct in their trait values. Third, we determined if species or genetic richness affected the multivariate trait variation among the measured individuals. Finally, we determined if species and genetic richness affected individual tree growth and stand-level biomass. We hypothesized that in addition to species identity, within-species seed-family identity would explain a part of the overall trait variation. We expected that increasing species and genetic richness would both have a positive effect on tree growth and stand biomass, although a smaller effect was expected for genetic than for species richness, owing to smaller ecological differences expected between genotypes within species.

## Materials and Methods

### Experimental site

A species × genetic tree diversity experiment was set up as part of the so-called BEF-China experiment (Bruelheide et al., 2014) located near Xingangshan, Dexing City, Jiangxi Province, China. The region is typical of the Chinese subtropics with a mean annual temperature of 16.7°C and a mean annual precipitation of 1821 mm. In 2010, 24 plots of the BEF-China experiment were used for the genetic diversity experiment (Fig. 1; Hahn et al., 2017). In short, each square plot with an area of 0.067 ha (1 mu) was planted with 400 individuals belonging to one or all of the four tree species Alniphyllum fortunei, Cinnamomum camphora, Daphniphyllum oldhamii and Idesia polycarpa. These species were chosen based on the fact that two are deciduous (A. fortunei and I. polycarpa) and two are evergreen (C. camphora and D. oldhamii). In addition, the taxonomic differences among species, the quantity of seedlings from different seed families (see later for more information about seed families) and the species' expected survival contributed to the selection of these four species. The species richness was one or four per plot. Within each plot the planted individuals belonged to one or four different seed families per species, resulting in a full factorial design of species × genetic richness (Fig. 1). Here, 'genetic richness' refers to within-species number of seed families and was created by collecting seeds and producing seedlings from different mother trees. Seed families therefore represented something between fulland half-sib families, depending on how many father trees pollinated the seeds of a single mother tree. Assuming that there are genetic differences between mother trees and thus seed families, we therefore use the term 'genetic richness' as shorthad for 'seedfamily richness'. The seed families were collected in the Gutianshan National Nature Reserve, Zhejiang Province, China, which is close to the BEF-China experiment (40 km away).

Because of the availability of only 24 plots (1 mu each) for this experiment, subplots of 0.25 mu were used as our main experimental unit. All treatments with a single species, that is species monocultures, were established per subplot, and every seed family occurred once as seed-family monoculture in a subplot and one or two times in species monoculture with four seed families per subplot (Supporting Information Table S1; Fig. 1). Species mixtures were planted on whole 1 mu plots, providing four subplots of identical species × seed family composition (Table S1).



Fig. 1 Overview of the designed genetic tree experiment within BEF-China. (a) In total there were 24 plots. Each square plot with an area of 1 mu (0.067 ha) was planted with 400 individuals belonging to one or all of the four tree species Alniphyllum fortunei, Cinnamomum camphora, Daphniphyllum oldhamii and Idesia polycarpa. The species richness per plot was one or four and genetic richness per species was also one or four, resulting in a full factorial design of species × genetic richness (represented by 1.1, 1.4, 4.1 and 4.4 levels; see legend). (b) Each 1 mu plot was subdivided into four subplots, which were use to create different seed-family monocultures in the 1.1 richness and different seed-family compositions of four seed families in the 1.4 richness level. In total there were two 1 mu plots per species and a half 1 mu plot per species for the 1.1 and 1.4 richness levels respectively (see Supporting Information Table S1 for species-seed family compositions). Within four-species mixtures the four subplots per 1 mu plot had identical seed families per species compositions. In total there were eight and six 1 mu plots for the 4.1 and 4.4 richness levels, respectively (Table S1). Different species are represented by tree shapes, while different seed families are represented by tree coloring. The asterisk (\*) indicates the plot that was not established as a result of logistic constraints.

Using subplots for species monocultures allowed us to reduce the total number of plots and seedlings to be planted, and using 1 mu plots for species mixtures allowed us to have enough replicates of species and seed families within each plot. Seed-family compositions were not replicated in multiple plots in order to have a larger number of different seed-family compositions. Different compositions at each richness level were considered the relevant degree of replication. Individual trees came from nine seed families in A. fortunei and C. camphora, eight seed families in I. polycarpa and 15 seed families in D. oldhamii (Table S1).

### Functional traits

Eight functional traits were measured in 2017 and 2018 for 547 individuals across all seed families and species in the four different species × genetic richness combinations (see Table S2 for number of individuals per seed family per species per richness level). For each individual we measured wood density (WD), leaf area (LA), specific leaf area (SLA), Chl content (CHL), leaf nitrogen content (LN), leaf carbon content (LC), stomatal density

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(SD) and stomatal aperture (SA). These traits were chosen because of their association with growth strategies and functions within the community (Wright et al., 2004; Diaz et al., 2016; Bruelheide et al., 2018). In this study we did not measure ecological functions directly, but rather associated variation in trait values with variation in ecological functions.

Wood density  $(g cm^{-3})$  was estimated on one branch segment per tree by dividing the oven-dried weight (80°C, 48 h) through the water-displaced fresh volume. We used branch-WD to represent stem-WD because the two are generally highly correlated (Swenson & Enquist, 2008). Four to five leaves were sampled per tree, scanned for area (LA, in cm<sup>2</sup>) and dried for 48 h at 70°C for DW, after which SLA  $(cm^2 g^{-1})$  was calculated. Chl content per area was estimated based on three measurements per leaf with a SPAD meter (SPAD-502; Minolta Camera, Tokyo, Japan). LN and LC (%) were determined using the CHNOS Elemental Analyzer Vario EL III (Elementar Analysensysteme GmbH, Hanau, Germany). For each individual tree three stomatal imprints were made and analyzed under a microscope (Nikon 80i) to calculate SD (count  $mm^{-2}$ ) and for a total of five stomata the inner stomatal

length and width were measured to calculate SA  $(\mu m^2)$  based on an elliptical shape. All leaves used for trait measurements were fresh and healthy without any dirt, physical damage or fungal infections.

### Multivariate hypervolumes

To estimate the total multivariate trait variation among individuals per richness level, multivariate trait hypervolumes were calculated using multidimensional kernel density estimation within the 'hypervolume\_gaussian' function of the HYPERVOLUME package in R (Blonder et al., 2014, 2018), assuming default settings (among others, Silverman bandwidth estimation and 95% probability threshold). The size of a hypervolume depends on the number of observations and dimensions, which in our case are the individuals and traits, respectively. In addition, an accurate approximation of the hypervolume size is sensitive to correlations among dimensions and to the ratio between observations and dimensions (Blonder et al., 2018 recommended to use no more different dimensions than the number of log(observations)). Therefore we used only three trait dimensions and constructed various hypervolumes with an equal number of observations to properly compare hypervolumes across species and richness levels. To avoid biases based on the selection of traits, we used the first three uncorrelated axes of a principal component analysis created by the eight traits (together explaining 71% of variation; Fig. S1). The values of these three axes were assumed to provide a more holistic representation of ecological function than any set of three particular traits. Nevertheless, we also performed additional hypervolume simulations to explore the effect of species and genetic richness on multivariate trait variation using particular traits (Figs S2, S3).

To assess potential effects of variability for comparing sizes of hypervolumes across richness levels, we simulated 500 hypervolumes per richness level using bootstrapping (similar to that done by Bittebiere et al., 2019). We randomly sampled individuals, with replacement per species and richness level depending on the particular comparison. First, to test if the hypervolume of all species together or species-specific hypervolumes were affected by species richness, we sampled 29 observations per species from species monocultures versus species mixtures (ignoring the genetic richness levels). The 29 observations per species were pooled together per species monoculture versus species mixtures and these 116 observations were used to calculate the hypervolume size of the four species together. For the species separately, each hypervolume was based on the 29 sampled observations. We used 29 observations because that was the lowest number of observations in species monocultures (see Table S2). Second, to test if the hypervolume was affected by genetic richness (i.e. number of seed families per species), nine observations per species from each species × genetic richness level were sampled. To test the effect of genetic richness in species monocultures, hypervolumes for each species separately were calculated using these nine observations. To test the effect of genetic richness on top of species richness, the nine samples per species were pooled to calculate hypervolumes for all species together. Nine values were sampled because that was the lowest number of observations (for *I. polycarpa* monoculture with one seed family; see Table S2).

### Individual tree growth and stand-level tree biomass

Individual tree biomass (kg) was calculated for all trees within the experiment, excluding the outer two rows of trees per 1 mu plot to avoid border effects. Individual tree biomass was calculated as  $H \times BA \times CV$ , where H is the height (in m), BA is the basal area at the ground (in m<sup>2</sup>) and CV is a correction factor for stem shape and WD, estimated by Huang et al. (2018). We tested if this method would be considerably different from using individual measured WD in the estimation of tree biomass. For the 547 individuals for which we measured WD directly, we calculated biomass by multiplying the volume estimate with WD. Because the correlation between the two methods was high  $(R^2 = 0.97)$ , we decided to use the Huang et al. (2018) method throughout for consistency. Individual tree growth was calculated as the relative biomass growth rate between 2013 and 2018, that is, the difference between the log-transformed biomasses from the two dates. Stand-level tree biomass (Mg ha<sup>-1</sup>) was calculated per subplot (quarter of the 1 mu plot) as the sum of all alive trees.

### Statistical analysis

To determine how trait and growth variation among individual trees was affected by explanatory factors, ANOVAs were fitted with sequential order of the terms species identity, seed-family identity, species × genetic richness combination (or its contrasts, i.e. the factorial species richness + genetic richness + interaction), interactions of these and, finally, the 1 mu plot. Similar ANOVAs with sequentially fitted factors were performed for each species separately to estimate the relative contribution of each explanatory factor on trait variation for each species. Species effects were tested against seed families and richness effects against 1 mu plot variation, while all other effects were tested against the residual variation. All ANOVAs were performed with the 'aov' function in R. Note that this procedure corresponds to mixed-model analysis of variance but allows greater flexibility in assessing contributions of random effects than other implementations of mixed-model analyses, such as those based on residual maximum likelihood methods (for a detailed explanation of this model choice, see Schmid et al., 2017).

To determine which pairs of species or seed families within species had significantly different trait values, *post hoc* analyses after one-way ANOVAs were performed, taking all individuals together or per species independently ('TukeyHSD' function in R). All trait values, except for Chl values, were log-transformed before analysis to increase data normality. In addition, we checked the normality and homoscedasticity of the residuals of the performed ANOVAs.

The effect of species and genetic richness, their interaction and species identity on individual tree growth was calculated with an ANOVA with sequential order of species richness, genetic richness, the interaction of species  $\times$  genetic richness, species identity and the interaction of species identity with species  $\times$  genetic richness. Stand-level biomass was tested with an ANOVA with sequential order of the terms species richness, genetic richness, the interaction of species  $\times$  genetic richness, monoculture species

and the interaction between monoculture species and genetic richness. For both analyses, the 1 mu plot functioned as error term.

### **Results**

#### Partitioning functional trait variation

Analyzing the total variation in the eight functional traits measured on 547 individuals we observed that a large part was a result of variation among the four tree species (22.61-59.09%; Fig. 2; Table S2) and a further substantial part (2.69-11.45%) was a result of variation among seed families within tree species, supporting our assumption that seed families differed genetically. The amount of trait variation explained by species and seed families varied from trait to trait. For example, compared with the other traits, LN varied relatively less among species but more among seed families within species (Fig. 2).

The effects of species and genetic richness combined explained only a small part of total observed trait variation (0.16-3.85%), with species richness generally explaining more variation than genetic richness (Table S2). These small main effects were related to different responses of species and seed families to the richness treatment. The corresponding interaction terms

species × richness and seed family × richness explained 1.33-6.58% and 4.87%-11.23% of variation, respectively (Fig. 2; Table S3). In the separate analysis for each of the four species, the relative contribution of seed families to the total within-species variation in the measured traits was quite large, but richness effects and interactions among the two, that is variation in the responses of individual seed families to richness, were of similar importance (Fig. 2b-e). On average, differences among seed families explained 7.2%, 9.3%, 11.7% and 11.9% of trait variation in A. fortunei, C. camphora, D. oldhamii and I. polycarpa, respectively. Among the structural traits (WD, LA and SLA), seed families explained most variation in LA for three species (the exception was D. oldhamii where seed families explained most variation in SLA). Among the chemical traits (CHL, LN and LC), seed families explained most variation in LN, namely 20% in I. polycarpa and D. oldhamii. In addition, for all four species, seed families explained more variation in SA than in SD.

The relative percentage of variation explained by richness was highly variable and ranged from 30% of the variation in SLA for I. polycarpa to 0.9% of variation in SA for D. oldhamii (Fig. 2be). In addition, the different responses of seed families to richness explained 4.6-23.7% of trait variation in the different species. This is illustrated in the reaction-norm plots shown in Fig. S4. Although no seed family showed a consistent up or down change



#### Variance explained (%)

Fig. 2 Contributions of the different explanatory terms to the total variation in eight measured traits. (a-e) Shown are the percentages of explained variation (increments of total sum-of-squares as terms are sequentially fitted to the ANOVA model from left to right) per trait for all species together (a) and for the four species analyzed separately (b-e). Note that for the latter the terms 'species' (dark green) and 'species × richness' (dark turquoise) do not exist. See legend for color key of terms 'species' (species identity), 'Seed family' (seed family identity), 'richness' (species × genetic richness combination, four levels), 'species x richness' (species specific responses to richness), 'seed families x richness' (seed-family specific responses to richness), 'plot' (identity of 1 mu plot) and 'residuals' (remaining variation). Exact values for percentage explained variance for all species together (a), and related statistics, are shown in Supporting Information Table S3.

in trait value with richness, their reaction norms illustrate that seed families differed in their trait-specific plastic responses to richness.

# Trait differences between species and seed families within species

The extent to which species differed in their traits varied among traits; for example, *I. polycarpa* had larger LA and *D. oldhamii* had higher CHL compared with the other species (Fig. 3a; Table S4). For other traits, some of the species had similar values, such as *I. polycarpa* and *A. fortunei* for SLA and *C. camphora* and *D. oldhamii* for LN. Considering all eight traits together demonstrated that the four species differed in multivariate trait space (Fig. 3a). This could also be seen in the ordinations obtained with the first three axes of the principal component analysis (together explaining 71%; Fig. S1).

Comparing the eight traits among seed families within species resulted in various significant pairwise seed-family comparisons (Table S4). Using the same polygon plots as for the comparison among species (Fig. 3a), it can be seen that some seed families are distinct from other seed families (Fig. 3b–e). Compared with other seed families, seed family AF.103 (see Table S2 for a list of all seed families) had distinct values for LA, LN and SA in *A. fortunei* and CC.110 had distinct values for LN and SA in *C. camphora.* In *I. polycarpa* IP.10 had distinct values for LA and

LN. The differences between seed families were most pronounced for chemical traits in *D. oldhamii*, with distinct values in CHL for DO.19, in LN for DO.36 and in LC for DO.27 and other seed families.

# Multivariate trait hypervolumes created by individuals from different richness levels

Despite the relatively small main effect of richness on trait variation across the four species, richness did affect trait values differently between species and seed families, indicated by the interaction effects of species or seed families with richness (Table S3). The multivariate hypervolume created by individuals from species mixtures was larger than that created by individuals from monocultures when the four species were combined (0.0065 vs 0.0043 SD<sup>3</sup>; P<0.001; Fig. 4a). Hypervolume sizes were also higher for the separate species, especially for A. fortunei and I. polycarpa (P<0.001 for all species; Fig. 4b; see Fig. S5 for an example of hypervolumes per species in species monocultures versus species mixtures). This analysis disregarded the difference in genetic richness. The increased hypervolume size represents higher differences among the measured individuals when individuals interact with heterospecific individuals in species mixtures compared with conspecific individuals in species monocultures. The volume centroids of the species shifted slightly between the species-richness levels one and four, which suggests that on



**Fig. 3** Illustration of relative mean trait values of species and seed families within species. Species are identified with different colors and seed families by various shades of colors and identification numbers (see legends and Supporting Information Table S2 for the full list of seed families). (a–e) The polygons connect scaled values (mean = 0 and SD = 1 after log-transformation if required) for each trait for the four species (a) and for the seed families within the species *Alniphyllum fortunei* (b), *Cinnamonum camphora* (c), *Daphniphyllum oldhamii* (d) and *Idesia polycarpa* (e). For trait means of seed families (b– e), trait values were scaled for each species separately. See Table S4 for significant differences between species and seed-family pairs per trait.



**Fig. 4** Multivariate trait hypervolume sizes of all species combined and each species separately, calculated using individuals from low or high (one vs four) species richness. (a) Hypervolume sizes of individuals from all four species from species monocultures vs species mixtures; (b) hypervolume sizes of individuals per species separate from species monocultures vs mixtures; (c) visualization of hypervolumes created by individuals from all species together from species monocultures vs mixtures (the two hypervolumes have a size of 0.0037 and 0.0071 SD<sup>3</sup> for species monoculture and mixture, respectively). For each value on the x-axis 500 bootstrapped volumes were obtained, calculated with 29 randomly sampled individuals, with replacement, per species per species richness (b) or the summed 116 sampled individuals per species richness (a). See Supporting Information Table S2 for the number of individuals available per sampling (per species per species richness). AF, *Alniphyllum fortunei*; CC, *Cinnamonum camphora*; DO, *Daphniphyllum oldhamii*; IP, *Idesia polycarpa*. Individual open dots represent individual hypervolume sizes; black solid items are mean values  $\pm$  1 SD. See the Materials and Methods section for detailed explanation of calculations regarding the multivariate hypervolumes.

average the species became more distinct in the species mixtures compared with monocultures (Fig. S6).

Increasing the genetic richness had a different effect on the multivariate hypervolume size (Fig. 5). In species monocultures, the hypervolumes created by individuals from plots with a genetic richness of four seed families were smaller or larger than the hypervolumes created by individuals from plots with a genetic richness of one seed family. Volumes of A. fortunei and I. polycarpa were smaller; for genetic richness four vs one the hypervolume sizes were 0.0010 vs 0.0015 and 0.0007 vs 0.0014 SD<sup>3</sup> for A. fortunei and I. polycarpa, respectively, while volumes of C. camphora and D. oldhamii were slightly larger  $(0.00051 \text{ vs } 0.00045 \text{ and } 0.00068 \text{ vs } 0.00061 \text{ SD}^3 \text{ for}$ C. camphora and D. oldhamii, respectively) (P<0.001 for all species; Fig. 5a). In species mixtures, the total hypervolume size created by individuals from all four species with a genetic richness of four seed families was similar to the hypervolume size created by individuals from species mixtures with a genetic richness of one seed family (both 0.0074  $SD^3$ ; P=0.54; Fig. 5b). Moreover, the results from the multiple specific trait combinations showed that the effect of richness on the hypervolume size was similar across all trait combinations, while the effect of genetic richness in species mixtures was slightly different (Figs S2, S3).

#### Individual tree growth and stand-level biomass

Individual growth of the 547 individuals was, just as the traits, primarily determined by species, followed by seed families and their interaction with richness (Table S3), demonstrating that

growth was also highly variable among species and seed families and that both showed variation in their responses to species  $\times$  genetic richness combinations (Fig. S4). However, tree growth was not significantly affected in the same direction by richness as shown by the small effects of the species and genetic richness combination (Table S3). In addition, across all individuals within the experiment (total 2073 individuals) individual growth was not significantly affected by species or genetic richness (Fig. 6a,b; Table S5). Nevertheless, growth was different among species and both the species and genetic richness effects were species-specific, as indicated by the interaction terms (Fig. 6a,b; Table S5). At the community level, standing tree biomass after 8 yr of growth was significantly higher in the four-species mixture than in the average monoculture, while stand-level tree biomass was not affected by genetic richness (Fig. 6c; Table S5). Additionally, there were significant differences in stand volumes between the four species in monoculture, but no species-specific genetic richness effect (Fig. 6d).

#### Discussion

We analyzed trait variation among species and within species among seed families in a full-factorial species  $\times$  genetic richness tree biodiversity experiment and assessed if trait variation played a role in the effect of species and genetic richness on individual tree growth and stand-level tree biomass after 8 yr of community development. We observed that seed families explained a substantial amount of trait variation, on top of the amount explained by differences between species. In addition, trait values changed plastically in response to richness, with variation in these



**Fig. 5** Multivariate trait hypervolume sizes of each species separately or all species combined in species monocultures or mixtures, calculated using individuals from low or high genetic richness. (a) Hypervolume size per species separately in low and high genetic richness (one vs four seed families per species) in species monocultures. (b) Hypervolume size of all species together in low and high genetic richness (one vs four seed families per species) in species mixtures. For each value on the x-axis, 500 bootstrapped volumes were obtained, calculated with nine randomly sampled individuals, with replacement, per species per genetic richness in species monoculture (a) or 36 sampled individuals per genetic richness in species mixtures (b). The 36 individuals consisted of nine individuals per species. See Supporting Information Table S2 for the number of individuals available per sampling (per species per species × genetic richness). AF, *Alniphyllum fortunei*; CC, *Cinnamonum camphora*; DO, *Daphniphyllum oldhamii*; IP, *Idesia polycarpa*. Individual open dots represent individual hypervolume sizes; black solid items are mean values  $\pm 1$  SD.

responses both among species and among seed families within species. Multivariate trait variation among individuals was increased in the four-species mixture compared with the monocultures, while it was slightly decreased or stayed constant with increasing within-species genetic richness. Overall, species richness had a positive effect on stand-level tree biomass, while an effect of genetic richness was absent.

#### Trait variation among species and seed families

Trait variation was mainly determined by differences between species, but differences between seed families within species also explained a substantial amount of variation, which fits with previous studies showing the relative contribution of intraspecific trait variation to total trait variation (Albert et al., 2010; Hulshof & Swenson, 2010; Poorter et al., 2018). Intraspecific trait variation was created by both within-genetic components (seed family) and plastic responses to environmental conditions. Previous work assigned intraspecific trait variation to overall plasticity (Lipowsky et al., 2015; Roscher et al., 2018a; Benavides et al., 2019) without separation of genetic vs environmental components. Here, we attempted such a separation by explicitly including a within-species genetic component (seed family identity) of trait variation and also analyzing effects of species and genetic richness on trait variation. These findings add to an increasing body of literature demonstrating that while species' mean trait values can be useful to assess large-scale variation (Wright et al., 2004; Diaz et al., 2016; Bruelheide et al., 2018), within-species

trait variation must be included if effects of functional diversity on plant growth and ecosystem functioning are analyzed (Hughes *et al.*, 2008; Bolnick *et al.*, 2011; Violle *et al.*, 2012; Siefert *et al.*, 2015; Des Roches *et al.*, 2018).

Leaf area, LC and SA were the three traits most distinct between species. Leaf area variation was mainly a result of the much bigger leaves of I. polycarpa compared with the other species, while LC and SA were mainly determined by the difference between I. polycarpa and C. camphora. LN was mostly explained by variation among seed families, especially seed families from D. oldhamii. This fits with observations of Albert et al. (2010), who showed that, compared with other traits, intraspecific variation was highest for LN. Trait variation among seed families may be maintained by the heterogeneous neighborhoods encountered by individual trees in these species-rich subtropical forests (Ang et al., 2016) and by sibling competition and Janzen-Connell effects within species (Eck et al., 2019). Indeed, in a recent study it was found that species-rich experimental forests maintained a higher genetic diversity in two dipterocarp species than did monocultures (Ang et al., 2016).

A limitation of our study was that the expectation of differences among seed families was based on the assumption of higher genetic relatedness among individuals within than between seed families; individuals from the same seed family came from the same mother tree and were therefore considered to be at least half-sibs. Future research should analyze the genetic relatedness among our individuals in more detail to better assess the genetic contribution to phenotypic variation.



**Fig. 6** Individual tree growth and stand-level biomass per species  $\times$  genetic richness combination. (a, b) Relative individual tree biomass growth for each species  $\times$  genetic richness level, for all individuals together (a, n = 2073) or per species separately (b). (c, d) Stand-level biomass for each species  $\times$  genetic richness level, for all plots (c, n = 92) and for the species monocultures separately (d). Species  $\times$  genetic richness level contains values of one vs four species and one vs four seed families. Pink dots and blue triangles represent individual trees (a, b) or subplots (c, d) with one or four seed families per species, respectively. Large black items are mean values  $\pm 1$  SE per species  $\times$  genetic richness combination. AF, *Alniphyllum fortunei*; CC, *Cinnamomum camphora*; DO, *Daphniphyllum oldhamii*; IP, *Idesia polycarpa*. See Supporting Information Table S5 for the related statistics.

# Univariate and multivariate trait variation as a result of species and genetic richness

In addition to the trait variation among species and seed families, we also found significant trait variation within species and seed families that was a result of plasticity in response to species  $\times$  genetic richness treatments. Trait variation as a result of responses to species richness has been observed in other BEF-related studies

(Roscher *et al.*, 2011, 2018a; Lipowsky *et al.*, 2015; Benavides *et al.*, 2019; Bongers *et al.*, 2020) and has been shown to result in positive effects on productivity (Zuppinger-Dingley *et al.*, 2014; Zhu *et al.*, 2015). However, in these previous studies it was not tested if these phenotypic responses were a result of genetic differences or plasticity within genetic groups to environmental conditions (plasticity in the strict sense of its definition; see Bradshaw, 1965), because the genetic relationship between plants was not known.

The observed trait plasticity in response to richness led to higher multivariate trait diversity among individuals from species mixtures than from species monocultures, which was represented by increased multivariate hypervolume sizes. This effect was mainly a result of A. fortunei and I. polycarpa, because these species increased their hypervolume sizes and also slightly shifted their volume centroids. By contrast, increasing the genetic richness from one to four seed families affected the multivariate trait variation, depending on the species identity in species monocultures, while it had no effect on multivariate trait variation in the four-species mixture. However, increasing multivariate trait variation with increasing genetic richness was observed in the species mixture when using different sets of traits (Fig. S3). This could be related to the role of specific traits and illustrates that a true holistic interpretation of the ecological functions of individuals within a community remains challenging. Within this study, we believe that the use of principal component axes instead of sets of traits is a more appropriate method to estimate variation in ecological functions among individuals. Nevertheless, the fact that A. fortunei and I. polycarpa are both deciduous species and C. camphora and D. oldhamii are evergreen species suggests that this single trait could explain the difference in trait plasticity that determined multivariate trait variation change in response to both species and genetic richness (Valladares et al., 2000; Böhnke & Bruelheide, 2013).

Changes in multivariate trait variation as a result of species or genetic richness may be related to various processes. Higher species richness can increase the local environmental heterogeneity among all individuals and thus leads to increased differences among individuals. While higher genetic richness has less effect on local environmental heterogeneity it could reduce sibling competition and therefore increase asymmetries and differences among individuals (Willson et al., 1987; Weiner, 1990). By contrast, increased genetic richness can also reduce asymmetric competition and thus reduce individual variation. The different changes of the multivariate trait variation among individuals with increasing genetic richness illustrates that different processes can be at play within and among species. Further experiments would be needed to determine the relative roles of different drivers, for example competitive asymmetry or niche overlap, influencing differences between individuals. Linking this kind of knowledge to trait and growth variation will be an additional step towards understanding the ecological and functional mechanisms underlying BEF relationships.

The increased multivariate trait variation among individuals with increasing species richness as a result of species-specific trait shifts would be consistent with the hypothesis that individuals

and species try to minimize niche overlap to avoid competition (see e.g. Zuppinger-Dingley et al., 2014). The observation that species-specific multivariate trait variation increased with increasing species richness has also recently been shown for trees (Benavides et al., 2019) but is opposite to some previous findings regarding grasses (Zuppinger-Dingley et al., 2014; van Moorsel et al., 2018). The increased multivariate trait variation within species in species mixtures could be a result of larger heterogeneity of local environmental conditions and neighbor interactions compared with species monocultures. Decreased multivariate trait variation within species can be related to minimizing niche overlap among species, which could be functional in decreasing resource competition, as expected based on niche theory in diverse systems (Turnbull et al., 2016). Nevertheless, to determine if species or individuals shift their functional niche in diverse systems, more detailed studies are required that measure time- or space-specific traits (Niklaus et al., 2017; Williams et al., 2017) or actual resource uptake (von Felten et al., 2009; Ashton et al., 2010).

# Species richness, but not genetic richness within species, increased stand-level productivity

Over an 8 yr growth period, species richness positively affected stand-level tree biomass, but against our expectations no significant effect of genetic richness within species was observed. As expected, total trait variation among individuals in species mixtures was larger than the sum of the variation of the individuals from monocultures. This suggests niche differentiation among species and could be a mechanism contributing to the increased stand-level productivity in four species mixtures compared with the average monocultures (Zuppinger-Dingley *et al.*, 2014).

Although seed families did significantly contribute to trait variation within species, increased multivariate trait variation among individuals from more genetically rich communities was not observed within species mixtures and was only observed for two out of the four species in separate monocultures. This lack of effect could therefore explain the absence of strong genetic effects on individual growth and community-level productivity. Our results extend those of Fischer *et al.* (2017), who also found significant trait and growth differences between genotypes but no relationship between genetic richness and stand productivity within the single species *Populus fremontii.* The absence of an overall increased stand-level biomass in genetically rich communities across all four species could also be the result of opposite growth responses of the four species to the species × genetic richness treatment combinations.

In this context it should be noted that *A. fortunei* and *I. polycarpa* had decreased stand-level biomass and decreased multivariate trait variation under higher genetic richness, whereas *C. camphora* and *D. oldhamii* showed the opposite responses. These observations do match our hypothesis that the size of multivariate trait variation links to community productivity. Two recent studies that also test for such a relation between genetically based functional trait variation within species and species-level functioning in one case found positive effects and trait variation for *A. thaliana* (Wuest & Niklaus, 2018) and, in another case, no effect despite trait variation for rice (Montazeaud *et al.*, 2018). Thus, although there are several studies that report positive effects of genetic richness on (mostly single-species) ecosystem functioning (Schmid, 1994; Crutsinger *et al.*, 2006; Kotowska *et al.*, 2010; Prieto *et al.*, 2015; Cao *et al.*, 2018), it remains unknown to what extent these effects can be related to genetically caused functional trait variation within species. In addition, it remains to be studied if such effects could be extrapolated from single-species to multispecies communities.

There could also be a difference between trees and other organisms with regard to effects of genetic richness on ecosystem functioning and a difference between the type of ecosystem function considered. No or weak effects of within-species genetic richness on tree or plant growth and productivity have been reported in several other studies (Crawford & Rudgers, 2012; Prieto et al., 2015; Schöb et al., 2015; Hahn et al., 2017; Zeng et al., 2017). By contrast, strong within-plant-species genetic richness effects on fungal pathogens or herbivore abundance or diversity have been reported several times (Schmid, 1994; Crutsinger et al., 2006; Johnson et al., 2006; Barton et al., 2015; Campos-Navarrete et al., 2015; Cao et al., 2018; Barantal et al., 2019). For the same plots as in our study, Cao et al. (2018) also found a positive trend of increasing within-plant genetic richness on top of species richness for higher-level trophic community interactions, which is an important ecosystem process in forests.

### Conclusion

While previous studies found that increasing functional diversity with increasing species richness can explain parts of positive BEF relationships, the extent to which genetically caused functional trait diversity has similar effects has so far remained unknown. Here we explicitly separated the variation in functional traits between and within species into genetic components and species × genetic richness-induced plasticities. While species richness effects on trait variation could be related to increased standlevel functional diversity and tree biomass, genetic richness had a limited effect on increasing trait variation in species monocultures or species mixtures and consequently we did not find significant effects of genetic richness on stand-level productivity. Our research presents, to our knowledge, one of the few cases where the effects of genetic diversity within species and richness-driven plasticity on functional trait variation and stand-level forest characteristics have been analyzed.

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## **Author contributions**

KM, BS, HB, WD and XL designed the experiment and FJB, SL, WD, CZH and HY collected data. FJB, BS and XL performed data analyses. FJB, BS and XL wrote the paper. All authors contributed to the final preparation of the manuscript.

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### References

- Albert CH, Thuiller W, Yoccoz NG, Douzet R, Aubert S, Lavorel S. 2010. A multi-trait approach reveals the structure and the relative importance of intravs. interspecific variability in plant traits. *Functional Ecology* 6: 1192–1201.
- Ang CC, O'Brien MJ, Ng KKS, Lee PC, Hector A, Schmid B, Shimizu KK. 2016. Genetic diversity of two tropical tree species of the Dipterocarpaceae following logging and restoration in Borneo: high genetic diversity in plots with high species diversity. *Plant Ecology & Diversity* 9: 459–469.
- Ashton IW, Miller AE, Bowman WD, Suding KN. 2010. Niche complementarity due to plasticity in resource use: plant partitioning of chemical N forms. *Ecology* 91: 3252–3260.
- Balvanera P, Pfisterer AB, Buchmann N, He JS, Nakashizuka T, Raffaelli D, Schmid B. 2006. Quantifying the evidence for biodiversity effects on ecosystem functioning and services. *Ecology Letters* 9: 1146–1156.
- Barantal S, Castagneyrol B, Durka W, Iason G, Morath S, Koricheva J. 2019. Contrasting effects of tree species and genetic diversity on the leaf-miner communities associated with silver birch. *Oecologia* **189**: 687–697.
- Barton KE, Valkama E, Vehviläinen H, Ruohomäki K, Knight TM, Koricheva J. 2015. Additive and non-additive effects of birch genotypic diversity on arthropod herbivory in a long-term field experiment. *Oikos* 124: 697–706.
- Benavides R, Valladares F, Wirth C, Müller S, Scherer-lorenzen M. 2019. Intraspecific trait variability of trees is related to canopy species richness in European forests. *Perspectives in Plant Ecology, Evolution and Systematics* 36: 24–32.
- Bittebiere AK, Saiz H, Mony C. 2019. New insights from multidimensional trait space responses to competition in two clonal plant species. *Functional Ecology* 33: 297–307.
- Blonder B, Lamanna C, Violle C, Enquist BJ. 2014. The n-dimensional hypervolume. *Global Ecology and Biogeography* 23: 595–609.
- Blonder B, Morrow CB, Maitner B, Harris DJ, Lamanna C, Violle C, Enquist BJ, Kerkhoff AJ. 2018. New approaches for delineating n-dimensional hypervolumes. *Methods in Ecology and Evolution* 9: 305–319.
- Böhnke M, Bruelheide H. 2013. How do evergreen and deciduous species respond to shade?-Tolerance and plasticity of subtropical tree and shrub species of South-East China. *Environmental and Experimental Botany* 87: 179–190.
- Bolnick DI, Amarasekare P, Araújo MS, Bürger R, Levine JM, Novak M, Rudolf VHW, Schreiber SJ, Urban MC, Vasseur DA. 2011. Why intraspecific

trait variation matters to community ecology. *Trends in Ecology and Evolution* 26: 183–192.

- Bongers FJ, Schmid B, Sun Z, Li Y, Hardtle W, von Oheimb G, Li Y, Li S, Staab M, Ma K *et al.* 2020. Growth-trait relationships in subtropical forest are stronger at higher diversity. *Journal of Ecology* 108: 256–266.
- Booth RE, Grime JP. 2003. Effects of genetic impoverishment on plant community diversity. *Journal of Ecology* **91**: 721–730.
- Bradshaw AD. 1965. Evolutionary significance of phenotypic plasticity in plants. *Advances in Genetics* 13: 115–155.
- Bruelheide H, Dengler J, Purschke O, Lenoir J, Jiménez-alfaro B, Hennekens SM, Botta-Dukay Z, Chytry M, Field R, Jansen F *et al.* 2018. Global trait – environment relationships of plant communities. *Nature Ecology and Evolution* 2: 1906–1917.
- Bruelheide H, Nadrowski K, Assmann T, Bauhus J, Both S, Buscot F, Chen X, Ding B, Durka W, Erfmeier A *et al.* 2014. Designing forest biodiversity experiments: general considerations illustrated by a new large experiment in subtropical China. *Methods in Ecology and Evolution* 5: 74–89.
- Campos-Navarrete MJ, Munguía-Rosas MA, Abdala-Roberts L, Quinto J, Parra-Tabla V. 2015. Effects of tree genotypic diversity and species diversity on the arthropod community associated with big-leaf mahogany. *Biotropica* 47: 579–587.
- Cao HX, Klein AM, Zhu C, Staab M, Durka W, Fischer M, Fornoff F. 2018. Intra- and interspecific tree diversity promotes multitrophic plant–Hemiptera– ant interactions in a forest diversity experiment. *Basic and Applied Ecology* 29: 89–97.
- Cook-Patton SC, Mcart SH, Parachnowitsch AL, Thaler JS, Agrawal AA. 2011. A direct comparison of the consequences of plant genotypic and species diversity on communities and ecosystem function. *Ecology* 92: 915–923.
- Crawford KM, Rudgers JA. 2012. Plant species diversity and genetic diversity within a dominant species interactively affect plant community biomass. *Journal of Ecology* 100: 1512–1521.
- Crutsinger GM, Collins MD, Fordyce JA, Gompert Z, Nice CC, Sanders NJ. 2006. Plant genotypic diversity predicts community structure and governs an ecosystem process. *Science* 313: 966–968.
- Des Roches S, Post DM, Turley NE, Bailey JK, Hendry AP, Kinnison MT, Schweitzer JA, Palkovacs EP. 2018. The ecological importance of intraspecific variation. *Nature Ecology and Evolution* 2: 57–64.
- Diaz S, Kattge J, Cornelissen JHC, Wright IJ, Lavorel S, Dray S, Reu B, Kleyer M, Wirth C, Prentice IC *et al.* 2016. The global spectrum of plant form and function. *Nature* 529: 167–171.
- Drummond EBM, Vellend M. 2012. Genotypic diversity effects on the performance of *Taraxacum officinale* populations increase with time and environmental favorability. *PLoS ONE7*: e30314.
- Eck JL, Stump SM, Delavaux CS, Mangan SA, Comita LS. 2019. Evidence of within-species specialization by soil microbes and the implications for plant community diversity. *Proceedings of the National Academy of Sciences, USA* 116: 7371–7376.
- Fischer DG, Wimp GM, Hersch-Green E, Bangert RK, LeRoy CJ, Bailey JK, Schewitzer JA, Dirks C, Hart SC, Allan GJ *et al.* 2017. Tree genetics strongly affect forest productivity, but intraspecific diversity–productivity relationships do not. *Functional Ecology* 31: 520–529.
- Hahn CZ, Niklaus PA, Bruelheide H, Michalski SG, Shi M, Yang X, Zeng X, Fischer M, Durka W. 2017. Opposing intraspecific vs. interspecific diversity effects on herbivory and growth in subtropical experimental tree assemblages. *Journal of Plant Ecology* 10: 244–251.
- Hector A, Schmid B, Beierkuhnlein C, Caldeira MC, Diemer M, Dimitrakopoulos PG, Finn JA, Freitas H, Giller PS, Good J et al. 1999. Plant diversity and productivity experiments in European grasslands. *Science* 286: 1123–1127.
- Huang Y, Chen Y, Castro-Izaguirre N, Baruffol M, Brezzi M, Lang A, Li Y, Hardtle W, von Oheimb G, Yang X *et al.* 2018. Impacts of species richness on productivity in a large-scale subtropical forest experiment. *Science* 362: 80–83.
- Hughes AR, Inouye BD, Johnson MTJ, Underwood N, Vellend M. 2008. Ecological consequences of genetic diversity. *Ecology Letters* 11: 609–623.
- Hughes AR, Stachowicz JJ. 2004. Genetic diversity enhances the resistance of a seagrass ecosystem to disturbance. *Proceedings of the National Academy of Sciences, USA* 101: 8998–9002.

Hulshof CM, Swenson NG. 2010. Variation in leaf functional trait values within and across individuals and species: an example from a Costa Rican dry forest. *Functional Ecology* 1: 217–223.

Isbell F, Calcagno V, Hector A, Connolly J, Harpole WS, Reich PB, Scherer-Lorenzen M, Schmid B, Tilman D, van Ruijven J *et al.* 2011. High plant diversity is needed to maintain ecosystem services. *Nature* 477: 199–202.

Johnson MTJ, Lajeunesse MJ, Agrawal AA. 2006. Additive and interactive effects of plant genotypic diversity on arthropod communities and plant fitness. *Ecology Letters* 9: 24–34.

Kattge J, Bönisch G, Diaz S, Lavorel S, Prentice IC, Leadley P, Tautenhahn S, Werner GDA, Aakala T, Abedi M *et al.* 2020. Twelve years of TRY – towards a third generation of plant trait data assimilation and sharing. *Global Change Biology* 26: 119–188.

Kotowska AM, Cahill JF, Keddie BA. 2010. Plant genetic diversity yields increased plant productivity and herbivore performance. *Journal of Ecology* 98: 237–245.

Kraft NJB, Valencia R, Ackerly DD. 2008. Functional traits and niche-based tree community assembly in an Amazonian forest. *Science* 322: 580–582.

Kröber W, Li Y, Härdtle W, Ma K, Schmid B, Schmidt K, Scholten T, Seidler G, von Oheimb G, Welk E *et al.* 2015. Early subtropical forest growth is driven by community mean trait values and functional diversity rather than the abiotic environment. *Ecology and Evolution* 5: 3541–3556.

Latzel V, Allan E, Bortolini Silveira A, Colot V, Fischer M, Bossdorf O. 2013. Epigenetic diversity increases the productivity and stability of plant populations. *Nature Communications* 4: 2875.

Liang J, Crowther TW, Picard N, Wiser S, Zhou M, Alberti G, Schulze E, McGuire AD, Bozzato F, Pretzsch H *et al.* 2016. Positive biodiversity– productivity relationship predominant in global forests. *Science* 354: aaf8957.

Lipowsky A, Roscher C, Schumacher J, Michalski SG, Gubsch M, Buchmann N, Schulze E, Schmid B. 2015. Plasticity of functional traits of forb species in response to biodiversity. *Perspectives in Plant Ecology, Evolution and Systematics* 17: 66–77.

McGill BJ, Enquist BJ, Weiher E, Westoby M. 2006. Rebuilding community ecology from functional traits. *Trends in Ecology and Evolution* 21: 178–185.

Montazeaud G, Violle C, Fréville H, Luquet D, Ahmadi N, Courtois B, Bouhaba I, Fort F. 2018. Crop mixtures: does niche complementarity hold for belowground resources? An experimental test using rice genotypic pairs. *Plant and Soil* 424: 187–202.

Niklaus PA, Baruffol M, He JS, Ma K, Schmid B. 2017. Can niche plasticity promote biodiversity–productivity relationships through increased complementarity? *Ecology* 98: 1104–1116.

Poorter L, Castilho CV, Schietti J, Oliveira RS, Costa FRC. 2018. Can traits predict individual growth performance? A test in a hyperdiverse tropical forest. *New Phytologist* 219: 109–121.

Prieto I, Violle C, Barre P, Durand JL, Ghesquiere M, Litrico I. 2015. Complementary effects of species and genetic diversity on productivity and stability of sown grasslands. *Nature Plants* 1: 15033.

Roscher C, Gubsch M, Lipowsky A, Schumacher J, Weigelt A, Buchmann N, Schulze E, Schmid B. 2018a. Trait means, trait plasticity and trait differences to other species jointly explain species performances in grasslands of varying diversity. *Oikos* 127: 855–865.

Roscher C, Schmid B, Buchmann N, Weigelt A, Schulze ED. 2011. Legume species differ in the responses of their functional traits to plant diversity. *Oecologia* 165: 437–452.

Roscher C, Schumacher J, Gubsch M, Lipowsky A, Weigelt A, Buchmann N, Schulze E, Schmid B. 2018b. Interspecific trait differences rather than intraspecific trait variation increase the extent and filling of community trait space with increasing plant diversity in experimental grasslands. *Perspectives in Plant Ecology, Evolution and Systematics* 33: 42–50.

Roscher C, Schumacher J, Schmid B, Schulze E-D. 2015. Contrasting effects of intraspecific trait variation on trait-based niches and performance of legumes in plant mixtures. *PLoS ONE* 10: e0119786.

Schmid B. 1994. Effects of genetic diversity in experimental stands of *Solidago altissima* – evidence for the potential role of pathogens as selective agents in plant populations. *Journal of Ecology* 82: 165.

Schmid B, Baruffol M, Wang Z, Niklaus PA. 2017. A guide to analyzing biodiversity experiments. *Journal of Plant Ecology* **10**: 91–110.

Schöb C, Kerle S, Karley AJ, Morcillo L, Pakeman RJ, Newton AC, Brooker RW. 2015. Intraspecific genetic diversity and composition modify species-level diversity-productivity relationships. *New Phytologist* 205: 720–730.

Shipley B, de Bello F, Cornelissen HSC, Laliberte E, Laughlin DC, Reich PB. 2016. Reinforcing loose foundation stones in trait-based plant ecology. *Oecologia* 4: 923–931.

Siefert A, Violle C, Chalmandrier L, Albert CH, Taudiere A, Fajardo A, Aarssen LW, Baraloto C, Carlucci MB, Cianciaruso MV *et al.* 2015. A global metaanalysis of the relative extent of intraspecific trait variation in plant communities. *Ecology Letters* 18: 1406–1419.

Swenson NG, Enquist BJ. 2008. The relationship between stem and branch wood specific gravity and the ability of each measure to predict leaf area. *American Journal of Botany* 95: 516–519.

Tilman D, Isbell F, Cowles JM. 2014. Biodiversity and ecosystem function. Annual Review of Ecology, Evolution, and Systematics 45: 471–493.

Tilman D, Wedin D, Knops J. 1996. Productivity and sustainability influenced by biodiversity in grassland ecosystems. *Nature* **379**: 718–720.

Tobner CM, Paquette A, Gravel D, Reich PB, Williams LJ, Messier C. 2016. Functional identity is the main driver of diversity effects in young tree communities. *Ecology Letters* 19: 638–647.

Turnbull L, Isbell F, Purves DW, Loreau M, Hector A. 2016. Understanding the value of plant diversity for ecosystem functioning through niche theory. *Proceedings of the Royal Society B: Biological Sciences* 283: 20160536.

Valladares F, Wright SJ, Lasso E, Kitajima K, Pearcy RW. 2000. Plastic phenotypic response to light of 16 congeneric shrubs from a panamanian rainforest. *Ecology* 81: 1925–1936.

van Moorsel SJ, Hahl T, Wagg C, De Deyn GB, Flynn DFB, Zuppinger-Dingley D, Schmid B. 2018. Community evolution increases plant productivity at low diversity. *Ecology Letters* 21: 128–137.

Verheyen K, Vanhellemont M, Auge H, Baeten L, Baraloto C, Barsoum N, Bilodeau-Gauthier S, Bruelheide H, Castagneyrol B, Godbold D *et al.* 2016. Contributions of a global network of tree diversity experiments to sustainable forest plantations. *Ambio* 45: 29–41.

Violle C, Enquist BJ, McGill BJ, Jiang L, Albert CH, Hulshof C, Jung V, Messier J. 2012. The return of the variance: intraspecific variability in community ecology. *Trends in Ecology and Evolution* 4: 244–252.

von Felten S, Hector A, Buchmann N, Niklaus PA, Schmid B, Scherer-Lorenzen M. 2009. Belowground nitrogen partitioning in experimental grassland plant communities of varying species richness. *Ecology* 90: 1389– 1399.

Wagg C, Ebeling A, Roscher C, Ravenek J, Bachmann D, Eisenhauer N, Mommer L, Buchmann N, Hillebrand H, Schmid B et al. 2017. Functional trait dissimilarity drives both species complementarity and competitive disparity. *Functional Ecology* 31: 2320–2329.

Weiner J. 1990. Asymmetric competition in plant populations. Trends in Ecology and Evolution 5: 360–364.

Weisser WW, Roscher C, Meyer ST, Ebeling A, Luo G, Allan E, Bessler H, Barnard RL, Buchmann N, Buscot F et al. 2017. Biodiversity effects on ecosystem functioning in a 15-year grassland experiment: patterns, mechanisms, and open questions. *Basic and Applied Ecology* 23: 1–73.

Williams LJ, Paquette A, Cavender-Bares J, Messier C, Reich PB. 2017. Spatial complementarity in tree crowns explains overyielding in species mixtures. *Nature Ecology and Evolution* 1: 0063.

Willson MF, Hoppes WG, Goldman DA, Thomas PA, Katusic-Malmborg PL, Bothwell JL. 1987. Sibling competition in plants: an experimental study. *American Naturalist* 129: 304–311.

Wright IJ, Reich PB, Westoby M, Ackerly DD, Baruch Z, Bongers F, Cavender-Bares J, Chapin T, Cornelissen JHC, Diemer M et al. 2004. The worldwide leaf economics spectrum. *Nature* 428: 821–827.

Wuest SE, Niklaus PA. 2018. A plant biodiversity effect resolved to a single chromosomal region. *Nature Ecology and Evolution* 2: 1933–1939.

Yang J, Cao M, Swenson NG. 2018. Why functional traits do not predict tree demographic rates. *Trends in Ecology and Evolution* 33: 326–336.

Zeng X, Durka W, Fischer M. 2017. Species-specific effects of genetic diversity and species diversity of experimental communities on early tree performance. *Journal of Plant Ecology* 10: 252–258.

- Zhu J, van der Werf W, Anten NPR, Vos J, Evers JB. 2015. The contribution of phenotypic plasticity to complementary light capture in plant mixtures. *New Phytologist* 207: 1213–1222.
- Zuppinger-Dingley D, Schmid B, Petermann JS, Yadav V, De Deyn GB, Flynn DF. 2014. Selection for niche differentiation in plant communities increases biodiversity effects. *Nature* 515: 108–111.

# **Supporting Information**

Additional Supporting Information may be found online in the Supporting Information section at the end of the article.

**Fig. S1** Visualization of the loading of the eight traits on the first three axes of the principal component ordinations performed using 547 individuals.

**Fig. S2** Additional analyses showing multivariate trait hypervolume sizes of all species combined and for each species separately, calculated using individuals from low or high (one vs four) species richness.

**Fig. S3** Additional analyses showing multivariate trait hypervolume sizes of each species separately in species monoculture or all species combined in species mixtures, calculated using individuals from low or high genetic richness.

Fig. S4 Trait values and growth rate reaction-norms of seed families within species for species  $\times$  genetic richness.

Fig. S5 Visualization of one of the 500 simulated hypervolumes per species separately in species monoculture versus species mixtures.

Fig. S6 Dimension-specific hypervolume centroids per species within species monocultures and species mixtures.

**Table S1** Overview of the species and seed-family compositionper plot for each species × genetic richness combination.

**Table S2** Number of individuals per species per seed family that were measured for functional traits and growth for each species  $\times$  genetic richness combination.

**Table S3** Summary of linear model analyses of species, seed-family and richness effects on eight traits and relative tree growth, considering factors in hierarchical structure.

**Table S4** List of species pairs that were nonsignificant and seed-family pairs within species that were significantly different for a specific trait (P < 0.05).

**Table S5** Summary results of ANOVAs testing the effect of species and genetic richness on individual tree growth (2013–2018) and stand-level tree biomass (2018).

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